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Evaluation of a lifetime-based optode to measure oxygen in aquatic systems

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Abstract:

In this paper we evaluate the performance of a commercially available life-time based optode and compare it with data obtained with other methods. A set of 11 different tests including targeted laboratory evaluations and field studies were performed covering a wide range of situations from shallow coastal waters and waste water treatment plants to abyssal depths. The principal conclusion reached is that this method, due to high accuracy ($\pm 2 \mu M$); long-term stability (more than 20 months); lower fouling sensitivity; no pressure hysteresis and limited cross sensitivity, is overall more suitable for oxygen monitoring than other methods.

Keywords: aquatic systems, oxygen, measure, lifetime-based optode

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2

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18	situations from shallow coastal waters and waste water treatment plants to abyssal depths. The
19	principal conclusion reached is that this method, due to high accuracy ($\pm 2 \mu M$); long-term
20	stability (more than 20 months); lower fouling sensitivity; no pressure hysteresis and limited
21	cross sensitivity, is overall more suitable for oxygen monitoring than other methods.
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23	Introduction

1	Most chemical and biological processes are influenced by changes in dissolved oxygen
2	concentrations. Oxygen is therefore a prime parameter to measure in a variety of applications
3	ranging from industrial processes to environmental studies.
4	The standard method to analyze oxygen content in water is a two-step wet chemistry
5	precipitation of the dissolved oxygen followed by a titration. The method was first described
6	by Winkler (1888) and has since then remained the overall standard. Winkler titration is the
7	method which is almost always performed in the laboratory on collected water samples. The
8	collection and handling of water samples can induce errors and the analytical work is time
9	consuming and demands meticulous care. It is therefore not a suitable method to obtain <i>in-situ</i>
10	data with high spatial and temporal resolution.
11	For in-situ measurements of oxygen, electrochemical sensors (often called Clark type
12	sensors after a US patent, Clark, 1959; Kanwisher, 1959) are the most common.
13	Electrochemical sensors have been developed and used in a wide range of applications and
14	therefore vary in size and design from micrometer sized glass electrodes to more robust macro
15	electrodes with sensing tips of several centimeters (for a recent review see Glud et al., 2000
16	and references therein). Performance studies of different types of electrochemical sensors
17	have been presented in e.g. Briggs and Viney (1964); Atwood et al. (1977); Hitchman (1978);
18	Gneiger and Forstner (1983); Short and Shell (1984); Berntsson et al. (1997); Gundersen et
19	al., (1998) and Glud et al. (2000). Regardless of the design, these studies have shown that the
20	Clark sensor requires frequent (at least monthly) calibration to obtain accurate measurements
21	of dissolved oxygen.
22	Optodes (also called optrodes) may provide a more suitable method than
23	electrochemical sensors for direct measurement of dissolved oxygen. Optode technology has

- been known for years (e.g. Kautsky, 1939) but it is relatively new to the aquatic research (e.g.
- 25 Klimant et al., 1995; Glud et al., 1999; Wenzhöffer et al., 2001). The fundamental principle is

1	based on the ability of selected substances to act as dynamic luminescence quenchers. In the
2	case of oxygen, if a ruthenium-complex is illuminated with a blue light it will be excited and
3	emit a red luminescent light with an intensity, or lifetime, that depends on the ambient oxygen
4	concentration. It is important to distinguish between three different principles in detecting the
5	red luminescence: Intensity (how strong the luminescence is), life-time (how quickly the
6	luminescence dies out) and phase shift (in principle also a life time based measurement, see
7	below Measurement Principle). Intensity based measurements are technically easier to do, but
8	they can drift over time. The different signal detection techniques are summarized by
9	Wolfbeis (1991), Demas et al. (1999) and Glud et al. (2000) along with a wide range of
10	applications. Klimant et al. (1995) and Stokes and Romero (1999) described intensity based
11	oxygen optodes and their use. The function and use of lifetime based optodes was described
12	by Holst et al. (1995) and Klimant et al. (1996).
13	Optode technology has an advantage over conventional sensors in that it can also be
14	used to assess oxygen distributions in two dimensions (e.g. Holst et al., 1997, 1998; Glud et
15	al., 1999, 2001, 2005) as well as detection of other substances (e.g. Klimant et al., 2000;
16	Huber et al., 2000, 2001a, 2001b; Hulth et al., 2002; Bültzingslöwen et al., 2002).
17	In this paper we evaluate a commercially available lifetime-based oxygen optode and
18	compare its performance with different electrochemical sensors and Winkler titrations. A
19	novelty with this sensor, compared to the optodes used in other studies, is that it combines the
20	benefits of using a platinum porphyrine luminescence dye with digital signal processing
21	electronics (DSP). The construction and working principles of this sensor is described briefly
22	and its accuracy and precision evaluated in the laboratory. The influence of simultaneous

- changes in oxygen, salinity, temperature, stirring, pressure and pH were tested by using
- 24 multivariate statistical methods. Data from extended evaluations of pressure behavior as well

as response time are presented and discussed. Finally a range of field application examples
 are given to demonstrate the ability and stability of oxygen optodes in different field settings.

4 <u>Measurement principle</u>

5

The particular type of sensor that has been used in this evaluation is commercially available
from Aanderaa Instruments, Norway (Oxygen Optode model 3830 and model 3930, Tengberg
et al., 2003). The sensor is based on oxygen luminescence quenching of a platinum
porphyrine complex. The lifetime and hence the oxygen measurement is made by a so called
phase shift detection of the returning, oxygen quenched red luminescence. The relationship
between oxygen concentration and the luminescent decay time can be described by the *Stern-Volmer* equation:

13

14
$$[O_2] = \frac{1}{K_{SV}} \left\{ \frac{\tau_0}{\tau} - 1 \right\}$$

15

16 Where: $\tau = \text{decay time}$, $\tau_0 = \text{decay time in the absence of oxygen and } K_{SV} = \text{Stern-}$ 17 Volmer constant (the quenching efficiency). The foil is excited with a blue-green light 18 modulated at 5 kHz. The decay time is a direct function of the phase of the received red 19 luminescent light which is used directly for oxygen detection, without calculating the decay 20 time. The basic working principles of dynamic luminescence quenching, lifetime-based 21 optodes and phase shift detection can be found in e.g. Klimant et al. (1996); Demas et al. 22 (1999); Glud et al. (2000).

The sensor housing is made of Titanium, rated to 600 bar pressure, with a diameter of
36 mm and a total length of 86 mm. This housing includes an optical part, a temperature

1	sensor (Fig. 1) and the necessary electronics (a microprocessor with digital signal processing
2	capacity) to process signals and output absolute temperature compensated oxygen readings (in
3	μM or % saturation). An advantage of using digital signal processing, over analog technology,
4	is enhanced possibilities of signal filtering and less risk of drift in the electronics with changes
5	in temperature and over time.
6	The sensing foil is composed of an oxygen sensitive luminescent substance
7	(luminophore) that is embedded in a polymer layer which is coated onto a thin film of
8	polyester support (Fig. 1). The most commonly used oxygen lumniphores have been
9	ruthenium complexes (e.g. Klimant et al., 1996; Stokes & Romero, 1999) but for this sensor
10	an oxygen-sensitive luminophore based on a platinum porphyrine complex, commercial
11	available from PreSens GmbH (Regensburg, Germany) was used mainly due to its longer
12	lifetime (about a factor 5 longer than for ruthenium). A longer lifetime makes it easier to
13	detect the signal and to obtain stable readings. Another advantage with platinum porphyrine is
14	that it is less sensitive to photobleaching.
15	We tested two types of foils, with and without, a gas permeable protective black
16	silicon layer (Fig. 1). The silicon layer also acts as an optical isolation layer to avoid potential
17	influence from fluorescent/luminescent material in the surrounding water and/or direct
18	incoming sunlight, when measuring in the photic zone. The disadvantage of this layer is that
19	the sensor response time becomes longer.
20	
21	Results from sensor tests and trials
22	
23	To evaluate how suitable optodes are for measurements of dissolved oxygen, 11
24	different tests (see Table 1) using about 70 different sensors were selected from an initial set
25	of 40 datasets.

1

a) Calibration performance

3

2

4 The response of an optode is giving highest sensitivity at low oxygen concentrations, 5 and follows the Stern-Volmer equation (e.g. Demas et al., 1999). The response is also affected 6 by temperature, which requires compensation in calibrations/measurements. To evaluate the 7 effect of calibration on sensor accuracy two different procedures were tested (Table 1). The 8 first was a 30-point calibration (5 different temperatures and 6 different oxygen 9 concentrations) of a batch of foils (normally produced in batches of 100) to obtain foil-10 specific calibration constants. These are stored in the sensor processor when the foil is 11 mounted. Then a two point calibration is performed at 0 % and 100 % air-saturation, taking 12 into account the prevailing air pressure. The batch constants and the two-point calibration are 13 together used to fit the Stern-Volmer equation to the particularities of the foil and the sensor. 14 This is the calibration procedure that is presently used when sensors are produced. 15 A second set of calibrations were performed on 20 different sensors to assess if and by 16 how much the accuracy could be improved by making an individual sensor specific 30-point 17 calibration with the foil mounted from the start. This method is more time consuming, since 18 every sensor has to go through an elaborate calibration procedure, but it is likely to better 19 account for individualities of sensors and foils. 20 The individual 30 point sensor calibrations enhanced the accuracy by a factor of 21 approximately 3 compared to a batch calibration with a subsequent two point calibration 22 (Table 1). We have not been able to pinpoint the reason(s) for the improvements but we 23 suspect that the lower accuracy from batch calibrations was caused by a combination of 24 factors: 1. Individual temperature behavior of the foil; 2. Differences in the lifetime between 25 foils, even though the foils are produced in batches of 100; 3. Differences in the electronic

1	and optical properties between sensors, even if the sensors are produced in batches of 25. The
2	obtained accuracy ($\pm 5\mu M$) using a batch-calibrated sensor should be sufficient for most
3	applications, especially in the coastal environment were variations in dissolved oxygen are
4	often large. The absolute precision (resolution) of the sensors was the same regardless of the
5	calibration method. As expected the sensor performance was better (± 0.5 $\mu M)$ at lower
6	oxygen concentrations. At the higher end (300-500 $\mu M)$ precision was approximately \pm 1 $\mu M.$
7	A major advantage of the optode is that it is more accurate at lower concentrations for which
8	Winkler titrations are more difficult, due to the risk of contamination.
9	
10	b) Physical and chemical cross sensitivity
11	
12	The solubility of oxygen in water is dependent on salinity and temperature (e.g. Weiss, 1970;
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23 factors affected the sensor readings in a complex interrelated manner.

In this study we have used the same methods as Berntsson et al. (1997) but with higher
 pressure levels of 1, 250 and 500 bar (Table 1). To be able to perform these tests under
 controlled conditions, an experimental chamber that can be placed into a pressure tank was
 constructed (Fig. 2).

5 The outcome was that stirring had no effect, which was expected from earlier work by 6 Klimant et al. (1995) and since the sensor is not consuming any oxygen. This is an advantage 7 and also explains why optodes are less sensitive to fouling (see field data below) as compared 8 to the oxygen consuming electrochemical sensors. Pressure had an effect (about 4 % lower 9 response per 100 bar of pressure) but was fully reversible and predictable (i.e. all sensors had 10 the same response). Temperature also has an influence on the optical measurements but since 11 the optodes described here are equipped with temperature sensors this effect is automatically 12 compensated for by the internal processor and no remaining influence of temperature could be 13 distinguished, which also indicates that the temperature compensation was done correctly. 14 Figure 3A gives raw data from the three different sensors that were run in parallel during the 15 multivariate evaluations. Figure 3B shows the data after a general salinity (according to 16 Garcia and Gordon, 1992) and pressure compensation.

A set of separate more detailed pressure tests were performed, using the same equipment as described in Fig. 2, to evaluate the effects of high pressure and rapid pressure changes. These evaluations confirmed the above described pressure effect of 4 % lower response per 100 bar and also demonstrated that rapid pressure cycling from 3 to 405 bar and back did not leave any remaining pressure effects (hysteresis) on the sensors (Fig. 4). Pressure hysteresis is a know problem for electrochemical sensors which limits the usefulness of this technology at high pressure (e.g. above 100 bar).

Potentially other substances present in natural waters could interfere with the oxygen
 measurements. Contamination by hydrogen sulfide (H₂S) is a major concern when using

electrochemical sensors in oxygen depleted environments. H_2S has no influence on the response of the optodes and it will not damage them in any way (Klimant et al., 1995). The only interferences (cross-sensitivity) are found with gaseous sulfur dioxide (SO₂) and gaseous chlorine (Cl₂). The sensors can also be used in methanol- and ethanol-water mixtures as well as in pure methanol and ethanol. It is recommended to avoid other organic solvents, such as acetone, chloroform or dichloromethane, which may swell the sensor's foil matrix (data not shown).

8

9 c) Response time

10

11 A fast response time is a requirement for sensors that are used on profiling CTD 12 instruments (typically descending/ascending at rates of 0.5 m/s), on towed vehicles (often 13 towed at 2-5 m/s) and when high frequency sampling is required to resolve rapid oxygen 14 fluctuations for example close to the bottom (e.g. Berg et al., 2003). Electrochemical 15 microelectrodes (e.g. Revsbech, 1989) are unmatched with regard to fast response time and a 16 well-designed sensor can have a 90 % response time (t_{90}) of around 0.1 s (Glud et al., 2000). 17 These sensor are however not rugged and difficult to handle and they suffer from the typical 18 limitations of electrochemical sensors (except possibly stirring sensitivity since the oxygen 19 consumption is low). Microoptodes are easier to make and more robust than the 20 electrochemical microelectrodes and these have shown response times (t_{90}) of around 2 s. 21 Larger electrodes are more rugged but the response times are generally longer. A typical 22 electrochemical macro-sensor has a t_{90} of 30-120 s but this can be improved by using thinner 23 membranes (which also makes the sensor more stirring sensitive and noisy). The response 24 time of the optodes used in this study were tested with two different foils one which was

1	optically isolated ($t_{90} = 47$ s) and one without optical isolation with a $t_{90} = 10$ s. It should be
2	noted that there is no standard way of testing the response time and the results obtained are to
3	a large extent dependent on how the tests were done. In our case we choose to simply move
4	the sensors, without wiping the water off, between two containers one which was air saturated
5	and the other depleted in oxygen, through stripping with N_2 gas. If the sensor is wiped off
6	before being submerged into a new container the thin water film in contact with the foil is
7	eliminated and the response time drops by approximately a factor 3. We have, however,
8	chosen to use the conservative values here since we believe that this test method is more
9	representative to what is happening in the field, e.g. when profiling through the water column.
10	While a t_{90} of 10 s should be sufficient for most applications a t_{90} of 47 s is probably to slow
11	for many profiling applications, especially when passing through sharp gradients in the water
12	column. It is likely that the response time of these sensors could be improved by optimizing
13	the foil and sensor design for this particular purpose.

14

15 *d)* Long-term stability

16

The amplitude of the red luminescent signal (see Fig. 1) should not be of importance if a life time based detection method is used. So if the foil is bleached/degraded over time, which it will be, or the optical properties of the system change, for example with changes in temperature the response should not be affected. To evaluate the long-term stability several laboratory and field evaluations were performed. Two are presented here each one which lasted for about 600 days (Table 1).

23 The first dataset was collected from an optode that was mounted on a profiling Argo 24 float. These floats are autonomous and freely drifting in the oceans. Today there are about 25 1900 units in operation (http://www-argo.ucsd.edu/). The floats can change their buoyancy



1	can depend on local variations since they are all drifting at different locations and/or on
2	different calibrations of the float mounted optodes. This data set is not an absolute evidence
3	that the sensors have not drifted over the 600 day deployment period but the results indicate
4	that no or little (less than 5 %) drift has occurred.
5	Other data sets from continuous field measurements for periods of 10-16 months have
6	also indicated no drift (not presented here) which give reason to believe that the stability of
7	these sensors is at least 600 days and probably longer.
8	It should be noted that all the above presented stability data were collected in
9	environment with low fouling. When fouling is affecting the sensor such long stabilities can
10	not be expected (see below).
11	
12	e) Sensitivity to biofouling
13	
14	Biofouling and bacterial growth is a major obstacle to long-term monitoring in the

15 aquatic environment. The types of fouling that occur vary greatly from one environment to 16 another and so do the effects of fouling. It is well known that the electrochemical measuring 17 principle is sensitive to fouling since oxygen is consumed during measurements. If the supply 18 of oxygen is impeded this might results in a sensor drift. Various technical solutions such as 19 wipers, pulsed sensors and slow response sensors with thicker membranes can improve the 20 fouling resistance. Optical sensors are likely to be less sensitive to fouling since they do not 21 consume any of the measured analyte. In order to test this, three different evaluations were 22 done with the aim of covering a wide variety of fouling situations (Table 1).

A network of environmental on-line monitoring stations has been operational in rivers around Paris (river L'Orge) for many years. One of the most important parameters to monitor at these stations is oxygen which is measured with electrochemical sensors. Oxygen has also

1 been the most labor-demanding parameter since drifting slime blocks the membrane of the 2 sensors and rarely allows them to be operational for more than one week. Therefore weekly 3 service intervals are conducted. During service the oxygen sensors are taken up, cleaned and a 4 recalibrated (in two points). With the aim of prolonging the time between services an oxygen 5 optode was tested for three weeks in parallel with one of the electrochemical sensors in 6 normal use. The electrochemical sensor was submitted to its regular weekly service intervals 7 while the optode was left untouched for the 20 days test period. Measured oxygen 8 concentrations in the river follow the daily variations in temperature (Fig. 7). When the water 9 becomes colder during night the oxygen solubility increases leading to higher concentrations. 10 During daytime the temperature increases and the oxygen level drops. In spite of becoming 11 covered with slime only after 2-3 days the optode was stable for the whole test period. In 12 general the electrochemical sensor started to drift towards lower values after 2-3 days. After 13 cleaning and recalibration it returned to readings which were similar to the optode recordings 14 (Fig. 7). In this particular case fouling did not have any detectable influence on the optode. 15 There are, however, other river examples (data not shown here) in which the slime possibly 16 contained more active microorganisms and the fouling started to affect the naked optode after 17 14 days (see also example from waste water treatment plants below).

18 City waste waters have a high content of organic material, the microbiological activity 19 is elevated and the oxygen concentration is a critical element for proper processing. An 20 accurate monitoring of oxygen in waste water treatment plants is necessary to optimize the 21 relation between cleaning efficiency and energy used for water aeration. Several comparative 22 tests (Table 1) were made with optodes in two different waste water treatment plants (in 23 Västerås, Sweden and at the Attleboro waste water treatment plant, Providence, USA). In the 24 aeration basins, the residence time is relatively short (hours) and the bubbling ensures mixing 25 of the water. The fouling consists of high organic content slime. Unprotected optodes were

1 placed in parallel with special process-adapted electrochemical sensor systems with 2 continuous cleaning. Unprotected optodes gave correct readings on average for 14 days. After 3 that the microbiological oxygen consumption of the organic material that had accumulated in 4 the small depression in front of the foil (see Fig. 1) resulted in signal drift. The use of a 5 protective copper plate had no effect, on the contrary the mounting of the plate made the 6 particles accumulate faster. To solve this problem a spray nozzle (normally used to clean cars) 7 was fitted and directed towards the sensor foil. The nozzle was connected to a freshwater 8 garden hose. By turning on the water the sensor was sprayed/cleaned once every 1-2 days and 9 it was continuously operated in this mode for 90 days (data not shown) without any visible 10 fouling effects and without any drift.

11 To increase the agricultural reuse of livestock waste as fertilizers and to decrease the 12 environmental effects when spreading these on the fields different types of bioreactors have 13 been developed and studied (e.g. Skjelhaugen, 1999). In order to optimize the efficiency of 14 the bioreactors and to study the ongoing processes it is desired to monitor the oxygen levels 15 inside the 5-12 % (by weight) livestock waste / water slurries. Previously it has been difficult 16 to use electrochemical sensors here since H_2S is often present and since these sensors are 17 known to be stirring sensitive. An optode was used in several studies measuring both in the 18 headspace above the slurries and directly in the slurries. Measurements in the headspace gave 19 consistent results but when measuring in the slurries with waste concentrations above 6 % it 20 was not possible to conclude anything on the efficiency of the aeration. When introduced into 21 the slurry the sensor rapidly dropped to readings close to 0 and stayed at these levels in-spite 22 of high levels of aeration. We think that the reason for this behavior is caused by an improper 23 mixing of the slurry in front of the sensor. Since the content of organic materials is elevated 24 the oxygen is rapidly consumed and if the aeration is not capable to achieve a proper mixing / 25 aeration in front of the sensor it will measure to low values.

1 In the coastal marine environment the fouling is often a major impediment to 2 monitoring. Unlike the cases described above (rivers and waste water) the fouling often 3 consists of plants (e.g. algae, seaweed etc.) and/or animals (e.g. shell building barnacles). 4 Several trials were done to prevent the effects of fouling on the here described sensor. The so 5 far most efficient method was to wrap a beryllium-copper alloy net (used for domestic 6 cleaning) around the sensor. In environments with heavy fouling (shallow coastal waters in 7 Chesapeake Bay and the Gulf of Mexico) this method prolonged the service interval from 8 approximately 7-10 days to 40-60 days. It is important to prevent the beryllium-copper net to 9 enter into contact with the metal of the sensor since this will create a galvanic element which 10 makes the mesh disappear faster. It also makes it significantly less efficient in preventing 11 fouling. Electronic isolation was obtained by wrapping tape around the sensor before adding 12 the alloy net. Other solutions such as a fixing a fine meshed copper plate in front of the foil 13 has proven inefficient in most situations since the mesh was quickly clogged and prevented a 14 proper circulation.

15

16 f) Other field applications

17

We selected two examples of data that were collected in the marine environment with the aim of covering a wide range of measuring situations from shallow depth to the deep sea (5500 m), from tropical surface waters with temperatures approaching 30°C to cold abyssal surroundings with temperatures around 2°C and with oxygen levels ranging from 20 to 230 μ M. In both cases the data registered by the optodes corresponded well (within the nominal limits of accuracy) with independent measurements of oxygen obtained through Winkler titration.

Figure 8 presents a subset of optode data that were collected from a shipboard
underway seawater pumping system during a 35-day expedition of the German R/V Meteor to
the tropical Atlantic Ocean. The two meridional sections along $\sim 26.5^{\circ}$ W (southbound) and
\sim 24 °W (northbound) shows very similar oxygen concentrations at 5 m depth in the surface
mixed layer. The corresponding saturation level is 101.5–104.5 %. The observed variability is
mainly due to small-scale patchiness of phytoplankton (confirmed by separate samples) but
also exhibits some diurnal cyclicity. Oxygen concentration from Winkler titration of samples
taken regularly from 5 m hydro cast samples along the two transects are in good agreement
with the optode data (offset = $0.9 \ \mu$ M, rms = $1.1 \ \mu$ M). Winkler samples were also taken
regularly from the underway pumping system but these discrete samples show frequent
contamination by small air-bubbles and are therefore of inferior quality (data not shown).
On-line profiling with cable operated CTD (Conductivity, Temperature and Depth)
instruments has been common practice in oceanography for several decades. The combination
of real-time data access with the possibility of closing water sample bottles at depths of
interest makes such instruments useful in many types of applications. CTDs have been
equipped with a whole suite of sensors, e.g. to measure particles in the water, chlorophyll
content, pH, current velocity etc. Oxygen, measured with electrochemical sensors, has also
been one of the most common parameters to be included on CTD instruments. These sensors
demand special attention and frequent recalibrations to be able to give reliable data.
Furthermore, when making deep-sea profiles residual pressure effects (hysteresis) frequently
causes these sensors to drift and to give unreliable readings.
Examples of data from one CTD mounted optode measurement, along with Winkler
analyzed data from discrete samples, are presented in Fig. 9. This Figure shows an oxygen
optode profile down to 5500 m off Guinea in the Equatorial Atlantic. Oxygen concentrations

25 obtained by Winkler titrations on collected water samples are also included in the figure. The

1	average difference between the Winkler values and the optode readings was $2.1\pm5.1~\mu M.$ No
2	remaining pressure effects (hysteresis) were discovered and the sensor was reading 252 μM in
3	the air before the deployment and 253 μ M after.
4	
5	Discussion, comments and recommendations
6	
7	The data presented above strongly suggest that the lifetime-based methodology and the
8	oxygen optode sensors evaluated here are well-suited for measurements of dissolved oxygen
9	in the aquatic environment. The sensor specifications and performance can be summarized as
10	follows:
11	
12	• Precision ($\pm 1 \ \mu M$) and accuracy ($\pm 2 \ \mu M$) similar to what can be achieved by Winkler
13	titration.
14	• Long-term stability for more than 600 days.
15	• No stirring sensitivity since the sensor does not consume any oxygen
16	• Temperature has an influence on the luminescence quenching and has to be corrected for.
17	The sensors evaluated here had built-in temperature sensors. The corrections applied are
18	adequate and performed internally and automatically.
19	• Salinity has to be corrected for by using standard formulas given in the literature (Garcia
20	and Gordon, 1992).
21	• Pressure has an effect on the sensor response by lowering the signal with 4 % per 1000 m
22	(100 bar) of water depth. This effect is fully and instantaneously reversible (no
23	hysteresis).

• The 90 % response time (t_{90}) is 47 s for optically isolated and 10 s for non-isolated foils.

1	• This and other studies (Klimant et al., 1995) have demonstrated that there exists no cross
2	sensitivity to, and the sensors do not get damaged by, hydrogen sulfide (H ₂ S), carbon
3	dioxide (CO ₂), ammonia (NH ₃), pH, any ionic species like sulfide (S^{2-}), sulfate (SO_4^{2-}) or
4	chloride (Cl ⁻). The sensors can also be used in methanol-water and ethanol-water mixtures
5	as well as in pure methanol and ethanol.
6	• Interferences (cross-sensitivity) were found only for gaseous sulfur dioxide (SO ₂) and
7	gaseous chlorine (Cl ₂).
8	• Biofouling resistance depends on the environment. In the marine environment longer
9	deployment times (months) can be achieved by wrapping a beryllium-copper alloy net
10	(used for domestic cleaning) around the sensor. In waste water flushing or a redesign of
11	the sensor is needed to achieve longer deployment times.
12	
13	The performance of this method in general and of this type of sensor in particular has been
14	demonstrated here through numerous laboratory and field examples. When comparing to the
15	general behavior of electrochemical sensors it is clear that the lifetime-based optical
16	technology is superior in every aspect, except for the fast response time that has been
17	demonstrated by electrochemical microelectrodes. With future development there is potential
18	to improve the response time of the foils / sensors but probably not to the level achievable
19	with microelectrodes ($t_{90} = 0.1$ s). Also when comparing this technology to intensity-based
20	optodes, the lifetime-based measurements have advantages mainly in terms of stability.
21	The active chemical compound used in this sensor is a platinum porphyrine complex
22	which has the advantage to yield a longer and more easily detectable lifetimes than the more
23	commonly used ruthenium complexes (e.g. Klimant et al., 1996; Stokes and Romero, 1999).
24	Ruthenium-based sensors are likely to generally perform similar to the sensors tested here but

the accuracy and precision is expected to be lower, due to the shorter lifetimes. The pressure behavior might also be different.

3 A crucial point for the long-term stability of this type of sensors is how strongly the 4 sensing compounds are bound to the support layer and how rapidly they bleach with time 5 (number of excitations). If the sensing substance dissolves with time or if it bleaches rapidly 6 the long-term stability will be compromised. The technology presented here appears to be 7 stable for years (not yet fully demonstrated) but this is chemistry dependent and the same 8 stability can not automatically be assumed for other chemistries (e.g. ruthenium complexes or 9 platinum porphyrine complexes with different ligands). Also for the electronics, the 10 mechanical and optical systems differences between other makes of sensors should be 11 expected.

Use of this sensor in waste water treatment plants and in live stock waste has demonstrated that this design, with the foil placed in a shallow depression, is not optimal for these types of applications. Organic material accumulates in the depression, the response time becomes longer and the bacteria change the local oxygen conditions. A leveled mounting of the foil would most likely improve the ability of the sensor in this type of environments since the accumulation of organic material would become lower.

18 In conclusion it has been demonstrated here that the lifetime-based sensor technology 19 provides a new potential for accurate long-term measurements of oxygen. This opens new 20 possibilities not only for environmental monitoring and research in lakes, rivers and the 21 coastal environment where oxygen variations are in general large and rapid. It also gives the 22 possibility to use oxygen as a tracer in oceanographic studies and to monitor the open ocean 23 and deep sea environments where changes might take years/decades (e.g. Körtzinger et al., 24 2004a,b). The oxygen response of an optode is exponential, yielding highest sensitivity at low 25 concentrations. A high sensitivity at low concentrations is important in environments such as

- 1 ocean oxygen minimum zones, found in e.g. the Pacific Ocean and Arabian Sea. A slight
- 2 change in the ambient oxygen level can make a drastic change in the benthic community, due
- 3 too different tolerance levels for hypoxia (Levin, 2003).

1	
2	References
3	
4	Atwood D. K., W. F. Kinard, M. J. Barcelona, and E. C. Johnson. 1977. Comparison of
5	polarographic electrode and Winkler titration determination of dissolved oxygen in
6	oceanographic samples. Deep-Sea Res. 24: 311-313.
7	Berg P., H. Røy, F. Janssen, V. Meyer, B. B. Jørgensen, M. Huettel, and D. de Beer. 2003.
8	Oxygen uptake by aquatic sediments measured with a novel non-invasive eddy-
9	correlation technique. Mar. Ecol. Prog. Ser. 261: 75-83.
10	Berntsson M., A. Tengberg, P. O. J. Hall, and M. Josefsson. 1997. Multivariate experimental
11	methodology applied to the calibration of a Clark type oxygen sensor. Analyt. Chim.
12	Acta. 355: 43-53.
13	Bourget E., P. L. Ardisson, L. Lapointe, and G. Daigle. 2003. Environmental factors as
14	predictors of epibenthic assemblage biomass in the St. Lawrence system. Estuar. Coast.
15	and Shelf Sci. 57: 641-652.
16	Briggs R., and M. Viney. 1964. The design and performance of temperature compensated
17	electrodes for oxygen measurement. J. Sci. Instrument. 41: 78-83.
18	Bültzingslöwen (von) C., A.K. McEvoy, C. McDonagh, B.D. MacCraith, I. Klimant, C.
19	Krause, and O. S. Wolfbeis. 2002. Sol-gel optical carbon dioxide sensor employing dual
20	luminophore referencing for application in food packaging, The Analyst. 127: 1478-
21	1483.
22	Clark L. C. JR. 1959. Electrochemical device for chemical analysis. US Patent number
23	2,913,386.
24	Demas J. N., B. A. De Graff, and P. Coleman. 1999. Oxygen Sensors Based on Luminescence
25	Quenching. Analyt. Chem. 71: 793A-800A.

1	Francois, R., S. Honjo, R. Krishfield, and S. Manganini. 2002. Factors controlling the flux of
2	organic carbon to the bathypelagic zone of the ocean. Glob. Biogeochem. Cycl., 16: art.
3	no1087.
4	Garcia H. E., and L. I. Gordon. 1992. Oxygen solubility in seawater: Better fitting equations
5	Limnol. Oceanogr. 37:1307-1312.
6	Glud R. N., J. K. Gundersen, and N. B. Ramsing. 2000. Electrochemical and optical oxygen
7	microsensors for in situ measurements. In situ monitoring of aquatic systems: Chemical
8	analysis and speciation. John Wiley & Sons Ltd (eds J. Buffle & G. Horvai). Chapter 2:
9	19-73.
10	Glud R. N., I. Klimant, G. Holst, O. Kohls, V. Meyer, M. Kühl and J. K. Gundersen 1999. Adaptation, test
11	and in situ measurements with O2 microopt(r)odes on benthic landers. Deep-Sea Res. I 46: 171-183.
12	Glud R. N., M. Kühl, O. Kohls, and N. B. Ramsing. 1999. Heterogeneity of oxygen production and
13	consumption in a photosynthetic microbial mat as studied by planar optodes. J. Phycol. 35: 270-279.
14	Glud R. N., A. Tengberg, M. Kühl, P. O. J. Hall, and I. Klimant. 2001. An in situ instrument for planar O2
15	optode measurements at benthic interfaces. Limnol. Oceanogr. 46: 2073-2080.
16	Glud R. N., F. Wenzhofer, A. Tengberg, M. Middelboe, K. Oguri and H. Kitazato. 2005. Benthic oxygen
17	distribution in central Sagami Bay, Japan: In situ measurements by microelectrodes and planar optodes.
18	Deep Sea Res. I, accepted.
19	Gneigner E., and H. Forstner. 1983. Polarographic oxygen sensors. Springer, Heidelberg, Germany.
20	Gundersen J. K., N. B. Ramsing, and R. N. Glud. 1998. Predicting the signal of O ₂
21	microsensors from physical dimensions, salinity and O ₂ concentration. Limnol.
22	Oceanogr. 43: 1932-1940.
23	Haus F., O. Boissel, and G. A. Junter. 2003. Multiple regression modelling of mineral base oil
24	biodegradability based on their physical properties and overall chemical composition.
25	Chemosphere, 50: 939-948.

1	Holst G., R. N. Glud, M. Kühl and I. Klimant. 1997. A microoptode array for fine-scale measurements of
2	oxygen distribution. Sens. Actuat. 38-39: 122-129.
3	Holst G., O. Kohls, I. Klimant, B. König, M. Kühl, and T. Richter. 1998. A modular luminescence
4	lifetime imaging system for mapping oxygen distribution in biological samples. Sens. Actuat. 51: 163-
5	170.
6	Holst G., M. Kühl, and I. Klimant. 1995. A novel measuring system for oxygen microoptodes based on a
7	phase modulation technique. Proceedings SPIE 2508, 45, 387-398.
8	Hitchman M. L. 1978. Measurements of dissolved oxygen. In: Chemical Analysis, 49, Wiley,
9	New York, USA, pp. 255.
10	Huber C., I. Klimant, C. Krause, T. Werner, T. Mayr, and O.S. Wolfbeis. 2000. Optical
11	sensor for seawater salinity. Fresenius J. Analyt. Chem. 368: 196-202.
12	Huber C., I. Klimant, C. Krause, T. Werner, and O. S. Wolfbeis. 2001a. Nitrate-selective
13	optical sensor applying lipophilic fluorescent potential-sensitive dye, Analyt. Chim.
14	Acta. 449: 81-93.
15	Huber C., I. Klimant, C. Krause, and O. S. Wolfbeis. 2001b. Dual lifetime referencing as
16	applied to a chloride optical sensor. Anal. Chem. 73: 2097-2103.
17	Hulth S., R. C. Aller, P. Engström, and E. Selander. 2002. A pH plate fluorosensor (optode)
18	for early diagenetic studies of marine sediments, Limnol. Oceanogr. 47: 212-220.
19	Kanwisher J1959. A polarographic oxygen electrode. Limnol. Oceanogr. 4: 210-218.
20	Kautsky H. 1939. Quenching of luminescence by oxygen. Transac. Faraday Soc. 35:216-219.
21	Klimant I., C. Huber, G. Liebsch, G. Neurauter, A. Stangelmayer, and O. S. Wolfbeis. 2000.
22	Dual Lifetime Referencing (DLR) - a New Scheme for Converting Fluorescence
23	Intensity into a Frequency-Domain or Time-Domain Information. In Fluorescence
24	Spectroscopy: New Methods and Applications, Valeur B. & Brochon C. eds. Springer
25	Berlin (2000).

1	Klimant I., M. Kühl, R. N. Glud, and G. Holst. 1996. Optical measurement of oxygen and temperature in
2	microscale: strategies and biological applications. Sens. Actu. B, 000-000: 1-9.
3	Klimant I., V. Meyer, and M. Kohls. 1995. Fibre-optic oxygen microsensors, a new tool in aquatic
4	biology. Limnol. Oceanogr. 40: 1159-1165.
5	Körtzinger, A., J. Schimanski, and U. Send. 2004a. High-quality oxygen measurements from profiling
6	floats: A promising new technique. J. Atmos. Ocean. Techn., 22: 302-308.
7	Körtzinger, A., J. Schimanski, U. Send, and D.W.R. Wallace. 2004b. The ocean takes a deep breath.
8	Science, 306: 1337.
9	Levin, L. A. 2003. Oxygen minimum zone benthos: Adaptation and community response to hypoxia.
10	Oceanogr. Mar. Biol., 41: 1-45.
11	Revsbech N. P. 1989. An oxygen microsensor with a guard cathode. Limnol. Oceanogr. 34: 474-478.
12	Sayles F. L., W. R. Martin, and W. G. Deuser. 1994. Response of benthic oxygen demand to
13	particulate organic carbon supply in the deep sea near Bermuda. Nature. 371: 686-689.
14	Short D. L., and G. S. G. Shell. 1984. Fundamentals of Clark membrane configuration oxygen
15	sensors: some confusion clarified. J. Phys. E. Sci. Instrum. 17: 1085-1092.
16	Skjelhaugen O. J. 1999. Thermophilic aerobic reactor for processing organic liquid wastes.
17	Water Res. 33(7): 1593-1602.
18	Stokes M. D., and G. N. Romero. 1999. An optical oxygen sensor and reaction vessel for
19	high-pressure applications. Limnol. Oceanogr. 44: 189-195.
20	Tengberg A, J. Hovdenes, D. Barranger, O. Brocandel, R. Diaz, J. Sarkkula, C. Huber and A.
21	Stangelmayer. 2003. Optodes to measure oxygen in the aquatic environment. Sea Tech.,
22	44(2): 10-15.
23	Weiss, R. F. 1970. The Solubility of Nitrogen, Oxygen, and Argon in Water and Sea Water.
24	Deep-Sea Res., 17: 721-735.

- 1 Wenzhöfer F., O. Holby and O. Kohls 2001. Deep penetrating benthic oxygen profiles
- 2 measured in situ by oxygen optodes. Deep-Sea Res. I 48: 1741-1755.
- 3 Winkler L. W. 1888. Die Bestimmung des im Wasser gelosten Sauerstoffes. Ber. Dtsch.
- 4 Chem. Ges. Berlin. 21: 2843-2846.
- 5 Wolfbeis O.S. 1991. Fiber optic chemical sensors and biosensors. Volumes I+II, CRC Press,
- 6 Boca Raton.

1 Table 1: Description of different tests and their major conclusions.

Test [#]	Test;	Location;	Main physical	Comparisons	Major	Data
(number	(test period)	Equipment	settings*	with other	conclusions	in
of	· • /	• •	O=O ₂ ; T=Temp	methods		Fig #
sensors			pH=pH; P=Pres			Ū
used)			Sa=Salinity			
, i i i i i i i i i i i i i i i i i i i			St=Stirring			
1	Accuracy of	Lab;	O:0-500; T:2-40	Conc.	Batch calib:Accuracy $\pm 5 \mu M$	Not
(about 20)	different calibration	Mass flow	pH:7; P:1	is known	Individual calib: Accuracy $\pm 2\mu M$	shown
	methods	controllers,	Sa:0	from the	Precission $\pm 1 \mu M$	
	(5 days)	thermobath	St:from gas mix	gas mix		
2	Multivariate	Lab;	O:0,115,230; T:6,13,20	Winkler	Sal influence, corrected	Fig.2
(3)	calibration	Incubation	pH:5,6.5,8; P:1,250,500		with standard equations;	Fig.3A
	(10 days)	device,	Sa:0,17,34		Press effect 4% per 100 bar,	Fig.3B
		press. chamb.	St:0,25,50 RPM		linear; No sensor is individual	
3	Pressure	Lab;	O:380; T:7	No	Press effect 4% per 100 bar,	Fig.2
(3)	Cycling	Incubation	pH:7; P:3-405		linear, no remaining pressure	Fig.4
	(1 day)	device,	Sa:0		effects (hysteresis).	-
		press. chamb.	St:0 RPM		No sensor is individual.	
4	Response time	Lab;	O:0,284; T:20	No	Isolation: t_{67} =23s, t_{90} =47s, t_{99} =95s	Not
(3)	of isolated & non	Air & N ₂	pH:7; P:1		No isolation:t ₆₇ =6s, t ₉₀ =10s, t ₉₉ =18s	shown
	isolated foils	bubbled	Sa:0			
	(1 day)	water	St:from bubbling			
5	Long term	Off Canada	O:295; T:3	Water	No drift detected	Fig.5
(1)	stability I	in Labrador	pH:ND; P:180	horizon	Average conc. = $295 \mu\text{M}$	
. /	(600 days)	Current;	Sa:35	with const.	Standard dev. = $0.7 \mu M$	
	(***	On Argo float	St:natural circulation	oxvgen	values = 80 samples	
6	Long term	Off Guinea	O:228-253; T:28	Surface	The floats are located at different	Fig.6
(37)	stability II	in surface	pH:ND; P:1	zone	positions. They give different	0
()	(600 days)	water;	Sa:35	close to	saturation values 94-104 %. No	
		On Rafos floats	St:natural circulation	100 % sat	indication of drift.	
7	River fouling	Orge River	O:170-300; T:15-22	Electro-	Electrochemical sensor	Fig.7
(1)	sensitivity	(France);	pH:ND; P:1	chemical	affected by fouling after 2 days;	č
(-)	(20 days)	Monitoring	Sa:0	sensor	Naked Optode no influence	
	(= • • • • • • • • • •	station	St:natural circulation		of fouling within 20 days	
8	Waste water	Various	O:0-150; T:10-25	Several	Naked optode stable 14 days.	Not
(2)	fouling	waste water	pH:ND: P:1	Electro-	Copper plate has no effect.	shown
(-)	sensitivity	treatment	Sa:0	chemical	With water jet stable	
	(90 days)	plants	St:natural circulation	sensors	more than 90 days	
9	Slurry of	Bioreactor	O:0-90; T:35	No	Above 6 % Total Solids	Not
(1)	livestock		pH:ND; P:2		the mixing in the slurry is	shown
. /	waste		Sa:0		not sufficent to	· ·
	(2 hours)		St:no		achieve good readings	
10	On-line	R/V Meteor	O:190-215: T:26-30	Winkler	50400 values, no spikes	Fig.8
(1)	ship system	expedition;	pH:ND; P:2		Aver. diff. Winkler-Optode 0.9 μ M,	
	accuracy	Brasil-Africa:	Sa:31-36		Standard dev. = 1.1 µM	
	(35 days)	Flowthrough	St:throughflow		Winkler values = 72	
11	Profiling	Off Guinea:	0.17-233. T.2-28	Winkler	High corr Winkler and Ontode	Fig 9
(1)	on CTD	Optode logger	pH:ND: P:1-550	() miller	Average difference = 2.1 µM	1.8
()	accuracy	on CTD	Sa:33-26		Standard dev = $5.1 \mu M$	
	(10 hours)	with bottles	St:natural circulation		Winkler values = 14	

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* Oxygen Concentrations given in µmol/l; Temperature in degrees C; Pressure in bar and Salinity in ppt

1 Figure legends

2

3	Figure 1. Optical design and an outside view of the evaluated optode-based oxygen sensor.
4	Figure 2. Set-up that was used for multivariate calibration and for pressure testing of three
5	optodes in parallel. Temperature is measured by the optodes and pressure by a separate
6	pressure sensor placed in the center of the top plate.
7	Figure 3. Results from multivariate calibration experiments. Oxygen readings obtained with
8	three parallel optodes plotted against oxygen readings obtained by Winkler titrations.
9	Figure 3A shows the raw data and figure 3B the data after pressure (4 % per 100 bar)
10	and salinity compensation (according to Garcia and Gordon, 1992).
11	Figure 4. Oxygen readings from three optodes during pressure cycling between 3 and 405 bar.
12	The pressurization (from 3-405 bar) took approximately 4 minutes. The depressurization
13	lasted approximately 1 minute.
14	Figure 5. Five hundred and eighty days of oxygen and density data collected from a free
15	drifting Argo float off the east coast of Canada in the Labrador Current. The presented
16	data shows oxygen readings as the float passes at 1800 m depth. As long as the float was
17	in the Labrador Current (the first 400 days) it is considered that salinity and temperature
18	and oxygen readings at this depth should be constant. The data demonstrates no drift
19	over the given time period but as the float moves out of the Labrador Current the
20	oxygen and density readings shift slightly. The average oxygen readings during the
21	whole period were $295.0 \pm 0.7 \ \mu \text{mol L}^{-1}$.
22	Figure 6. Five day average of oxygen saturation values recorded with optodes mounted on 37
23	Rafos floats drifting at the surface at different locations off Guinea (East Atlantic). The
24	data were recorded after each of the floats had accomplished a 600 day deployment in
25	deeper layer. The optodes were mounted approximately 2 m below the surface. Since all

1	the floats are drifting in oligotrophic waters with low production/consumption oxygen
2	saturation values around 100 % should be expected.
3	Figure 7. Oxygen data from a comparison between an optode and an electrochemical sensor at
4	a monitoring station in the River l'Orge (France). Two occasions when the
5	electrochemical sensor was taken up, cleaned and recalibrated are marked with arrows
6	in the figure. The optode was never cleaned during this experiment.
7	Figure 8. Tropical surface ocean oxygen data collected during a cruise to the tropical Atlantic
8	Ocean. Optode measurements were made on seawater provided by an underway
9	pumping system whereas Winkler titrations were made on water samples taken with
10	Niskin bottles from regular hydrocasts.
11	Figure 9. Optode (continuous line) and Winkler data (dots) from CTD deployment to 5500 m

12 depth in the equatorial Atlantic off Guinea.





1 Fig. 3A & 3B





















1	Appendix 1
2	
3	Salinity compensation of oxygen readings from optode and calculation of % saturation at a
4	given salinity and temperature (Garcia and Gordon, 1992).
5	
6	Input variables:
7	$[O_2] = O_2$ concentration in μ M, from the optode
8	t = Temperature in degrees Celsius, from the optode temperature sensor
9	S = Salinity in ppt, from a parallel salinity measurement
10	
11	Empirical constants:
12	$A_0 = 2.00856$
13	$A_1 = 3.22400$
14	$A_2 = 3.99063$
15	$A_3 = 4.80299$
16	$A_4 = 9.78188e-1$
17	$A_5 = 1.71069$
18	$B_0 = -6.24097e-3$
19	$B_1 = -6.93498e-3$
20	$B_2 = -6.90358e-3$
21	$B_3 = -4.29155e-3$
22	$C_0 = -3.11680e-7$
23	
24	Intermediate calculations:
25	

1
$$T_s = \text{Scaled Temperature} = \ln \left[\frac{298.15 - t}{273.15 + t} \right]$$

2
3 $\alpha = S(B_0 + B_1 \cdot T_s + B_2 \cdot T_s^2 + B_3 \cdot T_s^3) + C_0 S^2$
4
5 $\beta = A_0 + A_1 T_s + A_2 T_s^2 + A_3 T_s^3 + A_4 T_s^4 + A_5 T_s^5$
6
7
8 Final results:
9
10 Salinity compensated oxygen concentration
11
12 $O_2 [\mu M] = [O_2] * e^{\alpha}$
13
14 100 % oxygen saturation with air at a pressure of 1013 mbar:
15
16 $O_{2,100\% \text{ sat}} [\mu M] = e^{(\beta + \alpha)} / 2.2414$
17
18 percent saturation in the given water
19
20 $O_{2,\text{sat}} [\%] = O_2[\mu M] / O_{2,100\% \text{ sat}} [\mu M]$