Evolutionary Applications August 2009, Volume 2 Issue 3, Pages 356 - 370 http://dx.doi.org/10.1111/j.1752-4571.2009.00075.x © 2009 Wiley Blackwell Publishing, Inc.

The definitive version is available at http://www3.interscience.wiley.com/

Size-selective fishing gear and life history evolution in the Northeast Arctic cod

Christian Jørgensen^{1,*}, Bruno Ernande² and Øyvind Fiksen¹

¹ Department of Biology, University of Bergen, Bergen, Norway

² Laboratoire Ressources Halieutiques, IFREMER, Port-en-bessin, France

*: Corresponding author : Christian Jørgensen, Tel.: (+47) 55584618; fax: (+47) 55584450; email address : <u>christian.jorgensen@bio.uib.no</u>

Abstract:

Industrial fishing has been identified as a cause for life history changes in many harvested stocks, mainly because of the intense fishing mortality and its size-selectivity. Because these changes are potentially evolutionary, we investigate evolutionarily stable life-histories and yield in an energy-allocation state-dependent model for Northeast Arctic cod Gadus morhua. We focus on the evolutionary effects of size-selective fishing because regulation of gear selectivity may be an efficient management tool. Trawling, which harvests fish above a certain size, leads to early maturation except when fishing is low and confined to mature fish. Gillnets, where small and large fish escape, lead to late maturation for low to moderate harvest rates, but when harvest rates increase maturation age suddenly drops. This is because bell-shaped selectivity has two size-refuges, for fish that are below and above the harvestable size-classes. Depending on the harvest rate it either pays to grow through the harvestable slot and mature above it, or mature small below it. Sustainable yield on the evolutionary time-scale is highest when fishing is done by trawling, but only for a small parameter region. Fishing with gillnets is better able to withstand life-history evolution, and maintains yield over a wider range of fishing intensities.

Keywords: energy allocation • evolutionary modeling • fisheries management • fishing-induced changes • life history evolution

Introduction

Exploitation of living resources can lead to evolutionary changes in harvested populations of plants (Law and Salick 2005) and mammals (Coltman et al. 2003), but because of the grand scale of commercial fishing most examples come from fish. The reviews by e.g., Jørgensen et al. (2007), Kuparinen and Merilä (2007), Fenberg and Roy (2008), and Hutchings and Fraser (2008) list phenotypic evidence of morphology and life history traits that changed over time in wild 22 populations. Because these changes cannot be totally explained by environmental factors, part of

23 the change is thought to represent contemporary evolution. The rates of change are furthermore 24 rapid, and in general comparable to those observed in breeding programs (Reznick and 25 Ghalambor 2001, Jørgensen et al. 2007). In a comparative study of anthropogenic causes for 26 contemporary evolution, harvesting was found to result in quicker evolutionary change than 27 other human influences (Darimont et al. 2009). Heritable changes in a number of behavioral and 28 life history traits have also been observed in populations harvested experimentally (Conover and 29 Munch 2002, Walsh et al. 2006, Biro and Post 2008). The problem of fishing-induced evolution 30 requires attention because it might be widespread, as most commercially harvested fish stocks 31 experience intense exploitation rates, with fishing mortality being up to four times higher than 32 the natural mortality (Mertz and Myers 1998). In general, theoretical models predict less of an 33 evolutionary response to harvesting if fishing mortality is lower (Law and Grey 1989, Ernande et 34 al. 2004, Brown et al. 2008). This prediction is supported also by fisheries data, for example in 35 Pink salmon where phenotypic change that could not be explained by environment was slower in 36 regions with lower fishing mortality or less selective gear (Ricker 1981).

37 Although there is general agreement that the world's fishing fleets are vastly oversized 38 for a sustainable harvest practice, it has turned out difficult to down-regulate this overcapacity. 39 The reason is a fundamental conflict between the short-term rewards that motivate individual 40 players in the fishing industry versus the long-term goals of sustainability that would insure the 41 viability of the very same industry. Any management solution has to acknowledge and address 42 this trade-off (the following description of the state of fisheries management draws on Clark 43 2006). Catch quotas are often negotiated higher than advised because of lobbying by the fishing 44 industry looking to capitalize quickly on investments in infrastructure and vessels, or small-scale 45 fishermen eager to cover their expenses while competing with other fishermen for the same

46 shared resource. A similar fate befalls effort control regulations, where technology creep leads to 47 increasing ability to catch fish although the numbers of vessels or days at sea remain constant. 48 The attempts by politicians and managers of more drastic methods, such as buyback programs to 49 take vessels out of the fishery, have also failed. Fishermen are like other business owners, and 50 either expect such buyback programs to take place and expand beforehand, or sell only the most 51 inefficient boats so that there are only minor effects on overall fleet capacity. Individually owned 52 quotas that give fishermen a long-term perspective have been proposed as a promising option to 53 break the trade-off between short-term profit and long-term sustainability (see Costello et al. 54 2008). Under such regulations it becomes profitable for a fisherman to spare some fish since he 55 has ownership of a fixed share of future harvest. Such individually transferable quotas (ITQs) 56 avoid the tragedy of the commons but run into other problems because the public basically 57 abandons ownership of a natural resource for free; at least the question of compensation remains 58 largely unresolved. Given the immense difficulties in controlling effort and reducing harvest 59 pressure, it seems that general advice of the type 'reduce fishing mortality to one quarter to 60 lessen the evolutionary impact' is unlikely to be effectively implemented in the near future. An 61 alternative avenue may be to manage fishing gear and its size-selectivity (Law and Rowell 1993, 62 Law 2000).

Most kinds of fishing gear do not catch all fish with equal probability but are selective for certain types of fish in one way or another. Often, this selectivity is based on body size. For example, small fish may slip through gillnets and large fish avoid getting caught, while fish with a girth close to the mesh size are most effectively harvested (Hamley 1975). For trawls, sorting grids and mesh size in the codend let small fish escape while larger fish are harvested. The pattern of size-selectivity may have large consequences for fishing-induced evolution, because 69 the fish that survive and can pass on their genes to the next generation differ between gear types. 70 For example, it has often been stated that if only fish above a certain size threshold are harvested, 71 then it would become optimal to grow slower (Miller 1957) and mature earlier (Law and Grey 72 1989). These qualitative expectations have been confirmed by theoretical models (e.g., Favro et 73 al. 1979, Law and Grey 1989, Ernande et al. 2004, Gårdmark and Dieckmann 2006) and found in 74 harvesting experiments (Edley and Law 1988, Conover and Munch 2002). The most 75 comprehensive experiment to date is on Atlantic silversides Menidia menidia (Conover and 76 Munch 2002, Walsh et al. 2006). In that study, harvesting of the largest individuals, which is 77 analogous to trawling, led to heritable changes towards smaller fish, slower growth, reduced 78 fecundity, poorer viability of larvae, and diminished yield. Because of the potential for 79 detrimental evolutionary effects caused by trawling-like size-selectivity, several authors have 80 suggested that the bell-shaped selectivity curves of gillnets may be better from a sustainability 81 perspective (Law and Rowell 1993, Law 2007). With a bell-shaped selectivity curve, fish that 82 survive as they grow through the sizes vulnerable to the fishery may successfully reproduce, 83 potentially repeatedly and at a large size. This may weaken selection toward early maturation, 84 and thus lead to less of an evolutionary response compared to trawling (Law 2007).

Because there already is a tradition for mesh-size and gear-type regulation in fisheries management as well as routines for enforcement, it seems worthwhile considering gear regulation as a tool to manage evolutionary trait changes generated by fishing practices. This raises the question: what would a desirable harvesting regime look like from the perspective of an evolutionarily concerned fisheries manager? Ultimately, the goals of management are to be decided through a democratic and political process during which the views of the public, stakeholders, and interest groups are duly heard and considered (Jørgensen et al. 2007). From a

92 biological perspective and for the sake of illustration, however, it can be worthwhile to focus on 93 two relatively conservative aims initially. The first one is that the harvesting practice leads to 94 little evolutionary change relative to the pre-harvesting situation. Currently, little is known about 95 the potential consequences of fishing-induced evolution, but because life history traits are 96 affected and these are central to population dynamics, many stock characteristics such as 97 productivity, yield, and resilience might be altered. In general, the manager's tasks of predicting 98 stock development and planning harvesting schemes would be easier if stock properties stayed as 99 constant as possible. It might therefore be good for a manager to have a stock that undergoes 100 little evolutionary change, so that sensitivity to environmental factors and responses to 101 harvesting can be known or learnt from the stock's past behavior. Ecosystem relationships are 102 also more likely to remain the same if the changes in stock characteristics are small. A second 103 property that our evolutionary concerned manager might desire is that her choice of gear type is 104 robust to excess harvesting, as the overall harvest rate has proven difficult to control as discussed 105 above. By expecting that harvest rates might be higher than planned, she should choose a gear 106 type that has minor consequences for the evolutionary outcome if harvest levels were to increase. 107 To our knowledge, the potential evolutionary outcome of a fishery with a bell-shaped 108 selectivity curves has not yet been investigated. We therefore chose to study the effects of gear 109 type and its size-selectivity on expected evolutionary trait changes. The life history model we use 110 is designed for the Northeast Arctic stock of Atlantic cod Gadus morhua and is rich in ecological 111 and physiological detail (Jørgensen and Fiksen 2006). Primarily we contrast different types of

112 gillnets and trawls in search of evolutionarily desirable harvesting practices.

113 Model description

114 Life history model

115 The life history model we used finds the lifelong pattern of energy allocation between growth 116 and reproduction that would maximize lifetime expected fecundity. The allocation decision 117 depends on the individual's state, that is, its age (in months, thus also including seasonal 118 patterns), body length (in cm), level of stored energy (percent of full energy stores), and the 119 current state of the feeding environment that undergoes autocorrelated temporal variability (a 120 graphical overview of the model is given in Figure 1). As such, the model explicitly includes 121 phenotypically plastic response to the internal (age, body length, level of stored energy) and the 122 external (food availability) environment. The model optimizes the state-dependent energy 123 allocation, with the result that age and size at maturation, growth rates, skipped spawning, and 124 the level of energy stored are emergent properties arising from the optimal pattern of energy 125 allocation. The description below is based on Jørgensen and Fiksen (2006) which is referred to 126 for further details. The model has previously been used to investigate the potential for fishing-127 induced evolution to cause changes in life history traits and skipped spawning (Jørgensen et al. 128 2006) and in migration distance and large-scale geographical distribution (Jørgensen et al. 2008). 129 Although selection from size-selective fishing gear has been quantified in earlier studies (e.g., 130 Law and Rowell 1993) this is to our knowledge the first paper to systematically investigate the 131 potential evolutionary consequences of size-selectivity harvesting on fish life histories.

Each month the individuals receive an amount of food that is stochastic and autocorrelated in time. Food intake scales allometrically with length L [cm] as $L^{2.41}$ (based on Jobling 1988), and after energy to cover metabolic rate and basic activity has been spent, the remainder is available for allocation between somatic growth and storage as lipids and proteinsfor future reproduction.

137 The Northeast Arctic cod stock uses the Barents Sea as feeding area but spawns along the 138 Norwegian coast, with the main spawning taking place in Lofoten after a migration of around 139 800 km. We assume it takes five months in total to migrate south to the spawning area, spawn 140 (cod may produce up to 20 batches of eggs that each needs to mature; Kjesbu et al. 1996), and to 141 migrate back north again. The energetic cost of migration is taken from the energy stores, and 142 during spawning and migration cod only eat enough to cover their standard metabolism. The 143 weight-specific energetic cost of migration decreases with fish size (see Ware 1978 and 144 Alexander 2003 for general treatments of size-dependent swimming costs in fish). What is left of 145 their energy stores after migration is used to produce eggs that are spawned. We model only 146 females to avoid the problems of sexual selection and frequency-dependent competition among 147 males.

148 Natural mortality M(L) is negatively size-dependent so that it is highest for small fish and then stabilizes at 0.25 year⁻¹ for larger fish. Fishing takes place both at the spawning grounds 149 150 (where there are only mature fish) and at the feeding grounds (where immature fish are all year 151 round and mature fish the seven months they are not migrating or spawning). Life history theory 152 has shown that these two types of fishery act in opposite directions on age at maturation (Law 153 and Grey 1989): while the feeder fishery favors early-maturing fish that manage to reproduce 154 while still alive, a high mortality on the spawning grounds favor fish that are large when they 155 risk that mortality to reproduce and therefore selects for late maturation. Studies have shown that 156 maturation is much more sensitive to the mortality in the feeder fishery than the harvest rate at 157 the spawning grounds (Law and Grey 1989, Jørgensen et al. 2006).

158 **Evolutionary modeling approach**

159 Optimal energy allocation patterns, i.e., energy allocation patterns that maximize individual 160 fitness and thus result in optimal emergent life history strategies, were found using state-161 dependent dynamic programming (Houston and McNamara 1999, Clark and Mangel 2000). As 162 fitness measure we used the expected lifetime reproductive success R_0 , i.e. the expected number of offspring produced in a lifetime. Theoretical studies have shown that evolution optimizes R_0 163 164 as long as density dependence affecting the population acts only multiplicatively on the expected 165 lifetime production of offspring (Mylius and Diekmann 1995), which is the case in our model 166 (see section below on yield calculation). We therefore maximize reproductive value V, which at 167 birth corresponds to R_0 (Houston and McNamara 1999). For each age, reproductive value V has a 168 future component that the individual can achieve if it survives, and this expected residual 169 reproductive value depends on the new state the individual is in. Every year, the individual can 170 also add to its reproductive value by reproducing. The algorithm finds the optimal allocation 171 pattern by iterating backwards, starting at the maximum age of 25 years at which all individuals 172 die and have no residual reproductive value. The model then compares the fitness consequences 173 of all allocation values α between reproduction and growth and stores the allocation value that 174 maximizes fitness. This is repeated for all lengths L, sizes of energy store E, and environmental 175 food availability F, before the model moves one time-step backwards and repeats the process, assuming that energy allocation is optimal for the remainder of its life. The logic can be 176 177 condensed to the dynamic programming equation, which can be written as:

178
$$V(A, L, E, F) = \max_{\alpha} \left[B(E) + S \sum_{F'} P(F' \mid F) \cdot V(A+1, L', E', F') \right].$$

179 Here the left-hand side is the reproductive value for a given state-combination, and it is found by 180 choosing the allocation α that maximizes fitness. Fitness is expressed within the square brackets, and it consists of *i*) the current fecundity *B* which is a function of the energy available for reproduction, and *ii*) the residual fitness if the individual survives (with probability *S*). The residual fitness depends on the new state at age A+1 (one month older) when the individual has grown to length *L'* and its energy store is *E'* (these state changes follow from α). The summation over *F'* finds the expected fitness over the possible levels of the feeding environment. Since *F* is autocorrelated in time, the conditional probability P(F'|F) gives the probability of having food availability *F'* in the next month if the current level is *F*.

Optimization approaches rely on a static fitness measure, which is valid for populations experiencing simple density-dependent processes but means that frequency-dependent consequences on fitness cannot be included. Examples of such effects that need to be ignored are density dependence acting on growth, or mate competition based on relative size structure. The environmental influence is modeled as fluctuating and autocorrelated, but the environmental variance is constant over time.

194 The advantage of using state-dependent dynamic programming over other evolutionary 195 modeling approaches, like optimality models (Stearns 1992) or selection gradient approaches 196 (Abrams 2001) such as quantitative genetics (Lande 1976) or adaptive dynamics (Dieckmann 197 and Law 1996), is that state dependence can be incorporated in great detail and that individual-198 level processes can be relatively complex. As such, the methodology needs to make fewer 199 assumptions about how phenotypic plasticity should be constrained. The biological complexity 200 of state-dependent strategies and phenotypic plasticity that dynamic programming models can 201 deal with is generally out of reach for the other approaches. One limitation is that, although it is 202 fully accounted for, the selection gradient cannot be computed explicitly.

When the optimal energy allocation pattern has been found for a given fishing regime, we simulate the population dynamics of this life history strategy using a state-structured population model in discrete time to record the emergent life history traits and long-term yield. The results we show are recorded during such forward simulations of optimal life history strategies.

The details of the model including the dynamic programming equation were published in Jørgensen and Fiksen (2006), and we refer the reader to that paper for further details on the physiological and ecological mechanisms included.

210 Genetic assumptions

211 The methodology we use finds phenotypes that optimize individual fitness given selection 212 pressures generated by the ecological setting, which here originates from fishing and its size-213 selectivity but also from natural mortality, physiological constraints and environmental stochasticity. These optimal phenotypes are evolutionarily stable strategies. As such, they are 214 215 evolutionary endpoints resulting from long-term evolution as dictated by selection pressures 216 only. The underlying genetic assumptions of the model are thus according to the streetcar theory 217 of evolution (Hammerstein 1996): there are no genetic constraints that delimit the long term 218 evolutionary outcome. In other words, traits are heritable and genetic correlations do not prevent 219 evolution to proceed (note that ecological and physiological constraints are specified in the 220 model processes, though). The modeling approach does not rely on any particular value for 221 heritability for the traits in question, as the heritability parameter would only scale the rate at 222 which the optimal phenotype is approached but not change the long-term evolutionary endpoint 223 itself. It is worthwhile highlighting here that the model actually finds optimal phenotypically plastic (or state-dependent) phenotypes and therefore incorporates explicitly both plasticity and 224

evolution by considering the long-term evolution of plastic responses or multidimensionalreaction norms.

227 Gear selectivity

228 In previous versions of this model, fishing mortality affected all individuals equally, regardless 229 of their size or other individual states. The addition in this paper is that we introduce fishing 230 selectivity curves U(L) that depend on an individual's length L and take values between 0 (the 231 fish is unaffected by the fishery) and 1 (the fish is maximally selected by the fishery). The 232 fishing mortality F(L) an individual of length L experiences is thus $F(L)=U(L)\cdot f_{\text{max}}$, where f_{max} is the maximum annual harvest [year⁻¹] when selectivity U(L)=1. Total mortality is Z(L) = M(L) + M(L)233 F(L) and monthly survival probability P(L) is then $P(L) = e^{-Z(L)/12}$. The results we present use 234 different values of f_{max} for the feeder fishery and the spawner fishery, as indicated on the graphs. 235 The focus is on drawing the management 'map' of how a choice of fisheries mortalities in the 236 237 two fisheries might affect the evolutionary outcome in the longer term.

238 Because we use optimization, the methodology constrains us to use a fixed fishing 239 intensity and size-selectivity while we find the optimal life history response to it. An inherent 240 assumption is therefore that the fishing fleet continues to use the same gear while fish evolve life 241 histories that allow them to escape the fishing mortality. This can result in a realized fishing 242 mortality that is much lower than f_{max} as life histories evolve to sizes that are less vulnerable to 243 the fishery. Alternative approaches could have been to determine a fixed total allowable catch or a harvest control rule with a given size-selectivity, and study the consequences of that 244 245 management regime. In an optimization framework that would require an iterative procedure that first finds the optimal life history strategy, then simulates the catch resulting from this strategy, 246 247 before modifying the fishing regime and repeating these steps in the next iteration, this until convergence. Such studies are better tackled through models where ecology and evolution take place on the same time-scale such as selection gradient approaches (Abrams 2001) or individualbased evolutionary models (Strand et al. 2002), although these models cannot include as much individual detail. We amend this by showing the fisheries yield that results from a given optimal life history, so that it becomes easier to interpret the evolutionary endpoint of a given harvest regime together with the long-term fisheries yield it would result in.

Size-selectivity curves for Atlantic cod have been determined for trawls and gillnets by Huse et al. (2000). Their findings agree with earlier studies where gillnets show bell-shaped selectivity curves as fish with a certain girth are captured with a higher probability than smaller fish that can slip through or larger fish that don't get far enough through the mesh to get stuck (reviewed by Hamley 1975). We modeled gillnet selectivity U_G as a Gaussian function around a size of maximum selectivity L_{max} :

260
$$U_G(L) = \exp\left(\frac{-(L-L_{\max})^2}{2\sigma^2}\right).$$

261 Huse et al. (2000) found that the width parameter σ of the selectivity curves was 14% of the 262 mean for the three mesh sizes they investigated. A mesh size regulation implemented in a real 263 fishery would probably catch a wider size-range of fish than what was obtained in one scientific study (Huse et al. 2000), because of differences between boats, variable fishing practices, and 264 265 variation in location and timing of fishing. We thus chose to double the width of the selectivity 266 curve, such that $\sigma = 0.28 \cdot L_{\text{max}}$. This makes the size-selectivity curves more similar to the empirical examples shown in Hilborn and Minte-Vera (2008). The conclusions and the 267 268 qualitative results are the same if a narrower selectivity curve is used, but the quantitative predictions change somewhat. For cod, longlines (hook and bait) have a similar bell-shaped sizeselectivity to gillnets (Huse et al. 2000).

In trawls, small fish can escape through the mesh in the codend or through specially designed sorting grids, whereas fish larger than a certain size are retained (Millar and Fryer 1999) resulting in sigmoid selectivity curves. To facilitate comparison, we modeled the trawl selectivity curve U_T as the left half of the same Gaussian function as gillnets up to a size of L_{max} , from which on selectivity was kept at 1:

276
$$U_T(L) = \begin{cases} U_G(L), & L < L_{\max} \\ 1, & L \ge L_{\max} \end{cases}$$

We are aware of the tradition in fisheries science of using logistic functions for sigmoid selectivity curves. However for the sake of comparison, we preferred to keep the formulation of the selectivity curves as similar as possible for gillnet and trawl.

280 For both types of fishing gear, we show results for L_{max} of 70 cm, 90 cm, and 110 cm; 281 selectivity curves for gillnet and trawl with these parameters are shown in Figure 2. The main 282 aim of this study is to contrast the effects of gillnetting and trawling on fishing-induced life 283 history evolution and its consequences for yield. To achieve this comparison, we applied the 284 same fishing size-selectivity, either sigmoid or bell-shaped, at both the spawning and the feeding 285 grounds. In contrast, the existing fishery for Northeast Arctic cod is dominated by trawling at the 286 feeding grounds and longlines and gillnets at the spawning grounds. Therefore, we also ran the 287 model with sigmoid selectivity, $U_T(L)$, at the feeding grounds and bell-shaped selectivity, $U_G(L)$, at the spawning grounds, both with $L_{max} = 90$ cm, to see how a fishery with mixed gear types 288 289 compares to fishing with only one gear type.

290 Yield calculations

The population dynamics of individuals following the optimal life history strategies were simulated in a structured population model. Population size was regulated by a Beverton-Holt density-dependent function for recruit survival, which on its general form can be written as:

294
$$N_2(t+2) = \frac{\lambda B(t)}{1+\gamma B(t)}$$

Here $N_2(t+2)$ is the number of recruits introduced into the structured population model at age 2 at time t+2, and B(t) is the population's total egg production at time t. The two parameters λ and γ determine the strength of density dependence, where λ is the recruit survival at low population densities and the ratio λ/γ is the asymptotic recruitment level. By choosing parameters so that $\lambda = \gamma$ the asymptotic recruitment level becomes $\lambda/\gamma = 1$, implying that abundance N is scaled relative to this asymptotic level.

301 We used two versions of Beverton-Holt parameters in the population dynamics 302 simulations, corresponding to different assumptions about how recruitment density dependence 303 may change with population size. First, we assumed that the Beverton-Holt relationship remained constant whatever population size ($\lambda = \gamma = 5.45 \cdot 10^{-7}$; Fig. A1a). This would correspond 304 305 to the situation where recruitment density dependence is regulated by purely external factors 306 such as habitat availability, or by food (prey) abundance and predation that depend on static 307 populations which do not respond to the size of our focal population (Walters and Korman 308 1999). In a second scenario, we assumed that the Beverton-Holt density dependent recruitment 309 curve had constant curvature around the equilibrium population biomass. This would correspond 310 to the assumption that density dependence is felt in the same way by individuals whatever the 311 size of their population. This would occur if prey and predator populations were changing in size

with our focal species, or if range contractions (or expansions) led to constant strength of predator-prey interactions as our focal population becomes smaller (or larger). This scenario was implemented using a calibrating procedure. First, the population dynamics was run with a constant recruitment $\overline{N}_2 = 0.6$. We then recorded the annual egg production in the population over time to estimate its mean \overline{B} . The value of λ (= γ) that would let the Beverton-Holt curve go through the point (\overline{B} , \overline{N}_2) was then found by solving:

318
$$\overline{N}_2 = \frac{\lambda \overline{B}}{1 + \lambda \overline{B}} ,$$

319 which, when rearranged, gives

320
$$\lambda = \frac{\overline{N}_2}{(1 - \overline{N}_2)\overline{B}} .$$

321 This value of λ was then used for the population dynamics simulations. Examples of this 322 rescaling are shown in Figure A1b.

323 **Results**

324 Our results depict the long-term effects of fishing-induced life history evolution on the stock as a function of fishing on all fish, i.e. immature and mature individuals (the feeder fishery), versus 325 326 fishing only on mature fish (the spawner fishery). The results will be shown as 3D-surfaces 327 representing age at maturation emerging from optimal energy allocation strategies as a function 328 of maximum fishing mortality f_{max} in the feeder fishery (on the left-to-right axis) and in the 329 spawner fishery (on the front-to-back axis). Age at maturation was chosen because this central 330 life history trait is linked to population dynamics, stock productivity, and the stock's size-331 structure. It is also the trait for which most empirical evidence suggesting fishing-induced evolution has been analyzed and published (Jørgensen et al. 2007). We also show similar 3D-332

surfaces for yield. For a stock like the Northeast Arctic cod, where the mature component of the population is geographically separated for parts of the year and the two fisheries can be managed relatively independently, the surfaces represent a decision landscape for our evolutionarily concerned fisheries manager. In fish stocks where fishing intensity cannot be distributed between mature and immature individuals, the manager should look along the diagonal line where the fishing intensity is the same in the spawner and the feeder fishery.

339 We first illustrate the general effect of size-selectivity on evolution of maturation age 340 (Fig. 3). If fishing is un-selective for size (fish of all sizes are harvested with equal probability), 341 then the optimal age at maturation decreases strongly as mortality goes up in the feeder fishery, 342 while it increases slightly as the mortality in the spawner fishery becomes more intense (Fig. 3a). 343 This result agrees with the general prediction from life history theory (see also Law and Grey 344 1989, Ernande et al. 2004, and Jørgensen et al. 2006). This is in contrast to maturation evolution 345 if the fishery is based purely on trawling, where almost all combinations of harvest rates lead to 346 early maturation (Fig. 3b; sigmoid size-selectivity, $L_{max} = 70$ cm). The only exception is when 347 there is no or little fishing in the feeder fishery and just some fishing at the spawning grounds. If 348 the fish were harvested by gillnets, yet another situation occurs (Fig. 3c; bell-shaped size-349 selectivity, $L_{max} = 90$ cm). Now there are two plateaus for optimal age at maturation depending 350 on the exploitation rate in the feeder fishery: when feeder fishery mortality is light to moderate 351 then late maturation is optimal, whereas there is a sudden drop to early maturation if harvest 352 rates become more intense at the feeding grounds. The spawner fishery has a weak effect of 353 raising age at maturation.

The evolutionary outcome for age at maturation in a gillnet fishery depends on its mesh size, which corresponds to the fish length for which selectivity is maximal, L_{max} (Fig. 4a-c). 356 Common for the different mesh sizes is that age at maturation is most sensitive to the fishing 357 mortality in the feeder fishery. Below a threshold value for the intensity of the feeder fishery, the 358 optimal life history matures late (at ~12.8 years irrespective of mesh size). This threshold fishing mortality declines from ~1.4 year⁻¹ to ~0.8 year⁻¹ to ~0.4 year⁻¹ as the mesh size goes up from 70 359 360 to 90 to 110 cm, respectively. A stock adapted to a fishery with smaller mesh sizes can thus 361 withstand higher fishing rates before selection leads to early maturation ages. On the other hand, 362 if fishing were so intense that it exceeded this threshold and caused evolution toward earlier 363 maturation, then a smaller mesh size would lead to a larger drop in maturation age. This is seen 364 as the level of the rightmost plateau which increases with mesh size in the panels of Figure 4a-c, 365 where age at maturation is ~5.7 years when L_{max} is 70 cm (Fig. 4a), 7.0 years when L_{max} is 90 cm 366 (Fig. 4b), and 8.3 years when L_{max} is 110 cm (Fig. 4c).

The outcome is different when fishing is conducted with a gear type that has a sigmoid size-selectivity, for example trawls (Fig. 4d-f). Except when fishing is confined to the spawning grounds and is conducted at low intensities, it is optimal for the cod to mature below the sizes at which vulnerability to harvesting is maximal. The age and corresponding sizes are: 5.6 years and 55 cm when L_{max} is 70 cm (Fig. 4d), 7.0 years and 72 cm when L_{max} is 90 cm (Fig. 4e), and 8.3 years and 88 cm when L_{max} is 110 cm (Fig. 4f).

The long-term equilibrium yield depends on the fisheries selectivity first because it determines which fish are caught, and second because the harvesting regime leads to life history evolution that changes size distributions and population dynamics and thereby which fish are available for the fishery. Whatever the scenario considered for Beverton-Holt density-dependent recruitment (Fig. 7 and 8), maximum yield obtained with sigmoid size-selectivity is higher than with bell-shaped size-selectivity. However, high levels of yield for trawl selectivity are limited to a sharp peak around a small range of harvest rates in the spawner fishery only. Harvesting in the
feeder fishery or at higher intensities in the spawner fishery leads to smaller yield than with
gillnet selectivity.

In one of our scenarios for yield, we rescaled the Beverton-Holt recruitment curve for each optimal life history strategy so that the equilibrium population always produced the same number of recruits (see Fig. A1b in the online appendix). Under these conditions, the stock did not go extinct even at the highest exploitation rates (Fig. 5). In this case, yield when fishing is done by gillnets appears insensitive to the harvest rate in the spawner fishery and shows a domeshaped relationship with harvest rate in the feeder fishery, peaking at $f_{\text{max}} \approx 1.5 \text{ year}^{-1}$ when L_{max} is 70 cm, $f_{\text{max}} \approx 0.8 \text{ year}^{-1}$ 90 cm, and $f_{\text{max}} \approx 0.4 \text{ year}^{-1}$ when L_{max} is 110 cm.

389 When the Beverton-Holt stock recruitment curve was kept constant, harvest could bring 390 about stock extinction (Fig. A2 in the online appendix). This happens with bell-shaped selectivity when the feeder fishery is intense ($f_{max} > 1.0 \text{ year}^{-1}$; Fig. A2a,b) and with sigmoid 391 size-selectivity even when fishing is kept at low levels ($f_{\text{max}} > 0.2 \text{ year}^{-1}$ in the feeder fishery and 392 $f_{\text{max}} > 0.5 \text{ year}^{-1}$ in the spawner fishery; Fig. A2d,e). When mesh size is large (L_{max} =110 cm; Fig. 393 394 A2c,f), fishing does not lead to extinction as the stock is able to evolve life histories that mature and reproduce sufficiently at sizes smaller than those vulnerable to fishing, regardless of whether 395 396 size-selectivity is bell-shaped or sigmoid. Otherwise, the surfaces that depict yield (Fig. A2) have 397 the same qualitative characteristics as with a rescaled Beverton-Holt curve (Fig. 5). The main 398 quantitative differences are that for bell-shaped size-selectivity, the maximum yield is obtained 399 at lower fishing intensities, and for sigmoid size-selectivity, the peak of high levels of yield 400 extends over a narrower range of harvest rates in the spawner fishery.

401 Finally, we ran the model with mixed gear types to better reflect how the fishery really 402 operates nowadays. We used bell-shaped size-selectivity curves in Lofoten, as fishing for the 403 spawning fish is done mostly with longlines and gillnets that both have similar size-selectivity 404 (Huse et al. 2000). For the feeder fishery in the Barents Sea, mostly done by trawling, we used 405 sigmoid size-selectivity. Both selectivity curves had a maximum selectivity size of $L_{max} = 90$ cm. 406 The resulting optimal age at maturation is similar to that when only sigmoid selectivity curves 407 (trawling) were used in both fisheries, except for very low harvest rates in the feeder fishery (Fig. 408 A3).

409 **Discussion**

410 In this paper we used a life history optimization model for fish to investigate how the sizeselectivity of fishing gear may affect fishing-induced evolution. We focused on two outcomes. 411 412 First, we assessed the degree of expected life history evolution in a given harvest regime, 413 quantified as the mean age at maturation emerging from the optimal state-dependent energy 414 allocation pattern that would result from long-term evolution as dictated by selective pressures. 415 Since we used state-dependent optimization, the energy allocation strategy that maximizes 416 lifetime reproductive value corresponds to the evolutionarily stable strategies in the environment 417 set by fishing, and the methodology includes phenotypic plasticity. Second, we quantified the 418 effects of fishing-induced evolution on the long-term equilibrium yield. The harvesting strategy 419 that maximizes long-term equilibrium yield after life-history evolution has been called the 420 evolutionarily stable optimal harvesting strategy (ESOHS, Law and Grey 1989). Instead of 421 focusing only on the single harvesting regime that optimizes yield, we show results for varying 422 fishing mortalities in both the feeder and the spawner fishery to paint decision landscapes for 423 how a harvesting strategy would affect long-term outcomes.

424 Maturation evolution under bell-shaped versus sigmoid size-selectivity

425 Our results show that whether fishing gear has bell-shaped or sigmoid size-selectivity has 426 surprisingly strong effects on the qualitative outcomes both in terms of expected evolution of life 427 history traits and in terms of the resulting fisheries yield. The ages and sizes at maturation that 428 are optimal at intense fishing rates with gillnets are evolutionarily favored at even low harvest 429 rates when trawling. It is also interesting how the gradual response in maturation age seen under 430 random harvest disappears when fishing is size-selective, leading to a few life history outcomes 431 that cover wide regions of parameter space and with sharp transitions between them. This can be 432 explained by keeping in mind that fitness is defined as the expected lifetime reproductive 433 success, which depends on the fecundity at age and the survival probability until that age.

With sigmoid size-selectivity when trawling, all fish above a certain size suffer reduced survival, which quickly erodes the advantages of maturing large with high fecundity. Even at low fishing intensities, fish that mature at sizes below the vulnerable ones have high survival until maturation, and will also survive better for the consecutive spawning seasons if they do not grow into the harvestable size classes. There is a size-refuge for small fish, which quickly becomes the best option even at low harvest intensities.

In the case of bell-shaped size-selectivity, such as with gillnets, the situation is more complex because there are size refugia for both small and large fish, while fish of intermediate size are harvested. Under intense fishing, few fish would survive as they grow through the harvestable size slot. Because the larger size-refuge is then practically unreachable due to the high fishing pressure, the optimal solution is to mature at sizes below those vulnerable, i.e. in the smaller size-refuge, as for trawling. If fishing is less intense, however, it can still pay to grow large and mature in the large size-refuge, although there is some probability of dying on the way. By maturing large, fecundity will increase but survival until maturation will be reduced. Depending on which effect is strongest, the expected lifetime reproductive success can be higher if maturation takes place at sizes either above or below the harvestable size slot. This explains the two plateaus for late and early maturation seen in optimal strategies with gillnet harvesting. Also, the transition between late and early maturation is so sudden in terms of increasing harvest rates because the optimal strategy is either to mature before or after the harvestable slot, but not in the middle of it.

454 Ecological advantages of bell-shaped selectivity

455 Bell-shaped size-selectivity curves retain some of the older and larger fish in the population. The 456 fraction retained and its size composition depends on harvest rate and the exact shape and width 457 of the size-selectivity curve. There are, as Law (2007) pointed out, both good ecological and good evolutionary reasons to prefer bell-shaped over sigmoid size-selectivity curves (see also 458 459 Berkeley et al. 2004b, Birkeland and Dayton 2005). As an example of a beneficial ecological 460 effect, a diverse age-structure has been shown to lead to enhanced recruitment in Icelandic cod 461 (Marteinsdottir and Thorarinsson 1998). In our model, the main ecological advantages of larger 462 fish are that they have higher fecundity because of sheer size, and they also have more cost-463 effective migration to and from the spawning sites. We did not specifically implement parental 464 effects, which have the potential to add further advantages to a diverse stock structure. One example of such an effect is the increased viability of larvae spawned by large or old mothers in 465 466 rockfish (Berkeley et al. 2004a). Similar correlations between parent size and offspring traits 467 were found also in the artificial harvesting experiment reported in Walsh et al. (2006).

468 Another interesting feature of a broad age- or size-distribution in the population is that it 469 buffers the population dynamical effects of environmental factors such as climate (Brander 470 2008). Ottersen et al. (2006) studied recruitment as a proxy for population dynamics of the 471 Northeast Arctic cod, together with regional climate indices such as the North Atlantic 472 Oscillation and sea temperature through a transect in the Barents Sea. Their main finding was 473 that the correlation between recruitment and climate grew stronger over time, in parallel with the 474 truncation of the population's age- and size-structure. As the fish matured earlier, they became 475 more tightly tied to climate. Along similar lines, Hsieh et al. (2008) reported that the 476 geographical ranges of fished species fluctuated more with climate than unfished species, and 477 they related this tightened coupling to the truncated population structure caused by fishing. For 478 the Northeast Arctic cod, one possible mechanism can be the long spawning migration, which is 479 relatively cheaper in energetic terms for larger fish. With a similar model as here but allowing 480 migration distance to vary, Jørgensen et al. (2009) showed that because a population adapted to 481 fishing will contain more early-maturing and smaller fish, the optimal migration distances are 482 shorter, and they will spawn along a reduced geographical range compared to the pristine pre-483 fishing state. Because fish then would sample the environment over a narrower geographical 484 range, one could expect that spawning areas that are particularly good in a given year may not be 485 visited and that the population as a whole does not buffer climatic variation to the same degree as 486 before. Any such effect where large or old individuals have beneficial consequences for 487 population dynamics and recruitment would be preserved better with a bell-shaped size-488 selectivity curve whereby some large fish are retained, compared to sigmoid size-selectivity 489 curves where also the big ones are fished out.

490 **Fishing at the spawning grounds**

491 The standard prediction from life history models without size-selective harvesting mortality is 492 that mortality at the spawning grounds, which removes only mature fish, will increase age at 493 maturation. This can be seen in the scenario with un-selective harvest (Fig. 3a) and has been 494 shown also in other studies (Law and Grey 1989, Ernande et al. 2004, Gårdmark and Dieckmann 495 2006, Jørgensen et al. 2006). The underlying life history logic is that because fecundity increases 496 with size, it becomes profitable to have large gonads when one accepts the extra mortality 497 associated with spawning. Conversely, if mortality at the spawning grounds is lower than 498 elsewhere, it would select for earlier maturation as spawning then offers a refuge from 499 harvesting.

500 With size-selective harvesting, this result holds also under the bell-shaped size-selectivity 501 typical for a gillnet fishery: although the effect is weaker, optimal maturation age goes up as the 502 spawner fishery is increased (most easily seen in the sharp transition phase between the two 503 plateaus in Fig. 4a-c). In contrast, under sigmoid size-selectivity, optimal age at maturation 504 increases with increasing harvest mortality at the spawning grounds only up to a certain point, 505 from which it suddenly drops to early maturation. The sigmoid size-selectivity creates a size-506 refuge for small fish also at the spawning grounds, and this refuge becomes more important as 507 harvest rates go up. With little selection for early maturation in the feeder fishery and low 508 harvest rates at the spawning grounds, the benefit of maturing large still outweighs early 509 maturation (left corner of panels in Fig. 4d-f). However, increasing harvest rates in both the 510 feeder and spawner fishery removes this size-advantage, consequently leading to early 511 maturation below the harvestable size.

512 Effects of size-selectivity on evolutionary stable yield

513 The model predicts that trawling can give the highest yield if it selects also small fish. High yield 514 results from the same harvest rates that cause late maturation, but even a minor increase in 515 harvest intensity causes early maturation to become optimal (Fig. 4c-f, 5, A2). Thus, high evolutionarily stable yield is only achieved if three conditions are met: harvest rates have to be very low, mainly confined to the spawning grounds, and strictly controlled. Each of these three conditions are hard to meet with current fisheries management, meaning that the optimal solution is likely impossible in practice. If excess harvesting happens intermittently, it may be sufficient to cause maturation evolution and long-term loss of yield.

521 Gillnets produce more stable outcomes with respect to variation in fishing rates so that 522 larger regions of parameter space give good yield (Fig. 5; see also Fig. A2). From a management 523 perspective this means that a gillnet fishery is more robust to errors in the realized fishing 524 mortality. Such errors can stem from uncertainty in stock assessments, politicized quota setting, 525 or poor enforcement of management regulations. The stock's sensitivity to evolution may also 526 vary with factors that were not included in the model. The major problem with a fishery based on 527 gear types with bell-shaped size-selectivity is that if harvest rates become too high in the feeder 528 fishery, then early maturation can become optimal, resulting also in a drop in the evolutionarily 529 stable yield.

530 **Comparing stock-recruitment curves**

531 The two alternatives we investigated for the stock-recruitment relationship represent different 532 scenarios for how the ecosystem will respond to changes in population abundance and 533 demography. It is probably unlikely that the stock-recruitment curve stays exactly the same as 534 the stock undergoes large changes (first scenario). It is probably also unlikely that the ecosystem 535 fully compensates so that the curves becomes steeper as egg production diminishes, but with the 536 same carrying capacity (second scenario). The ecosystem response will likely lie somewhere in 537 between our two scenarios, so that they bracket the potential outcomes (although at least in 538 theory, the stock-recruitment curve could also become less steep as adaptations take place). The major difference between the two scenarios is that a constant stock-recruitment relationshipcauses stock extinction at intense harvest levels.

541 Beyond stock extinction, the qualitative differences between the stock-recruitment 542 mechanisms are minor and discrepancies are mostly quantitative. Studying yield consequences of 543 fishing-induced changes is extremely difficult because it relies on critical assumptions about 544 density dependence and how the rest of the ecosystem will react to harvest-induced changes in 545 stock structure and life history traits (e.g., Abrams and Matsuda 2005, Gårdmark et al. 2003). For 546 instance, we omitted other types of density dependence than the Beverton-Holt recruit survival, 547 and many are known to act in the wild (e.g. on growth, Lorenzen and Enberg 2002; see also 548 Enberg et al. this volume and the role of size-dependent growth in Arlinghaus et al. this volume). 549 The predictions we report for yield must therefore be interpreted with caution. Fully frequency-550 dependent models are better suited to include several sources of density dependence, and we 551 look forward to studies that address effects on yield in further detail.

552 Robust management

If our evolutionarily concerned fisheries manager could manage gear size-selectivity, she would probably choose bell-shaped size-selectivity. One reason is that sigmoid size-selectivity easily leads to stock extinction or extremely low yield, depending on the stock-recruitment scenario, as soon as harvest rates increase to moderate levels. In contrast, population dynamics is much more robust under gillnet selectivity, which yields favorable outcomes over a larger area of parameter space.

However, when implementing a gillnet fishery, our manager is faced with a trade-off with respect to the choice of mesh size. For small mesh size, the population is evolutionarily relatively insensitive to harvest rates, and only with extreme fishing rates at the feeding grounds 562 will it become optimal for cod to mature early. Yield is also high across many combinations of 563 harvest rates on the feeding and spawning grounds. However, small mesh size also means that if 564 harvest rates are so high as to lead to evolution, then the resulting life history strategy matures at 565 a very early age. When mesh size is large, the late maturation plateau shrinks, evolution toward 566 the early maturation plateau may take place at lower fishing intensities, but the early maturation 567 plateau corresponds to a later maturation age. By choosing larger mesh sizes, the manager thus 568 needs to exercise more caution in controlling the applied harvest rates, but she risks less of a 569 reduction in maturation age if she fails. The optimal level of this trade-off depends on how 570 important it is for management to prevent evolution and to what degree the fisheries management 571 institution can influence quotas and control the fishing fleet.

572 **Comparison with current fishing regime**

Currently, Northeast Arctic cod is mostly fished by trawlers in the Barents Sea, represented by 573 the feeder fishery in our model, with a harvest rate around 0.5 year⁻¹. The potential to market 574 575 fresh fish year round concentrates effort on the feeder fishery rather than the spawner fishery, where fish are only present for a few months. In contrast, the spawner fishery consists mostly of 576 long-liners and gill-netters. Our model predicts that with these gear types, any harvest rate above 577 0.2 year^{-1} in the feeder fishery has the potential to cause early maturing fish, regardless of the 578 579 mortality in the spawner fishery. Maturation ages have indeed declined in the Northeast Arctic 580 cod (Jørgensen 1990), and the observed temporal trend can be related to changes in the 581 probabilistic maturation reaction norm (Heino et al. 2002), which suggests that its cause is 582 evolutionary. The observed changes in the stock are consistent with the predictions from our 583 model, although we cannot assess evolutionary rates because optimization methods find 584 evolutionary endpoints but do not answer whether these endpoints are attainable or at what rates they may be attained. On this issue, it is worth noting that the rate of change of maturation age in Northeast Arctic cod is comparable to observed rates in many other stocks thought to undergo fishing-induced evolution (quantified in the online appendix of Jørgensen et al. 2007). It would be interesting to look into what gear types have been used in these fisheries, how the use of gear and its selectivity have changed over time, what evolutionary outcomes one could expect, and finally compare those predicted outcomes to the observed life history changes for these stocks.

591 **References**

592 Abrams, P. A. 2001. Modelling the adaptive dynamics of traits involved in inter- and

593 intraspecific interactions: an assessment of three methods. Ecology Letters 4:166-175.

- Abrams, P. A., and H. Matsuda. 2005. The effect of adaptive change in the prey on the dynamics
 of an exploited predator population. Canadian Journal of Fisheries and Aquatic Sciences
 62:758-766.
- Alexander, R. M. 2003. Principles of animal locomotion. Princeton and Oxford, Princeton
 University Press.
- Arlinghaus, R., S. Matsumura, and U. Dieckmann. 2009. Quantifying selection pressures caused
 by recreational fishing. Evolutionary Applications 00:000-000.
- Berkeley, S. A., C. Chapman, and S. M. Sogard. 2004. Maternal age as a determinant of larval
 growth and survival in a marine fish, *Sebastes melanops*. Ecology 85:1258-1264.
- Berkeley, S. A., M. A. Hixon, R. J. Larson, and M. S. Love. 2004. Fisheries Sustainability via
 protection of age structure and spatial distribution of fish populations. Fisheries 29:23-32.
- biotection of age structure and spatial distribution of fish populations. Fisheries 27.25-52.
- Birkeland, C., and P. K. Dayton. 2005. The importance in fishery management of leaving the big
- ones. Trends in Ecology & Evolution 20:356-358.

607	Biro, P. A., and J. R. Post. 2008. Rapid depletion of genotypes with fast growth and bold
608	personality traits from harvested fish populations. Proceedings of the National Academy
609	of Sciences 105:2919-2922.
610	Brander, K. 2008. Tackling the old familiar problems of pollution, habitat alteration and
611	overfishing will help with adapting to climate change. Marine Pollution Bulletin 56:1957-
612	1958.
613	Brown, C. J., A. J. Hobday, P. E. Ziegler, and D. C. Welsford. 2008. Darwinian fisheries science
614	needs to consider realistic fishing pressures over evolutionary time scales. Marine
615	Ecology Progress Series 369:257-266.
616	Clark, C. W. 2006. The worldwide crisis in fisheries. Cambridge, UK, Cambridge University
617	Press.

618 Clark, C. W., and M. Mangel. 2000. Dynamic state variable models in ecology. New York,
619 Oxford University Press.

620 Coltman, D. W., P. O'Donoghue, J. T. Jorgenson, J. T. Hogg, C. Strobeck, and M. Festa-

- Bianchet. 2003. Undesirable evolutionary consequences of trophy hunting. Nature426:655-658.
- 623 Conover, D. O., and S. B. Munch. 2002. Sustaining fisheries yields over evolutionary time
 624 scales. Science 297:94-96.
- 625 Costello, C., S. D. Gaines, and J. Lynham. 2008. Can catch shares prevent fisheries collapse?
 626 Science 321:1678-1681.
- 627 Darimont, C. T., S. M. Carlson, M. T. Kinnison, P. C. Paquet, T. E. Reimchen, and C. C.
- 628 Wilmers. 2009. Human predators outpace other agents of trait change in the wild.
- 629 Proceedings of the National Academy of Sciences 106:952-954.

- Dieckmann, U., and R. Law. 1996. The dynamical theory of coevolution: a derivation from
 stochastic ecological processes. Journal of Mathematical Biology 34:579-612.
- 632 Edley, M. T., and R. Law. 1988. Evolution of life histories and yields in experimental
- 633 populations of *Daphnia magna*. Biological Journal of the Linnean Society 34:309-326.
- Enberg, K., C. Jørgensen, E. Dunlop, M. Heino, and U. Dieckmann. 2009. Implications of
- 635 fisheries-induced evolution for stock rebuilding and recovery. Evolutionary Applications636 00:000-000.
- 637 Ernande, B., U. Dieckmann, and M. Heino. 2004. Adaptive changes in harvested populations:
- plasticity and evolution of age and size at maturation. Proceedings of the Royal Society
 of London Series B-Biological Sciences 271:415-423.
- Favro, L. D., P. K. Kuo, and J. F. MacDonald. 1979. Population-genetic study of the effect of
 selective fishing on the growth rate of trout. Journal of the Fisheries Research Board of
 Canada 36:552-531.
- Fenberg, P. B., and K. Roy. 2008. Ecological and evolutionary consequences of size-selective
 harvesting: how much do we know? Molecular Ecology 17:209-220.
- 645 Gårdmark, A., U. Dieckmann, and P. Lundberg. 2003. Life-history evolution in harvested
 646 populations: the role of natural predation. Evolutionary Ecology Research 5:239-257.
- 647 Gårdmark, A., and U. Dieckmann. 2006. Disparate maturation adaptations to size-dependent
 648 mortality. Proceedings of the Royal Society B-Biological Sciences 273:2185-2192.
- Hamley, J. M. 1975. Review of gillnet selectivity. Journal of the Fisheries Research Board ofCanada 3:1943-1969.
- Hammerstein, P. 1996. Darwinian adaptation, population genetics and the streetcar theory of
 evolution. Journal of Mathematical Biology 34:511-532.

- Heino, M., U. Dieckmann, and O. R. Godø. 2002. Reaction norm analysis of fisheries-induced
 adaptive change and the case of the Northeast Arctic cod. ICES CM 2002 Y:14.
- Hilborn, R., and C. V. Minte-Vera. 2008. Fisheries-induced changes in growth rates in marine
 fisheries: are they significant? Bulletin of Marine Science 83:95-105.
- Houston, A. I., and J. M. McNamara. 1999. Models of adaptive behaviour: an approach based on
 state. Cambridge, UK, Cambridge University Press.
- Huse, I., S. Løkkeborg, and A. V. Soldal. 2000. Relative selectivity in trawl, longline and gillnet
 fisheries for cod and haddock. ICES Journal of Marine Science 57:1271-1282.
- Hutchings, J. A., and D. J. Fraser. 2008. The nature of fisheries- and farming-induced evolution.
 Molecular Ecology 17:294-313.
- Jobling, M. 1988. A review of the physiological and nutritional energetics of cod, *Gadus morhua*L, with particular reference to growth under farmed conditions. Aquaculture 70:1-19.
- Jørgensen, C., B. Ernande, Ø. Fiksen, and U. Dieckmann. 2006. The logic of skipped spawning
 in fish. Canadian Journal of Fisheries and Aquatic Sciences 63:200-211.
- 667 Jørgensen, C., E. S. Dunlop, A. F. Opdal, and Ø. Fiksen. 2008. The evolution of spawning
- migrations: the role of individual state, population structure, and fishing-inducedchanges. Ecology 89:3436-3448.
- Jørgensen, C., K. Enberg, E. S. Dunlop, R. Arlinghaus, D. S. Boukal, K. Brander, B. Ernande et
 al. 2007. Managing the world's evolving fish stocks. Science 318:1247-1248.
- Jørgensen, C., and Ø. Fiksen. 2006. State-dependent energy allocation in cod (*Gadus morhua*).
- 673 Canadian Journal of Fisheries and Aquatic Sciences 63:186-199.
- Jørgensen, T. 1990. Long-term changes in age at sexual maturity of Northeast Arctic cod (*Gadus morhua* L.). Journal Du Conseil 46:235-248.

676	Kjesbu, O. S., P. Solemdal, P. Bratland, and M. Fonn. 1996. Variation in annual egg production
677	in individual captive Atlantic cod (Gadus morhua). Canadian Journal of Fisheries and
678	Aquatic Sciences 53:610-620.
679	Kuparinen, A., and J. Merilä. 2007. Detecting and managing fisheries-induced evolution. Trends
680	in Ecology & Evolution 22:652-659.
681	Lande, R. 1976. Natural selection and random genetic drift in phenotypic evolution. Evolution
682	30:314-334.
683	Law, R. 2000. Fishing, selection, and phenotypic evolution. ICES Journal of Marine Science
684	57:659-668.
685	Law, R. 2007. Fisheries-induced evolution: present status and future directions. Marine Ecology-
686	Progress Series 335:271-277.
687	Law, R., and C. A. Rowell. 1993. Cohort-structured populations, selection responses, and
688	exploitation of the North Sea cod, Pages 155-173 in T. K. Stokes, J. M. McGlade, and R.
689	Law, eds. The exploitation of evolving resources. Lecture Notes in Biomathematics.
690	Berlin, Germany, Springer-Verlag.
691	Law, R., and D. R. Grey. 1989. Evolution of yields from populations with age-specific cropping.
692	Evolutionary Ecology 3:343-359.
693	Law, W., and J. Salick. 2005. Human-induced dwarfing of Himalayan snow lotus, Saussurea
694	laniceps (Asteraceae). Proceedings of the National Academy of Sciences of the United
695	States of America 102:10218-10220.
696	Lorenzen, K., and K. Enberg. 2002. Density-dependent growth as a key mechanism in the
697	regulation of fish populations: evidence from among-population comparisons.

698 Proceedings of the Royal Society B-Biological Sciences 269:49-54.

- 699 Marteinsdottir, G., and K. Thorarinsson. 1998. Improving the stock-recruitment relationship in
- 700Icelandic cod (Gadus morhua) by including age diversity of spawners. Canadian Journal
- 701 of Fisheries and Aquatic Sciences 55:1372-1377.
- Mertz, G., and R. A. Myers. 1998. A simplified formulation for fish production. Canadian
 Journal of Fisheries and Aquatic Sciences 55:478-484.
- 704 Millar, R. B., and R. J. Fryer. 1999. Estimating the size-selection curves of towed gears, traps,
- nets and hooks. Reviews in Fish Biology and Fisheries 9:89-116.
- Miller, R. B. 1957. Have the genetic patterns of fishing been altered by introductions or by
- selective fishing? Journal of the Fisheries Research Board of Canada 14:797-806.
- Mylius, S. D., and O. Diekmann. 1995. On evolutionarily stable life histories, optimization and
 the need to be specific about density dependence. Oikos 74:218-224.
- 710 Reznick, D. N., and C. K. Ghalambor. 2001. The population ecology of contemporary
- 711
 adaptations: what empirical studies reveal about the conditions that promote adaptive

 712
 112
 112
 102
- 712 evolution. Genetica 112-113:183-198.
- Ricker, W. E. 1981. Changes in the average size and average age of Pacific salmon. Canadian
 Journal of Fisheries and Aquatic Sciences 38:1636-1656.
- 715 Stearns, S. C. 1992. The evolution of life histories. Oxford, Oxford University Press.
- Strand, E., G. Huse, and J. Giske. 2002. Artificial evolution of life history and behavior.
 American Naturalist 159:624-644.
- 718 Swain, D. P., A. F. Sinclair, and J. M. Hanson. 2008. Evolution of growth in Gulf of St
- 719 Lawrence cod: reply to Heino *et al.* Proceedings of the Royal Society B: Biological
- 720 Sciences 275:1113-1115.

721	Walsh, M. R., S. B. Munch, S. Chiba, and D. O. Conover. 2006. Maladaptive changes in
722	multiple traits caused by fishing: impediments to population recovery. Ecology Letters
723	9:142-148.
724	Walters, C., and J. Korman. 1999. Linking recruitment to trophic factors: revisiting the
725	Beverton-Holt recruitment model from a life history and multispecies perspective.

726 Reviews in Fish Biology and Fisheries 9: 187-202.

- Ware, D. M. 1978. Bioenergetics of pelagic fish: theoretical change in swimming speed and 727
- ration with body size. Journal of the Fisheries Research Board of Canada 35:220-228. 728

729 Figures

730



731

732

733 Figure 1. Graphical representation of the life-history energy-allocation model, parameterized for 734 the Northeast Arctic cod (Gadus morhua). The central process is energy allocation toward 735 growth or stores/reproduction. The energy allocation can take independent values depending on 736 the individuals state (age, size, level of energy stored, and current food availability). The model 737 is then solved with different forcing, here with focus on changing the size-selectivity and 738 intensity of the fishing mortality. Black arrows denote energy flow. In periods when food intake 739 is insufficient for metabolic demands, energy requirements can be met by stores (dotted line). 740 The figure is modified from Jørgensen and Fiksen (2006).

741 742



Figure 2. Size-selectivity curves used to impose selection on Northeast Arctic cod life histories. At maximum selectivity, fish of that size are harvested at the rate specified by the parameter f_{max} . a) Bell-shaped size-selectivity curves for gillnets. The peaks of the Gaussian functions are at 70 cm (dotted line), 90 cm (thin black line), and 110 cm (thick grey line). In each case the standard deviation is 28% of the mean. b) Sigmoid trawl selectivity was modeled based on the same probability distributions as for gillnets, but with maximum selectivity for all fish lengths larger than the peak.



Figure 3. The effects of size-selective fishing gear on optimal age at maturation in the Northeast Arctic cod Gadus morhua. The leftto-right axis is the fishing mortality f_{max} [year⁻ ¹] in the feeder fishery at the sizes where selectivity is 1, and the front-to-back axis is the fishing mortality f_{max} [year⁻¹] at the spawning grounds. a) Unselective fishing mortality; all sizes have the same probability of being caught (selectivity U is always 1). b) Sigmoid trawl selectivity, where the size-selectivity of the fishery increases with the fish' body size (here at maximum from $L_{\text{max}} = 70$ cm onwards). c) Bell-shaped size-selectivity is typical for gillnets (here with maximum selectivity at $L_{\rm max} = 90 \, {\rm cm}$).





Figure 4. Optimal age at maturation in the Northeast Arctic cod, shown for increasing fishing mortality rates f_{max} [year⁻¹] in the feeder and spawning fishery. a-c) with bell-shaped sizeselectivity curves as is typical for gillnets, and d-f) with sigmoid size-selectivity curves for example with trawls. Maximum selectivity L_{max} is at: a,d) 70 cm; b,e) 90 cm; and c,f) 110 cm. See legend to figure 3 for further explanation of axes.





Figure 5. Long-term equilibrium yield as in Figure 6, but with the assumption that the Beverton-Holt total egg production-recruitment curve is rescaled so that the equilibrium population always produces the same number of recruits. The recruitment curve has thus increased in steepness as harvest becomes stronger, as in Fig. A1b). Bell-shaped size-selectivity: a) $L_{\text{max}} = 70$ cm; b) L_{max} = 90 cm; c) $L_{\text{max}} = 110$ cm. Sigmoid size-selectivity: d) $L_{\text{max}} = 70$ cm; e) $L_{\text{max}} = 90$ cm; f) $L_{\text{max}} = 110$ cm. Further legend is given in Figure 3.

758	
759	Supplementary online appendix to
760	
761	Size-selective fishing gear and life history evolution in the Northeast
762	Arctic cod
763	Christian Jørgensen, Bruno Ernande, and Øyvind Fiksen
764	
765	Contents:
766	Figure A1: Rescaling of Beverton-Holt recruitment functions used in yield calculations.
767	Figure A2: Yield with constant recruitment function.
768	Figure A3: Age at maturation for optimal life histories with combination of trawls and gillnets.





769

771 Figure A1. Beverton-Holt stock recruitment curves used in two scenarios to calculate long-term 772 fisheries yield at evolutionary equilibrium when cod life histories evolve in response to size-773 selective fishing gear. a) In the first scenario we assume that the stock recruitment curve remains 774 constant. The black curve shows the range of total egg production and corresponding recruitment 775 for the stock prior to harvest, with the thin line representing the 90% range and the thick line the 776 50% range. The dotted line is the continuation of this curve to the origin. After harvesting and 777 resulting life history evolution, the stock produces fewer eggs annually and recruitment is 778 correspondingly lower (the lines are displaced slightly to improve readability; light grey lines harvest rate 0.1 year⁻¹ in both the feeder and spawner fishery; dark grey lines harvest rate 0.3 779 year⁻¹; all fish were harvested with equal probability regardless of size as in Fig. 3a of the main 780 781 manuscript). b) In the second scenario we rescaled the steepness of the Beverton-Holt curves so 782 that a population with lower egg production produces the same number of recruits on average 783 also after life histories have evolve (shading corresponds to harvest rates as in panel a).





Figure A2. Long-term equilibrium yield for the optimal life history strategies under different harvest levels when the Beverton-Holt total egg production-recruitment curve is assumed to remain constant with the parameters as in the pre-harvest situation. Bell-shaped size-selectivity: a) $L_{\text{max}} = 70$ cm; b) $L_{\text{max}} = 90$ cm; c) $L_{\text{max}} = 110$ cm. Sigmoid size-selectivity: d) $L_{\text{max}} = 70$ cm; e) $L_{\text{max}} = 90$ cm; f) $L_{\text{max}} = 110$ cm. For further explanation of axes see Fig. 3 in the main article.



791

Figure A3. Optimal age at maturation in when gear selectivity reflects the gear types used in the fishery for Northeast Arctic cod, that is trawling with sigmoid size-selectivity at the feeding grounds and gillnets and longline with bell-shaped selectivity curves at the spawning grounds. For both gear types we used $L_{max} = 90$ cm. The axes are further explained in the legend to Fig. 3 of the main article.

790