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Relationships between oyster mortality patterns and environmental data from monitoring databases along the coasts of France

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Abstract:

Oyster mortality was monitored at multiple sites along the French coastline (~ 5 500 km) between 1993 and 2005. Mortality data for one- and two-year-old oysters were collected over 10-12 years in 39 oyster culture sites bordering 3 different "seas": the English Channel, Bay of Biscaye and Mediterranean. Combining these data with records from environmental monitoring databases, 11 of these sites had consistent chronological data sets including both environmental data and mortality records. Mortality in one-year-old oysters was clearly summer mortality (49% of their annual mortality) whereas mortality of two-year-olds occurred mostly in spring (51%). Analysis of variance revealed that "coastal area" was the main influence on mortality of one-year-olds (77.5%% of the variance) and that "year studied" was the main influence on mortality of two-year-olds (60.6% of the variance). The highest mortalities occurred in Marennes and in several sites in Brittany for both age groups, and in Veys Bay (Normandy) for two-year-old oysters only. Environmental parameters were then analysed to investigate which of these might influence summer mortality. Principal Component Analyses revealed that environmental factors such as chlorophyll a (food resource indicator) and salinity (watershed effect) influence oyster mortality. Chlorophyll a concentration (10% of the variance), water temperature (7% of the variance) and turbidity (5% of the variance) are the main significant factors for the mortality of one-year-olds, while salinity and chlorophyll a have more effect on the mortality of two-year-old oysters (respectively 5% and 4% of the variance).

Keywords: Pacific oyster; Crassostrea gigas; Environment; Mortality

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Introduction

Since the beginning of the 1960s, the number of reports of "abnormal" mortality (mortality > 30% of the population) in the oyster *Crassostrea gigas* has increased throughout the world (Beattie et al., 1980; Farley, 1992; Imaï et al., 1965; Perdue, 1983; Sinderman, 1976). The first reports of an episode of such mortality were made in Japan and date from as early as 1915 (Takeuchi et al., 1960). The west coast of North America was particularly affected from the 1960s to the 1980s, when summer mortality events destroyed up to 60% of *C. gigas* livestock (Glude, 1975; Koganezawa, 1975). Severe mortalities in France (affecting more than 30 % of the cultivated population) occurred in Marennes-Oléron Bay in 1976-77 (Heral, et al., 1978), in Arcachon Bay (south-west Atlantic coast) in the 1980s and 1990s (Maurer et al., 1986), in Marennes-Oléron in 1988 (Bodoy et al., 1990) and 1993 (Lodato, 1997), in west Brittany and north-west Normandy in 1994-95 (Goulletquer et al., 1998; Fleury et al., 2001), and in Normandy in 2001 (Costil et al., 2005).

Several causal factors have been proposed to explain summer mortalities: these include water temperature, salinity, "exceptional" climatic conditions (Koganezawa, 1975; Ventilla, 1984). Physiological and/or metabolic disturbances have often been observed to accompany mortality episodes (Mori, 1979; Perdue et al., 1981; Maurer et al., 1986; Soletchnik et al., 2005); mortality events are sometimes associated with the over-maturation of the gonads (Perdue et al., 1981). Eutrophication has been implicated in some cases of mortality (Cho and Kim 1977; Tamate et al., 1965) and many studies recorded a decrease in dissolved oxygen (DO) (Cheney et al., 2000) associated with oyster mortality. Rainfall and floods can also be involved in processes that influence mortality (Bodoy et al, 1990; Calvo et al., 1999): either because of drastic freshwater discharges, which bring about a decrease in DO (Cheney et al., 2000; Cho and Kim, 1977), or because of pollution by solutes or particulate matter carried by this water. In some cases, mortality events are clearly of pathogenic origin (Elston et al., 1987; Meyers et al., 1990; Farley, 1992; Renault et al., 1995; Waechter et al., 2002; Le Roux et al., 2002).

Summer mortalities can affect animals of both one and two years old (Maurer et al., 1986), and both diploids and triploids (Calvo et al., 1999). Different susceptibilities have been

reported between *C. gigas* stocks from different origins and between different hatchery-bred families (Beattie et al., 1980; Soletchnik et al., 2002; Ernande et al., 2004; Degre mont et al., 2005), which indicate there is genetic variability for resistance to summer mortality.

Considered together, these studies indicate that for the most part summer mortalities cannot be explained by a single factor, but rather by the combination of environmental (biotic and abiotic) and internal (genetic, physiological and immunological) parameters. The present study is part of the French national MOREST project on oyster mortality. MOREST attempted from 2002 to 2006 to classify the importance of different factors involved in summer mortalities.

Summer mortalities have been progressively integrated into surveys of oyster culture in France (Bodoy et al, 1990; Le Bec and Mazurié, 1992) and elsewhere in the world (Brown and Hartwick, 1988). In 1992, IFREMER (Institut Français pour la Recherche et l'Exploitation de la MER) merged several of its regional regular surveys into a national monitoring program, REMORA, for coordinated collection of growth and mortality data (Fleury et al., 2001). The REPHY data collection, started in 1984 to protect consumers from toxic algae in seafood, holds records of hydrological parameters that could be valuable for understanding the environmental context of summer mortality episodes (Belin, 1998; Beliaeff et al., 2001).

Biological results obtained from the REMORA database in Brittany from 1993 to 1998 were presented in a previous publication (Fleury et al., 2001). In the present study this approach was broadened, with data collected over a longer time period, 1993-2005, along the entire French coastline. This paper investigates oyster mortality in 9 "Coastal Areas" including 20 "Farming Areas" and 39 experimental "Sites", and considers two age classes: one-year-old oysters (studied since 1995) and two-year-old oysters (studied since 1993). Patterns of mortality are described according to age, site, year and season. Further analysis then combines data from environmental databases (REPHY and Météo France) with data from REMORA, to attempt to identify causal factors. Genetic effects have already been found to account for around 45% of summer mortality variation in the MOREST experiments (Degremont et al., 2005), this paper aims to investigate how environmental factors influence mortality.

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Material and methods

Oyster mortality survey (REMORA) and environmental databanks (REPHY and Méteo France).

This study examined the environmental factors influencing mortality bringing together mortality data from the REMORA database and environmental data from REPHY and Météo France. The laboratories working on REMORA collect data on oyster rearing performances using the same standardized methods (Fleury et al., 2001). Two oyster groups, the first 7-8 months old ("one-year-old oysters", around 1 g mean whole weight) and the second 20 months old ("two-year-old oysters", 25-35 g after calibration), were placed in 39 "sites" (Fig 1.) in February of each year (from 1993 to 2005). Each site was stocked with one oyster meshed bag (traditionally used by oyster farmers) for each oyster group. Both age classes were renewed each year: yearling oysters from spat collecting area of Arcachon Bay, and two-year-old oysters from one common site in Brittany, most often the Morbihan Gulf. Oysters were monitored from March (year "n") to April of the following year (year "n+1"). Three scorings were made per cycle, in June (S1), September (S2) and November-December (S3), corresponding to spring, summer and autumn (Fleury et al, 2001). At each sampling, mortality was scored by an exhaustive counting of live and dead oysters. This survey covered three different seas (the English Channel, Biscaye Bay from Atlantic ocean and Mediterranean sea). Twenty oyster farming areas were then delimited (according to traditional production basins). Within these 20 farming areas, distributed between latitude 43° and 50° North on the French coastline, 39 sites were defined (eg BV01, BV02, and BV03 in the coastal area of Veys bay). Thirty-five of these sites are located within intertidal zones, at levels which are emerged about 20% of the time. The four other sites remain immersed continually. Three of these were in a Mediterranean lagoon (TH - Fig. 1) and one in a subtidal deep-water oyster culture area in Quiberon Bay (QB02 - Fig. 1). For variance analyses with necessarily replicates, the 39 sites are grouped in 9 coastal areas ecologically coherent (eg. sites CO01, CO02, CO03 from west Cotentin are grouped with eastern west Brittany sites (Cancale CA02 and Paimpol PL03); table 1 and Figure 1 and 3).

The REPHY network has been monitoring phytoplankton and phycotoxins along French coasts since 1984. Its main role is to detect any algal blooming that could become a public health hazard through the consumption of contaminated seafood (Belin, 1998). Every fortnight, at each of the 43 sampling points of the network, phytoplankton species are identified and counted, and measurements are made of water temperature (°C), salinity (PSU), turbidity (NTU), dissolved oxygen (%), chlorophyll a (µg L⁻¹) and pheophytin (µg L⁻¹).

Daily data of air temperature (°C), rainfall (mm), sun exposure (hours) and radiation (Joules per cm²) measured at stations close to the different coastal areas of this study were obtained from Météo France.

Monthly values were calculated for all environmental and biological parameters for analysis purpose. Complete temporal series (8-10 years monitoring) were selected from REPHY to be combined with the REMORA data set. Out of the 20 coastal areas studied, only 11 had consistent datasets for both biological and environmental parameters allowing multivariate analyses to be performed (Fig.1 and Table 2).

Parameters of the study

The spatial scale of the analyses was either site or coastal area. Mortality was measured over spring (S1), summer (S2) and fall (S3) seasons on the same groups of oysters grown in bags. Oyster mortality was calculated per bag as: [(initial number – final number) / initial number] for the seasons S1, S2 and S3, where initial number was the number in the bag at the start of each season (i.e. the survivors from the previous seasons for S1 and S2). When mortality and environmental data were combined in the second part of the study, the mortality measurement used was that which combined both seasons (S1 + S2). The environmental data set contained monthly averages of water temperature (°C), salinity, turbidity (NTU), dissolved oxygen (%), chlorophyll a (μ g L⁻¹) and pheophytin (μ g L⁻¹) from the REPHY network and mean air temperature (°C), monthly cumulated rainfalls (mm) and sun exposure (hours) from the Météo France network.

Statistical analysis

Environmental and mortality data were log transformed and normalized for statistical analysis. A Principal Component Analysis (PCA) was performed to present hydrological characteristics of 11 farming areas on a monthly basis and over the 12 years. PCAs and linear regressions were performed on non-transformed mortality an environmental data to highlight relations between mortality and environmental parameters. Environmental data were averaged on different periods (ie. whole year centred on winter (September n-1 to September n; labelled 9-9), autumn-winter (September n-1 to March n, labelled 9-3), winter (December n-1 to March n, labelled 12-3), winter-spring (December n-1 to June n, labelled 12-6), spring (March n to June n; labelled 3-6), spring-summer (March n to September n; labelled 3-9) and summer (June n to September n, labelled 6-9)) to point out the parameter and the period of year which explain the most oyster mortalities. Autumn of the precedent year (from September n-1 to December n-1, 9-12) can not be used because of lack of data. These analyses were performed with the Statgraphic V.5.1 software and the XLStat-Pro v.7.1 software.

Results

Mortality of the two classes of oysters

Seasonal effects on mortality

Mortality levels, monitored over the twelve years, for the two age groups, at the 39 rearing sites, are given in Fig. 2. Age had no significant effect on overall mortality levels, which were in the range of 5-14% for one-year-old oysters and 7-13% for two-year-olds across all seasons (p=0.48, Table 3). The "season" factor alone had highly significant effect on oyster mortality (p<0.0001, Table 3). Mean mortality levels were $2.74\%(\pm 0.21)$, $5.31\%(\pm 0.21)$ and $5.77\%(\pm 0.21)$ respectively during fall, spring and summer (Table 4). The interaction between age and season was also highly significant, despite the fact that age effect alone was not significant. The highest mortality in one-year-old oysters (7.26%) occurred in summer,

whereas the highest mortality of two-year-old oysters occurred in spring (6.61%, Table 4). This difference in seasonal mortality with age explains the significant interaction between age and season (Table 3).

Effects of location and year on oyster mortality

Analyses of variance were then performed using 9 coastal areas (Fig. 1; Table 1). Effects of "year" and "coastal area" on oyster mortality were thus compared on a finer scale (Table 5). One-year-old oysters were more sensitive to the "coastal area" effect than were two-year-olds. The "coastal area" variance component was 77.5 % for this class of age but only 33.5% for the two-year-old oysters (Table 5). However, "year" effect accounted for 60.6% of variance in summer mortality of two-year-old oysters but only 14.5% in one-year-olds.

For the one-year-old oysters, high mortality values (15-30%) were mostly concentrated at Marennes-Oléron Bay: mortality reached 40% at the station NMA02 in the central zone of Marennes-Oléron Bay. Mor Bras (QB02, PF02) and North of Finistère (MX02, mortality significantly higher than other sites within this area) have quite elevated mortalities too (Table 6 and Fig.2). In one-year-olds, the lowest mortality rates (ranging from 7.2 to 9%) were reached during the years 2000, 1996 and 2002. The highest mortality rate (17.5%) was reached in 1995; it was significantly higher than other years (p<0.05).

For two-year-old oysters though, year of rearing appeared to account for the clearest differences in mortality (Fig. 2). In two-year-olds, mortalities encountered in 1999, 1995 and 2000, ranging from 13.7% to 19.9%, were significantly higher than in other years (p<0.05) (Table 6). At the other end of the scale, 2002 and 2005 had the lowest mortality rates (4.4%-4.9%). Most recent years show moderate levels of mortality (2002 to 2005 are within the five lowest levels of mortality)(Fig. 2). Some sites also seem to be particularly affected, like BV02 (Veys Bay, significantly higher than other sites within Normandy coastal area), NM03 (Poitou-Charentes) and PB02 (significantly higher than other sites within Mor Bras coastal area) (Fig. 2).

In a few sites, high mortality affected both one- and two-year-old oysters (MX02, PB02; NM03; SM06 and NM01) (Fig. 2). Only in one site though (BV02) was mortality of two-year-olds higher (28.2%) than that of one-year-olds (9.1%). Many sites did not present high mortalities in either age group: Normano-breton Gulf, Bourgneuf, Arcachon Basin and Thau Lagoon.

Relationships between summer mortalities and environmental conditions

The mortality results therefore showed the main effects to be "site" and "year" for one and two-year-old oysters respectively. The next step was to analyse the effect of environmental parameters on oyster mortality more precisely in the eleven sites where data was available for both (Table 2).

Environmental characteristics of the coastal areas

Monthly averages of hydrological and meteorological parameters are shown as box and whisker plots for 11 of the farming areas (Fig. 3). Mean air temperature was around 12°C from the English Channel to Brest, and then rose from 13°C in South Brittany to 16°C on the Thau lagoon (Mediterranean sea) (Fig. 3a). The pattern was almost the same for water temperature (Fig. 3b). Rainfall ranged between 20 and 100 mm per month (first two quartiles) for all the coastal areas, apart from North of Finistère where rainfall was 60-140 mm (Fig. 3c). Rainfall showed an inverse relationship with sun exposure, which had the lowest number of hours per month (median 120) for the two sites in this area (Morlaix Bay and Brest roadstead). Sun exposure was slightly longer further to the North (Veys Bay and Normano-breton Gulf), but rose up to 180 in the south part of the Bay of Biscay, and 220 on the Mediterranean (Fig. 3e). In this latter coastal area, salinity increased up to 40 due to summer evaporation (Fig. 3d). Most Atlantic sites (BR, QB, BO, NMO, SMO, AR), exposed to large fluvial discharges (Loire, Gironde), present typical estuarine cycles with salinity decreasing to 26-28 during winter (at high tide). English Channel sites (CA, PL, MX) at the opposite present higher and more stable salinities (Fig. 3d). Suspended particulate matter

was gauged via chlorophyll a concentration, turbidity and pheophytin (Fig. 3f, g, h). The highest chlorophyll a concentrations were found in Veys Bay, Brest roadstead, Marennes-Oléron Bay and Thau lagoon (Fig. 3f). The highest turbidity levels were observed in Bourgneuf Bay, in Marennes-Oléron Bay and in Cancale Bay (Fig. 3g).

A Principal Component Analysis (PCA) was used to display mean monthly hydrological change in the coastal areas (Fig. 4). Morlaix river (MX) and the bays of Paimpol (PL) and Cancale (CA) in North Brittany, and Quiberon Bay (QB) in South Brittany all presented a similar pattern with a low chlorophyll a and pheophytin concentration in the water throughout the year (group 1). Monthly values were mainly distributed along the first axis of temperature, salinity and turbidity, showing a slight increase in temperature and salinity from January to July-August and a decrease during fall and early winter. Arcachon basin (AR) showed a similar pattern, but with a higher level of chlorophyll a and pheophytin (group 2). Thau lagoon (TH) was characterized by the highest summer temperature and salinity (group 3). High seasonal change in salinity occurred in the Bourgneuf Bay (BO) (group 4), and in Marennes-Oléron Bay (SM and NM) (group 5) where it was associated with the spring bloom (chlorophyll a increase). Brest roadsted (BR) and Veys Bay (BV) were characterized by high chlorophyll a seasonality in a low range of salinity (group 6).

Relationships between environmental parameters and summer mortality

Two Principal Component Analyses (PCAs), one for each age group, were performed with the same set of data (11 sites along the French coast where environmental and mortality data are available on the 1993-2005 period, Table 2) and revealed relationships between some environmental conditions and oysters mortality.

Environmental data used in these PCAs are chlorophyll a (CHLOROA), pheophytin (PHEO)

environmental data used in these PCAs are chlorophyll a (CHLOROA), pheophytin (PHEO) and turbidity (TURB) to stand for availability and level of trophic resource, water temperature (TEMP) and then salinity (SALI) to account for freshwater inputs on the rearing site, with different means (table 7 and figure 5).

One-year-old oysters mortality is significantly linked to chlorophyll a level of summer (6-9, 12% variability explained), spring-summer (3-9, 9% variability explained), whole year centred

on winter (9-9, 8% variability explained) and autumn-winter (9-3, 6% variability explained). It is significantly linked to water temperature too, especially with the one of spring-summer (3-9, 7% variability explained), spring (3-6, 6% variability explained), winter-spring (12-6, 5% variability explained) and whole year centred on winter (9-9, 5 % variability explained). It is significantly linked to summer turbidity level too (6-9, 5% variability explained).

260 Two-year-old oysters mortality is significantly inversely linked to autumn-winter salinity (9-3,

5% variability explained) and to autumn-winter chlorophyll a level (4% variability explained).

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Using the same parameters as used in the PCAs linear models can be adjusted with the progressive regression procedure (XL-StatPro v. 7.1) to point out the best parameter to explain oysters mortality (table 8).

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Model equation for one-year-old oysters spring and summer mortality (1y-SSM):

$$1y$$
-SSM = $7,54.10^{-02} + 1,75.10^{-02}$ CHLOROA 6-9

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Model equation for two-year-old oysters spring and summer mortality (2y-SSM):

$$2y$$
-SSM = $0.16 - 9.96.10^{-03*}$ SALI $9-3 - 4.28.10^{-02*}$ CHLOROA $9-3 + 2.22.10^{-02*}$ TEMP $9-9$

These models are both significant (p< 0.05) and they respectively explain 11.9% and 16.3%

of mortality variability.

275 A polynomial model has been tested on one-year-olds mortality data versus CHLOROA 6-9

to test the hypothesis that a lack of chlorophyll a in summer could be a factor of mortality.

This model is significant (p < 0.05) and explain 16.2% of mortality variability.

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For the one-year-old oysters, the strongest parameters responsible for mortality seems to be

the one of spring-summer period (level of chlorophyll a, turbidity and temperature); for the

281 two-year-old oysters, the low salinity measurements of the autumn-winter period seems to

be the strongest parameter linked to oysters mortality;

The two "models" of one-year-old and two-year-old oysters mortality are quite different

284 regarding the period concerned for the effect (spring-summer for the one-year-old oysters

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and previous autumn-winter for the two-year-old oysters), and the parameters responsible for mortality (eg. level of chlorophyll a, water temperature and turbidity for one-year-old oysters, and salinity, water temperature and chlorophyll a (as a reducer of mortality) for two-year-olds).

Discussion

Oyster mortalities have often been attributed to "summer mortality" phenomena (Arakawa et al., 1971; Beattie et al., 1980; Cheney et al., 1998; Maurer and Comps, 1986). The present survey provides a better understanding of these events along the French coasts because mortality was monitored in spring, summer and fall. Most oyster mortalities (~ 80 %) occurred during the spring-summer period. A significant difference was observed between "spring" mortality of two-year-old oysters and "summer" mortality of one-year-old oysters. Differences in levels and periods of mortality between the age groups might be due to an effect of the maturation process, which influences summer mortality and is age-dependent (Maurer and Comps, 1986; Mori, 1979).

In this study, only 6% of the data points from juveniles showed mortality over 30%. The worst sites for "chronic" summer mortality of "juveniles" (one-year-olds) in the study were in Marennes-Oléron Bay. Mor Bras was also heavily affected, with mortalities in the Gulf of Morbihan, Quiberon Bay and Vilaine estuary. Bay of Morlaix was also affected. For two-year-old oysters, mortalities up to 30% were confined to 2% of the data points recorded during the survey. Globally, the same sites were concerned as for the mortality of one year olds, with Veys Bay (BV02) in addition.

Temperature is the environmental parameter most frequently implicated in mortality events (Cheney et al., 2000; Friedman et al., 1997; Lipovsky and Chew, 1972; Meyers et al., 1990; Shafee and Sabatie, 1986), even if no direct relationship has been established. In this study, water temperature is significantly linked to one-year-old oyster mortality on different periods of year, principally in spring-summer, even if this correlation does not explain much of the mortality variability; water temperature effect can be understood through its action on sexual maturation (which could weaken oysters towards stresses), or as a factor of stress.

On the other hand, the intensity of oyster summer mortalities was not related to the positive north-south thermal gradient along the French coastline. Water temperatures in Thau lagoon on the Mediterranean coast rose as high as 28-30°C without any lethal consequence. Middle latitude sites in the gulf of Biscay were much more affected by oyster summer mortalities than those to the north or south, which confirms that temperature is not the only causal factor. This shows the need for specific investigations in each coastal area and for a comparative typology study, to provide a better understanding of the summer mortality phenomenon.

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High trophic level has sometimes been suspected to contribute to summer mortality events in laboratory experiments (Lipovsky and Chew, 1972) or those made in the field (Tamate et al., 1965). Chlorophyll a is often considered as a trophic resource indicator (Cigarria, 1999; Heral et al., 1984; Soletchnik et al., 1996). This effect is probably related to the investment of energy from trophic resources into sexual maturation. For example, a comparison was made in Japan between the oligotrophic Onagawa Bay, where summer mortalities was negligible, and the eutrophic Matsushima Bay, where over-maturation of oocytes led to physiological and metabolic disorders inducing high mortality (Mori et al., 1965; Tamate et al., 1965). In the present study, summer chlorophyll a level showed a slight but significant correlation with juvenile mortality rate. Even though only a small amount of the variability was explained by chlorophyll a levels, this path should be investigated so as to understand how this food resource indicator can have a negative effect on survival of oneyear-old oysters, and which period of year is the most "risky". Chlorophyll a of the autumnwinter period is significantly linked to two-year-old oysters mortality, but inversely; a lack of food in autumn-winter could provoke a physiologic weakness at the beginning of spring that could enhance the stress due to sexual maturation. Previous analysis of REMORA and REPHY data had identified lowered pheophytin level as a factor involved in juvenile mortality. Heral et al., (1984) showed that food requirement differed between one- and twoyear-old oysters. Growth correlated better with pheophytin in one-year-olds and better with chlorophyll a in two-year-olds.

A low autumn-winter salinity tended to enhance mortalities in this study, as a reflect of rainfall and watershed discharge that happened during these seasons. It was specially

implicated in some mortalities of two-year-olds in ecosystems strongly influenced by river plumes. Freshwater discharge constitutes a stress factor or transports anthropic pollution (Mene squen, 1992). Two hypothesis can be put forth: a toxicity of the freshwater discharged in the farming areas, that would weaken oysters making them more sensitive to stresses during spring-summer period, or an effect through the trophic resource which production could be enhanced by inputs of nutrients. In France, susceptible sites showing more than 15% mortality of cupped oysters, such as Veys Bay, Morlaix Bay, Pen Bé and Penerf bays are directly influenced by river plumes. Some of these sites have already suffered oyster mortality crises, like Pen Bé marked by 30% mortality in 1988 and 1989 which reached 55% in the inner part of the bay (Mazurié pers. com.). Terrestrial pollution was suspected to be a likely cause in this case. Further south, Marennes-Oléron Bay also had a high mortality crisis in 1988 with estimated losses up to 7,800 t (Bodoy et al., 1990). Although pathological and toxicological analyses were made, none identified a probable cause. However exceptional rainfall was recorded at this time, accompanied by high temperatures and a delay in the seasonality of phytoplanktonic blooms (Bodoy et al., 1990). Exceptional floods and high temperatures were also implicated in mortality of Ostrea edulis during summer 1997 (Calvo et al., 1999). In Veys Bay, freshwater discharges include intense and short term discharges which go beyond the normal scale of measurement and are thus not properly shown by hydrological recordings made in this area (Costil et al., 2005). This example highlights the importance of scaling and the need for correct spatio-temporal sampling in these ecosystems. The English Channel coast (apart from Veys Bay) sites which are little influenced by freshwater discharge did not suffer significant summer mortalities. However Bourgneuf Bay (Loire influence) presented seasonal freshening (low salinity) without enduring mortality events. Geographical mapping of mortality was therefore not directly related to river influences. In south-west Florida Water management practices have drastically altered natural water quality conditions within estuaries by affecting salinity and nutrient influx (Savarese and Volety, 2001). The eastern oyster, Crassostrea virginica, is used as an indicator of ecosystem health in this area to establish target water quality conditions for restoration efforts (Savarese and Volety, 2001).

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Seasonal occurrence of mortality and effects of location and year were all dependent on the age of the oysters in this study. Over the twelve years of mortality monitoring, 1995 and 1999 (years of high mortality) contrasted with 1993 and recent years (2002-2005) (years of low mortality along the French coastline), due to interannual differences in meteorological and hydrological conditions. However these conditions also interacted with the spatial scale of the survey. Whatever the geographic scale considered, mortality differed between sites always being higher in certain: Veys Bay, Morlaix Bay, several sites in Mor Bras (Auray River, Quiberon Bay, Penbé Bay), and Marennes-Oléron Bay. A simple typology developed from hydrological parameters in the farming areas did not isolate any specific pattern that would identify a "risky" ecosystem for oyster culture. However, locations on the Normanobreton Gulf, which had lower variations in their temperature cycles and lower freshwater input, were less affected by mortality events.

On a worldwide scale, the reason for summer mortalities has been identified as a complex of multiple stress factors acting together to induce physiological disorders and metabolic disturbance (Tamate et al., 1965; Cheney et al., 2000). Mortality events result from combined factors in a different manner from one ecosystem to another. The models obtained from the present study, despite the low amount of variability explained, allow some general trends to be derived for this first integrated study of oyster mortality over the whole French coastline. They point out the fact that the environmental factors responsible for oysters mortality are different according to age of oysters, probably in relation with metabolic differences between the two ages. They also open paths for further investigation, including research into additional stress parameters which could bring about mortality events.

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Tables

- Table 1. REMORA Network: 39 experimental sites references in 20 Farming Areas and 9 Coastal Areas along the French coasts. Bold types written coastal areas and sites are those combining mortality and environmental data bank.
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- Table 7. Extracts from the correlation matrix of the Principal Component Analyses performed on mortality and environmental data; 1y-SSM stands for 1-year-old oyster spring and summer mortality; 2y-SSM stands for 2-year-old oyster spring and summer mortality.
- Table 8: Linear models analyses: R², significance and parameters analyses and estimations and equation for two and one-year-old oysters summer mortality; 1y-SSM stands for 1-year-old oyster spring and summer mortality; 2y-SSM stands for 2-year-old oyster spring and summer mortality.

Table 1

Coastal area	Farming area	Sites		
Norway and AN	Veys bay (BV)	BV01; BV02 ; BV03		
Normandy (N)	Saint Vast-La-Hougue bay (SV)	SV01; SV02;;SV03		
	West of Cotentin (CO)	CO01; CO02; CO03		
Normano-breton Gulf (GNB)	Cancale bay (CA)	CA02		
	Paimpol bay (PL)	PL03		
	Morlaix bay (MX)	MX02		
North of Finistère (FN)	Aber Benoit (AB)	AB02		
	Brest roadstead (BR)	BR03		
Not used	Etel river (EL)	EL02		
	Quiberon bay (QB)	QB01; QB02		
	Auray river (AY)	AY02		
Mor Bras (MB)	Gulf of Morbihan (GM)	GM02		
	Penerf bay (PF)	PF02		
	Pen-Bé bay (PB)	PB02		
Pays de Loire (PL)	Bourgneuf bay (BO)	BO01; BO02; BO03; BO0		
Ré Island (RE)	Ré island (RE)	RE01; RE02		
Maranna Oláran hay (MA)	North part of Marennes-Oléron bay (NM)	NM01; NM02 ; NM03		
Marennes-Oléron bay (MA)	South part of Marennes-Oléron bay (SM)	SM04; SM05; SM06		
Aquitaine (AQ)	Arcachon basin (AR)	AR01; AR02; AR03		
Mediterranean coest (ME)	Thau lagoon(TH)	TH01 ; TH02; TH03		

Table 2

Farming area (from North to South)		Hydrology - reference station	Meteorology - reference station
Veys Bay	BV	Roches de Grandcamp	Engles-PeB-BV
Cancale Bay	CA	Cancale nord	Dinard
Paimpol Bay	PL	Bréhat	Saint Brieuc
Morlaix Bay	MX	Pen al Lann	Landivisiau
Brest Roadstead	BR	Lanvéoc	Landivisiau
Quiberon Bay	QB	Men er Roue	Vannes
Bourgneuf bay	во	Bois de la Chaise	Bourgneuf
North Marennes-Oléron bay	NMA	Boyard	La Rochelle
South Marennes-Oléron bay	SMA	Auger	La Rochelle
Arcachon Basin	AR	Teychan	Cap Ferret
Thau lagoon	TH	Bouzigues	Montpellier

Table 3

Source	Df	Mortality		
		Mean squares	P value	Variance
				Component (%)
Age	1	0.007	0.4857	0.4
season	2	0.971	0.0000	51.2
Age x Season	2	0.904	0.0000	47.7
residual	30	0.013		0.7

Table 4

Age	Season	N	Mortality	Groups*
1	Spring	311	4.02 ± 0.31	*
	Summer	311	7.26 \pm 0.32	*
	Autumn	309	3.50 ± 0.32	*
2	Spring	410	6.61 ± 0.28	*
	Summer	406	4.28 ± 0.28	*
	Autumn	408	1.98 ± 0.28	*
1	All seasons	931	4.92 ± 0.18	*
2	All seasons	1224	4.29 ± 0.16	*
All ages	Spring	721	5.31 ± 0.21	*
	Summer	717	5.77 \pm 0.21	*
	Autumn	717	2.74 \pm 0.21	*

^{*}Bonferroni tests (Statgraphic V.5) were performed to distinguish significantly different groups (p < 0.05).

Table 5

٨٥٥	Course	df	Mortality		
Age	Source		Mean square	Pr > F	Variance component (%)
1 year old	Coastal Area	7	0.820	< 0.0001	77.5%
	Year	10	0.153	< 0.0001	14.5%
	Coastal Area*Year	70	0.050	0.015	4.8%
	Error	264	0.034		3.2%
2 years old	Coastal Area	8	0.261	< 0.0001	33.5%
	Year	12	0.472	< 0.0001	60.6%
	Coastal Area*Year	96	0.026	0.063	3.3%
	Error	351	0.020		2.6%

Table 6

Age	Coastal area	N	Mortality	SE		Groups	S*	
1 year old	MA	63	0.225	0.017	*			
	MB	60	0.132	0.014		*		
	FN	30	0.107	0.017		*	*	
	RE	20	0.105	0.016		*	*	
	AQ	22	0.089	0.018			*	*
	GNB	55	0.078	0.009			*	*
	PL	42	0.065	0.007				*
	N	60	0.061	0.006				*

Age	Year	N	Mortality	SE				(Group	S			
2 years old	1999	35	0.199	0.018	*								
	1995	37	0.174	0.016	*	*							
	2000	37	0.137	0.009		*							
	1994	36	0.113	0.013			*						
	1997	36	0.096	0.012			*	*					
	2001	36	0.096	0.009			*	*					
	1998	36	0.090	0.013			*	*	*				
	1996	37	0.082	0.007				*	*	*			
	2004	36	0.078	0.014					*	*	*		
	1993	35	0.065	0.007						*	*	*	
	2003	36	0.056	0.008							*	*	*
	2005	34	0.049	0.004								*	*
	2002	36	0.044	0.005									*

*Least Significance Difference tests (LSD ; XLStatPro 7.1) were performed. Groups were significantly defined at p < 0.05 level

Table 7

Correlation matrix	1y-SSM	Correlation matrix	2y-SSM
1y-SSM	1	2y-SSM	1
CHLOROA 9-9	0.290	CHLOROA 9-9	0.011
PHEO 9-9	-0.020	PHEO 9-9	-0.009
SALI 9-9	-0.137	SALI 9-9	-0.177
TEMP 9-9	0.221	TEMP 9-9	0.087
TURB 9-9	0.170	TURB 9-9	-0.043
CHLOROA 3-6	0.153	CHLOROA 3-6	0.040
PHEO 3-6	-0.150	PHEO 3-6	0.081
SALI 3-6	-0.125	SALI 3-6	-0.125
TEMP 3-6	0.244	TEMP 3-6	0.007
TURB 3-6	0.166	TURB 3-6	-0.041
CHLOROA 3-9	0.304	CHLOROA 3-9	0.043
PHEO 3-9	-0.183	PHEO 3-9	0.054
SALI 3-9	-0.108	SALI 3-9	-0.157
TEMP 3-9	0.266	TEMP 3-9	-0.010
TURB 3-9	0.176	TURB 3-9	-0.015
CHLOROA 12-6	0.154	CHLOROA 12-6	0.002
PHEO 12-6	-0.121	PHEO 12-6	0.033
SALI 12-6	-0.131	SALI 12-6	-0.136
TEMP 12-6	0.234	TEMP 12-6	0.054
TURB 12-6	0.148	TURB 12-6	-0.055
CHLOROA 6-9	0.346	CHLOROA 6-9	0.067
PHEO 6-9	-0.180	PHEO 6-9	0.050
SALI 6-9	-0.070	SALI 6-9	-0.172
TEMP 6-9	0.202	TEMP 6-9	-0.130
TURB 6-9	0.217	TURB 6-9	0.047
CHLOROA 9-3	0.238	CHLOROA 9-3	-0.194
PHEO 9-3	0.069	PHEO 9-3	-0.038
SALI 9-3	-0.154	SALI 9-3	-0.217
TEMP 9-3	0.132	TEMP 9-3	0.088
TURB 9-3	0.157	TURB 9-3	0.002
CHLOROA 12-3	0.204	CHLOROA 12-3	-0.138
PHEO 12-3	-0.061	PHEO 12-3	-0.036
SALI 12-3	-0.168	SALI 12-3	-0.147
TEMP 12-3	0.186	TEMP 12-3	0.140
TURB 12-3	0.116	TURB 12-3	-0.040

*In bold, significant correlations to the p< 0.005 level

Legend	
2y-SSM	2-year-old oysters spring and summer mortality
1y-SSM	1-year-old oysters spring and summer mortality
CHLOROA 9-9	mean chlorophyll a from september n-1 to september n
PHEO 9-9	mean pheophytin from september n-1 to september n
SALI 9-9	mean salinity from september n-1 to september n
TEMP 9-9	mean temperature from september n-1 to september n
TURB 9-9	mean turbidity from september n-1 to september n
CHLOROA 3-9	mean chlorophyll a from march n to september n
PHEO 3-9	mean pheophytin from march n to september n
SALI 3-9	mean salinity from march n to september n
TEMP 3-9	mean temperature from march n to september n
TURB 3-9	mean turbidity from march n to september n
CHLOROA 9-3	mean chlorophyll a from september n-1 to march n
PHEO 9-3	mean pheophytin from september n-1 to march n
SALI 9-3	mean salinity from september n-1 to march n
TEMP 9-3	mean temperature from september n-1 to march n
TURB 9-3	mean turbidity from september n-1 to march n
CHLOROA 3-6	mean chlorophyll a from march n to june n
PHEO 3-6	mean pheophytin from march n to june n
SALI 3-6	mean salinity from march n to june n
TEMP 3-6	mean temperature from march n to june n
TURB 3-6	mean turbidity from march n to june n
CHLOROA 12-3	mean chlorophyll a from december n-1 to march n
PHEO 12-3	mean pheophytin from december n-1 to march n
SALI 12-3	mean salinity from december n-1 to march n
TEMP 12-3	mean temperature from december n-1 to march n
TURB 12-3	mean turbidity from december n-1 to march n
CHLOROA 12-6	mean chlorophyll a from december n-1 to june n
PHEO 12-6	mean pheophytin from december n-1 to june n
SALI 12-6	mean salinity from december n-1 to june n
TEMP 12-6	mean temperature from december n-1 to june n
TURB 12-6	mean turbidity from december n-1 to june n
CHLOROA 6-9	mean chlorophyll a from june n to september n
PHEO 6-9	mean pheophytin from june n to september n
SALI 6-9	mean salinity from june n to september n
TEMP 6-9	mean temperature from june n to september n
TURB 6-9	mean turbidity from june n to september n

One-year-old oysters summer mortality (1y-SSM)

Adjustment coefficents

R	0.346
R ²	0.119
R²aj.	0.109
SCR	0.779

Evaluation of the value of the information brought by the variables (H0 = Y = Moy(Y)):

Evaluation of	tric value	or tite information	brought by the	variables (110 - 1 - 1VIOy(1)).
Source	df	Sum of squares	Mean square	F - Fisher	Pr > F
Model	1	0.106	0.106	11.941	0.001
Error	88	0.779	0.009		
Total	89	0.884			

Model analysis (Type III SS):

	. /	, .			
Source	df	Sum of squares	Mean square	F - Fisher	Pr > F
CHLOROA 6-9	1	0.106	0.106	11.941	0.001

Model parameters

Parameter	Value	SD	t - Student	Pr > t	Lower limit 95 %	Upper limit 95 %
Constant	0.075	0.016	4.809	< 0.0001	0.044	0.107
CHLOROA 6-9	0	0.005	3.456	0.001	0.007	0.028

Model equation : 1y-SSM = $7.54.10^{-02} + 1.75.10^{-02} *CHLOROA 6-9$

Two-year-old oysters summer mortality (2y-SSM)

Adjustmen	t coefficents
R	0.404
R ²	0.163
R²aj.	0.139
SCR	0.499

Evaluation of the value of the information brought by the variables (H0 = Y=Moy(Y)):

Source	df	Sum of squares	Mean square	F - Fisher	Pr > F
Model	3	0.097	0.032	6.743	0.000
Error	104	0.499	0.005		
Total	107	0.596			

Model analysis (Type III SS):

Source	df	Sum of squares	Mean square	F - Fisher	Pr > F
SALI 9-3	1	0.020	0.020	4.180	0.043
CHLOROA 9-3	1	0.069	0.069	14.344	0.000
TEMP 9-9	1	0.035	0.035	7.339	0.008

Model parameters

Parameter	Value	SD	t - Student	Pr > t	Lower limit 95 %	Upper limit 95 %
Constant	0.161	0.222	0.729	0.468	-0.278	0.601
SALI 9-3	-0.010	0.005	-2.045	0.043	-0.020	0.000
CHLOROA 9-3	-0.043	0.011	-3.787	0.000	-0.065	-0.020
TEMP 9-9	0.022	0.008	2.709	0.008	0.006	0.038

Model equation : 2y-SSM = $0.16 - 9.96.10^{-03}$ *SALI 9-3 - 4.28.10⁻⁰² *CHLOROA 9-3 + 2.22.10⁻⁰² *TEMP 9-9

1 Figure captions

3 Figure 1. Mortality survey of Crassostrea gigas in 39 sites (black type) among which 11 where 4 environmental data are available (black type on gray circle, table 2), 20 "farming areas" 5 (white type and dotted lines) and 9 coastal areas: Normandy, Normano-breton Gulf and 6 North of Finistere on English Channel coast, Mor Bras, Pays de Loire, Re Island, Marennes-7 Oléron Bay and Aquitaine on the French Atlantic coast and Languedoc on Mediterranean 8 sea. Farming areas: Veys bay(BV), Saint Vaast-La-Hougue bay (SV), West of Cotentin (CO), 9 Cancale bay(CA), Paimpol bay (PL), Morlaix bay (MX), Aber Benoit (AB), Brest roadstead 10 (BR), Etel river (EL), Quiberon bay (QB), Auray river (AY), Gulf of Morbihan (GM), Penerf 11 bay (PF), Pen-Bé bay (PB), Bourgneuf bay (BO), Ré island (RE), North part of Marennes-12 Oléron bay (NM), South part of Marennes-Oléron bay (SM), Arcachon basin (AR), Thau 13 lagoon (TH). 14 15 Figure 2. Mortality rate of one- and two-year-old oysters on French coasts between 1995 and 16 2005 (one-year-olds) or 1993 and 2005 (two-year-olds). Coastal areas: Veys bay (BV), 17 Saint Vaast-La-Hougue bay (SV), West of Cotentin (CO), Cancale bay (CA), Paimpol bay 18 (PL), Morlaix bay (MX), Aber Benoit (AB), Brest roadstead (BR), Etel river (EL), Quiberon 19 bay (QB), Auray river (AY), Gulf of Morbihan (GM), Penerf bay (PF), Pen-Bé bay (PB), 20 Bourgneuf bay (BO), Ré island (RE), North part of Marennes-Oléron bay (NM), South part of 21 Marennes-Oléron bay (SM), Arcachon basin (AR), Thau lagoon (TH). 22 23 Figure 3. Mortality rate for one-year-old (A) and two-year-old (B) oysters, according to the 24 rearing sites. Distribution of the mean monthly values in quartiles (Box and Whisker plot 25 presentation). Farming areas: Veys bay (BV), Saint Vaast-La-Hougue bay (SV), West of 26 Cotentin (CO), Cancale bay (CA), Paimpol bay (PL), Morlaix bay (MX), Aber Benoit (AB), 27 Brest roadstead (BR), Etel river (EL), Quiberon bay (QB), Auray river (AY), Gulf of Morbihan 28 (GM), Penerf bay (PF), Pen-Bé bay (PB), Bourgneuf bay (BO), Ré island (RE), North part of 29 Marennes Oléron bay (NM), South part of Marennes-Oléron bay (SM), Arcachon basin (AR), 30 Thau lagoon (TH).

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33 Figure 4. Seasonal and hydrological characteristics of the oyster culture sites along the French
34
      coasts. General PCA with the descriptors (turbidity, .....) . The numbers '1' and '12' refer to
35
      months of the year. To improve typology, coastal areas are presented separately: Veys
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      bay (BV), Cancal bay (CA), Paimpol bay(PL), Morlaix bay (MX), Brest roadstead (BR),
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      Quiberon bay (QB), Bourgneuf bay (BO), North part of Marennes-Oléron bay (NM), South
38
      part of Marennes-Oléron bay (SM), Arcachon basin (AR), Thau lagoon (TH).
39
40 Figure 5. Principal Component Analyses variables representations (summer mortality of oysters
41
      and environmental parameters); A: One-year-old oysters summer mortality (labelled 1y-SM)
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      and environmental parameters (CHLOROA, PHEO, SALI, TURB, TEMP); B: Two-years-old
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      oysters summer mortality (labelled 2y-SM) and environmental parameters (CHLOROA,
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      PHEO, SALI, TURB, TEMP)
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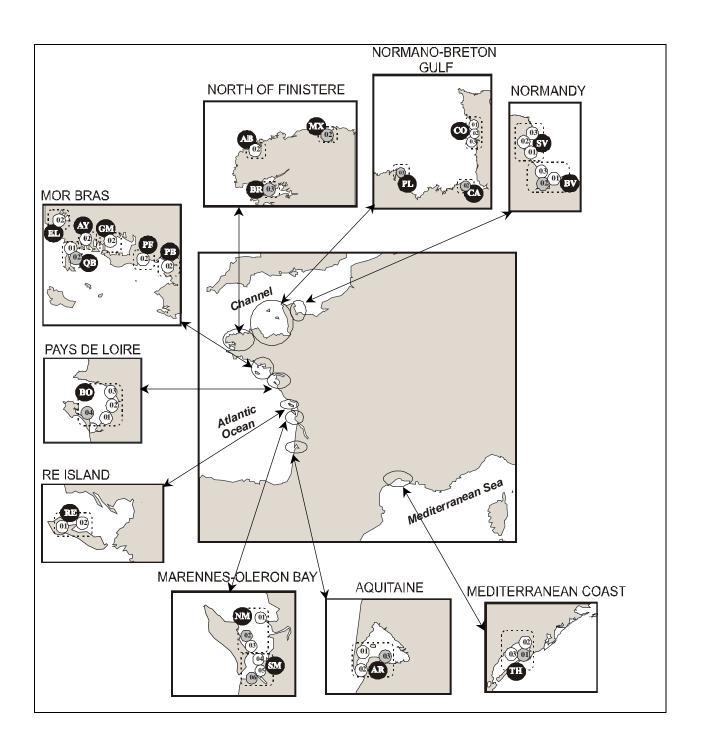


Figure 1

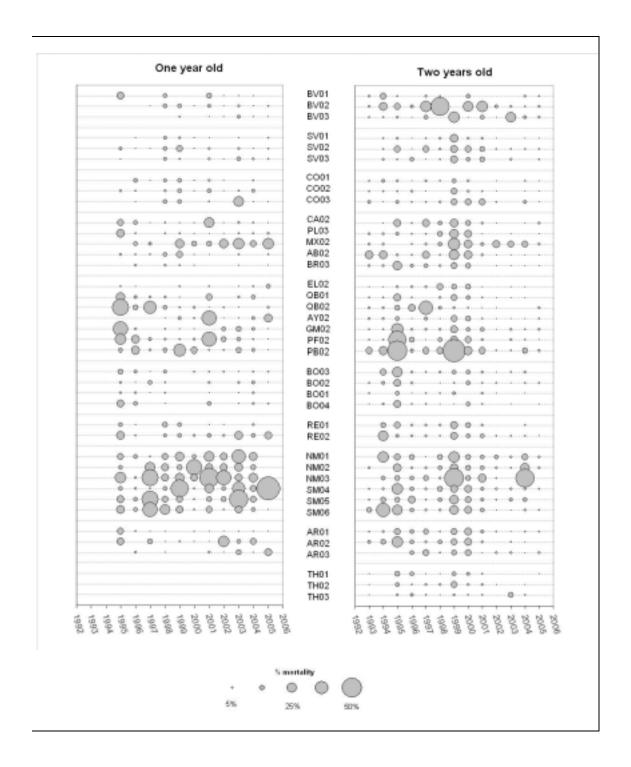


Figure 2

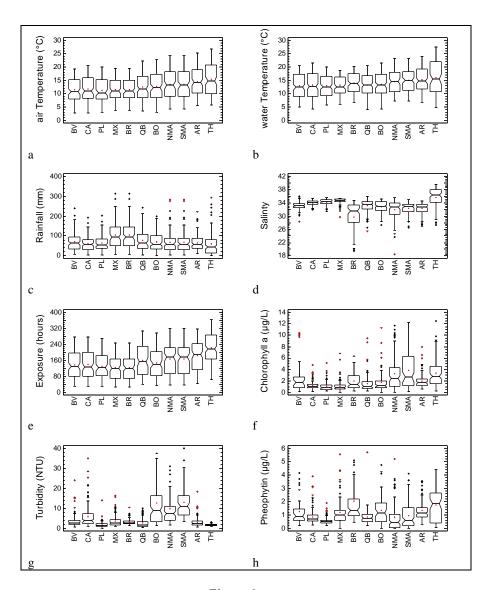


Figure 3

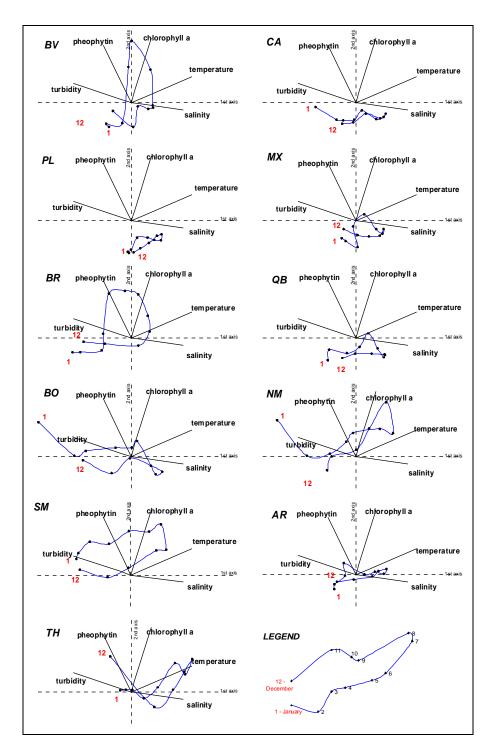


Figure 4

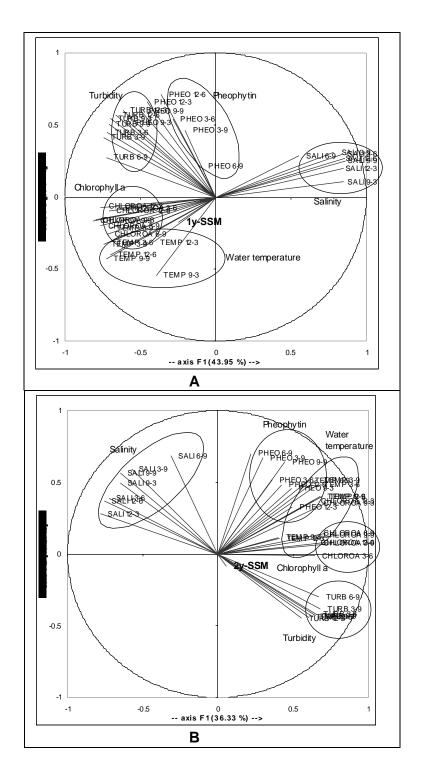


Figure 5

* Response to Reviewers

Response to reviewers (Manuscript AQUA-D-06-0007)

Dear reviewers,

We have considered and taken into account many of your remarks.

We have taken the opportunity of this revision to add two years, available from the two monitoring bases, environment (Rephy) and mortality (Remora): the new period is so 1993-2005. As you 'll see, the main factors correlated to mortalities remain unchanged.

We also focused on the explanatory seawater parameters CHLOROA, SALI, PHEO, TEMP, TURB, instead of more external parameters like meteorology (only considered graphically).

We have tried to redefine and present more clearly the three geographic levels: 9 coastal areals, 20 farming areas, and 39 sites (11 of which are documented environmentally). Doing so, we realized a slight change in the limits of the 9 "coastal areas" (west Cotentin grouped with North-East Brittany) which looks more rationale, ecologically.

Concerning the statistical analysis:

- the method of "autocorrelations" seemed to us less adapted than standard PCAs and multiple regressions.
- the analyses coupling environmental data (grouped systematically by 3, 6, 9 or 12 months) and mortality have been redone completely. The PCAs allow to describe globally the relations between variables. Multiple regressions are based on a "progressive regression" that combines ascendant and descendant stepwise regression.

Finally, no factor alone explains much of the mortalities, and the ways of action are not always clear, but this crossing between independent monitoring data bases looked seemed to us useful.

Regards

J. Mazurie