

A comparative analysis of *Alexandrium catenella/tamarensis* blooms in Annaba Bay (Algeria) and Thau lagoon (France); phosphorus limitation as a trigger

Analyse comparative des efflorescences d'*Alexandrium catenella/tamarensis* en baie d'Annaba (Algérie) et dans l'étang de Thau (France) : limitation en phosphore comme élément déclencheur

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Abstract:

Environmental conditions ultimately leading to blooms of the toxic dinoflagellate *Alexandrium catenella/tamarensis* were investigated at two Mediterranean sites (Annaba Bay, Algeria and Thau lagoon, France). Three years were examined in details: 1992 (a pre-*Alexandrium* period), 2002 (a year with the first bloom in Annaba) and 2010 (a year with a major bloom in Annaba). Most conditions were similar, but ammonium concentrations were much higher in Annaba (up to 100 μM) than in Thau (up to 10 μM). First records of *A. catenella/tamarensis* were in 1995 for Thau and 2002 for Annaba, and coincided with soluble reactive phosphorus (SRP) decreasing below a concentration of about 1 μM . No other environmental variable could be related to those blooms. Thus, it is likely that the large reductions in SRP at both sites led to phosphorus limitation of a certain number of phytoplankton species and favored the development of *A. catenella/tamarensis*.

Résumé

Les conditions du milieu amenant à des efflorescences du dinoflagellé toxique *Alexandrium catenella/tamarensis* ont été suivies sur deux sites méditerranéens (la baie d'Annaba, Algérie et l'étang de Thau, France). Trois années ont été examinées en détail : 1992 (période pre-*Alexandrium*), 2002 (première efflorescence à Annaba) et 2010 (efflorescence majeure à Annaba). La plupart des conditions étaient comparables, sauf pour l'ammonium, avec des concentrations bien plus fortes à Annaba (jusqu'à 100 μM) qu'à Thau (jusqu'à 10 μM). Les premières observations d'*A. catenella/tamarensis* ont été réalisées en 1995 à Thau et 2002 à Annaba, et ont coïncidé avec le passage des concentrations en phosphore réactif dissous (PRD) sous la barre des 1 μM . Aucune autre variable du milieu n'a pu être reliée à l'apparition de ces efflorescences. Il est donc probable que

les fortes diminutions en PRD aux deux sites aient entraîné une limitation par le phosphore d'un certain nombre d'espèces phytoplanctoniques et favorisé le développement d'*A. catenella/tamarensis*.

Keywords : *Alexandrium catenella/tamarensis* ; Blooms ; Algeria ; France ; Mediterranean

Mots clés : *Alexandrium catenella/tamarensis* ; Efflorescences ; Algérie ; France ; Méditerranée

1. Introduction

The genus *Alexandrium* belongs to one of the most important genera of harmful algal blooms (HABs) producing Paralytic Shellfish Poisoning (PSP) toxins [1]. *Alexandrium* species have a worldwide distribution, being present in coastal zones, continental shelf waters, in temperate as well as tropical areas [2]. The species *A. catenella*, *A. tamarensis* and *A. fundyense* are the most toxic of the genus *Alexandrium*. Together, these species make up the *A. tamarensis* complex [1] and [3]. Such species are often observed around the Mediterranean region [2]. In the northern part, *A. catenella* has first been reported in the Balearic islands in 1983 [4], and later on the French [5], Spanish [6] and Italian [7] coasts. In the southern part, *Alexandrium* species have been reported in Morocco [8], [9], [10] and [11], Tunisia [12] and [13], and Algeria [14].

A. catenella blooms in inorganic nutrient rich [6] and [15] and nutrient poor [16] and [17] waters. This is from the viewpoint of inorganic nutrients. The *A. tamarensis* complex has been classified as a « frontal zone taxon » or « mixing-drift » group [18]. and therefore characteristic of areas that are unstable and intermediate between high and low inorganic nutrients. Very little is known generally about dissolved organic nutrients at those sites. It is therefore difficult to find recurrent patterns of bloom development for this mixotrophic species in relation with trophic conditions. Comparative studies of HABs ecosystems have been recognized as a way to deal with such difficulties and could “reveal fundamental processes governing population development” [19].

76 In this study, we compare *A.catenella/tamarensis* dynamics in two contrasted sites in the
77 Mediterranean. The first one is the relatively eutrophic [14] Bay of Annaba (Algeria) and the
78 second is the Thau lagoon (France) that has recently gone through a period of
79 oligotrophication [20]. We try to relate those blooms to environmental factors in those two
80 coastal sites on both sides of the Mediterranean Sea, and deduce possible causes of their
81 recent development and recurring presence. We study 3 different years over an 18 year
82 period, one being “pre-Alexandrium”, one coinciding with the first *A. catenella* bloom in
83 Annaba and the last one with a very large *A. catenella* bloom in Annaba. As the Thau lagoon
84 blooms were connected with oligotrophication, we selected a period over which the same
85 phenomenon appeared to take place in Annaba.

86

87 2. Material and methods

88

89 2.1. Study sites

90 The Bay of Annaba is located in the eastern part of Algeria (Fig. 1), between cape
91 Rosa (36°38'N, 8°15'E) and Cape de Garde (36°38'N, 7°16'E), with a maximal depth of 50
92 m. The main temporary rivers (oueds) are Oued Seybouse (second river in Algeria), Oued
93 Boudjemâa carrying urban and industrial effluents, and Oued Kouba carrying sewage waters.
94 *Alexandrium catenella* has bloomed periodically since 2002 in this bay [14].

95 The Thau lagoon is a shallow marine lagoon located on the French Mediterranean
96 coast (43°24'N, 3°36'E) covering 75 km² (Fig. 2). The mean depth is about 4 m, with a
97 maximum of 10 m. The lagoon is connected to the sea by 3 narrow channels. Three oyster
98 farming areas are located along the northwestern shore. The lagoon represents 10% of French
99 annual oyster production and is the main oyster production center on the Mediterranean. Since

100 1998, it has experienced recurrent blooms of *Alexandrium catenella* that periodically threaten
101 economic activities [20].

102 2.2. Sampling

103 Sampling stations for Annaba Bay

104 Station 1 (36°54.073'N, 7°46.929'E) has a depth of 5 m and is sheltered from dominant (NW)
105 winds, with a rocky/sandy bottom.

106 Station 2 (36°53.976'N, 7°47.111'E) is deeper (13 m) than station 1 and is located in front of
107 the commercial harbor. It is subject to major land influence through input from Oued
108 Seybouse and Oued Boudjemaa and is considered eutrophic relative to station 1. Bottom
109 sediments are silty. The sampling frequency was twice a month.

110 Sampling stations for Thau lagoon

111 Station B (43°26.070'N, 3°39.920'E) is located at one of the deepest part of the lagoon (8 m)

112 Station A5 (43°26.916'N, 3°40.300'E) is located in the Angle Creek and is shallower (2 m)
113 and more sheltered than Station B. The sampling frequency varied from 1 to 8 samples per
114 month for physical variables. For chemical and biological variables, sampling was carried out
115 at least twice a month at station B.

116 2.3. Physical variables

117 In Annaba, surface temperature and salinity were measured with a WTW 191 multiparameter
118 probe (precision 0.1°C for temperature and 0.05 for salinity).

119 For Thau, the Ifremer observation network provided records of surface water temperature and
120 salinity (monthly means).

121 2.4. Nutrients

122 At both sites, samples for ammonium determination were immediately fixed and measured at
123 the laboratory [21]. Samples for the other nutrients were kept in the cold and dark, then frozen
124 at -20°C until analysis. Nitrate [22], nitrite [23] and soluble reactive phosphorus (SRP)[24]

125 were measured after thawing at room temperature. Detection limits were 0.05, 0.01, 0.05 and
126 0.02 μM for nitrate, nitrite, ammonium and SRP respectively.

127 2.5. Biological variables

128 Chlorophyll

129 In Annaba, chlorophyll a was measured by spectrophotometry [25]. In Thau lagoon,
130 chlorophyll a was estimated from 90% acetone extracts and fluorimetry [26] or
131 spectrofluorometry [27].

132 Microphytoplankton.

133 In Annaba, for the qualitative survey of phytoplankton, horizontal net tows were carried out at
134 50 cm below the surface, with a 20 μm mesh size net. During *Alexandrium* blooms and for the
135 quantitative study, bucket sampling was used, and 40 liters were filtered on a 20 μm mesh
136 size net. Samples were then resuspended in a 100 ml volume, then subsamples of 1 or 5 ml
137 were used depending on the cell density. Samples were immediately fixed with buffered
138 formaldehyde (5% final concentration). Identification was by electron microscopy in 2002
139 [14] and by optical microscopy in 2010, using morphological criteria. Both 2 cell and 4 cell
140 chains were observed.

141 In Thau lagoon, at least one liter of sea water was taken with a sampling bottle. Additions of
142 0.2 ml to 0.4 ml Lugol's solution per 100 ml was carried out for quantitative analysis. A
143 counting chamber was filled with 10 ml of fixed sample for sedimentation [28].

144 Phytoplankton cells greater than 10 μm equivalent cell diameter were counted by optical
145 microscopy.

146 2.6. Statistical analyses

147 Variables were compared by Kruskal-Wallis one-way analysis of variance by ranks and
148 Dunn's multiple comparison tests.

149

150 3. Results

151

152 In 1992, *A. catenella/tamarensis* had not yet been recorded either in Thau lagoon [5] or in

153 Algeria [14].

154 3.1. *Alexandrium catenella/tamarensis* blooms

155 *Alexandrium catenella* was first observed in Annaba in March 2002 and in Thau in July 1995.

156 Temporal changes of cell densities during 2002 are shown for station 2 in Annaba and

157 stations A5 and B at Thau (Fig. 3). Highest monthly mean cell densities (117000 cells L⁻¹) in

158 Annaba were recorded in March with a secondary peak in December (200 cells L⁻¹). In Thau,

159 monthly mean cell densities were generally higher at A5 (maximum of 66700 cells L⁻¹ in

160 May) than at station B (maximum of 5400 cells L⁻¹ also in May), with a secondary peak in

161 August (9600 cells L⁻¹).

162 In 2010, 2 stations were sampled in Annaba (Fig. 4) with higher cell densities at Station 2 (up

163 to 681000 cells L⁻¹ in May). At station 1, monthly mean cell densities never exceeded 73000

164 cells L⁻¹ (also in May). For individual dates, a record peak of 1.38 x 10⁶ cells L⁻¹ was recorded

165 in May 2010 at the surface of station 2. This was the highest cell density ever recorded so far

166 in Annaba.

167 In Thau, during the same year, no bloom developed in the spring, but cell densities reached a

168 maximum of 4400 cells L⁻¹ in September at station A5 and 1500 cells L⁻¹ also in September at

169 station B.

170 In 2002, there appeared to be no synchrony in blooms between sites (Fig. 3), but in 2010,

171 there was a bimodal distribution at both sites, with maxima in May-June and September-

172 October (Fig. 4).

173

174 3.2. Environmental factors

175 Table 1 summarizes the ranges of physical and chemical variables at both sites. Sea surface
176 temperatures were higher in Annaba than in Thau. The range of salinities was larger in Thau
177 than in Annaba, with maximal values being greater in Thau lagoon. Maximal nitrate,
178 ammonium and SRP concentrations were observed in Annaba, and were sometimes greater
179 than in Thau by an order of magnitude, such as for ammonium. Concerning phytoplankton
180 biomass, the range of chl a values was about the same.

181 Table 2 summarizes the changes in environmental parameters occurring over an 18 year
182 period. The year 1992 can be considered as a “pre-Alexandrium” period for both sites. No
183 significant changes in either annual mean SST, salinity (not shown) or ammonium occurred
184 over this period at either site. In Annaba, the most spectacular changes were due to nitrate
185 concentrations that decreased by a factor of 50 (significant difference with $p < 0.001$) and
186 specifically SRP concentrations that decreased by a factor of 250 (significant difference with
187 $p < 0.001$), from a mean annual value of 17.4 μM in 1992 to 0.07 μM in 2010. In Thau, the
188 most dramatic decrease in SRP occurred between 1972 and 1992 (from about 10 to 1 μM in
189 summer and from 3 μM to undetectable levels in winter). Between 1992 and 2010, the
190 decrease was significant ($p < 0.05$).

191

192 4. Discussion

193

194 Although nitrate decreased dramatically between 1992 and 2010 in Annaba, nitrogen
195 limitation is unlikely because ammonium levels remained high (Table 2). The decrease in
196 SRP was due to the discontinuation of phosphogypsum outputs from a local fertilizer plant in
197 Annaba (M. Retima, pers. com.), and from implementation of sewage treatment plants in
198 Thau [29]. Concerning *A. catenella* ability to use low SRP levels, laboratory studies show that
199 the half-saturation constant for SRP uptake was variable in cultures, depending on the growth

200 rate, but could reach values as low as 0.03 μM for strain ACT03 and 0.01 μM for strain TL01
201 [30]. This indicates that *A. catenella* could be a very strong competitor at the low SRP
202 concentrations recently observed in Annaba Bay. An alternative to its periodic dominance is
203 dissolved organic phosphorus use. Alkaline phosphatase activity was induced in *A. catenella*
204 at SRP levels between 0.4 and 1 μM [30] that are intermediate values between those observed
205 in 2002 and 2010 in Annaba Bay.

206 Thus it seems very likely that the appearance of blooms of *A. catenella* in Annaba Bay is due
207 to SRP reaching very low levels where other members of the phytoplankton community
208 cannot compete for the acquisition of this limiting resource.

209 This situation is similar to that experienced in Thau lagoon [20] where SRP reached
210 undetectable winter levels in 1992-1993 and *A. catenella* was first reported in 1995.

211 A parallel can also be made with the Seto Inland Sea (Japan) where SRP decreased from 0.7
212 to 0.2 μM between 1978 and 1984 [31]. *A. catenella* vegetative cells were first reported in this
213 area in 1982 [32], so that *A. catenella* probably started to bloom when SRP reached levels
214 between 0.7 and 0.2 μM . In the Gulf of Olbia (Sardinia, Italy), the first *A. catenella* bloom
215 was reported in 1999 [33] under conditions of SRP below 1 μM that were lower than
216 observed in 1992-1993, a situation also similar to the results presented here.

217

218 5. Conclusions

219

220 This inter-site comparison between the Northern and Southern coasts of the Mediterranean
221 has revealed that one of the causes of the development of *A. catenella/tamarense* is the onset
222 of inorganic phosphorus limitation in both environments that allowed this organism to
223 develop periodically and produce recurrent blooms. No other environmental variable could be
224 related to the emergence of this organism. Such an explanation is an alternative to the general

225 concept that eutrophication leads to harmful algal blooms and may require a revision of
226 coastal water management policies.

227

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333 No conflict of interest affecting the authors exists

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Table 1. General physical and chemical features at both sites. Range of values (monthly means) over several years (1992, 2002 and 2010 for Annaba ; 1992 to 2010 for Thau). S : salinity ; SRP : soluble reactive phosphorus ; SST : sea surface temperature ; U : undetectable

Site	SST (°C)	S	NO ₃ (μM)	NH ₄ (μM)	SRP (μM)	Chl a (μg/liter)
Annaba	14.1-28.6	35.3-38.1	U-65	U-110	U-43	0.8-32.8
Thau	5.0-25.6	29.5-39.2	U-20	U-10	U-10	0.3-37.4

Table 2. Changes in environmental factors at two coastal sites North and South of the Mediterranean for 3 selected years over an 18 year period.

Seasonal distributions are available at http://www.st.nmfs.noaa.gov/nauplius/media/time-series/site_mediterranean-thau-lagoon-phy/copepodite/index.html for Thau and in [34] and [14] for Annaba respectively in 1992 and 2002.

NA: not available; P: phosphorus; SD: standard deviation; SST: sea surface temperature.

Year	SST			nitrate			ammonium			soluble reactive P			Chlorophyll a		
	Mean	SD	n	mean	SD	n	mean	SD	n	mean	SD	n	mean	SD	n
Annaba	(°C)			(µM)			(µM)			(µM)			(µg/liter)		
1992	21.6	4.7	8	25.8	20.5	8	49.9	37.7	8	17.4	14	8	10.3	9.5	6
2002	19.5	4.1	12	1.9	3.0	12	11.6	15.3	12	1.2	1.8	12	9.0	11.3	11
2010	20.7	4.8	11	0.5	0.9	12	23.6	24.3	11	0.07	0.13	11	2.4	1.4	11
Thau															
1992	18.4	7.1	12	0.5	0.5	11	1.4	3.1	10	1.2	1.2	10	1.5	0.6	9
2002	15.9	5.6	12	0.2	0.3	3	1.1	0.7	10	0.5	0.2	10	3.0	1.5	12
2010	15.1	6.9	12	0.2	0.1	3	0.7	0.6	3	0.2	0.1	3	2.2	1.1	11

346 Figure legends

347

348 1. Station positions in Annaba Bay, Algeria.

349 2. Station positions in Thau lagoon, France.

350 3. Changes in monthly mean *Alexandrium catenella/tamarensis* cell densities during

351 2002. Diamonds : Thau station A5 ; squares : Thau station B ; triangles : Annaba

352 station 2. Detection limits: 100 cells L⁻¹.

353 4. Changes in monthly mean *Alexandrium catenella/tamarensis* cell densities during 2010.

354 Diamonds : Thau station A5 ; Squares : Thau station B ; triangles : Annaba station 2 ;

355 circles : Annaba station 1. Detection limits: 100 cells L⁻¹.

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