

Histopathological lesions and DNA adducts in the liver of European flounder (*Platichthys flesus*) collected in the Seine estuary versus two reference estuarine systems on the French Atlantic coast

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Abstract:

An epidemiological survey was conducted in the Seine estuary and in two smaller and relatively preserved estuaries on the French Atlantic coast in order to estimate the occurrence of liver lesions in European flounder, *Platichthys flesus*, and also to seek putative risk factors for the recorded pathologies. Four hundred and seventy-eight fish of both sexes and of different size ranges were sampled in the three studied areas, 338 of which in the Seine estuary. All fish were examined for histopathological liver lesions, while DNA adducts and otoliths were analyzed on a subsample. Five categories of hepatic lesions were recorded with the following prevalence for the Seine estuary: 36.7 % inflammations, 8 % parasites (mainly encysted nematodes), 6.5 % foci of cellular alteration (FCA), 5.3 % foci of necrosis or regeneration (FNR), and 1.5 % tumors. Inflammation occurrence increased according to age, contrary to parasitic infestations and FCA which were more prevalent in young fish, notably those of <1 year old (group 0). Tumors were only observed in females of more than two winters. Females exhibited a higher prevalence of tumors (3.0 %) and FCA (6.5 %) than males (0 and 2.6 %, respectively). Parasitic and infectious lesions and FNR were equally distributed in males and females. The prevalence of FNR was also shown to vary according to sampling season, with significantly more occurrences of liver necrosis in the fish collected in summer than in spring. Spatial differences were observed with a higher occurrence of encysted parasites in flounders from the upper Seine estuary, while inflammations predominated in flounders living downstream. Temporal trends were also noted, with an increased prevalence of parasitic infestations, inflammations, and FCA in the 2002–2003 period in comparison to the 1996–1997 one. The three flounder populations from

the Seine estuary (Normandy), Ster estuary (Brittany), and Bay of Veys (Normandy) showed different spectra of hepatic lesions. Flounders from the Bay of Veys had relatively few liver lesions as compared to flounders from the two other estuaries. Flounders from the Ster estuary exhibited the highest prevalence of parasites (37.2 %) and inflammations (51.1 %). Finally, FCA and liver tumors occurred at very similar levels in both flounder populations from the Seine and the Ster estuaries. Group 0 flounders inhabiting the upper Seine estuary were more prone to parasitic and pre-neoplastic hepatic lesions and had higher levels of liver DNA adducts than the older ones living downstream. It was postulated that group 0 European flounders may serve as valuable bioindicators for assessing the quality of estuarine waters and the health status of euryhaline fish populations.

Keywords: Epidemiological study ; European flounder ; Seine estuary ; Liver histopathology ; Necrosis ; FCA ; Tumors ; Inflammations ; Parasites ; DNA adducts ; Pollution

Abbreviations:

OSPAR: Oslo-Paris convention on the protection of the marine environment of the North-East Atlantic.

DRZ: diagonal radioactive zone.

EPCV: European College of Veterinary Pathologists

1. Introduction

Estuarine and coastal marine areas play a crucial ecological role as nursery and/or feeding grounds for a large variety of aquatic species, notably fish (Beck et al. 2001). Since nearly 60 % of the human worldwide population lives and works within 100 miles of a coast (Hinrichsen 1998), these areas are even more impacted by human activities and, particularly, by chemicals. It is now currently accepted that sediments from estuaries and coastal areas are major reservoirs for a large variety of persistent organic or inorganic pollutants and secondary

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4 74 sources of pollution for aquatic ecosystems (Harris et al., 1996). An increase in the occurrence
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6 75 of toxicopathic lesions and/or infectious or parasitic diseases in aquatic organisms chronically
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8 76 exposed to pollutants has extensively been reviewed in recent years (Harmon and Wiley,
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10 77 2010; Mearns et al., 2011). Increased susceptibility to pathologies can directly or indirectly
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12 78 impair the survival or biotic performances of individual organisms which, in turn, can affect
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14 79 the abundance, age structure, genetic diversity, and the reproduction of wild populations
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16 80 (Marchand et al., 2004; Benejam et al., 2010; Brooks et al., 2012).

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20 81 The liver of vertebrates, including fish, is the main target for toxicants because of its high
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22 82 vascularization, its high lipid content and its role in the organic xenobiotic biotransformation
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24 83 and metabolism of sex hormones (Hinton et al., 2001). It has been shown that fish exposure to
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26 84 organic and metallic pollutants can lead to a wide range of toxicopathic lesions, including
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28 85 tumors, FCA (foci of cellular alteration) and several non-neoplastic liver lesions (Hinton et
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30 86 al., 2001). In the field, a strong relationship was reported between toxicopathic liver lesions
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32 87 and environmental contamination by persistent organic pollutants including PAH and
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34 88 organochlorinated hydrocarbons (Myers et al., 1998; Harshbarger and Clark, 1990; Myers et
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36 89 al., 2003). For all these reasons, toxicopathic liver lesions in fish liver are considered as
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38 90 sensitive and integrative biomarkers of pollutant exposure (Hinton et al., 2001) and their
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40 91 integration in a pollution monitoring program is now recommended by the OSPAR
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42 92 international organization (SGIMC, 2011).

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48 93 The European flounder *Platichthys flesus* (L.) is a coastal and estuarine flatfish species which
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50 94 is widespread along the European coast, in the North Sea, the Atlantic Ocean and the
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52 95 Mediterranean Sea. Because of their benthic way of life, their bottom-feeding behavior and
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54 96 their relative longevity, flounders are particularly exposed to sediment-trapped pollutants. In
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56 97 addition, this species has been recommended by OSPAR for pollution monitoring in the
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58 98 North-East Atlantic.

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4 99 Several epidemiological studies focusing on liver pathologies in the European Flounder have
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6 100 already been conducted in the North Sea and the Baltic Sea along the British, German and
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8 101 Dutch coasts (Bogovski et al., 1999; Vethaak, 1992; Koehler, 2004; Vethaak et Wester, 1996;
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10 102 Stentiford et al., 2003; Lang et al., 2006). But no or little data is available for the French
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12 103 Atlantic coast. Furthermore, risk factors for certain liver lesions are not totally well-
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14 104 understood and need further investigation.

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18 105 In the present study, flounders were collected on the French Atlantic coast, in three different
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20 106 estuaries having specific features in terms of size, hydrological conditions, human pressures
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22 107 and water quality.

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25 108 The Seine estuary is a large (50 km²), man-altered, macrotidal ecosystem on the French
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27 109 Atlantic coast. The Seine River catchment covers about 79,000 km² and is one of the most
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29 110 urbanized and industrialized area in France with about 25% of the metropolitan population
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31 111 (16 million inhabitants), 40% of the national economic activity and 30% of the national
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33 112 agricultural activity. Due to the high pollutant inputs and a relatively low water flow (250 to
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35 113 900 mm³/s), PAHs and PCBs are currently detected at high levels in sediments (Cachot et al.,
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37 114 2006), suspended matters (Cailleaud et al., 2007) and biota (Minier et al., 2006; Rocher et al.,
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39 115 2006; Cailleaud et al., 2007). Furthermore, sediments from the upper Seine estuary were
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41 116 shown to contain potent mutagenic and carcinogenic pollutants (Cachot et al., 2006) and
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43 117 adverse health effects were reported in various invertebrates (Minier et al., 2006; Rocher et
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45 118 al., 2006; Cailleaud et al., 2009) and fish species (Marchand et al., 2004; Gilliers et al., 2006;
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47 119 Cachot et al., 2007; Amara et al., 2009).

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53 120 The Bay of Veys is a large and shallow estuary of about 37 km² located in the western part of
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55 121 the Seine bay (east coast of Cotentin). Three main rivers including the Douve, the Taute and
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57 122 the Vire are connected to this bay. The global river catchment of 3,500 km² is poorly
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4 123 urbanized (about 70 inhabitants/km²) showing mainly agricultural activity (cattle breeding
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6 124 and dairy farming).

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9 125 The Ster of Lesconil estuary is a small (0.36 km²) and pristine area located in Brittany with a
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11 126 water catchment of about 100 km². This river receives reduced domestic and agricultural
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13 127 inputs and no industrial waste, and the levels of organic pollutants and metals currently found
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15 128 in mussels and European flounders are low (Marchand et al., 2003; 2004).

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18 129 The main objective of the present work is to make a first survey of liver pathologies and DNA
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20 130 adduct levels in feral European flounders from these three contrasted estuaries. In addition,
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22 131 this study aims at analyzing, for each of the various flounder liver pathologies, the putative
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24 132 risk factors including biotic traits (age and sex) and abiotic parameters such as localization,
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26 133 season and year of collection.

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33 135 **2. Materials and Methods**

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39 137 **2.1 Sampling sites**

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41 138 The epidemiological survey was conducted in three estuaries along the French Atlantic coast:
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43 139 the Seine estuary (Normandy, France), the Bay of Veys (Normandy, France) and the Ster of
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45 140 Lesconil estuary (Brittany, France) (Figure 1A).

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48 141 The Seine estuary extends over 170 km from Poses to Le Havre (Cap de la Hève). Three
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50 142 distinct areas can be distinguished within the Seine estuary (Figure 1B): the upper Seine
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52 143 estuary from Poses (upstream limit of the dynamic tide) to Caudebec which is the fresh water
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54 144 part of the estuary, the middle Seine estuary from Caudebec to Honfleur with mesohaline
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56 145 water, and the lower Seine estuary from Honfleur to the Seine Bay characterized by salt water
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58 146 and a high tidal range. A fourth area, named Antifer, localized in the Eastern part of the Seine

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4 147 bay along the Pays de Caux but under the direct influence of the Seine river plume was also
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6 148 sampled (Figure 1B).

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10 150 **2.2 Fish sampling and liver collection**

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12 151 Juvenile and adult European flounders (*Platichthys flesus*, L.) of both sexes were collected by
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14 152 trawling or netting in the course of 15 fishing campaigns between April 1996 and May 2004
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16 153 (Table 1). Immediately after sampling, alive fish were transferred to tanks filled with
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18 154 recirculated water prior to dissection. Amongst the 1505 captured fish, 478 fish were
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20 155 randomly selected (the first twenty individuals for each sampled sites or all the captured fish
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22 156 if less than twenty) for liver histopathological analysis. The fish had their spinal cord cut and
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24 157 abdominal cavity carefully opened with dissecting scissors, in order to have the integrity of
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26 158 their internal organs preserved. The fish was sized (total body length), sex was determined,
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28 159 and the biggest otolith (the sagittal) was recovered from randomly selected fish for age
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30 160 determination. The whole liver of small fish (up to 110 mm in length) was sampled for
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32 161 histological examination, while for bigger individuals, 1 cm³ slice encompassing the whole
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34 162 tissue thickness and width was collected using clean razor blades. The liver samples were
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36 163 immediately fixed in 10% formalin buffer solutions. DNA adducts were measured in the liver
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38 164 of flounders collected in the Seine estuary in September and October 1996 and in the Bay of
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40 165 Veys in April 1997. A small piece of liver of about 1 cm³ was recovered from 10 to 15
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42 166 individual fish of both genders and within the same size range. Immediately after collection,
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44 167 liver samples were pooled in a clean DNase-free microtube and deep-frozen in liquid
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46 168 nitrogen. Samples were then stored at -80°C prior to analysis.
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56 170 **2.3 Age determination**

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4 171 The biggest otolith (the sagittal) was removed from 87 randomly selected fish of different size
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6 172 and sex which had been caught in the Seine estuary. Otoliths were immersed in water and
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8 173 examined with a stereomicroscope (Leica Wild M8) under transmitted light. Translucent
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10 174 annuli corresponding to winter periods were counted.
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15 176 **2.4 Liver histopathology**

17 177 Formalin-fixed tissues were dehydrated by transferring them through a series of alcohols of
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19 178 increasing concentrations (80%, 2*1 hour, 35°C, 2 times; 95%, 2*1 hour, 35°C, 2 times) up to
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21 179 100% alcohol (4*1 hour, 4 times). Next, they were placed into methylcyclohexan which is
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23 180 miscible with both 100% alcohol and paraffin (methylcyclohexan, 3*1 hour, 35°C, 3 times).
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25 181 Then they were put into melted paraffin (3*1 hour, 58°C, 3 times). All these operations were
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27 182 performed using an Automat Tissue-Tek VIP 3000. The samples were then embedded in
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29 183 paraffin wax (Automat Tissue-Tek TEC-5) and sliced with a microtom Reichert-Jung 2030
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31 184 into 4-µm thick sections. The sections were stained using a routine hematoxyllin-eosin-
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33 185 saffron staining method (Automat Sakura DRS601). A representative section of each sample
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35 186 was examined in a double-blind manner by means of light microscopy by one of the ECVP-
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37 187 pathologists (European college of veterinary pathologists, either authors #2 and/or #3 because
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39 188 of the long-term study), but the critical slides were observed by the two pathologists. All
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41 189 lesions were assigned to five main categories recommended for monitoring the biological
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43 190 effects of contaminants in flatfish species: parasitic lesions, inflammations,
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45 191 necrosis/regeneration foci, foci of cellular alteration and neoplasms (Feist et al., 2004).
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47 192 Melano-macrophage aggregates were counted in ten optical fields at high magnification
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49 193 (x400), and only more than 2 aggregates for ten optical fields were considered as an abnormal
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51 194 accumulation.
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4 196 2.5 DNA adduct analysis

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6 197 The livers of 10 to 15 individuals were pooled, and high molecular weight DNA was
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8 198 extracted using the chloroform/isoamyl alcohol method. Livers were homogenized in 0.8 mL
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10 199 of a solution containing NaCl (0.1 M), EDTA (20 mM), and Tris-HCl, pH 8 (50 mM) (SET).
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12 200 One hundred μ L of SDS (20%) was added to the homogenate, and following incubation for
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14 201 10 min at 65°C, 800 μ L of potassium acetate (6M, pH5) was added. Then the reaction mixture
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16 202 was kept at 0°C for 30 min. After centrifugation for 25 min at 0°C (10000g), the supernatant
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18 203 was collected and its nucleic acid content was precipitated overnight at -20°C by adding 2
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20 204 volumes of cold ethanol. DNA pellets were collected, washed once with 1mL of 90% ethanol
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22 205 and dissolved in 500 μ L of SET (15 min at 37°C). The total extract was mixed with 10 μ L of a
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24 206 mixture of RNase A (20 mg/mL) and RNase T1 (10 000 U/mL) and incubated for 1 h at 37
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26 207 °C; this treatment was repeated twice. Samples were then treated with 20 mg/mL of
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28 208 proteinase K for 1 h at 37 °C. After digestion, 500 μ L of Rotiphenol was added. The mixture
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30 209 was then moderately shaken for 20 min at room temperature and centrifuged for 15min at
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32 210 15°C (10000g). The aqueous phase was collected after two extractions. After a final
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34 211 extraction with one volume of chloroform/isoamyl alcohol (24:1), the aqueous phase was
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36 212 collected and 50 μ L of sodium acetate (3M, pH 6) was added. The DNA was precipitated by
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38 213 the addition of two volumes of cold ethanol overnight at -20°C followed by centrifugation at
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40 214 10000g for 30 min. The DNA pellet was washed four times with 90% ethanol. The purity of
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42 215 the DNA was checked by recording UV spectra between 220 and 320 nm.

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44 216 DNA adducts were measured using the ³²P-postlabeling method with Nuclease P1 treatment
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46 217 as described by Reddy and Randerath (1986) with minor modifications as follow. DNA (4 μ g)
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48 218 was digested at 37°C for 4h with micrococcal nuclease (500mU), spleen phosphodiesterase
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50 219 (105mU) buffered with sodium succinate (200mM), and calcium chloride (100mM, pH 6).
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52 220 The digested DNA was then treated with a mixture containing nuclease P1 (4mg/mL), ZnCl₂

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4 221 (1mM) and sodium acetate (0.5M, pH 5) at 37°C for 45min. The reaction was stopped by
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6 222 adding 3 µL of Tris base. The DNA adducts were labeled as follows. Five µL of the reaction
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8 223 mixture containing 2 µL of bicine buffer [Bicine (800 µM), dithiothreitol (400 mM), MgCl₂
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10 224 (400 mM), and spermidine (400 mM) adjusted to pH 9.8 with NaOH], 9.6U of polynucleotide
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12 225 kinase T4, and 100 µCi of [³²P]ATP (specific activity 6000Ci/mmol) was added to the NP1
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14 226 digest and incubated at 37°C for 45 min. Normal nucleotides, pyrophosphate, and excess ATP
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16 227 were removed by chromatography on PEI/cellulose TLC plates in 2.3M NaH₂PO₄ buffer, pH
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18 228 5.7, overnight. The origin areas containing labeled adducted nucleotides were cut out and
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20 229 transferred to another PEI/cellulose TLC plate, which was run in 5.3M lithium formate and
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22 230 8.5M urea (pH 3.5) for 3 h. A further migration was performed after turning the plate 90°
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24 231 anticlockwise in 1 M LiCl, 0.5M Tris and 8M urea (pH 8) for 2 h. Finally, the chromatogram
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26 232 was washed in the same direction in 1.7M NaH₂PO₄, pH 6, for 2 h (D4). Autoradiography
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28 233 was carried out at -80 °C for 48 h in the presence of an intensifying screen. Radioactive spots
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30 234 were detected by autoradiography on Kodak super X-Ray film.

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36 235 Quantification of DNA adducts was obtained by storage phosphor imaging techniques. The
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38 236 screens were scanned using a Typhoon 9210 (Amersham). The software program used to
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40 237 process the data was ImageQuant (version 5.0). After background subtraction, the levels of
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42 238 DNA adducts were expressed as relative adduct labeling (RAL) for 10⁹ nucleotides. The
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44 239 sensitivity of the method allows detection of B[a]P adduct as low as 0.1 nucleotide/ 10¹⁰
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46 240 nucleotides.

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52 53 242 **2.6 Statistical analysis**

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55 243 All statistics were performed with the SAS System release 9.3 (SAS Institute Inc., Cary, NC,
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57 244 USA). Categorical data was tested using a Fisher's exact test or a Cochran-Armitage trend
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59 245 test. Bonferroni's correction was applied for pairwise comparisons. The age and length

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246 relationship was investigated using linear regressions (Proc reg). Logistic regressions were
247 used in order to elucidate the implication of period, season, sex or geographic localization on
248 histological responses (Proc logistic). A p-value below 0.05 was considered as significant.

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250 **3. Results**

251 **3.1 Age of the sampled fish**

252 Age determination was performed by counting growth rings on otoliths from 87 flounders
253 collected in the Seine estuary. Most collected fish had one to four winters and only 3 out of 87
254 (3.4%) were much older (Figure 2).

255 In the first three years of fish life (2 winters), a clear age-dependent increase in total body
256 length was observed (Figure 2). For older fish (3 winters and more), body length was no more
257 related to age. In the present study, fish were ranked into four groups according to body
258 length and estimated age. Group 0 included fish up to 110 mm in length and less than one
259 year old. Group 1 included fish sized 115 up to 200 mm and aged from one to less than two
260 years old. Group 2 included individuals of 205 up to 300 mm and aged two to less than three
261 years old. Finally, group 3+ included all fish above 300 mm with an expected age of three
262 years and more.

263 The different groups were not equally distributed throughout the Seine estuary, with group 0
264 mainly located in the upper Seine estuary, group 1 in the upper and median estuary, group 2 in
265 the median and lower estuary, and finally group 3+ only represented in the lower estuary and
266 in the Seine bay.

267 None of the flounders sampled in the Bay of Veys and in the Ster estuary belonged to group 0.
268 Because of age-structure differences between the three flounder populations,
269 histopathological data from the Bay of Veys and Ster estuary were only used for inter-site
270 comparisons with the Seine estuary.

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3.2 Description of flounder liver pathologies

The liver of the European flounder is composed of parenchymal cells (hepatic cells) and clusters of pancreatic cells along the branches of the portal vein. A hepatic cell has a polygonal body containing a clear spherical nucleus, usually with a single nucleolus. Large quantities of lipids and glycogen are usually observed in the cytoplasm and cause the clear vacuolar appearance of the cell.

Lesions, encountered during the liver pathology assessment were categorized into five main items (Figure 3).

(i) Encysted parasites (mainly nematodes) are either surrounded by inflammatory cells, that are mainly macrophages forming a granuloma, or only by a thin fibrous capsule (Figure 3A). Depending on the sectioning level, the center of the granuloma and the causative parasite were not always observable (Figure 3B). Less frequently, amoeba infections are associated with ill-delineated foci of necrosis with an influx of macrophages (Figure 3D).

(ii) Inflammatory lesions are characterized by the accumulation of inflammatory cells (lymphocytes and macrophages) mainly observed at the vicinity of major blood vessels. Different grades of inflammations were observed, from a few accumulated melano-macrophage aggregates spread in the liver parenchyma to severe diffuse hepatitis. As the melano-macrophage aggregates can occur under normal conditions in flounder as in other fishes, only more than two aggregates for ten optical fields were considered as an abnormal accumulation (Figure 3C). Diffuse inflammations were very rarely observed and all inflammation types, except host reaction to parasites as described in (i), were jointly recorded.

(iii) Foci of necrosis and regeneration (FNR) were jointly recorded because they can both result from cell death following parasitic infestation, bacterial or viral infection and exposure to toxicants. Focal necrosis is characterized by small groups of hepatocytes exhibiting strong

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4 296 eosinophilic or pale vacuolated cytoplasm and hyperchromatic-condensed nuclei (Figure 3D).
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6 297 On rare occasions, some foci of small hepatocytes having a high mitosis index formed new
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8 298 trabeculae and were identified as regenerative foci.
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10 299 (iv) Foci of cellular alteration (vacuolated cell foci, clear cell foci, eosinophilic foci,
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12 300 basophilic foci or mixed eosinophilic and basophilic foci) are focal areas exhibiting different
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14 301 staining degrees: some altered staining behaviors can be eosinophilic or basophilic, and some
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16 302 others can be poorly stained (clear cell foci) (Figure 3E, 3F and 3G). These foci do not
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18 303 compress the surrounding parenchyma. Some of these lesions (basophilic foci) are considered
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20 304 as putative pre-neoplastic lesions (Myers et al., 1987).
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24 305 (v) Tumors were identified as hepatocellular or pancreatic, and as adenoma (Figure 3H) or
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26 306 adenocarcinoma (Figure 3I). Bile duct tumors were not observed herein. Benign as well as
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28 307 malignant tumors are characterized by cellular atypia associated with architectural
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30 308 abnormalities such as tubular, acinar or solid growth patterns. The tumor growth is associated
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32 309 with a compression (benign) or infiltration (malignant) of the surrounding tissue.
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311 3.3 *Prevalence of flounder liver pathologies according to age in the Seine estuary*

312 Granulomas with encysted parasites were observed in 8% (27/338) of sampled flounders in
313 the Seine estuary. Parasite prevalence declined significantly according to age group ($p < 0.001$,
314 Fisher's exact test), with a higher value in group 0 fish (27%) than in older ones (3.7%, 7.2%,
315 4.6% for groups 1, 2, 3+ respectively) (Figure 4A).

316 Nearly 37% of the sampled flounders (124/338) exhibited liver parenchyma inflammations.
317 Inflammation prevalence significantly differed according to age ($p < 0.001$, Fisher's exact test)
318 and was globally higher in old fish of three years and more (58.3%) than in younger ones
319 (22.2 to 32.4%) (Figure 4A).

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4 320 Foci of necrosis and regeneration (FNR) were recorded in 5.3% (18/338) of the sampled
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6 321 flounders. No significant difference was observed according to age group with prevalence
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8 322 varying from 8.1% for group 0 to 4.6% for group 3+ (Figure 4A).
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11 323 More than 6% (22/338) of all sampled flounders exhibited foci of cellular alteration (FCA).
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13 324 FCA were observed in all groups but with significant prevalence differences according to fish
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15 325 age ($p < 0.05$, Fisher's exact test). Indeed, FCA prevalence was at least three times higher in
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17 326 group 0 (18.9%) than in other groups (5.6%, 5.8% and 3.7% for group 1, 2 and 3+
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19 327 respectively) (Figure 4B).
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23 328 Hepatic tumors were recorded in 5 out of 338 flounders (1.5%). Adenoma occurrences were
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25 329 recorded but no adenocarcinoma was observed. Although not significant, an increasing trend
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27 330 of tumor prevalence was observed with age from 0% for age groups 0 and 1 to 0.7% and 3.7%
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29 331 for groups 2 and 3+ respectively (Figure 4B).
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34 35 333 **3.4 Prevalence of flounder liver pathologies according to gender in the Seine estuary**

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38 334 Prevalence of parasitic infestations, inflammations and FNR did not vary significantly
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40 335 according to sex (Figure 5A). Although the differences were not significant, FCA were 2.5
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42 336 times more frequent in females (6.5%) than in males (2.6%) and tumors were only observed
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44 337 in females (3.0%) (Figure 5B). Moreover, when FCA and tumors were scored together,
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46 338 females appeared significantly more affected (9.5%) than males (2.6%) ($p < 0.05$, Fisher's
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48 339 exact test).
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53 54 55 341 **3.5 Spatial distribution of flounder liver pathologies in the Seine estuary**

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58 342 Prevalence of parasitic infestation was significantly different ($p < 0.001$, Fisher's exact test)
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60 343 according to the sampling sites (Table 2). Indeed, the infestation prevalence that was recorded
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344 in flounders from the upper part of the Seine estuary (18.4%) was significantly higher than in
345 fish collected in the median Seine estuary (2.2%) or Antifer (2.3%). In contrast, prevalence of
346 inflammations, FNR and tumor did not show any spatial difference. Although FCA frequency
347 was particularly high in flounders from the upper Seine estuary (14.3%), no statistical
348 difference was observed between the different sampling sites.

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350 **3.6 *Seasonal variations of flounder liver pathologies in the Seine estuary***

351 Since flounders were sampled all the year round in the Seine estuary, it was possible to
352 analyze the seasonal variations of flounder liver pathologies. No significant seasonal variation
353 was observed for FCA, tumors, parasitic infestations and inflammations. Only FNR ($p < 0.05$,
354 Fisher's exact test) did show seasonal fluctuations (Figure 6A and 6B). In flounders from the
355 Seine estuary, FNR occurred more frequently in the summer (11.3%) than in spring (1.0%)
356 (Fisher's exact test, $p = 0.004$).

357

358 **3.7 *Temporal evolution of flounder liver pathologies in the Seine estuary***

359 The epidemiological survey of liver pathologies was conducted for all the sampled sites,
360 except for the lower Seine estuary, over two distinct periods in 1996-1997 and four years later
361 in 2002-2003. Comparisons of liver pathology profiles between the two sampling periods
362 showed significant differences, including for the 2002-2003 period, a 4-fold increase in
363 parasitic infestations ($p < 0.01$, Fisher's exact test), a 2-fold increase in inflammations
364 ($p < 0.001$, Fisher's exact test) and a 5-fold increase in FCA ($p < 0.01$, Fisher's exact test)
365 (Figure 7A and 7B). In contrast, FNR occurrence did not vary significantly over the two
366 sampling periods ($p \geq 0.05$, Fisher's exact test). For tumor prevalence, the temporal decrease

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4 367 was almost significant ($p=0.06$, Fisher's exact test) with 3.4% of flounders exhibiting liver
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6 368 tumors in 1996-1997 versus 0% in 2002-2003 (Figure 7B).

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11 370 ***3.8 Comparison of liver pathologies between flounder populations from the three studied***
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14 371 ***estuaries***

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16 372 Flounders were sampled over the same period in three estuaries along the French Atlantic
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19 373 coast. An attempt was made to compare the profiles of liver pathologies between the three
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21 374 flounder populations (Table 3). Since the age structure was different in the three studied
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23 375 populations with an underrepresentation of younger fish in the Bay of Veys and Ster estuary,
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26 376 only adult fish of more than 200mm (groups 2 and 3+) were selected for this analysis.
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28 377 Prevalence of parasitic infestations and inflammations was significantly different between the
29
30 378 three flounder populations ($p<0.001$, Fisher's exact test in both cases). When recorded
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33 379 together, tumors and FCA occurrence was also almost significantly different between
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35 380 estuaries ($p=0.053$, Fisher's exact test). Prevalence of infestations was significantly higher
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37 381 ($p<0.001$) in the Ster estuary (37.6%) in comparison to the Seine estuary (5.8%) and the Bay
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39 382 of Veys (6.5%). Flounders from the Ster and the Seine populations were more prone ($p<0.01$)
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41 383 to liver inflammations (51.8% and 40.3% respectively) than individuals from the Bay of Veys
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43 384 (15.2%) (Table 3). Flounders from the Ster and the Seine estuaries in contrast to those from
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45 385 the Bay of Veys exhibited FCA (9.4% and 4.9%, respectively) and tumors (1.2 and 2.1%
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47 386 respectively). For FNR occurrence, no significant difference was noted for the three studied
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50 387 populations.

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56 389 ***3.9 Relationship between the different flounder liver pathologies***
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390 A strong positive correlation between parasitic infection and inflammation ($p < 0.001$, Fisher's
391 exact test) and a negative correlation between inflammation and FNR ($p < 0.001$, Fisher' exact
392 test) were observed. In fact 89% of fish from the Seine estuary bearing encysted parasites also
393 exhibited liver inflammation while only 25% of fish with FNR also showed inflammation. In
394 contrast, no obvious relationship was noted between the other liver lesions.

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396 ***3.10 DNA adducts in flounder liver***

397 DNA adducts were measured using the ^{32}P -postlabelling technique from pooled livers of
398 flounders collected in September 1996 in the Seine estuary and in April 1997 in the Bay of
399 Veys. Due to insufficient data available (two DNA adduct analyses per site), statistical
400 analysis could not be performed. Nevertheless some spatial trends could be observed. Firstly,
401 the total number of DNA adducts varied greatly from $21.1 \pm 8.0 \text{ RAL}/10^9$ nucleotides in the
402 Bay of Veys to $149 \pm 38.5 \text{ RAP}/10^9$ nucleotides in the upper Seine estuary (Figure 8).
403 Secondly, the DNA adduct profiles were totally different in the Seine estuary and in the Bay
404 of Veys. The total number of individual spots per site varied from 4 up to 13 in the Seine
405 estuary, and reached 43 in the Bay of Veys (Data not shown). In addition, amongst the 43
406 individual spots detected in the fish from the Bay of Veys, only 18 (42%) were shared with
407 flounders from the Seine estuary. Finally, most of the adducts (65 to 82% of RAL) detected in
408 flounders from the Seine estuary were located within the diagonal radioactive zone (DRZ)
409 (Figure 9) while all the DNA adducts detected in flounders from the Bay of Veys were
410 located outside the DRZ (Data not shown).

411 Some spatial trends can also be noted for the Seine estuary. A 10-fold decrease of total DNA
412 adducts in fish liver was observed between the upper Seine estuary and Antifer (Figure 8).
413 The total number of individual adducts declined from 13 spots in the upper Seine estuary to 4

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4 414 spots in the lower estuary and in Antifer, and no DRZ DNA adducts was detected in Antifer
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6 415 (Figure 9).

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10 11 417 **4 Discussion**

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14 418 In the present paper, pattern and occurrence of liver pathologies were investigated in
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16 419 European flounders from the Seine estuary and from two other reference estuaries on the
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18 420 French Atlantic coast. Implications of various biotic (age, sex) and abiotic (year, season,
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20 421 location) factors in the onset of the different liver pathologies were investigated. Liver DNA
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22 422 adduct levels were also measured to evaluate fish exposure to genotoxic pollutants.
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27 28 29 424 *4.1 Infectious and parasitic diseases*

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32 425 The most prevalent lesion type was inflammation of liver parenchyma, observed in nearly
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34 426 37% of the sampled fish. Inflammation foci and melano-macrophage aggregates were jointly
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36 427 recorded. Both lesions can be induced by multiple stress factors, including at least infectious
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38 428 and parasitic diseases and toxicant exposure (Wolf and Wolfe, 2005). This lesion type was
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40 429 more frequent in older fish but no sex-specific difference was observed. Vethaak and Wester
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42 430 did not show any relationship with fish gender or age (Vethaak and Wester, 1996). This
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44 431 apparent discrepancy could at least partially be explained by the fact that in the latter study
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46 432 only fish with a total length of 200mm or more, i.e. only fish of two winters or more were
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48 433 examined (Vethaak and Wester, 1996). In the present study, inflammations were slightly more
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50 434 frequently observed in late winter. Wethaak and Wester (1996) reported the same pattern with
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52 435 a higher occurrence of inflammatory lesions, including parasitic cysts, at the end of winter in
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54 436 flounders captured in spawning grounds. The authors hypothesized that poor conditions and
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56 437 spawning stress could favor bacterial and/or parasite infections.
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438 A clear site-specific profile was also reported with a higher prevalence of inflammatory
439 lesions in individuals from the Ster estuary than in those from the Seine estuary or the Bay of
440 Veys. Previous studies had already documented considerable variations of inflammation
441 occurrence according to geographical area and sampling period from about 12% in Dutch
442 coastal and estuarine waters (Vethaak and Wester, 1996) to 30% in coastal areas of the Baltic
443 sea (Lang et al., 2006) (Table 4).

444 The second most prevalent lesion type was the encysted parasite lesion observed in 13% of all
445 sampled flounders (8% in the Seine estuary). Only a few studies have reported parasitic cysts
446 in flounder liver (Vethaak, 1992; Vethaak and Wester, 1996; Lang et al., 2006; Dezfuli et al.,
447 2007). At least one larval and five adult nematode species have already been identified in
448 flounders (El-Darsh and Whitfield, 1999). In contrast to inflammatory lesions, parasitic cysts
449 occurred with a higher prevalence in flounders of group 0. This feature could putatively be
450 explained by an age-dependent sensitivity of European flounder to nematode infestation and
451 also by heterogeneous spatial distribution of the parasite in favor of the oligohaline or
452 mesohaline part of the estuary. Significant spatial and temporal differences were also
453 evidenced with a similar distribution pattern to that of liver inflammations. Indeed, prevalence
454 of parasitic cysts was higher in the Ster estuary than in the Seine estuary or the Bay of Veys.
455 Moreover, encysted parasite prevalence was slightly higher in flounders collected in late
456 winter.

457 The higher frequency of inflammatory lesions and parasitic cysts in the flounder population
458 inhabiting the Ster is probably not related to a significant chemical stress because pollutant
459 inputs in this estuary are low (Marchand et al., 2003,2004). This result could more likely be
460 explained by the localization of this estuary in the southern part of the flounder distribution
461 area and by a possible thermal stress. This thermal stress could directly affect the immune
462 system or the global health condition of the fish, leading to an increased sensitivity to

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4 463 infectious agents or parasites. It is noteworthy that relatively high frequencies of flounder
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6 464 liver lesions were detected in the Gulf of Biscay, as compared to English coasts and the North
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8 465 Sea (Laroche et al., 2012).
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14 467 ***4.2 Toxicopathic liver lesions***

16 468 Necrosis and regeneration foci were observed in 4.2% of the sampled fish (5.3% in the Seine
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18 469 estuary). No significant difference according to age or gender was observed. This lesion type
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20 470 did not show any significant inter-site variations within the Seine estuary or between the three
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22 471 studied estuaries. In contrast, a clear seasonality was observed with a drastic increase in FNR
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24 472 prevalence in the summer. Seasonal variations of hepatocellular necrosis have already been
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26 473 documented for winter flounders (*Pleuronectes americanus*) sampled along the US Northeast
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28 474 coast (Johnson et al., 1993). Numerous man-made chemicals are known to induce necrotic
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30 475 lesions in the liver of fish but naturally-occurring toxins such as microcystins can also be
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32 476 potent hepatotoxic agents (Wolf and Wolfe, 2005). It was recently documented that blooms of
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34 477 the cyanobacteria *Microcystis aeruginosa* producing the microcystin-LR can occur in
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36 478 estuarine waters in the summer time when surface water temperature is high and the stream
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38 479 flow is low (Lehman et al., 2008). Therefore it cannot be ruled out that the higher occurrence
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40 480 of FNR in the liver of flounders in the summer time could not be due to cyanobacterial
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42 481 blooms.
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49 482 Liver tumors and FCA were observed in 1.3% and 6.3% of all sampled fish respectively (1.5
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51 483 and 6.5% in the Seine estuary respectively). Tumors were restricted to females aged two years
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53 484 and more. The differential sensitivity of male and female flounders as regards liver tumors has
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55 485 already been reported (Koehler, 2004; Vethaak and Wester, 1996). Exposure to carcinogens
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57 486 likely takes place early in the life of flounders when larvae migrate to their estuarine habitats
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59 487 (Koehler, 2004). Since all the collected individuals live in the same habitat and feed on the
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488 same preys, they are likely exposed to the similar cocktail of carcinogens. Koehler proposed
489 that endogenous factors such as estrogen overproduction during vitellogenesis may explain
490 the higher susceptibility of female flounder to develop liver cancer. Indeed, tumor promotion
491 activity of 17- β -oestradiol had already been demonstrated in several fish species (Nunez et al.,
492 1989, Cooke and Hinton, 1999).

493 In the present study, adenoma was the predominant tumor type (1.0%) while adenocarcinoma
494 was only observed in a single flounder (0.2%). In fact, flounders bearing malignant tumors are
495 rarely observed in the field (Vethaak and Wester, 1996) likely because of quick death or
496 predation of sick individuals. Tumors and FCA were recorded at similar occurrences in the
497 liver of flounders inhabiting the Seine estuary as well as other estuaries, and coastal areas in
498 the North Sea and Baltic Sea (Table 4). Prevalence was somewhat higher along the German
499 coast and particularly in the Elbe estuary but only fish over 180 mm were analyzed (Koehler,
500 2004). It is noteworthy that FCA prevalence in flounders from the Seine estuary was recorded
501 at a higher level in group 0 than in the other age groups. Interestingly, Amara and
502 collaborators (2009) have also reported lower growth (daily otolith increments) and condition
503 factor (Fulton's K) for group 0 flounders from the Seine estuary in comparison to those
504 collected in two clean estuaries of the Eastern English Channel. Group 0 flounders are
505 predominantly located in the oligohaline part of the Seine estuary. It was previously
506 documented that sediments from this area were genotoxic resulting at least partially from a
507 high content of mutagenic and carcinogenic PAHs (Cachot et al., 2006). It was also
508 demonstrated that embryos of Japanese medaka exposed to an organic extract of sediment
509 from the upper Seine estuary had an increased mutation rate and developed FCA and
510 neoplasms (Cachot et al., 2007).

511 In the present work, DNA adducts were measured at high levels in the liver of juvenile
512 flounders inhabiting the upper Seine estuary. In addition, the majority of those DNA adducts

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513 were located in the DRZ. Several authors have reported that DRZ-related DNA adducts are
514 mainly induced by hydrophobic hydrocarbons including polyaromatic hydrocarbon
515 metabolites (Varanasi et al., 1989; Lyons et al., 1999; 2004). Indeed, DRZ-related DNA
516 adducts have been detected in the liver of fish exposed in controlled laboratory conditions to
517 various hydrophobic genotoxicants, including PAHs (Varanasi et al., 1989; Malmström et al.,
518 2000), and numerous field studies have documented a positive relationship between pollutant
519 exposure, notably to PAHs, and DNA adduct levels in fish liver (Varanasi et al., 1989;
520 Ericson et al., 1998; Rose et al., 2000; Shaw and Connell, 2001; Aas et al., 2001; Myers et al.,
521 2003). DNA adducts are currently considered as good molecular dosimeters of carcinogenic
522 exposure (Shaw and Connell, 2001) and experimental data supports the role of DNA adducts
523 in the initiation of chemical carcinogenesis (Miller and Miller, 1981).

524 It is thus tempting to build a scenario of exposure for flounders dwelling in the Seine estuary.
525 Juveniles, during their first year of life in the upper Seine estuary, are directly exposed to pro-
526 genotoxic and carcinogenic pollutants such as PAHs originating from contaminated sediments
527 and benthic preys. Chronic exposure to these compounds induces DNA damage and cytotoxic
528 effects which in turn, trigger structural changes in liver parenchyma, including inflammations,
529 necrosis and FCA. Since tumor development is a long lasting process, liver tumors appear
530 much later (at least two years later) when fish migrate to their marine habitat. Vethaak and
531 collaborators demonstrated through a 3-year mesocosm experiment with flounder juveniles
532 exposed to a contaminated dredged spoil that at least 30 months is needed for hepatocellular
533 adenoma development while FCA can emerge within 6 to 12 months (Vethaak et al., 1996).

534 From this study several conclusions can be drawn regarding the use of flounder as a sentinel
535 species for pollution monitoring in estuarine ecosystems. (i) Flounder juveniles settle in
536 estuaries just after metamorphosis and spend at least their first year of life there. (ii) Group 0
537 flounders exhibit a high susceptibility to pollutants, which was revealed in the present study

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538 by DNA adduct induction and the development of toxicopathic liver lesions including
539 necrosis and FCA. (iii) Tumor development is a long lasting process which does not allow
540 reliable measurement of site-specific pollution. Therefore, the measurement of biochemical
541 and histological markers in flounders from group 0 could represent a valuable approach for
542 pollution biomonitoring and fish health assessment in estuarine areas.

543

544 **4.3 Conclusion**

545 It was demonstrated herein that flounder liver lesions show high spatial and temporal
546 variations which can be related to various biotic (age, sex, etc.) and abiotic (season, location,
547 pollution, etc.) risk factors. Interestingly, group 0 flounders inhabiting the upper Seine estuary
548 faced a higher prevalence of encysted parasites, and foci of cellular alteration than older ones
549 living in the lower part of the Seine estuary. This result likely reflects differences in the
550 sensitivity of flounders to parasites, pathogens and chemicals, according to age, but also
551 different levels of stress according to season and location. Because of their high sensitivity to
552 multiple stress factors and to their first-year of growth exclusively in estuarine waters,
553 European flounders from group 0 represent valuable bio-indicators for pollution and fish
554 population health monitoring in estuarine ecosystems.

555

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722 **Figure captions**

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723 Figure 1: Maps showing the geographical position of the different sampling sites. Localization
724 of the three sampled estuaries (A) and the different sampling areas within the Seine estuary
725 (B).

726 Figure 2: Age-length relationship for European flounder specimens collected in the Seine
727 estuary.

728 Figure 3: Histological presentation of European flounder liver lesions. The liver is composed
729 of parenchymal cells (a cell that has a polygonal body containing a clear spherical nucleus and
730 clear appearance of the cytoplasm) and clusters of pancreatic cells along the branches of the
731 portal vein (large dichromatic cells). (A) Parasitic cyst at the vicinity of a pancreatic islet
732 surrounded by a thin rim of fibrous tissue corresponding to encapsulation (arrow). (B) Two
733 well-delineated granulomata slightly compressing the surrounding liver parenchyma (arrows)
734 are composed of large ill-delineated slightly basophilic mononuclear cells characteristic of
735 macrophages. (C) Melano-macrophagic aggregates: randomly scattered in the liver
736 parenchyma, they are composed of large fainted, rarely pigmented cells without
737 encapsulation. (D) A large focus of liquefactive necrosis with cellular debris and numerous
738 amoeba (not shown at this magnification). (E-G) Alteration foci are observed in the
739 parenchyma without compression and are composed of hepatic cells with modifications of
740 their trabecular organization, their size and their cytoplasmic staining: (E) Vacuolated cell
741 alteration focus, ill-delineated focus made of large macro-vacuolated cells (arrow); (F)
742 eosinophilic alteration focus: well-delineated focus made of hypereosinophilic cells; (G)
743 Small basophilic alteration focus: ill-delineated focus made of small hyperbasophilic cells
744 (arrow). (H) Pancreatic adenoma with papillary organization of tumoral cells, well-delineated,
745 encapsulated and compressing the surrounding area of the pancreatic parenchyma. (I) Large
746 infiltrative hepatocellular carcinoma with tubulo-papillar differentiation and a fibrous stroma,

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4 747 totally replacing liver parenchyma. Hematoxylin-Eosin-Saffron. (A, B, H, I) Bar = 100 μm ;
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6 748 (C-G) Bar = 50 μm .

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9 749 Figure 4: Prevalence (%) of liver pathologies in European flounders collected in the Seine
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11 750 estuary according to age group. (A) Parasitic, inflammatory, necrotic and regenerative lesions.
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13 751 (B) Pre-neoplastic and neoplastic lesions. Significant differences between age groups are
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15 752 indicated by different letters ($p < 0.05$, Fisher's exact test with Bonferroni's correction).

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18 753 Figure 5: Profile of liver pathologies in European flounders from the Seine estuary according
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20 754 to gender. (A) Parasitic, inflammatory, necrotic and regenerative lesions. (B) Pre-neoplastic
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22 755 and neoplastic lesions. Significant differences between genders are indicated by different
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24 756 letters ($p < 0.05$, Fisher's exact test with Bonferroni's correction).

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28 757 Figure 6: Seasonal variations of liver pathologies in European flounders from the Seine
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30 758 estuary. (A) Parasitic, inflammatory, necrotic and regenerative lesions. (B) Pre-neoplastic and
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32 759 neoplastic lesions. Significant differences between seasons are indicated by different letters
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34 760 ($p < 0.05$, Fisher's exact test with Bonferroni's correction).

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38 761 Figure 7: Temporal evolution of liver pathologies in European flounder from the Seine
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40 762 estuary. (A) Parasitic, inflammatory, necrotic and regenerative lesions. (B) Pre-neoplastic and
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42 763 neoplastic lesions. Significant differences between sampling sites are indicated by different
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44 764 letters ($p < 0.05$, Fisher's exact test with Bonferroni's correction).

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47 765 Figure 8: DNA adduct levels in the liver of European flounders collected in the Seine estuary
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49 766 (September and October 1996) and in the Bay of Veys (April 1997). Data are expressed as
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51 767 relative adduct labeling per 10^9 nucleotides (Mean \pm SD, N=2)

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55 768 Figure 9: Representative autoradiograms obtained for European flounders collected in the
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57 769 upper Seine estuary (A) or in Antifer (B). In the first autoradiogram, DNA adducts are

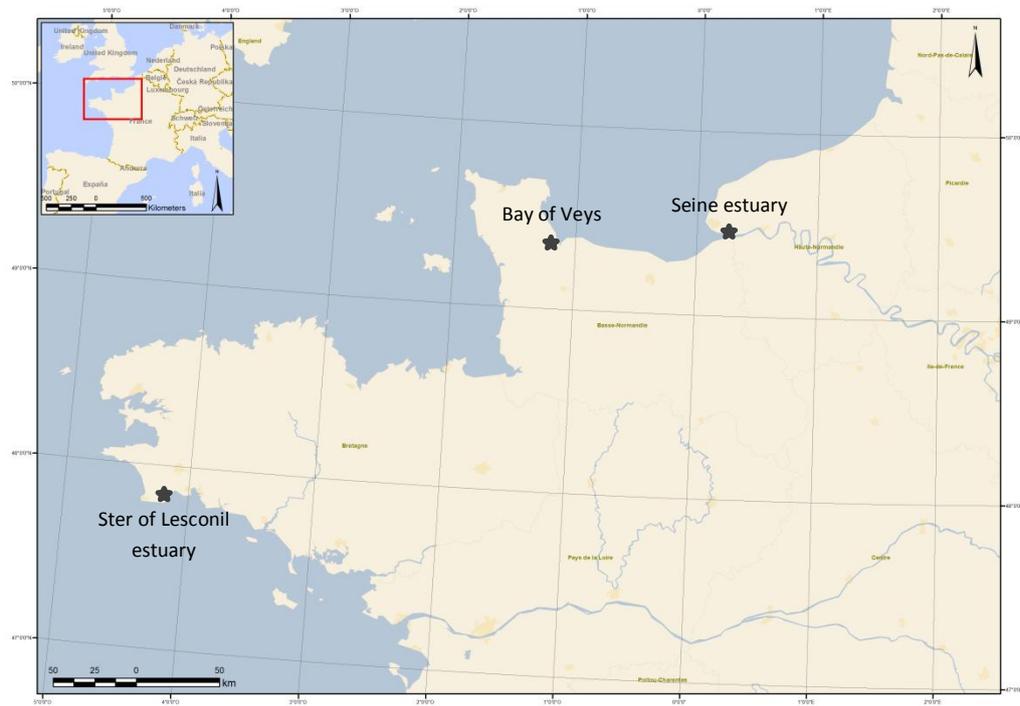
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770 predominantly located within the diagonal radioactive zone (DRZ) while in the second only

771 individual spots outside the DRZ are readily visible.

Figure 1

A



B

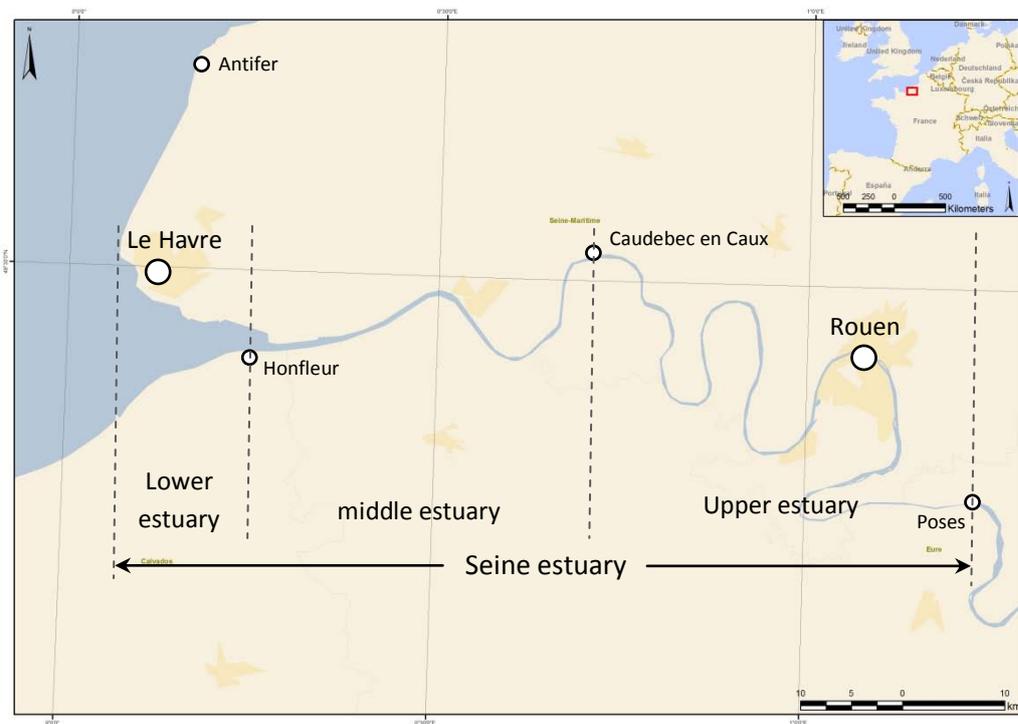


Figure 2

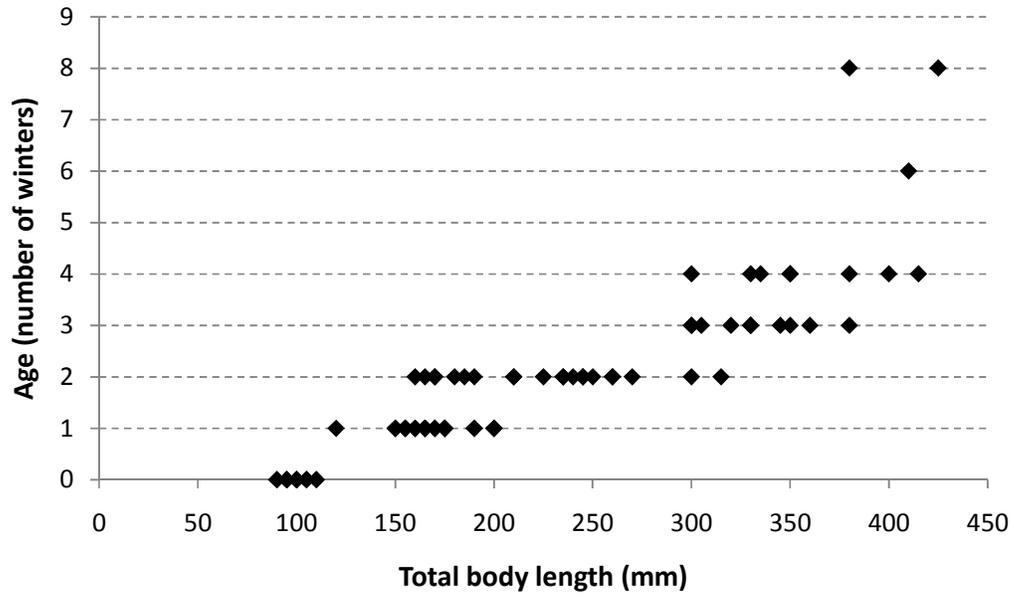


Figure 2

Figure 3

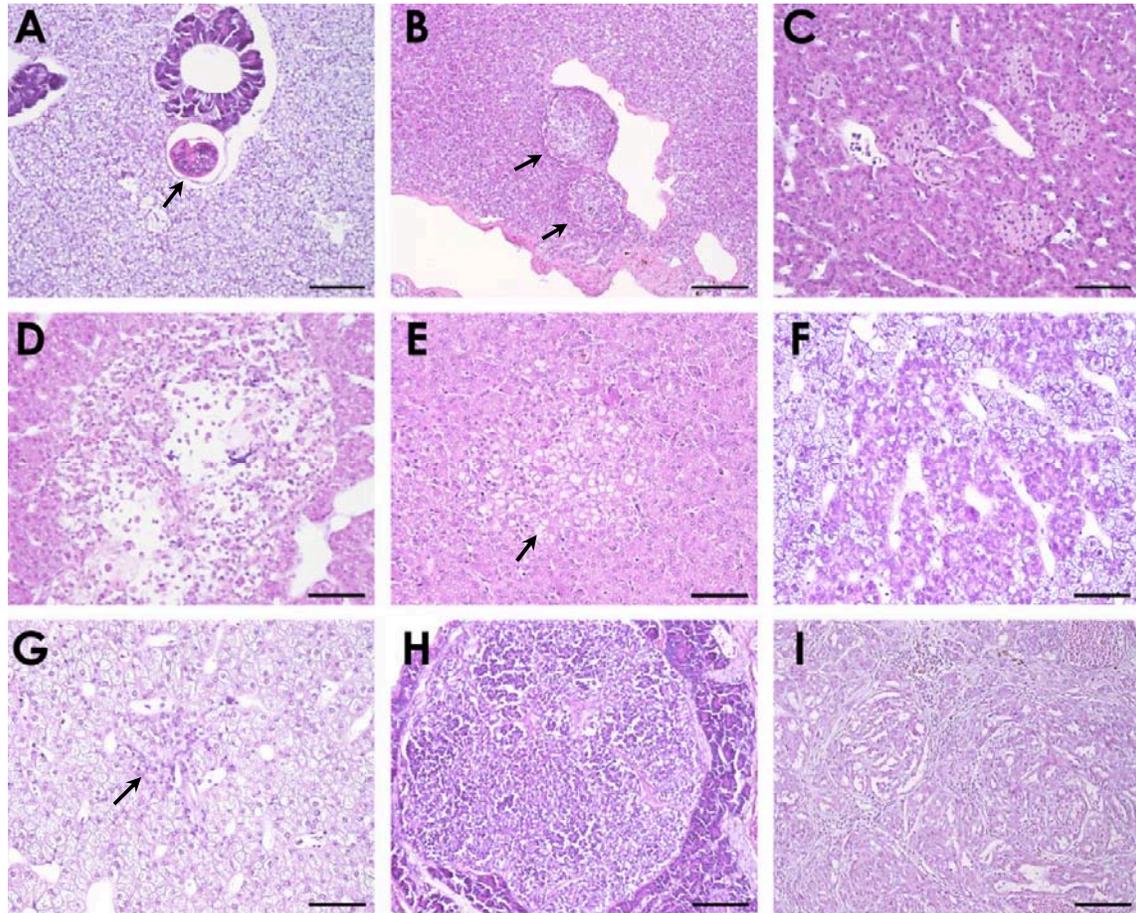
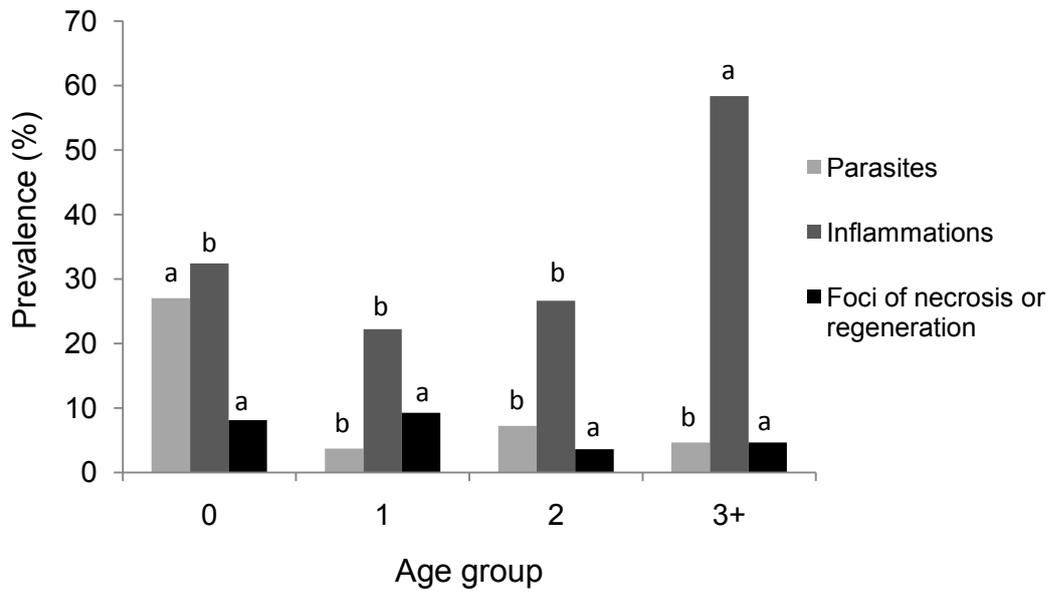


Figure 4

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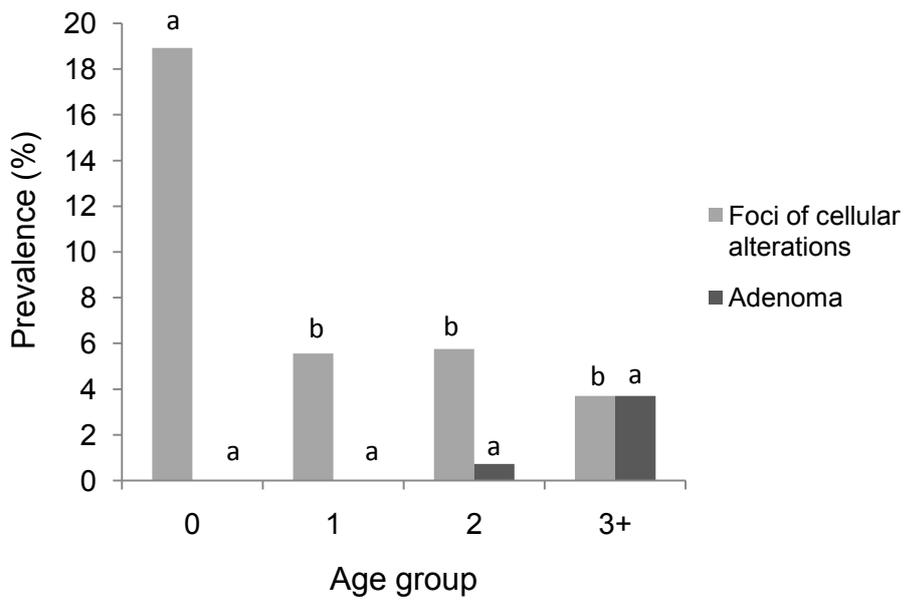
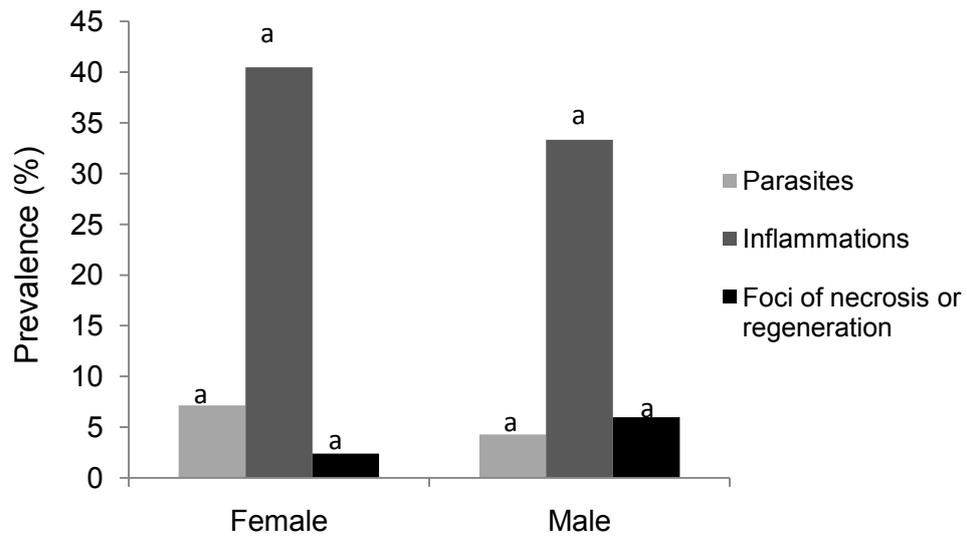


Figure 4

Figure 5

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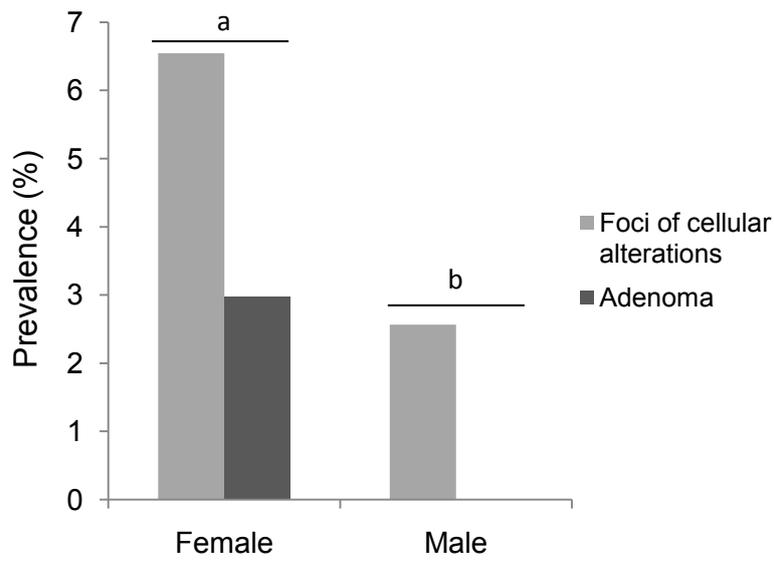
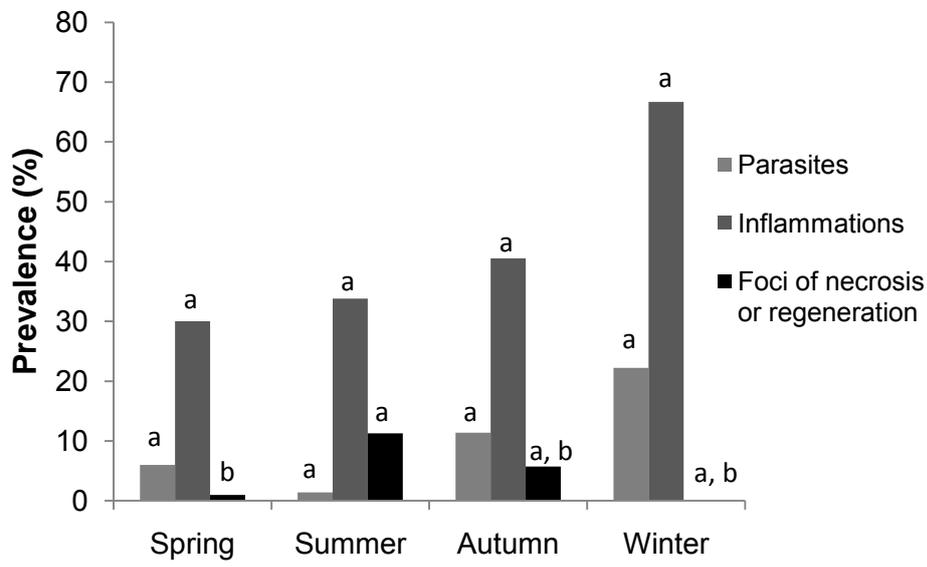


Figure 5

Figure 6

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B

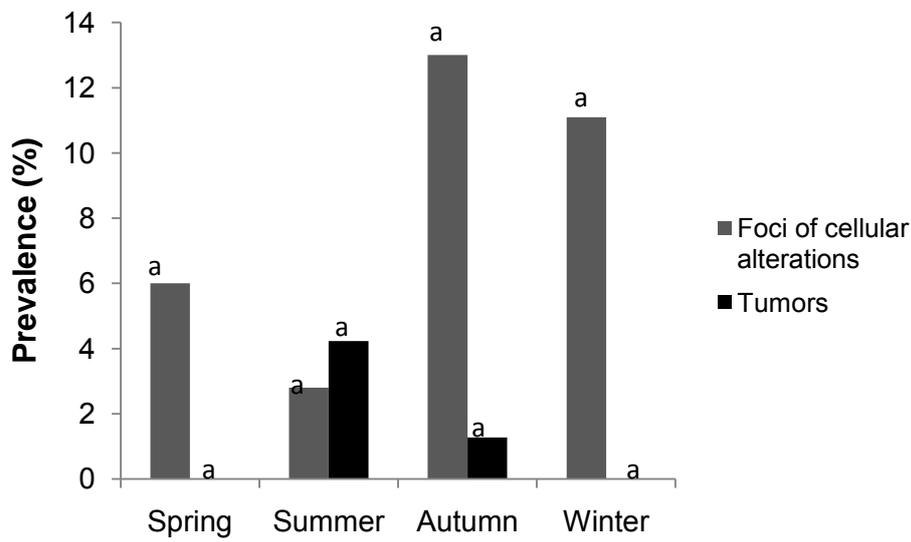
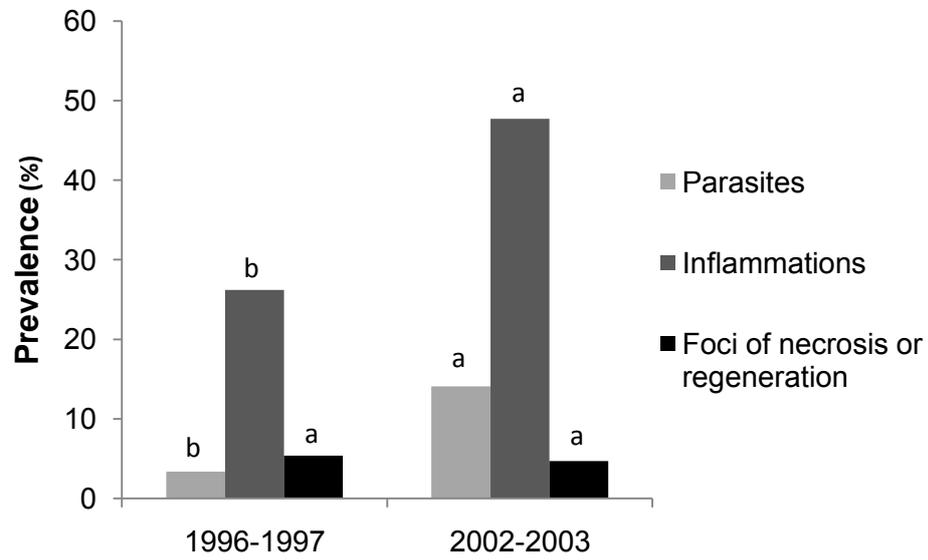


Figure 6

A



B

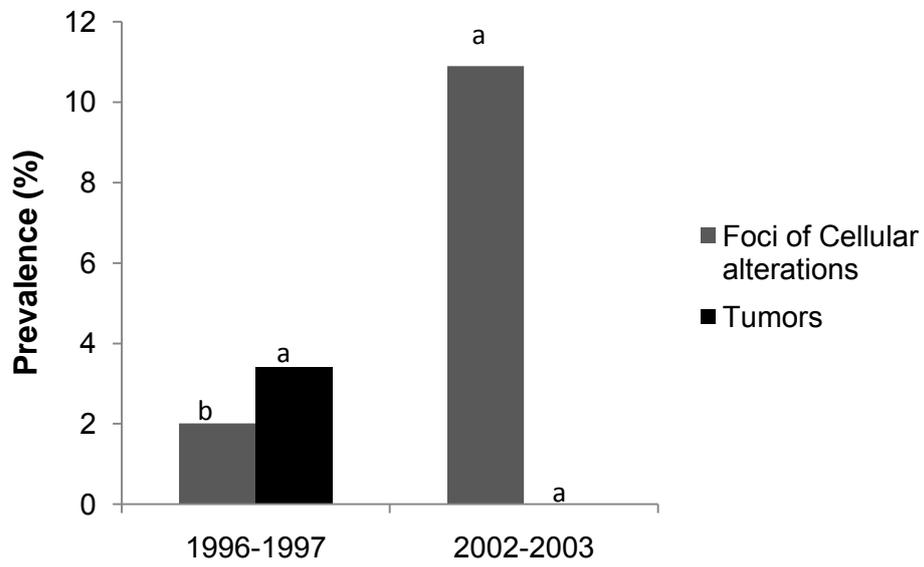


Figure 7

Figure 8

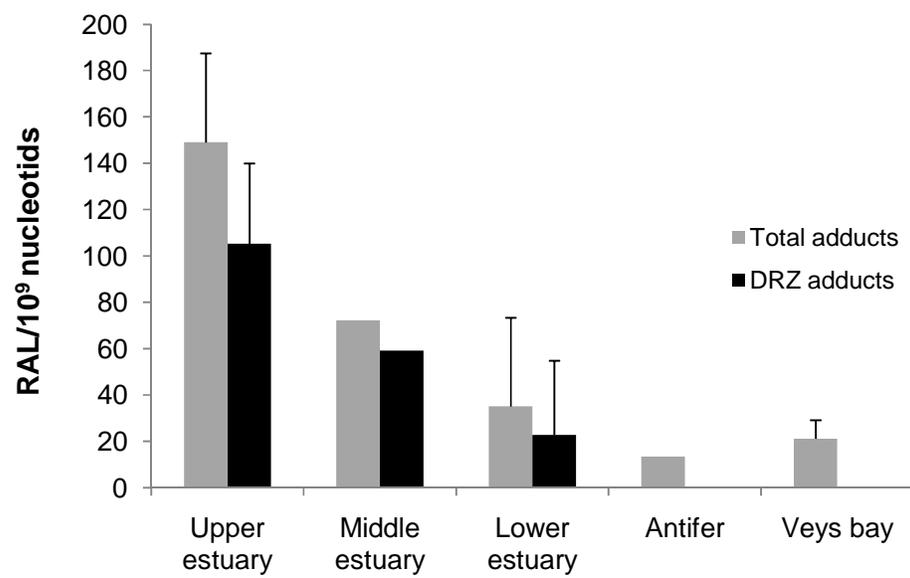


Figure 8

Figure 9

A



B



Figure 9

Table 1 : European flounder sampling campaigns in three estuaries from the French Atlantic coast

Sites	Number de fishing campaigns	Period	Number of fish collected	Sex ratio M/F	Total body length (cm)
Seine estuary	9	April 1996	42		
		Sept-Oct 1996	101		
		March 1997	51		
		March 2001	6		
		Sept 2001	10		
		Feb 2002	3		
		June 2002	23		
		Oct-Nov 2002	30		
		Nov 2003	72		
		Total	338	0.71	4.5 - 42.5
Bay of Veys	2	Sept 1996	4		
		April-May 1997	50		
		Total	54	1.41	15.5 - 46
Ster estuary	3	May 2002	28		
		Jan 2003	25		
		May 2004	33		
		Total	86	0.36	19.5 - 49.5

Table 2: Prevalence of liver pathologies in juveniles and adults of European flounder sampled at different locations in the Seine estuary

Sampling sites	Number						
	of fish	Parasites	Inflammations	FNR	FCA	Adenoma	Carcinoma
Upper part	49	18.4 a	26.5 a	10.2 a	14.3 a	0.0 a	0.0 a
Median part	45	2.2 b	26.7 a	4.4 a	4.4 a	0.0 a	0.0 a
Lower part	157	9.5 a	36.9 a	3.2 a	6.4 a	1.9 a	0.0 a
Antifer	87	2.3 b	47.1 a	6.9 a	3.4 a	2.3 a	0.0 a
Seine estuary	338	8.0	36.7	5.3	6.5	1.5	0.0

FCA : foci of cellular alteration, FNR : foci of necrosis or regeneration

Different letters indicate significant differences between sampling sites (Fisher's exact test with Bonferroni's correction)

Table 3 : Prevalence of liver pathologies in adults of European flounder (total length > 200mm) from three estuaries from the French Atlantic coast

Sampling area	Number of fish	Parasite	Inflammation	FNR	FCA	Adenoma	Carcinoma
Bay of Veys	46	6.5 b	15.2 b	0.0 a	0.0 a	0.0 a	0.0 a
Ster estuary	85	37.6 a	51.8 a	2.4 a	9.4 a	0.0 a	1.2 a
Seine estuary	243	5.8 b	40.3 a	3.7 a	4.9 a	2.1 a	0.0 a
<i>Median part</i>	38	2.6	28.9	0.0	5.3	0.0	0.0
<i>Lower part</i>	118	9.3	39.0	2.5	5.9	2.5	0.0
<i>Antifer</i>	87	2.3	47.1	6.9	3.5	2.3	0.0

FCA : foci of cellular alteration, FNR : foci of necrosis or regeneration

Different letters indicate significant differences between sampling sites (Fisher's exact test with Bonferroni's correction)

Table 4 : Prevalence of liver diseases in European flounders from different geographical areas of the North-east Atlantic

Sampling area	Year	Number of fish	Parasite	INF	FNR	FCA	Tumor	Reference
Dutch North Sea coast	1985-1989	315	-	11.7	5.4	11.4	0.3	Vethaak and Wester, 1996
Dutch Wadden Sea coast	1988	9608	5.0	-	-	<0.1	<0.1	Vethaak, 1992
Dutch Wadden Sea coast (Den Oever)	1996	240	-	12.5	-	1.7	0	Vethaak et al., 2011
German Wadden Sea coast	1995-2000	1468	-	-	-	19-34	4-38	Koehler, 2004
Baltic Sea	1994	3008	-	-	-	0.6	0.9	Bogovski et al. 1999
Baltic Sea coast	2001-2002	436*	17.7	28.9	-	4.6	0.7	Lang et al., 2006
UK estuaries	2000	204	-	20	-	17.6	1.5	Stentiford et al., 2003
UK estuaries	2006	81	0-6	10-67	5-22	0-28	0-11	Williams et al., 2011
Seine estuary	1996-2003	338	8.0	36.7	5.3	6.5	1.5	This study

INF : inflammation, FCA : Foci of Cellular Alteration, FNR : Foci of Necrosis or Regeneration. (*) Female only