Is benthic food web structure related to diversity of marine macrobenthic communities?

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Abstract:

Numerical structure and the organisation of food webs within macrozoobenthic communities has been assessed in the European waters (Svalbard, Barents Sea, Baltic Sea, North Sea, Atlantic Ocean and the Mediterranean Sea) to address the interactions between biodiversity and ecosystem functioning. Abundance and classical species diversity indices (S, H', J) of macrofaunal communities were related to principal attributes of food webs (relative trophic level and food chain length, FCL) that were determined from carbon and nitrogen stable isotope values. Structure of marine macrobenthos varies substantially at a geographical scale; total abundance ranges from 63 ind. m^{-2} to 34,517 ind. m^{-2}, species richness varies from 3 to 166 and the Shannon-Weaver diversity index from 0.26 to 3.26 while Pielou's evenness index is below 0.73. The major source of energy for macrobenthic communities is suspended particulate organic matter, consisting of phytoplankton and detrital particles, sediment particulate organic matter, and microphytobenthos in varying proportions. These food sources support the presence of suspension- and deposit-feeding communities, which dominate numerically on the sea floor. Benthic food webs include usually four to five trophic levels (FCL varies from 3.08 to 4.86). Most species are assigned to the second trophic level (primary consumers), fewer species are grouped in the third trophic level (secondary consumers), and benthic top predators are the least numerous. Most species cluster primarily at the lowest trophic level that is consistent with the typical organization of pyramidal food webs. Food chain length increases with biodiversity, highlighting a positive effect of more complex community structure on food web organisation. In more diverse benthic communities, energy is transferred through more trophic levels while species-poor communities sustain a shorter food chain.

1

1. Introduction

Interactions between biodiversity and ecosystem functioning in marine systems have attracted considerable attention for many years. Understanding the potential link of structural features of faunal communities and their functionality can provide clues to the assessment of threats to ecosystem state, productivity and services to humans due to biodiversity changes induced by climate change and anthropogenic activities. This study addresses biodiversity-functioning linkages by comparing benthic food web structure in marine and coastal ecosystems of a different biodiversity level and with diverse physical or human constraints at a pan-European scale. Ecosystem processes integrate the interactions between organisms and both the environment and each other, and biodiversity mediates these interactions (Heip et al., 1998). Diversity of habitats and species varies geographically across environmental and ecological gradients, structuring ecosystems and their functioning at local and global scales. Ecosystem metabolism is intimately linked to carbon and nitrogen fluxes from primary producers to consumers of higher trophic levels. This trophic transfer determines the productivity of ecosystems depending on the structure and efficiency of the food webs. Systems with high diversity and complex trophic interrelations such as the Mediterranean Sea or the Atlantic Ocean are considered to be stable and productive (Loreau, 2000), while in systems such as the Baltic Sea average diversity is low and food web structure relatively simple (Sokołowski, 2009). Despite such striking differences in their structures, the productivity of the food web in the Baltic is reported to be similar to that of the Atlantic (Elmgren and Hill, 1995) which contrasts the early "productivity hypothesis" of Elton (1927) assuming that more productive ecosystems have longer food chains. This would indicate that high biodiversity might not be an essential prerequisite for stable or highly productive ecosystems. However, the number of trophic interactions between species increases with a concomitant increase in species number, as do the number of material cycles and pathways within a food web. Other hypotheses argue that food chain length should increase with increasing ecosystem size and community organisation (Vander Zanden et al., 1999; Post et al., 2000) and that food webs with longer food-chain lengths might be more susceptible to shortening by disturbance (Jenkins et al., 1992). Furthermore, according to Post (2000) in isolated or evolutionarily young systems food web structure can be strongly influenced by colonisation history. Some evidence exists also that energy or resource availability at the base of food web and energetic efficiency of organisms within a food web (trophic efficiency) increase the complexity of food web and food chain length (Yodzis, 1984; Jenkins et al., 1992; Kaunzinger and Morin, 1998; Post, 2002). Although the importance of community

composition in explaining variation in food web structure has been debated in scientific community for long time, only limited progress has been made in understanding the relationship between biodiversity and trophic web organisation in natural ecological systems. Therefore the question arises; to what extent food web structure in an ecosystem (e.g. seagrass bed, mussel bed, subtidal sand, mudflats, etc.) is related to the diversity of the species pool.

To address this question, a large-scale comparative study was performed on the structure of food webs in macrobenthic systems that differ in the magnitude and complexity of species assemblages in different European marine and coastal waters. By relating data on basic food web principles with selected species diversity indices and total macrobenthic abundance, possible linkages between food web organisation and taxonomic diversity were unravelled.

2. Material and methods

2.1. Study sites

Data on taxonomic structure and carbon and nitrogen stable isotope ratios of macrobenthic fauna were obtained through original studies of the authors (own data) or other researchers and by conducting literature searches using available scientific databases such as ISI Web of Knowledge. Eight locations were selected to represent a range of environmental and ecological systems (e.g. the presence of tides, salinity gradients, shallow- and deep-water regions, the presence of macroplants) covering a suite of habitats from the European Arctic in the North (Spitsbergen and Barents Sea) through the Baltic Sea (Curonian Lagoon and the Gulf of Gdańsk), North Sea (Sylt-Rømø Bight and Westerschelde), the Atlantic coast of France (Marennes-Oleron Bay), to the Mediterranean Sea (Bay of Banyuls-sur-Mer) (Fig. 1, Table 1). In addition, at locations of high habitat diversity, two sites differing in environmental conditions and a structure of macrobenthic communities were chosen. Out of a total of 13 sites, five are BIOMARE project (Implementation and Networking of large-scale long-term Marine Biodiversity research in Europe) biodiversity sites (Warwick et al., 2003).

Fig. 1 and Table 1 here

2.2. Benthic community structure

Benthic communities were considered as the faunal (> 1mm) organisms inhabiting the top 10 cm of surficial sediments or epiphytic taxa living on the surface of bottom substratum. Special care was taken to use data from samples taken with quantitative gears (Van Veen grab or corer) of similar area (ca. 0.1 m^2) as sample size has been demonstrated to affect benthic

community structure parameters (Eleftheriou and McIntyre, 2005). In order to reduce variations in community structure due to temporal alterations of environmental variables and ecophysiological traits of organisms, data referring to spring or summer were extracted from the databases for most of the sites (Table 2). Since protists were not considered at the majority of sites, they were excluded from the analysis. On the basis of the collected data from direct field investigations or literature (Table 2), species richness (S, a total number of species per sample) and total macrobenthic abundance were calculated to provide descriptive statistics of benthic communities. In systems where predatory birds or mammals are considered to be the top predators, they were excluded from the analysis due to the lack of available data and to maintain consistency throughout the study. Taxonomic nomenclatures used followed the European Register of Marine Species (http://www.marbef.org/data/erms.php) and http://www.itis.gov. The diversity of macrobenthic communities was estimated using the Shannon-Weaver diversity index H' (Shannon and Weaver, 1963) while evenness was determined using the Pielou's J (Pielou, 1966).

2.3. Organisation of food webs

In order to define the organisation of macrobenthic food webs, data on stable isotopes of carbon ($^{13}C/^{12}C$) and nitrogen ($^{15}N/^{14}N$) in non-living organic and living components of each ecosystem were incorporated into the analyses. Because of the convincing presence of seasonal effects on isotope ratios (Martinez del Rio, 2007), only data collected in spring and summer were used. At most sites the isotope stable-isotope data were taken from literature while two sites, namely Hornsund and Brouage, were sampled in 2009 for the purpose of this study (Table 3). The samples were analysed for carbon and nitrogen stable isotope ratios following a routine procedure (Boutton, 1991). In brief, suspended particulate organic matter on Whatman GF/F filters (glass microfibre filters; 0.7 µm), <63 µm fraction of surficial sediments (0-5 cm), macrophytes and soft tissue of benthic animals were frozen, freeze-dried, homogenised and acidified with 1 M HCl overnight when necessary. Benthic organisms were grouped into three pools containing at least three, but for most species more, individuals in order to obtain a representative sample. $\delta^{13}C$ and $\delta^{15}N$ were determined using an Isoprime Micromass IRMS-EA (EUROVECTOR CHN analyser coupled with MICROMASS mass spectrometer) and the data obtained were expressed in the standard δ unit notation (‰). The composition of carbon isotopes has long been used as a tracer of food sources in marine and estuarine ecosystems (Hecky and Hesslein, 1995) while nitrogen isotope ratios indicate the trophic position in the food chain (Vander Zanden and Rasmussen, 2001). Nitrogen isotope

1 ratios were employed to estimate the relative trophic level (RTL) of a consumer in the food

2 web, using the model developed by Hobson and Welch (1992):

3 $$RTL = (\delta^{15}N_{consumer} - \delta^{15}N_{baseline})/3.4 + 2, \qquad (1)$$

4

5 where $\delta^{15}N_{consumer}$ is the nitrogen isotope ratio of a consumer, $\delta^{15}N_{baseline}$ is the nitrogen

6 isotope baseline of the primary consumer, 3.4 is the assumed ^{15}N trophic enrichment factor

7 (Post, 2002; Olive et al., 2003). Primary consumers have been agreed to provide an

8 appropriate baseline signal since they show lower temporal variability than organic matter

9 sources (Post, 2002; Vander Zanden and Fetzer, 2007). Due to possible among-site variations

10 in $\delta^{15}N_{baseline}$, the nitrogen isotope baseline has been calculated separately for each site as the

11 mean $\delta^{15}N$ of all primary consumers (Vander Zanden and Fetzer, 2007). The formula is a

12 simplified modification of the food web model of Post (2002) and allows for the contribution

13 of various organic matter sources in a diet of consumers, but rather gives rough estimates

14 since the weighting of each consumer, and thus its diet, is equal. However, such a model has

15 been successfully employed to define trophic levels of different species or groups of

16 organisms in marine, coastal and freshwater systems (Hobson et al., 1995; Vander Zanden et

17 al., 1999; Fredriksen, 2003; Grall et al., 2006). A RTL between 2 and 3 indicates the second

18 trophic level (primary consumers and omnivores) while RTLs>3 indicate consumers with

19 carnivorous diets (secondary, tertiary and higher level consumers).

20 Using carbon and nitrogen natural isotopes, the following principal attributes of food web

21 were determined separately at each site: (1) the source of organic matter fuelling benthic

22 biocenosis, (2) the relative trophic level (RTL) of each consumer species and (3) total food

23 chain length (*FCL*). The *FCL* has been employed following a definition of Duffy et al. (2007)

24 i.e. the average number of steps that energy is transferred as it moves from base resources to

25 top predators.

26 The origin of organic matter was assessed using carbon isotope ratios of all potential

27 carbon sources in a given system, including suspended particulate organic matter (SPOM),

28 sediment organic matter (SOM), microphytobenthos (bulk sample) and macrophytobenthos.

29 In addition, at the Barents Sea locations SPOM bound to ice was incorporated into analyses as

30 an important carbon source in the Arctic (Tamelander et al., 2006). Food chain length was

31 defined as the trophic position of the top predator i.e. taxon with the highest $\delta^{15}N$ value

32 (Vander Zanden and Fetzer, 2007):

33

$$FCL = (Max\delta^{15}N_{consumer} - \delta^{15}N_{baseline})/3.4 + 2, \qquad (2)$$

FCL is thus a measure of energy transfers (trophic levels) from the base to the top of the food web that integrates the assimilation of energy or mass flow through all the trophic pathways (Post, 2002). Functional relationships between pairs of variables, i.e. species richness, total macrobenthic abundance, diversity and evenness indices *vs. FCL* were assessed with regression analysis (Kendall's robust line-fit method). The significance of source type, sampling site and location on $\delta^{13}C$ and $\delta^{15}N$ was evaluated by analysis of variance (ANOVA) proceeded by analysis of normality (the Kolmogorov–Smirnov test for goodness of fit) and homogeneity of variances. The level of significance for all tests was set as $p < 0.05$ (Sokal and Rohlf, 1995). Statistical analyses were carried out using STATISTICA 8.1 (StatSoft, Inc., 1997).

3. Results

3.1. Macrobenthic community structure

A total of 499 faunal species were identified representing 33 classes. Polychaeta with totally 183 species was the most represented class, followed by Crustacea (110 species), Bivalvia (75 species) and Gastropoda (30 species). At two sites Curonian-mb and Curonian-bs, located in the Curonian Lagoon, Clitellata was the most numerous class, containing five and four species, respectively. Species richness (*S*) ranged from three at site Gdansk to 166 in the Barents Sea. At most sites the taxonomic diversity was dominated by polychaetes which constituted from 33.3% (Gdansk) up to 44.5 % (Banyuls) of the total number of macrofaunal species. Exceptions were found at the Baltic sites: at Puck Bay with a predominant role of bivalves (36.4%), Curonian-mb – gastropods (20.7%) and at Curonian-bs where equal contributions of 17.4 % were observed for bivalves, crustaceans and clitellates, respectively. Taxonomic diversity was also highly variable, with the Shannon-Weaver diversity index varying from 0.26 at site Oleron *Zostera*-up on the Atlantic coast of France to 3.26 at site St 4 in the Barents Sea. Evenness was not high with Pielou's index calculated to below 0.74 for all sites with the lowest value of 0.08 observed at site Oleron *Zostera*-uf (Tab. 2).

<u>Table 2 here</u>

High among-site variability was also clear for the total abundance of macrobenthic fauna which ranged from 63 ind. m^{-2} at site Gdansk to 71704 ind. m^{-2} at site Sylt-Rømø. In contrast

1 to the pattern of geographic variation in distribution of faunal species among classes, molluscs

2 were the most abundant group at eight sites (gastropods at five sites and bivalves at three

3 sites) comprising up to 96.2% of the total macrobenthic abundance followed by polychaetes

4 (four sites) with a percentage contribution of 85.8% (Fig. 2).

5 <u>Fig. 2 here</u>

6

7 *3.2. Stable nitrogen and carbon isotope ratios*

8 *3.2.1. Organic matter sources*

9 Five ecosystem components were considered the potential food sources to primary

10 consumers across all sampling sites i.e. sediment organic matter/detritus (SOM), suspended

11 particulate organic matter (SPOM) which include iceSPOM and phytoplankton,

12 microphytobenthos (bulk sample), macroalgae regrouped into nine genera (i.e. *Cladophora*,

13 *Fucus*, *Ulva*, *Ceramium*, *Gracilaria*, *Porphyra*, *Laminaria*, *Chorda*, *Acrosiphonia*) and

14 vascular plants representing five genera (i.e. *Zostera*, *Potamogeton*, *Spartina*, *Salicornia*,

15 *Posidonia*). These organic carbon sources differed statistically in isotope carbon ratios (δ^{13}C

16 spanning a range of 22.9% for all sites; *ANOVA*, $F_{4, 46} = 18.92$, $P < 0.001$) but not in stable

17 nitrogen ratios (δ^{15}N spanning a range of 8.7% for all sites) (Fig. 3). Regardless of geographic

18 position and environmental/biocenotic conditions, the most depleted of δ^{13}C was SPOM

19 (from -28.5% to -21.8%) with clearly lowered values of riverine suspended particulate

20 organic matter (down to -33.3% at site Curonian-mb) that can be distinguished from local

21 marine SPOM. Increased δ^{13}C was noticed for iceSPOM i.e. -21.7% and -20.0% at sites St 4

22 and St 16 in the Barents Sea, respectively. Macroalgae and microphytobenthos demonstrated

23 relatively enriched carbon isotope ratios varying from -21.8% to -13.2% and from -15.7% to -

24 14.4%, respectively. The isotope composition of SOM fell within a range of -24.7% to -

25 20.1% which reflects a combination of SPOM and phytal material in varying percentages

26 depending on the proximity of the two sources, and presumably also intense microbial

27 decomposition of sedimented organic material. Vascular plants had the most enriched carbon

28 isotope ratios from -15.1% to -8.4% at all sites, with the exception of the common glasswort,

29 *Salicornia* sp. at site Sylt-Rømø where δ^{13}C was low at about -28.6% (Fig. 3).

30 <u>Fig. 3 here</u>

31

32 The combined data on δ^{13}C and δ^{15}N revealed dual sources of organic matter to

33 macrobenthic communities at a site. The principal carbon source of the food webs comprised

34 SPOM and SOM (sediment detrital organic matter) which fuel benthic organisms at sites

1   Puck, Gdansk, Curonian-mb, Curonian-bs, Sylt-Rømø, Molenplaat, Banyuls, and St 16 and

2   Hornsund. At sites St 4 and St 16 in the Barents Sea iceSPOM appeared an important food

3   source to consumers on a local scale. In very shallow euphotic areas of dense benthic

4   vegetation (sites Oleron *Zostera*-uf and Oleron *Zostera*-lf) macroalgae and vascular plants can

5   contribute substantially to the organic matter pool which is exploited by benthic fauna. In

6   some instances however, δ^{13}C and δ^{15}N of benthic invertebrates cannot be linked to any

7   carbon sources analysed (e.g. insects and gastropods at site Curonian-bs, ophiuroids and

8   holothurians at St 4), suggesting the complex utilisation of other sources of organic matter

9   (e.g. decaying detritus, products of anaerobic metabolism of chemoautotrophic bacteria)

10 which may form separate trophic pathways.

11   Fig. 4 and Table 3 here

12

13 *3.2.2. Trophic positions of benthic consumers and food chain length*

14   A total of 327 macrobenthic faunal species regrouped into 30 classes were analysed for

15 stable isotope composition, representing from 17.4% (at site Curonian-bs) to 66.7% (at site

16 Puck), and from 47.1% (at site Curonian-bs) to 98.9% (at site Brouage) of species richness

17 and the total abundance at a given site, respectively. Low percentage contributions of species

18 number and abundance, that were measured for stable isotopes, occurred at the two sites St 4

19 and St 16 in the Barents Sea. This results from the presence of numerous small-sized

20 macrobenthic organisms of an even distribution and low individual dry weight (mainly

21 Crustacea, Polychaeta and Malacostraca) which provided insufficient material for isotope

22 analyses. Nevertheless, the sampling effort can be considered satisfactory since it covered a

23 wide array of macrobenthic species in each water-basin. The number of taxa that were

24 analysed per site varied from three at the site Gdansk to 121 at Banyuls (Table 3).

25   The carbon and nitrogen isotope composition of benthic macrofauna spanned a wide range

26 around the average of 20.4‰ (from -30.1‰ for freshwater leech *Helobdella stagnalis* in the

27 brackish Baltic Sea to -9.7‰ for *Idotea chelipes* on the Atlantic coast of France) and 19.8‰

28 (from 2.7‰ for the bivalve *Gari depressa* in the Mediterranean Sea to 22.5‰ for the

29 predatory polychaete *Eteone* in the North Sea) for δ^{13}C and δ^{15}N, respectively. Stable isotope

30 ratios showed apparent geographic variations among locations (*ANOVA*, $F_{6,\,399} = 110.07$, $P <$

31 0.001) and sites (*ANOVA*, $F_{12,\,399} = 74.72$, $P < 0.001$), reflecting spatial differences in organic

32 matter sources across sampling sites, but it can also indicate different feeding strategies

33 among species. Variation in δ^{13}C and δ^{15}N also differed among dominating systematic groups

34 within each site. The δ^{13}C ratio spanned the widest range for bivalves and crustaceans that

1 implies a broad food spectrum of these animals and reflects among-site differences in $\delta^{13}C$ of

2 carbon sources. In contrast, bivalves and gastropods demonstrated the least variable $\delta^{15}N$ as

3 they represent primary consumers feeding mainly on SPOM (suspension-feeders) and

4 epiphytic microalgae (herbivores), respectively. The exception was site Banyuls where

5 predatory gastropods of the genus *Nassarius* had a $\delta^{15}N$ up to 9.7‰.

6   The number of primary consumer species at a site ranged from two (Gdansk) to 36

7 (Banyuls) and was usually > 7 species at the vast majority of sites. The calculated site-specific

8 nitrogen isotope baseline ranged from 2.7‰ at site Banyuls up to 16.9‰ at site Molenplaat.

9 Food chain length (*FCL*) spanned at least three trophic levels and varied from 3.08 at the

10 Baltic site Gdansk up to 4.86 at site Curonian-mb (Table 3). Primary consumers (RTL

11 between 2.0 and 3.0) included mainly bivalves (filter-feeders), herbivorous gastropods,

12 crustaceans and polychaetes of mixed diet. At the majority of sites, secondary consumers

13 (RTL between 3.0 and 4.0) were comprised of larger predators, namely infaunal and epifaunal

14 crustaceans, polychaetes and fish. At site Banyuls in the Mediterranean Sea this trophic level

15 included also numerous carnivorous gastropods while at sites St 4 and St 16 in the Barents

16 Sea asteroids were abundant. Fish and asteroid species occupied the highest trophic position

17 of RTL >4.0. The number of species in subsequent trophic levels generally decreased with

18 trophic position (Fig. 5), shaping the hierarchical structure of the trophic pyramid. Most

19 species were assigned to the second level, fewer species were in the third trophic level, and

20 benthic top predators were the least numerous or even contained no species.

21   Fig. 5 here

22

23 *3.2.3. Relation between biodiversity and food web principles*

24   Relationships between total abundance of macrobenthos, species richness (*S*) and the

25 calculated diversity indices (*H'* and *J*) and *FCL* were determined, using the correlation

26 coefficient at the confidence level of 95 %, to assess potential links between the composition

27 of macrobenthic communities and food web structure. Food chain length was positively

28 correlated with species richness ($FCL=0.007S+3.702$, $R^2=0.48$, $P < 0.009$, n = 13) and the

29 Shannon-Weaver diversity index ($FCL=0.606H'+3.092$, $R^2=0.78$, $P < 0.001$, n = 13).

30

31

32 **4. Discussion**

33 *4.1. Structure of the marine benthos in European waters*

1   Large-scale studies of macrobenthic communities revealed substantial differences in

2   abundance and biodiversity level of the marine macrobenthos in European waters. The lowest

3   species richness (3-29) and density (63-5015 ind. m^{-2}) were observed in the non-tidal Baltic

4   Sea where low salinity (6.0-8.0), eutrophication and locally adverse oxygen conditions

5   strongly affect the number of benthic faunal taxa (Wolowicz et al., 2007; Renaud et al., 2009).

6   Another explanation includes the relatively young age of the Baltic which has been slowly

7   colonized by freshwater and marine organisms since the complete deglaciation only 8000

8   years ago (Bonsdorff, 2006). Larger numbers of taxa (27-32) and higher abundance of

9   macrobenthos (7152-15701 ind. m^{-2}) were noted in the fully-saline tidal flats of the North Sea

10   and the French Atlantic coast that are located between 45°N and 54°N. A similar number of

11   species in the coastal zone of intermediate latitudes was also reported in extensive studies of

12   diversity patterns of benthic shelf invertebrates by Renaud et al. (2009). Benthic assemblages

13   at two open-water deep Arctic sites in the Barents Sea demonstrated the highest species

14   richness (166), which fits well the pattern of increasing diversity with depth and towards the

15   pole. As the hypothesis on the increase of marine benthic biodiversity toward a pole has not

16   been fully confirmed (e.g. Włodarska-Kowalczuk et al., 2007; Renaud et al. 2006), the

17   species richness has been convincingly proved to follow unimodal trends with water depth

18   and a maximum at 180-220 m water depths (Renaud et al., 2009). Exceptionally high

19   numbers of benthic species (128) were shown at the most southern site Banyuls which is

20   located in the Bay of Banyuls-sur-Mer, a small open bay in the southwest of the Gulf of

21   Lions, France (northwestern Mediterranean Sea).

22

23   *4.2. Organisation of food webs - use of primary organic matter sources*

24   Irrespective of geographic location, the major sources of energy to macrobenthic

25   communities appeared to originate from suspended particulate organic matter, comprising

26   phytoplankton and detrital particles of floral and faunal origin (SPOM) in varying

27   proportions, and sediment organic matter (SOM). The significance of SPOM and SOM as the

28   main food source for macrobenthos can be also indicated by the dominance of suspension-

29   and deposit-feeding taxa including bivalves, gastropods (Mollusca) and polychaetes

30   (Annelida) over herbivorous taxa at the majority of sites (Fig. 2). These taxonomic groups are

31   hypothesized to contribute substantially to benthic-pelagic coupling, and also providing the

32   main energy transfer pathway within the benthic biocenoses. The dominating trophic role of

33   SPOM and planktonic production has also been noted in isotope investigations of benthic

34   food webs in a North Atlantic maerl bed (Grall et al., 2006), the Prince Edward Archipelago

1 (Kaehler et al., 2000), and in the Marennes-Oléron Bay (Riera, 1998). In the open waters of

2 the Barents Sea, but not in the Arctic fjord of Spitsbergen, iceSPOM which shows enriched

3 δ^{13}C and δ^{15}N values relative to SPOM (Tamelander et al., 2006), provides energy to benthic

4 invertebrates. In addition, on sandy and muddy tidal flats of the North Sea and the French

5 Atlantic coast, autochthonous sources of organic matter include microphytobenthos (Monbet,

6 1992). Epipelic diatoms have apparently a distinct carbon isotope composition from SPOM

7 and SOM that derives from their weaker discrimination against $^{13}CO_2$ due to relative lower

8 access to atmospheric CO_2 as it is in the water column for plankton algae. Benthic macroalgae

9 and vascular plants can be neglected as a potential carbon source in benthic food webs. Even

10 in the shallow-water euphotic areas of dense vegetation, only small groups of invertebrates

11 (some gastropods and crustaceans) had an isotope signature consistent with those of the

12 benthic macroalgae and vascular plants after trophic enrichment. Therefore, these

13 supplementary trophic links do not to support additional energy pathways to higher trophic

14 levels. A large part of macrophytal primary production probably supplies the sediment

15 organic matter pool which fuels secondary production *via* the detritus food chain.

16

17 *4.3. Vertical structure of food webs - food chain length*

18   The calculated site-specific nitrogen isotope baseline fell within a range of ca. from 6.0‰

19 to 9.4‰ at the majority of sites, except at Banyuls and Molenplaat where it was 2.7‰ and

20 16.9‰, respectively. The carbon and nitrogen stable isotope ratios of consumers generally

21 reflect the isotope composition of their food sources in a particular area with the known

22 increment factors (Ostrom and Fry, 1993). The similarity in δ^{13}C and δ^{15}N between animals

23 and plants from the same habitat was documented as early as in the first isotope studies of the

24 marine environment (Craig, 1953; Parker, 1964) and was further evidenced by a number of

25 later studies in various aquatic systems (e.g. Créach et al., 1997; Riera et al., 1999;

26 Fredriksen, 2003). The shift in δ^{13}C and δ^{15}N between a consumer and its food is the principal

27 and basic assumption in all ecological studies that aim at assessing trophic pathways in any

28 faunal community (Martinez del Rio, 2007). Close relationships in the carbon and nitrogen

29 composition of organic matter sources and animals also allows distinguishing different

30 populations that rely on isotopically divergent sources at local and larger scales. However,

31 recent reviews indicate that the fractionation of δ^{15}N might be affected by multiple factors

32 such as taxonomic position, the physiological condition of an animal (e.g. starvation), the

33 nitrogen contents of the food and the consumer (specifically protein and amino acid

concentrations), growth efficiency and ontogenic state (Vander Zanden and Rasmussen, 2001; McCutchan Jr. et al., 2003; Vanderklift and Ponsard, 2003).

Food chain length varied from 3.08 to 4.86 (Table 3) but these values might be underestimated due to the absence of migrating aquatic birds and marine mammals as top predators whose introduction into analyses can increase the *FCL* on average by 0.64 (Vander Zanden and Fetzer, 2007). The observed range of *FCL* is consistent with *FCL* estimates for estuarine, coastal and marine systems reported by Vander Zanden and Fetzer (2007), and generally agrees with classical food web theory, which states that food chains are usually limited to a maximum of three or four trophic levels in aquatic systems (Cohen et al., 1986; Briand and Cohen, 1987). Hastings and Conrad (1979) argued that the evolutionary stable length of a food chain would be three, and that the main determinant of the number of trophic levels is the quality of the organic matter energy sources (e.g. the C:nutrient ratio) and not their quantity, as is often suggested in discussions of food chain length. Omnivory may be a compromise between exploiting large quantities of low quality resources (at a low metabolic cost) and exploiting lesser quantities of high-quality food (at a high metabolic cost). Benthic omnivorous organisms can thus avoid nutrient deficiency (due to lower nutritional values of phytal and detrital materials of high C:nutrient ratio) through their access to a large reservoir of energy (Elser and Hessen, 2005). This feeding strategy has developed in the benthic fauna in different European waters. However, due to uneven sampling effort (differences in detailed species sampling) and the species-specific dynamics of isotope incorporation, the delineated species distribution among trophic levels (Fig. 5) provides only a rough proxy of hierarchical food web patterns. It cannot be excluded either that the observed *FCL*s could also be affected by ecosystem size (Vander Zanden and Fetzer, 2007) and the degree of environmental perturbations (Pimm, 1982).

4.4. Diversity of benthic communities vs. food chain length

Regression analyses of selected variables of macrobenthic communities and food web attributes revealed a positive correlation of species richness (*S*) and Shannon-Weaver diversity index (*H'*) with food chain length. It is noteworthy that the relation *H'-FCL* was stronger than the relation *S-FCL*, highlighting differences between the effect of the total number of species and of dominants (so called keystone species) on food web architecture. Since *H'* is biased toward the more dominant species which certainly participate more in the energy transfer (because of higher abundance and biomass) and the more rare community members (so called week interactors) contribute little to the Shannon-Weaver diversity index,

the *H'-FCL* relationship indicates strong trophic links of keystone taxa and their importance in food web shaping. This finding contrasts with models accounting for differences in interaction strength which imply that weak interactors play a crucial role by decreasing variations between consumers and resources and thus contribute to maintaining food web complexity and enhancing community stability (Benedetti-Cecchi., 2000).

Food chain length increases with biodiversity, indicating to the possible positive effect of community diversity and abundance on the number of links in the food chain from basal to top predatory species. High species number induces high functional complexity of the communities, where some functional groups (based on, for example, feeding type or mobility) are represented by many species, resulting in diverse cycles of energy and matter. In more diverse benthic assemblages energy is transferred through more trophic pathways which contain more species. Consequently, a larger number of species contained within different trophic levels implies more complex trophic interactions, increasing the stability of the food web - defined as the ability to withstand temporal variability/perturbations in a system. According to the stability-complexity theory, a high number of species and the resultant higher trophic connectance have a stabilizing effect on the ecosystem (MacArthur, 1955). Conversely, less diverse communities based on a small number of species (e.g. in habitats subject to disturbances, for example on oxygen-deficient bottom of the southern Baltic) sustain a simple food web structure with weak and rare interactions. Such communities show low redundancy (low internal trophic connections) and are considered to be more vulnerable to external perturbations which affect the trophic transfers between organisms (Baird et al., 2007; Scharler and Baird, 2005). This is consistent with the hypothesis on the limiting effect of simple community organisation and a high degree of disturbances on food chain length (Pimm, 1982; Post, 2002). However, identification which trophic interactions or which species are most critical to the stability of food webs in a system requires quantitative description of the flows of materials and energy between food web elements (Raffaelli, 2000). This modern energetic concept is based on estimates of biomass of all biocenotic compartments and turnover rates (rates of processes involved in organic matter trophic transformation) (Rooney et al., 2006) and uses energy (Joules) or elements (e.g. carbon, nitrogen) as the common currency for expressing the magnitude of particular connections. Another explanation of the causality of the observed relations between *H'* and *FCL* can include availability and quality of resources in a system. More divergent and more available organic matter resources in a system are expected to support more energy pathways. Similarly, Rooney et al. (2006) showed that structurally complex ecosystems depend on the

1   maintenance of the heterogeneity of distinct energy channels. In case of the use of the stable

2   isotope approach, however, this theory might be biased by incomplete topological webs (the

3   lack of species) and imperfect definitions of trophic links. A higher resolution study allowing

4   for species-specific diet composition of more biocenotic components and trophic interaction

5   within the entire food web, including small-bodied organisms and mobile predators (e.g. birds

6   and mammals), is required to define the main directions of carbon flows in a more detailed

7   manner.

8

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10   5. Conclusions

11   Large-scale European study of food web structure within macrobenthic communities

12   across a range of environmental parameters and biodiversity levels revealed important

13   patterns of vertical organisation of trophic links. Food chain length increases with species

14   richness and biodiversity (expressed by the Shannon-Wiener diversity index), highlighting the

15   importance of taxonomical diversity for food web organisation. In species-rich and diverse

16   communities energy is transferred through more trophic levels and along more numerous

17   trophic pathways which utilise subsequently more diverse carbon sources in a system.

18   Communities of a higher biodiversity level develop more complex trophic interactions which

19   support high stability of the food web. In contrast, benthic communities with a small number

20   of species support weak trophic connections and simple food web structure of a relatively

21   short food chain length that makes them more susceptible to external perturbations.

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References

Baird, D., Asmus, H., Asmus, R., 2007. Trophic dynamics of eight intertidal communities of the Sylt-Rømø Bight ecosystem, Northern Wadden Sea. Marine Ecology Progress Series 351, 25-41.

Benedetti-Cecchi, L., 2000. Variance in ecological consumer-resource interactions. Nature 407, 370-374.

Bonsdorff, E., 2006. Zoobenthic diversity-gradients in the Baltic Sea: continuous post-glacial succession in a stressed ecosystem. Journal of Experimental Marine Biology and Ecology 330, 383-391.

Boutton, T.W., 1991. Stable carbon isotope ratios of natural materials; I. Sample preparation and mass spectrometric analysis. In: Coleman, D.C., Fry, B. (Eds.), Carbon Isotope Techniques. Academic Press, San Diego, pp. 155-171.

Briand, F., Cohen, J.E., 1987. Environmental correlates of food chain length. Science 238, 956-960.

Carlier, A., Riera, P., Amoroux, J.M., Bodiou, J.-Y., Grémare, A., 2007. Benthic trophic network in the Bay of Banyuls-sur-Mer (northwest Mediterranean, France): An assessment based on stable carbon and nitrogen isotopes analysis. Estuarine Coastal Shelf Science 72, 1-15.

Cohen, J.E., Briand, F., Newman, C.M., 1986. A stochastic theory of community food webs. III. Predicted and observed lengths of food chains. Proceedings of the Royal Society B 228, pp. 317-353.

Craig, H., 1953. The geochemistry of the stable carbon isotopes. Geochimica et Cosmochimica Acta 3, 53-92.

Créach, V., Schricke, M.T., Bertru, G., Mariotti, A., 1997. Stable isotope and gut analyses to determine feeding relationships in saltmarsh macroconsumers. Estuarine Coastal Shelf Science 44, 599-611.

Duffy, J.E., Cardinale, B.J., France, K.E., McIntyre, P.B., Thébault, E., Loreau, M., 2007. The functional role of biodiversity in food webs: Incorporating trophic complexity. Ecology Letters 10, 522-538.

Eleftheriou, A., McIntyre, A.D., 2005. Methods for the study of marine benthos. 3rd edition. Blackwell Science, Oxford, 418 pp.

1   Elmgren, R., Hill, C., 1995. Ecosystem function at low biodiversity - the Baltic example. In:

2   Ormond, R.F.G., Gage, J.D., Angel, M.V. (Eds.), Marine Biodiversity Patterns and

3   Processes. Cambridge University Press, Cambridge. pp 319-336.

4   Elser, J.J., Hessen, D.O., 2005. Biosimplicity via stoichiometry: evolution of food-web

5   structure and processes. In: Belgrano, A., Scharler, U.M., Dunne, J., Ulanowicz, R.E.

6   (Eds.), Aquatic Food Webs. An Ecosystem Approach. Oxford University Press, New

7   York, pp. 7-18.

8   Fredriksen, S., 2003. Food web studies in a Norway kelp forest based on stable isotope (δ^{13}C

9   and δ^{15}N) analysis. Marine Ecology Progress Series 260, 71-81.

10   Grall, J., Le Loch, F., Guyonnet, B., Riera, P., 2006. Community structure and food web

11   based on stable isotopes (δ^{15}N and δ^{13}C) analysis of a North Eastern Atlantic maerl bed.

12   Journal of Experimental Marine Biology and Ecology 338, 1-15.

13   Hastings, A., Conrad, M., 1979. Length and evolutionary stability of food chains. Nature 140,

14   799-814.

15   Hecky, R.E., Hesslein, R.H., 1995. Contributions of benthic algae to lake food webs as

16   revealed by stable isotope analysis. Journal of the North American Benthological Society

17   14, 631-653.

18   Heip, C., Warwick, R., d'Ozouville, L., 1998. European Marine and Polar Science (EmaPS).

19   A European Science Plan on Marine Biodiversity. ESF (European Science Foundation)

20   Marine Board Report, Strasbourg, 48 pp.

21   Herman, P.M.J., Middelburg, J. J., Widdows, J., Lucas, C.H., Heip, C.H.R., 2000. Stable

22   isotopes as trophic tracers: Combining field sampling and manipulative labelling of food

23   resources for macrobenthos. Marine Ecology Progress Series 204, 79-92

24   Hobson, K.A., Welch, H.E., 1992. Determination of trophic relationships within a high Arctic

25   marine food web using δ^{13}C and δ^{15}N analysis. Marine Ecology Progress Series 84, 9-18.

26   Hobson, K.A., Ambrose, Jr. W.G., Renaud, P.E., 1995. Sources of primary production,

27   benthic-pelagic coupling, and trophic relationships within the Northeast Water Polynya:

28   insights from δ^{13}C and δ^{15}N analysis. Marine Ecology Progress Series 128, 1-10.

29   Kaehler, S., Pakhomov, E.A., McQuaid, C.D., 2000. Trophic structure of the marine food web

30   at the Prince Edward Islands (Southern Ocean) determined by δ^{13}C and δ^{15}N analysis.

31   Marine Ecology Progress Series 208, 13-20.

32   Lebreton N., 2009, Analyse de la structure et du fonctionnement du reseau trophique d'un

33   herbier par approche multi-traceurs: traçage isotopique naturel et profils acides gras.

34   Thesis, University of La Rochelle, La Rochelle, France, 246 pp., (in French), unpublished

1   Lesutienė, J., 2009. Food web of the Curonian Lagoon: organic matter sources and feeding of

2   mysids. Ph.D. Thesis, Klaipeda University, Klaipeda, Lithuania, 145 pp., unpublished

3   Loreau, M., 2000. Biodiversity and ecosystem functioning: recent theoretical advances. Oikos

4   91, 3-17.

5   MacArthur, R.H., 1955. Fluctuation of animal populations and a measure of community

6   stability. Ecology 36, 533-536.

7   Martinez del Rio, C., 2007. Isotopic ecology. In: Karasov, W.H., Martinez del Rio, C. (Eds.),

8   Physiological Ecology: How Animals Process Energy, Nutrients and Toxins. Princeton

9   University Press, Princeton, pp. 433-478.

10   McCutchan, Jr. J.H., Lewis Jr., W.M., Kendall, C., McGrath, C.C., 2003. Variation in trophic

11   shift for stable isotope ratios of carbon, nitrogen and sulfur. Oikos 102, 378-390.

12   Monbet, Y., 1992. Control of phytoplankton biomass in estuaries: A comparative analysis of

13   microtidal and macrotidal estuaries. Estuaries 15, 563-571.

14   Olive, P.J.W., Pinnegar, J.K., Polunin, N.V.C., Richards, G., Welch, R., 2003. Isotope

15   trophic-step fractionation: a dynamic equilibrium model. Journal of Animal Ecology 72,

16   608-617.

17   Ostrom, P.H., Fry, B., 1993. Sources and cycling of organic matter within modern and

18   prehistoric food webs. In: Engel, M.H., Macko, S.A. (Eds.), Organic Geochemistry:

19   Principles and Applications. Plenum Press, New York, pp. 785-797.

20   Parker, P.L., 1964. The biogeochemistry of the stable isotopes of carbon in a marine bay.

21   Geochimica et Cosmochimica Acta 28, 1155-1164.

22   Pielou, E.C., 1966. The measurement of diversity in different types of biological collections.

23   Journal of Theoretical Biology 13, 131-144.

24   Pimm, S.L., 1982. Food Webs. Chapman & Hall, London, 219 pp.

25   Post, D.M., 2002. Using stable isotopes to estimate trophic positions: models, methods, and

26   assumptions. Ecology 83, 703-718.

27   Raffaelli, D., 2000. Trends in research on shallow water food webs. Journal of Experimental

28   Marine Biology and Ecology 250, 223-232.

29   Renaud, P.E., Ambrose, W.G., Vanreusel, A., Clough, L.M., 2006. Nematode and

30   macrofaunal diversity in central Arctic Ocean benthos. Journal of Experimental Biology

31   and Ecology 330, 297–306.

32   Renaud, P.E., Webb, T.J., Bjørgesæter, A., Karakassis, I., Kędra, M., Kendall, M.A., Labrune,

33   C., Lampadariou, N., Somerfield, P.J., Włodarska-Kowalczuk, M., Vanden Berghe, E.,

34   Claus, S., Aleffi, I.F., Amouroux, J.M., Bryne, K.H., Cochrane, S.J., Dahle, S., Degraer,

S., Denisenko, S.G., Deprez, T., Dounas, C., Fleischer, D., Gil, J., Grémare, A., Janas, U., Mackie, A.S.Y., Palerud, R., Rumohr, H., Sardá, R., Speybroeck, J., Taboada, S., Van Hoey, G., Węsławski, J.M., Whomersley, P., Zettler, M.L., 2009. Continental-scale patterns in benthic invertebrate diversity: insights from the MacroBen database. Marine Ecology Progress Series 382, 239-252.

Riera, P., 1998. $\delta^{15}N$ of organic matter sources and benthic invertebrates along an estuarine gradient in Marennes-Oléron Bay (France): implications for the study of trophic structure. Marine Ecology Progress Series 166, 143-150.

Riera, P., Stal, L.S., Nieuwenhuize, J., Richard, P., Blanchard, G., Gentil, F., 1999. Determination of food sources for benthic invertebrates in a salt marsh (Aiguillon Bay, France) by carbon and nitrogen stable isotopes: importance of locally produced sources. Marine Ecology Progress Series 187, 301-307.

Rooney, N., McCann, K., Gellner, G., Moore, J.C., 2006. Structural asymmetry and the stability of diverse food webs. Nature 442, 265-269.

Scharler, U.M., Baird, D., 2005. A comparison of selected ecosystem attributes of three South African estuaries with different freshwater inflow regimes, using network analysis. Journal of Marine Systems 56, 283-308.

Shannon, C.E., Weaver, W., 1963. The Mathematical Theory of Communication. University of Illinois Press, Urbana, 117 pp.

Sokal, R.R., Rohlf, F.J., 1995. Biometry. W.H. Freeman and Company, New York, 887 pp.

Sokołowski, A., 2009, Tracing the flow of organic matter based upon dual stable isotope technique, and trophic transfer of trace metals in benthic food web of the Gulf of Gdańsk (the southern Baltic Sea). Wydawnictwo Uniwersytetu Gdańskiego, Gdańsk, 213 pp.

Tamelander, T., Renaud, P.E., Hop, H., Carroll, M.L., Ambrose Jr., W.G., Hobson, K., 2006. Trophic relationships and pelagic-benthic coupling during summer in the Barents Sea Marginal Ice Zone revealed by stable carbon and nitrogen isotope measurements. Marine Ecology Progress Series 310, 33-46.

Tamelander, T., Reigstad, M., Hop, H., Carroll, M., Wassmann, P., 2008. Pelagic and sympagic contribution of organic matter to zooplankton and vertical export in the Barents Sea marginal ice zone. Deep-Sea Research 55, 2330-2339.

Vander Zanden, M.J., Rasmussen, J.B., 2001. Variation in $\delta^{15}N$ and $\delta^{13}C$ trophic fractionation: implications for aquatic food web studies. Limnology and Oceanography 46, 2061-2066.

Vander Zanden, M.J., Fetzer, W.W., 2007. Global patterns of aquatic food chain length. Oikos 116, 1378-1388.

1   Vander Zanden, M.J., Shuter, B.J., Laster, N., Rasmussen, J.B., 1999. Patterns of food chain

2   length in lakes: stable isotope study. American Naturalist 154, 406-416.

3   Vanderklift, M.A., Ponsard, S., 2003. Sources of variation in consumer diet $\delta^{15}N$ enrichment: a

4   meta analysis. Oecologia 136, 169-182.

5   Warwick, R.M., Emblow, C., Féral, J.P., Hummel, H., van Avesaath, P., Heip, C., 2003.

6   European Marine Biodiversity Research Sites. NIOO-CEME, Yerseke, 136 pp.

7   Włodarska-Kowalczuk, M., Sicinski, J., Gromisz, S., Kendall, M.A., Dahle, S., 2007. Similar

8   soft-bottom polychaete diversity in Arctic and Antarctic marine inlets. Marine Biology

9   151, 607-616.

10   Wołowicz, M., Sokołowski, A., Lasota, R., 2007. Estuaries-a biological point of view.

11   Oceanological and Hydrobiological Studies 36, 113-130.

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Table 1. Characteristics of sampling sites used for comparative analyses of macrobenthic community structure and food web attributes in different ecosystems.

| Water-basin | Location | Site | Site abbreviation | Geographic position | | Salinity | Tides* | Sampling | | Type of habitat | BIOMARE site |
				Latitude	Longitude			season	depth[m]		
North Atlantic	Hornsund fjord	Hornsund	Hornsund	77°02.36'N	16°00.62'E	29.4	+	summer	60.0	muddy sand	ATBI, NFS
Barents Sea	open water,	Barents Sea, St 4	St 4	77°01.09'N	29°29.19'E	35.0	+	summer	222.0	muddy sand	
	Marginal Ice Zone	Barents Sea, St 16	St 16	77°05.20'N	28°33.00'E	35.0	+	spring	206.0	muddy sand	
Baltic Sea	the Gulf of Gdańsk	Puck Bay	Puck	54°41.05'N	18°31.75'E	6.2-7.0	-	spring	7.0	sandy bottom	LTBR
		site Gdansk	Gdansk	54°37.50'N	18°37.50'E	6.5-7.2	-	spring	30.0	muddy bottom	
	Curonian Lagoon	Curonian Lagoon-mussel bed	Curonian-mb	55°21.02'N	21°09.58'E	0.0-7.2	-	summer	0.7-3.4	mussel beds	NFS
		Curonian Lagoon-bare sediments	Curonian-bs	55°31.80'N	21°07.46'E	0.0-7.2	-	summer	0.7-2.3	bare sediments	
North Sea	Sylt-Rømø Bight	Sylt-Rømø Bight	Sylt-Rømø	55°00'N	8°25'E	28.0-32.0	+	spring	4.2**	sandy bottom	LTBR
	Westerschelde	Molenplaat	Molenplaat	51°26'N	3°57'E	20.0-25.0	+	autumn	0**	mudflat	
Bay of Biscay	Marennes-Oleron Bay	Brouage	Brouage	45°54.88'N	1°06.05'W	25.0-35.0	+	autumn	0	mudflat	
		Oleron *Zostera* bed-lower flat	Oleron *Zostera*-lf	45°55.05'N	1°12.25'W	28.0-34.5	+	spring	0	*Zostera* bed-lower flat	
		Oleron *Zostera* bed-upper flat	Oleron *Zostera*-uf	45°54.63'N	1° 12.88'W	28.0-34.5	+	spring	0	*Zostera* bed-upper flat	
Mediterranean Sea	Bay of Banyuls-sur-Mer	Banyuls	Banyuls	42°29.77'N	03°8.76'E	37.2	-	spring	17.4-34.0	well-sorted fine sand	LTBR

* "+" tidal area, "-" non-tidal area; ** intertidal zone

ATBI – All Taxon Biodiversity Inventory site, NFS – Normal Focal Site, LTBR – Long-Term Biodiversity Research site

Table 2. Characteristics of macrobenthic communities at sampling sites in European waters (in brackets – percentage contribution of a given class to the total number of species). Classes are arranged in order of a decreasing total number of species.

| Site | Total abundance (ind. m^{-2}) | No species | | | | | | Total no classes | H' | J | source |
		total	Polychaeta	Crustacea	Bivalvia	Gastropoda	Clitellata				
Hornsund	6040	48	20 (41.7)	3 (6.3)	13 (27.1)	4 (8.3)		7	2.35	0.61	own data
St 4	5550	166	69 (41.6)	46 (27.7)	11 (6.6)	7 (4.2)		21	3.26	0.64	own data
St 16	5328	166	71 (42.8)	44 (26.1)	11 (6.6)	9 (5.4)		19	1.61	0.32	own data
Puck	1133	11	2 (18.2)	2 (18.2)	4 (36.4)			4	1.46	0,66	a
Gdansk	63	3	1 (33.3)		1 (33.3)			3	0.71	0.64	a
Curonian-mb	5014	29	2 (6.9)	5 (17.2)	4 (13.8)	6 (20.7)	5 (17.2)	10	2.02	0.60	b
Curonian-bs	3913	23	2 (8.7)	4 (17.4)	4 (17.4)	3 (13.0)	4 (17.4)	10	1.76	0.56	b
Sylt-Rømø	71704	25	11 (44.0)	6 (24.0)	4 (16.0)	2 (8.0)		6	1.07	0.33	own data
Molenplaat	11069	19	8 (42.1)	5 (26.3)	3 (15.8)	1 (5.3)		6	2.15	0.73	c
Brouage	34517	30	11 (36.7)	4 (13.3)	9 (30.0)	3 (10.0)		7	1.57	0.46	own data
Oleron *Zostera*-lf	7152	27	10 (37.0)	6 (22.2)	5 (18.5)	3 (11.1)		7	0.49	0.15	own data
Oleron *Zostera*-uf	15701	32	11 (34.4)	4 (12.5)	9 (28.1)	2 (6.3)		9	0.26	0.08	own data
Banyuls	10928	128	57 (44.5)	21 (16.4)	27 (21.1)	5 (3.9)		12	2.86	0.59	own data

a) Sokołowski, 2009 b) Daunys, unpubl. data c) Herman, unpubl. data

Table 3. Number of species analysed for carbon and nitrogen stable isotope composition and their percentage contribution to species richness and the total abundance of benthic faunal communities along with selected features of food webs in different systems of European waters.

site	No species	Contribution to species richness	abundance	Nitrogen isotope baseline	FCL*	Principal organic matter source	source
Puck	14	66.7	78.8	8.4	3.64	SPOM, SOM	a
Gdansk	3	33.3	73.7	9.4	3.08	SPOM, SOM	a
Curonian-mb	18	24.1	72.5	7.0	4.86	SPOM	b
Curonian-bs	12	17.4	47.1	7.2	3.92	SPOM	b
Sylt-Rømø	28	60.0	84.6	10.9	3.65	SPOM, MPB**, macroalgae, vascular plants	own data
Molenplaat	16	63.2	87.4	16.9	4.12	SPOM, SOM	c
Brouage	32	60.0	98.7	8.3	4.12	SPOM, SOM, MPB	this study
Oleron *Zostera*-lf	30	48.1	97.7	6.8	3.64	MPB, *Zostera* detritus	d
Oleron *Zostera*-uf	30	40.6	98.9	7.2	3.47	MPB, SOM, *Zostera* detritus	d
Banyuls	121	20.3	14.6	2.7	4.75	SPOM, SOM	e
St 4	29	8.5	6.2	6.6	4.30	iceSPOM	f, g
St 16	29	4.2	1.3	9.6	4.38	iceSPOM, SPOM	f, g
Hornsund	21	22.9	58.4	6.0	4.46	SPOM, SOM	this study

* FLC – food chain length, ** MPB – microphytobenthos.

a) Sokołowski, 2009 b) Lesutienė, 2009 c) Herman et al., 2000 d) Lebreton, 2009 e) Carlier et al., 2007 f) Tamelander et al., 2006 g) Tamelander et al., 2008

Figure captions

Fig. 1.

Fig. 2.

Fig. 3.

Fig. 4.

Fig. 4. cont.

Fig. 5.