
Metallothionein mRNA induction is correlated with the decrease of DNA strand breaks in cadmium exposed zebra mussels

Françoise Vincent-Hubert^{a, b, *}, Amélie Châtel^{a, b}, Catherine Gourlay-Francé^{b, c}

^a IRSTEA Unité de Recherches Hydrosystèmes et Bioprocédés, 1 rue Pierre-Gilles de Gennes, CS 10030, 92761 Antony Cedex – France

^b IFREMER, Laboratoire de virologie - LNR, rue de l'île d'Yeu, BP 21105, F 44311 Nantes Cedex 03 – France

^c Université de Nantes, MMS, EA2160, Faculté de Pharmacie, 1 rue G. Veil, BP 53508, 44035 Nantes Cedex 1, France

*: Corresponding author : Françoise Vincent-Hubert, email address : francoise.vincent@ifremer.fr

Abstract:

We have previously shown that cadmium (Cd) and Benzo[a]pyrene (BaP) induced early DNA damages in zebra mussels, and that the level of DNA strand breaks (SB) returned to a basal level after 3 days of exposure to Cd. The aim of the present study was to go further in the mechanisms of Cd and BaP detoxification. For that purpose, expression of genes encoding for metallothionein (MT), Aryl Hydrocarbon Receptor (AHR), P-gP, catalase, glutathione S-transferase and Heat shock protein 70 (HSP70) proteins have been measured using RT-qPCR. Data reported here show that Cd is a strong inducer of MT and HSP70 genes, and that BaP is a strong inducer of P-gP and AHR genes. Exposure to Cd and BaP resulted in moderate changes in antioxidant enzymes mRNA. Since the increase of MT mRNA occurred when the DNA SB level returned to its basal level, we can suggest that MT is implicated in cadmium detoxification.

Highlights

► cadmium (Cd) and Benzo[a]pyrene (BaP) induced DNA damages in zebra mussels. ► Cd is a strong inducer of Metallothionein (MT) and HSP70 genes. ► Metallothionein might be implicated in cadmium detoxification. ► BaP is a strong inducer of P-gP and AHR genes. ► Cd and BaP induced in moderate changes in antioxidant enzymes mRNA.

Keywords : Benzo[a]pyrene ; cadmium–metallothionein ; Aryl Hydrocarbon Receptor ; antioxidant enzymes ; RT ; qPCR

47 **1. Introduction**

48

49 The zebra mussel is an organism of choice for the monitoring of metallic and organic
50 contaminants in freshwater ecosystems [1-4]. The sensitivity of zebra mussels to genotoxic
51 contaminants has been demonstrated through the induction of micronuclei, DNA strand-
52 breaks and DNA adducts [5-9]. Similarly, field studies revealed that DNA damages were
53 higher in mussels from polluted sites compared with reference site [10, 11]. However, little is
54 know about the regulation of proteins implicated in detoxification following DNA strand
55 breaks.

56 Recently, gene expression profiles of proteins implicated in detoxification and stress
57 were analyzed in zebra mussels exposed to cadmium and xenobiotics (metoprolol and
58 levonorgestrel). These first observations improved our knowledge in the regulation of these
59 genes [12-14]. These genes encode for proteins that are reliable biomarkers in field and
60 laboratory exposures to contaminants, including metals and organic pollutants. MT has been
61 suggested to be key elements causing the retention of Cd in mussels [15]. High level of MT
62 mRNA was observed in many organisms exposed to cadmium, including zebra mussel [16].
63 Heat shock proteins are widespread in plants, bacteria and animals and belong to chaperone
64 proteins, which are important for protein folding, protein transport and cell stabilisation [17,
65 18]). Superoxide dismutase (SOD) and Catalase (CAT) are antioxidant enzymes and reliable
66 biomarkers of reactive oxygen species [19], SOD is responsible for the reduction of the
67 superoxide radical into hydrogen peroxide, and CAT for catalysing hydrogen peroxide to
68 water. Glutathione S-transferase gene (GST) encodes a phase II metabolizing enzyme
69 known to catalyze the conjugation of glutathione with various electrophilic substances and
70 plays a role preventing oxidative damages [20]. The P-glycoprotein, P-gp, is a member of the
71 multidrug transporter proteins that are ATP dependent proteins which efflux a variety of
72 moderately hydrophobic compounds out of cells [21-23]. The HSP70, which is involved in
73 the processing of misfolded proteins due to different kinds of stress, is used as a very

74 general ecotoxicological endpoint for protein damage and subsequent protective
75 mechanisms [24].

76 The data presented here are the second part of a article published earlier in which we
77 first shown that zebra mussels exposed to environmentally relevant concentrations of Cd and
78 BaP displayed DNA damages [25]. Cadmium, as well as Benzo[a]pyrene (BaP), are model
79 environmental contaminants, classified as a human carcinogen by the IARC [26]. Cd shows
80 a co-genotoxic effects in combination with other mutagenic agents such as UV light,
81 alkylating agents and B[a]P in mammalian cells. Cadmium is assumed to be a weak
82 genotoxicant that amplifies the genotoxic effect of B[a]P [27, 28]. Therefore, based on these
83 studies, we wanted to determine whether Cd could have the same effect in mussels. We
84 have shown that Cd induced early DNA damages (DNA strand break and DNA oxidation),
85 and that BaP induced DNA damages only on the third day of exposure. The most surprising
86 is that the level of DNA returned to the basal level despite the continued presence of
87 cadmium.

88 The aim of the present study was to go further in the mechanisms of Cd and BaP
89 detoxification to determine why the level of DNA stand breaks return to its basal level. For
90 that purpose, we measured gene expression of several proteins described to be involved in
91 detoxification, such as metallothionein, AHR, P-gP, catalase, SOD and glutathione S-
92 transferase. HSP70 gene expression was also measured as a general indicator of stress.
93 Samples analyzed here are frozen samples of the first published study.

94
95
96
97
98
99
100
101

101 **2. Materials and Methods**

102

103 **2.1 Chemical reagents**

104 CdCl₂, BaP, DMSO, agarose, trypan-blue and PBS were purchased from Sigma (France).

105

106 **2.2 Mussel sampling and maintenance conditions**

107 Adult specimens of the zebra mussel *Dreissena polymorpha* (shell length 25±2 mm) were
108 collected in the East channel (Commercy, France), which is a reference site. Cd
109 concentration in mussels is in good agreement with those usually found in bivalves from
110 clean waters [29]; BaP was not detected [11, 30]. Mussels were detached from the rock by
111 cutting their byssus threads and carried to the laboratory in their original water. The mussels
112 were randomly placed in a 20-L aerated tank with about 100 specimens each, and
113 acclimatized to Valvert mineral water and temperature (15°C), a day/night lighting system
114 was applied. The mussels were fed every 3 days with algae (*Pseudokirchneriella*
115 *subcapitata*) and the water was changed every 2 days. The animals that had not become
116 attached to the tank were removed. The mussels were maintained in the above conditions for
117 8 days before the start of treatment since it was demonstrated that this time is needed to
118 reach a baseline level of DNA damage [31].

119

120 **2.3 In vivo exposure of zebra mussels**

121 We chose two genotoxicants with different modes of genotoxic action: benzo[a]pyrene, which
122 prevalent mechanism of action relies on DNA adducts, and cadmium for its pro-oxidant
123 properties and its role in the inhibition of DNA repair. In order to work with more
124 environmentally realistic concentrations, the Cd and BaP concentrations used here were low-
125 level concentrations equal to the lowest genotoxic concentration published for zebra mussels
126 or marine mussels [6, 32]. A stock solution of CdCl₂ and BaP was prepared in water or pure
127 DMSO, respectively. Preliminary experiments showed that DMSO in water (0.001%) did not
128 induce genotoxic effects.

129 Following overnight equilibrium of tanks with chemical compounds to avoid adsorption to the
130 wall of the tank, mussels were added after renewing the water. Ten mussels were exposed
131 to BaP at 10 µg /L, Cd at 10 µg /L and to a combination of Cd (10 µg/L) and BaP (1 µg/L)
132 dissolved in 8 L of water for 11 days at 15°C. In order to maintain the chemical concentration
133 constant, water was renewed every 2 days and then contaminated again with chemicals. For
134 each parameters analyzed, four mussels were sacrificed after 0, 12 h, 24 h and 3, 5 and 11
135 days of exposure for further analysis (bioaccumulation, DNA damages and gene expression).
136 Cadmium bioaccumulation, DNA damages and nuclear abnormalities data were already
137 published [25]. Gills of 4 mussels were pooled and frozen for gene expression analysis.

138

139 **2.4 RNA extraction, RT-PCR and qRT-PCR analysis**

140 Total RNA was isolated from frozen gills of *Dreissena polymorpha* using the phenolic reagent
141 TRIZOL (Invitrogen, France). RNA concentration and purity was measured by
142 spectrophotometric absorption (260/280 and 230/280 ratio), RNA quality was checked with
143 denaturing gel electrophoresis. First strand cDNA synthesis was carried out on 1 µg of total
144 RNA extract with oligo-dT primers according to Improm II Reverse Transcriptase kit
145 (Promega, France).

146 Real-time PCR assays for Ribosomal protein S3 (S3), Catalase (CAT), superoxide
147 dismutase (SOD), Glutathion S-transferase (GST), Metallothionein (MT), Heat-shock protein
148 70 (HSP70), Aryl Hydrocarbon receptor (AHR), and P-gp were run in a LightCycler 480 Real
149 Time PCR System (Biorad) using SYBR Green Power Master Mix (Invitrogen, France). PCR
150 reactions and primers pairs were those published by [12] except for Ribosomal protein S3
151 [14]. Determination of the transcript abundances in individual sample was conducted with the
152 comparative C_T method ($\Delta\Delta C_T$) in consideration of a calibrator sample (control sample).
153 Samples were normalized to Ribosomal S3 gene due to its constant expression [14]. At the
154 end of each PCR reaction a melting curve analysis was carried out to proof assay specificity.
155 PCR efficiency values for reference and tested genes were calculated as described [33], and
156 assumed to be close to 100% from these calculations.

157

158

159

160 **2.5 Statistic analysis**161 RT-qPCR results are given as mean values \pm S.D. of 3 values (4 mussels per condition

162 pooled and 3 repetitions of each test). The calculated values were compared among different

163 groups (model agents' concentration) using an Analysis of Variance (ANOVA) followed by a

164 Tukey post hoc test. Three levels were considered significant: $p < 0.05$ (*), $p < 0.01$ (**) and165 $p < 0.001$ (***). All statistical analysis was performed with R software.

166

Accepted Manuscript

166 **3. Results**

167

168 The data published here are the second part of a study published earlier [25] in which we
169 observed an increase of DNA strand breaks in gill cells of mussel exposed to
170 environmentally relevant concentrations of Cd and BaP (see Table 1, Supporting
171 information). The renewal of the water every 2 days ensured a constant cadmium
172 concentration over the 11-day experiment: 8.27 ± 0.18 and 8.23 ± 0.34 $\mu\text{g/L}$ in CdCl_2 and
173 $\text{CdCl}_2 + \text{B[a]P}$ media, respectively. Bioaccumulation of cadmium in soft tissues of mussels
174 exposed to CdCl_2 (10 $\mu\text{g/L}$) or $\text{CdCl}_2 + \text{B[a]P}$ (10 $\mu\text{g/L} + 1$ $\mu\text{g/L}$) increased over the 11 days
175 of exposure from 9 $\mu\text{g/g dw}$ after 24 hr of exposure to 45 $\mu\text{g/g dw}$ at the end on day 11 [25].

176

177 Strongest changes in the expression of detoxification genes were found for MT, HSP70 and
178 P-gp genes. A significant increase was noticed, up to 6.5, 10 and 17 fold, respectively after
179 5 days of exposure to BaP for P-gp ($p < 0.001$), after 3 days of exposure to Cd and Cd+BaP
180 for MT ($p < 0.001$), and after 11 days of exposure to Cd for HSP70 ($p < 0.001$).

181 It is important to note that DNA strand breaks induced by Cd return to the basal level when
182 MT gene expression started to increase. Indeed, MT mRNA expression increased gradually
183 from day 3 until day 11. The same trend was observed when mussels were exposed to
184 Cd+BaP. The presence of BaP in the mixture does not inhibit the effect of Cd on MT gene
185 induction.

186 Also, as expected, BaP is as well a strong inducer of AHR gene in zebra mussel, as we
187 observed an early and transient increase of the AHR mRNA expression as soon as 12 hr
188 after the beginning of the exposure to BaP alone ($p < 0.01$). and to the mixture Cd + BaP
189 ($p < 0.05$) AHR mRNA expression was slightly decreased at day 1 and day 5, respectively in
190 BaP ($p < 0.05$), Cd+BaP ($p < 0.001$) and in Cd exposed
191 mussels ($p < 0.01$).

192 P-gp was slightly up-regulated by Cd and BaP during the 11 days of exposure, with the
193 strongest induction at day 5 in Cd exposed mussels.

194 HSP70 was slightly up-regulated during the first days of exposure to Cd, and strongly up-
195 regulated at day 11 as already mentioned, indicating protein damage.

196 In contrast, the effect of Cd and BaP was less important on SOD, GST and CAT mRNA levels.

197 More precisely, CAT mRNA was two times decreased after 12 hr of exposure in BaP
198 exposed mussels ($p < 0.01$) and after 24 hr ($p < 0.001$) in Cd and Cd+BaP exposed mussels
199 compared to control. On the contrary, CAT gene was up-regulated, two times, after 3 and 5
200 days of exposure, respectively in Cd+BaP and Cd exposed mussels.

201 GST mRNA was 1.5 times increased after 12 h of exposure to Cd+BaP, and significantly
202 decrease after 3 days and 11 days of exposure to Cd, BaP and Cd+BaP.

203 SOD mRNA was about two times decreased in Cd exposed mussels from the first day until
204 the eleventh day but this was significant only on days 3 and 5. SOD mRNA was 1.5 times
205 increased after 12 hr, 5 and 11 days of exposure to Cd+BaP ($p < 0.001$ and $p < 0.05$) and after
206 3 days and 11 days of exposure to BaP ($p < 0.001$).

207

208

209

210

211

211 **4. Discussion**

212

213 The present study is the second part of a study in which we reported early genotoxic effects
214 of Cd and BaP in gill cells of zebra mussels. We wanted to determine whether stress-related
215 genes and detoxification genes are regulated at the transcriptional level in the hours
216 following genotoxic damage in gills.

217 Tissue-specific differences in gene expression following exposure to metals (Cd,
218 Cu, Hg) and BaP have been published previously [14] [34], hence we focused here on gills in
219 the hours following genotoxic damage. Indeed similar patterns of mRNA abundance were
220 observed between gills and digestive gland in Cd exposed mussels, while in BaP exposed
221 mussels, we noticed an early induction of *HSP70*, *PgP*, *AHR* and *SOD* mRNA levels in the
222 gills compared to the digestive glands. Therefore, it appears from these studies that gill
223 tissue represents an interesting model to investigate the molecular mechanisms of
224 detoxification [34].

225 We show here that, at environmentally relevant concentrations, Cd is a strong inducer
226 of MT and HSP genes, and that BaP is a strong inducer of *PgP* and *AHR* genes. Exposure to
227 Cd and BaP resulted in moderate changes in antioxidant enzymes mRNA. The mRNA level
228 of metallothionein increases when the level of DNA strand breaks returned to its basal level,
229 which is probably the most interesting result of this study

230 We confirmed that Cd (10 µg/L) is a strong inducer of MT gene, 8 times more than control.
231 Indeed, zebra mussels accumulate Cd at water concentrations as low as 9 µg/liter, excluding
232 the possibility of a homeostatic control [35], and as a consequence, MT protein is induced
233 [36]. Therefore it appears that MT mRNA induction and MT protein induction are early increase in
234 zebra mussel as Lecoœur et al. [37] observed an early increase of total MT biosynthesis after exposure
235 to Cd (2-20 µg/L). Only one isoform of MT gene has been reported for zebra mussel, the *Dp*
236 MT [38]. Engelken and Hildebrandt [38] first shown that cadmium was an inducer of MT gene
237 but this was for elevated concentration of cadmium; afterwards, an increase of this gene
238 expression has also been observed for lower Cd concentrations [14, 38-40] [16]. Cadmium

239 increases the synthesis rate of metallothionein messenger-RNA via transcriptional activation
240 of metal-responsive factors located in the upstream region of MT genes [41]. We used here
241 cadmium at environmentally relevant concentrations, measured in rivers strongly impacted
242 by mining activities [40].

243 Our data provide evidence on the exposure time required for the induction of MT gene by
244 cadmium: MT mRNA level gradually increases from the third day until the eleventh day of
245 exposure suggesting that during the first three days, the physiological concentration of MT
246 protein was not high enough to adsorb the Cd. A time-dependent increase was also reported
247 but for higher concentration of cadmium (20 µg/L) [14].

248 Interestingly, the increase of MT mRNA occurs as the DNA strand break level return to its
249 basal level, suggesting that DNA repair was efficient. However as a constant concentration of
250 cadmium in the water and an increase of bioaccumulated cadmium were measured, these
251 data also suggest a decrease in Cd bioavailability. Indeed, the MT protein concentration is
252 probably higher on the third day, compared to the first day; a higher level of MT protein can
253 lead to a decreased bioavailability of Cd and, as a consequence, Cd induced DNA damage
254 could be decreased too. Similar observations were recently reported by Qu & Waalkes [42];
255 these authors demonstrated that MT-competent cells activate MT in response to Cd, while
256 MT-deficient cells adapt to Cd primarily by turning on oxidant response systems.

257 Cd is known to induce reactive oxygen species (ROS) which in turn lead to DNA strand
258 breaks and Oxidative DNA damage [43], an effect that we also observed with zebra mussels
259 [25]. MT has two major functions in Cd toxicity inhibition: (i) MT detoxicates the metal by
260 direct binding and (ii) the cysteines in MT appear able to react directly with ROS, thus MT
261 may also act as an antioxidant independently of metal sequestration [45, 46]. Our data
262 highlight the role of MT in the inhibition of Cd genotoxicity.

263 One of the known protective mechanisms that aquatic animals have developed in
264 response to stresses is the induction of HSPs. We observed that Cd is a strong inducer of
265 HSP70 gene, on the contrary to Navaro et al. [14] who reported HSP70 gene induction only
266 in the digestive gland. Interestingly the strongest up-regulation of HSP mRNA, seventeen

267 times more than control, was observed after 11 days of constant exposure to Cd. HSP70 are
268 protein chaperones that are induced by various environmental stressors including organic
269 pollutants [47] and metals such as Hg, Cu, Cd [14]. HSP acts to prevent protein aggregation
270 and to maintain functional conformations. This transcriptional regulation of HSP70 is
271 probably necessary to enhance the tolerance to cadmium, as elevated levels of HSPs have
272 been proven to protect against the negative impact of metals on protein integrity [48, 49].
273 We confirmed that BaP (4.4 nM) is an inducer of HSP70 as we observed previously [34]. The
274 regulation of HSP70 gene in response to BaP remains unclear. For example, high
275 concentration of BaP (10 μ M) suppress the transcription of HSP70 gene in human
276 endothelial cell [50], and BaP (1 μ M) up-regulate HSP70 mRNA in bronchial cells, suggesting
277 a potential role of HSP70 in the NER DNA repair [51].

278 We observed that Cd up-regulated both MT and HSP70 genes. It is well known that
279 HSP70 and MT promoters possess anti-oxidant response element (ARE) which may provide
280 a mechanistic explanation to this correlated response [52, 53]. Cd has been described to
281 induce HSP70 and MT proteins in mammals and oysters [54, 55].

282 In Mammals, Cytochrome P450 enzymes are important in the metabolism of
283 xenobiotics, such as PAHs [56]). Induction of CYP1 is mediated mainly through a specific
284 cytosolic receptor, the aryl hydrocarbon receptor (AHR). Activated AHR is also a transcription
285 factor of other genes that encode phase I and II xenobiotic metabolizing enzymes [57]. Our
286 data confirmed that BaP is a strong inducer of AHR gene as we recently observed [34]. AHR
287 mRNA induction was lower in Cd+BaP exposed mussels, suggesting an interaction between
288 Cd and BaP, as observed for fish and Human hepatocytes. Indeed, it has been shown that
289 Cd decreases the induction of AHR by BaP in that biological models [58, 59]. AHR seems to
290 be implicated in the detoxification of BaP in zebra mussels; as the AHR serves as a
291 transcription factor for enzymes of the CYP450 family, our data suggest that BaP is
292 metabolised by CYP450, as it was recently proposed for pharmaceutical compounds [12].

293 We observed that exposure to Cd and BaP resulted in moderate changes in
294 antioxidant enzymes, characterized by a slight decrease of mRNA levels of SOD, GST and

295 CAT genes, which suggest a moderate regulation of these genes at the transcriptional level.
296 It is known that Cd induced a depletion of cellular GSH, which could explain the slight
297 decrease of anti-oxidant enzymes mRNA level. We previously observed that when BaP was
298 added in aquaria containing zebra mussels two times a day, animals presented an increase
299 in GST, CAT and SOD mRNA levels after 12h of exposure. In the present study, the same
300 concentration of BaP (renewed every two days) did not induce significant increase of mRNA
301 levels of those genes. As CAT, SOD and GST are the first enzymes induced after animal
302 exposure to xenobiotics, we can suggest that their mRNA levels probably increased before
303 the twelfth hour of exposure which would explain that we have not been able to detect their
304 increase. At the protein levels, data of the literature also appeared to be contradictory, hence
305 it is suggested that catalase activity was dependent on the animal tested, the nature of
306 chemicals and the intensity of exposure [60], which is confirmed in the present study. Hoarau
307 *et al.* [61] observed an inhibition of GST gene expression after marine mussel exposure to
308 BaP whereas mussels collected in a site highly contaminated with PAH exhibited an
309 induction of its expression.

310 The increase of P-gp expression observed in the present study is not surprising as
311 this protein is implicated in non specific excretion of xenobiotics, metabolites as well as
312 waste products derived from cell damage [62, 63]. The induction of P-gp by BaP was first
313 reported in Caco-2 cells [64, 65]. The inducibility of P-gp by BaP was also observed in blue
314 mussels at the protein level [65] and at the mRNA level [66]; in zebra mussel, we recently
315 shown an increase at the transcriptional level [34].

316

317 In conclusion, gill tissue represents an interesting model to investigate the expression of
318 genes encoding for stress related proteins. Cd and BaP induced a regulation at the
319 transcriptional level of genes implicated in either metabolism or detoxification such as AH-R,
320 HSP70, P-gp and MT. Cd seems to decrease the induction of AHR by BaP as already
321 described for fish and human hepatocytes. In Cd exposed mussels, the up-regulation of MT
322 mRNA is correlated with the restoration of the DNA SB basal level suggesting that MT is

323 implicated in cadmium detoxification. Future studies need to be done to clarify the interaction
324 of metals and PAHs on AHR gene expression.

325

326

327 **Conflict of interest:**

328 The authors declare that there are no conflicts of interest.

329

330 **Acknowledgements**

331 The authors thank the PIREN-Seine program and the ONEMA for financial support.

332

Accepted Manuscript

332 **Figure captions**

333

334 **Figure 1: Relative mRNA expression genes**

335 mRNA expression of genes in gills of zebra mussels exposed to Cd (10 µg/L), BaP (10µg/L)
336 and Cd+BaP (10 µg/L, 1 µg/L) for 12 hr, 24 hr, 3, 5 and 11 days. Determination of the
337 transcript abundances in individual sample was conducted with the comparative C_T method
338 ($\Delta\Delta C_T$) in consideration of a calibrator sample (control sample). Samples were normalized to
339 Ribosomal S3 gene due to its constant expression. Results (mean values \pm S.D.) were
340 compared among different groups using an ANOVA followed by a Tukey post hoc test. Three
341 levels were considered significant: $p < 0.05$ (*), $p < 0.01$ (**) and $p < 0.001$ (***).

342 **A:** metallothionein (MT), P-glycoprotein (P-gp), Aryl hydrocarbon receptor (AHR) and Heat
343 shock protein 70 (HSP70)

344 **B:** superoxide dismutase (SOD), Glutathion S transferase (GST) and catalase (CAT)

345

346

347

348

349 **Supporting information :**350 **Table 1: Tail DNA (%) in gill cell comets of *Dreissena polymorpha***

351 DNA strand breaks were measured with the comet assay and expressed as Tail DNA (%).
352 Control and various treatment groups of mussels were exposed to constant concentration of
353 dissolved CdCl₂ (10 µg/l), BaP (10 µg/l) and Cd+BaP (Cd=10 µg/L, BaP=1µg/l) for various
354 duration. Tail DNA % were reported as mean \pm standard error; * $p < 0.05$; ** $p < 0.01$;
355 *** $p < 0.001$.

356

356

357 **References**

358

359 [1] E. Guerlet, K. Ledy, A. Meyer, L. Giambérini, Towards a validation of a cellular biomarker
360 suite in native and transplanted zebra mussels: A 2-year integrative field study of seasonal
361 and pollution-induced variations, *Aquat. Toxicol.*, 81 (2007) 377-388.

362 [2] V. Marie, M. Baudrimont, A. Boudou, Cadmium and zinc bioaccumulation and
363 metallothionein response in two freshwater bivalves (*Corbicula fluminea* and *Dreissena*
364 *polymorpha*) transplanted along a polymetallic gradient, *Chemosphere*, 65 (2006) 609-617.

365 [3] J. Mersch, M.N. Beauvais, P. Nagel, Induction of micronuclei in haemocytes and gill cells
366 of zebra mussels, *Dreissena polymorpha*, exposed to clastogens, *Mutation Research -*
367 *Genetic Toxicology*, 371 (1996) 47-55.

368 [4] A. Binelli, C. Riva, D. Cogni, A. Provini, Genotoxic effects of p,p-DDT (1,1,1-trichloro-2,2-
369 bis-(chlorophenyl) ethane) and its metabolites in Zebra mussel (*D. polymorpha*) by SCGE
370 assay and micronucleus test, *Environ. Mol. Mutagen.*, 49 (2008) 406-415.

371 [5] M. Pavlica, G.I.V. Klobucar, N. Mojas, R. Erben, D. Papes, Detection of DNA damage in
372 haemocytes of zebra mussel using comet assay, *Mutation Research - Genetic Toxicology*
373 *and Environmental Mutagenesis*, 490 (2001) 209-214.

374 [6] A. Binelli, C. Riva, D. Cogni, A. Provini, Assessment of the genotoxic potential of
375 benzo(a)pyrene and pp-dichlorodiphenyldichloroethylene in Zebra mussel (*Dreissena*
376 *polymorpha*), *Mutation Research - Genetic Toxicology and Environmental Mutagenesis*, 649
377 (2008) 135-145.

378 [7] J. Le Goff, J. Gallois, L. Pelhuet, M.H. Devier, H. Budzinski, D. Pottier, V. Andre, J.
379 Cachot, DNA adduct measurements in zebra mussels, *Dreissena polymorpha*, Pallas -
380 Potential use for genotoxicant biomonitoring of fresh water ecosystems, *Aquat. Toxicol.*, 79
381 (2006) 55-64.

382 [8] Y. De Lafontaine, F. Gagné, C. Blaise, G. Costan, P. Gagnon, H.M. Chan, Biomarkers in
383 zebra mussels (*Dreissena polymorpha*) for the assessment and monitoring of water quality of
384 the St Lawrence River (Canada), *Aquat. Toxicol.*, 50 (2000) 51-71.

385 [9] A. Binelli, C. Riva, A. Provini, Biomarkers in Zebra mussel for monitoring and quality
386 assessment of Lake Maggiore (Italy), *Biomarkers*, 12 (2007) 349-368.

387 [10] A. Bourgeault, C. Gourlay-France, F. Vincent-Hubert, F. Palais, A. Geffard, S. Biagianti-
388 Risbourg, S. Pain-Devin, M.H. Tusseau-Vuillemin, Lessons from a transplantation of zebra
389 mussels into a small urban river: An integrated ecotoxicological assessment, *Environmental*
390 *Toxicology*, 25 (2010) 468-478.

391 [11] C. Michel, A. Bourgeault, C. Gourlay-France, F. Palais, A. Geffard, F. Vincent-Hubert,
392 Seasonal and PAH impact on DNA strand-break levels in gills of transplanted zebra mussels,
393 *Ecotoxicol. Environ. Saf.*, 92 (2013) 18-26.

394 [12] V. Contardo-Jara, S. Pflugmacher, G. Nützmann, W. Kloas, C. Wiegand, The beta-
395 receptor blocker metoprolol alters detoxification processes in the non-target organism
396 *Dreissena polymorpha*, *Environmental Pollution*, 158 (2010) 2059-2066.

397 [13] V. Contardo-Jara, C. Lorenz, S. Pflugmacher, G. Nützmann, W. Kloas, C. Wiegand,
398 Molecular effects and bioaccumulation of levonorgestrel in the non-target organism
399 *Dreissena polymorpha*, *Environmental Pollution*, 159 (2011) 38-44.

400 [14] A. Navarro, M. Faria, C. Barata, B. Pina, Transcriptional response of stress genes to
401 metal exposure in zebra mussel larvae and adults, *Environmental Pollution*, 159 (2011) 100-
402 107.

403 [15] A. Viarengo, M.N. Moore, G. Mancinelli, A. Mazzucotelli, R.K. Pipe, Significance of
404 metallothioneins and lysosomes in cadmium toxicity and homeostasis in the digestive gland
405 cells of mussels exposed to the metal in presence or absence of phenanthrene, *Marine*
406 *Environmental Research*, 17 (1985) 184-187.

407 [16] J.P. Bourdineaud, M. Baudrimont, P. Gonzalez, J.L. Moreau, Challenging the model for
408 induction of metallothionein gene expression, *Biochimie*, 88 (2006) 1787-1792.

- 409 [17] H.C. Schroder, R. Batel, S. Lauenroth, H.M.A. Hassanein, M. Lacorn, T. Simat, H.
410 Steinhart, W.E.G. Muller, Induction of DNA damage and expression of heat shock protein
411 HSP70 by polychlorinated biphenyls in the marine sponge *Suberites domuncula* Olivi,
412 *Journal of Experimental Marine Biology and Ecology*, 233 (1999) 285-300.
- 413 [18] I. Werner, D.E. Hinton, Field validation of hsp70 stress proteins as biomarkers in Asian
414 clam (*Potamocorbula amurensis*): is downregulation an indicator of stress?, *Biomarkers*, 4
415 (1999) 473-484.
- 416 [19] M.J. Bebianno, F. Geret, P. Hoarau, M.A. Serafim, M.R. Coelho, M. Gnassia-Barelli, M.
417 Romeo, *Biomarkers in Ruditapes decussatus: a potential bioindicator species*, *Biomarkers*, 9
418 (2004) 305-330.
- 419 [20] B. Ketterer, B. Coles, D.J. Meyer, The role of glutathione in detoxication *Environ. Health*
420 *Perspect.*, 49 (1983) 59-69.
- 421 [21] S. Pain, M. Parant, Response of multixenobiotic defence mechanism in *Dreissena*
422 *polymorpha* exposed to environmental stress, *Chemosphere*, 52 (2003) 1105-1113.
- 423 [22] S. Pain, M. Parant, Identification of multixenobiotic defence mechanism (MXR)
424 background activities in the freshwater bivalve *Dreissena polymorpha* as reference values for
425 its use as biomarker in contaminated ecosystems, *Chemosphere*, 67 (2007) 1258-1263.
- 426 [23] C. Minier, A. Abarnou, A. Jaouen-Madoulet, A.M. Le Guellec, R. Tutundjian, G.
427 Bocquene, F. Leboulenger, A pollution-monitoring pilot study involving contaminant and
428 biomarker measurements in the Seine Estuary, France, using zebra mussels (*Dreissena*
429 *polymorpha*), *Environmental Toxicology and Chemistry*, 25 (2006) 112-119.
- 430 [24] S. Lewis, R.D. Handy, B. Cordi, Z. Billinghamurst, M.H. Depledge, Stress proteins (HSP's):
431 Methods of detection and their use as an environmental biomarker, *Ecotoxicology*, 8 (1999)
432 351-368.
- 433 [25] F. Vincent-Hubert, Arini, A., Gourlay-Francé, C. , Early genotoxic effects in gill cells and
434 hemocytes of *dreissena polymorpha* exposed to cadmium, and a combination of B[a]P and
435 Cd Mutation Research - Genetic Toxicology and Environmental Mutagenesis, 723 (2011) 26-
436 35.
- 437 [26] M. Waisberg, P. Joseph, B. Hale, D. Beyersmann, Molecular and cellular mechanisms of
438 cadmium carcinogenesis, *Toxicology*, 192 (2003) 95-117.
- 439 [27] A. Hartmann, G. Speit, Effect of arsenic and cadmium on the persistence of mutagen-
440 induced DNA lesions in human cells, *Environ. Mol. Mutagen.*, 27 (1996) 98-104.
- 441 [28] S.E. Hook, R.F. Lee, Interactive effects of UV, benzo[a] pyrene, and cadmium on DNA
442 damage and repair in embryos of the grass shrimp *Palaemonetes pugio*, *Marine*
443 *Environmental Research*, 58 (2004) 735-739.
- 444 [29] J. Mersch, E. Morhain, C. Mouvet, Laboratory accumulation and depuration of copper
445 and cadmium in the freshwater mussel *Dreissena polymorpha* and the aquatic moss
446 *Rhynchosostegium riparioides*, *Chemosphere*, 27 (1993) 1475-1485.
- 447 [30] A. Bourgeault, C. Gourlay-France, C. Priadi, S. Ayrault, M.H. Tusseau-Vuillemin,
448 Bioavailability of particulate metal to zebra mussels: Biodynamic modelling shows that
449 assimilation efficiencies are site-specific, *Environmental Pollution*, 159 (2011) 3381-3389.
- 450 [31] I.V. Villela, I.M. de Oliveira, J. da Silva, J.A.P. Henriques, DNA damage and repair in
451 haemolymph cells of golden mussel (*Limnoperna fortunei*) exposed to environmental
452 contaminants, *Mutation Research - Genetic Toxicology and Environmental Mutagenesis*, 605
453 (2006) 78-86.
- 454 [32] C. Emmanouil, T.M.T. Sheehan, J.K. Chipman, Macromolecule oxidation and DNA
455 repair in mussel (*Mytilus edulis* L.) gill following exposure to Cd and Cr(VI), *Aquat. Toxicol.*,
456 82 (2007) 27-35.
- 457 [33] M.W. Pfaffl, A new mathematical model for relative quantification in real-time RT-PCR,
458 *Nucleic Acids Res.*, 29 (2001).
- 459 [34] A. Chatel, V. Faucet-Marquis, M. Perret, C. Gourlay-France, E. Uher, A. Pfohl-
460 Leszkowicz, F. Vincent-Hubert, Genotoxicity assessment and detoxification induction in
461 *Dreissena polymorpha* exposed to benzo a pyrene, *Mutagenesis*, 27 (2012) 703-711.

- 462 [35] C. Tessier, J.-S. Blais, Determination of Cadmium-Metallothioneins in Zebra Mussels
463 Exposed to Subchronic Concentrations of Cd²⁺, *Ecotoxicol. Environ. Saf.*, 33 (1996) 246-
464 252.
- 465 [36] D. Ivankovic, J. Pavicic, V. Beatovic, R.S. Klobucar, G.I.V. Klobucar, Inducibility of
466 Metallothionein Biosynthesis in the Whole Soft Tissue of Zebra Mussels *Dreissena*
467 *polymorpha* Exposed to Cadmium, Copper, and Pentachlorophenol, *Environmental*
468 *Toxicology*, 25 (2010) 198-211.
- 469 [37] S. Lecoecur, B. Videmann, P. Berny, Evaluation of metallothionein as a biomarker of
470 single and combined Cd/Cu exposure in *Dreissena polymorpha*, *Environ. Res.*, 94 (2004)
471 184-191.
- 472 [38] J. Engelken, A. Hildebrandt, cDNA cloning and cadmium-induced expression of
473 metallothionein mRNA in the zebra mussel *Dreissena polymorpha*, *Biochemistry and Cell*
474 *Biology-Biochimie Et Biologie Cellulaire*, 77 (1999) 237-241.
- 475 [39] V. Marie, P. Gonzalez, M. Baudrimont, J.-P. Bourdineaud, A. Boudou, Metallothionein
476 response to cadmium and zinc exposures compared in two freshwater bivalves, *Dreissena*
477 *polymorpha* and *Corbicula fluminea*, *Biometals*, 19 (2006) 399-407.
- 478 [40] S. Morin, T.T. Duong, A. Dabrin, A. Coynel, O. Herlory, M. Baudrimont, F. Delmas, G.
479 Durrieu, J. Schäfer, P. Winterton, G. Blanc, M. Coste, Long-term survey of heavy-metal
480 pollution, biofilm contamination and diatom community structure in the Riou Mort watershed,
481 South-West France, *Environmental Pollution*, 151 (2008) 532-542.
- 482 [41] S. Labbe, L. Larouche, D. Mailhot, C. Seguin, Purification of mouse MEP-1, a nuclear-
483 protein which binds to the metal regulatory elements of genes encoding metallothionein,
484 *Nucleic Acids Res.*, 21 (1993) 1549-1554.
- 485 [42] W. Qu, J. Pi, M.P. Waalkes, Metallothionein blocks oxidative DNA damage in vitro, *Arch.*
486 *Toxicol.*, 87 (2013) 311-321.
- 487 [43] M. Filipic, T.K. Hei, Mutagenicity of cadmium in mammalian cells: Implication of oxidative
488 DNA damage, *Mutation Research - Fundamental and Molecular Mechanisms of*
489 *Mutagenesis*, 546 (2004) 81-91.
- 490 [45] N. Chiaverini, M. De Ley, Protective effect of metallothionein on oxidative stress-induced
491 DNA damage, *Free Radic. Res.*, 44 (2010) 605-613.
- 492 [46] C.D. Klaassen, J. Liu, B.A. Diwan, Metallothionein protection of cadmium toxicity,
493 *Toxicol. Appl. Pharmacol.*, 238 (2009) 215-220.
- 494 [47] I. Boutet, A. Tanguy, D. Moraga, Response of the Pacific oyster *Crassostrea gigas* to
495 hydrocarbon contamination under experimental conditions, *Gene*, 329 (2004) 147-157.
- 496 [48] T. Haap, H.-R. Koehler, Cadmium tolerance in seven *Daphnia magna* clones is
497 associated with reduced hsp70 baseline levels and induction, *Aquatic Toxicology*, 94 (2009)
498 131-137.
- 499 [49] C. Singer, S. Zimmermann, B. Sures, Induction of heat shock proteins (hsp70) in the
500 zebra mussel (*Dreissena polymorpha*) following exposure to platinum group metals
501 (platinum, palladium and rhodium): Comparison with lead and cadmium exposures, *Aquat.*
502 *Toxicol.*, 75 (2005) 65-75.
- 503 [50] Z. Gong, J. Yang, M. Yang, F. Wang, Q. Wei, R.M. Tanguay, T. Wu, Benzo(a)pyrene
504 inhibits expression of inducible heat shock protein 70 in vascular endothelial cells, *Toxicology*
505 *Letters*, 166 (2006) 229-236.
- 506 [51] J. Yang, X. Liu, P. Niu, Y. Zou, Y. Duan, Correlations and co-localizations of Hsp70 with
507 XPA, XPG in human bronchial epithelia cells exposed to benzo a pyrene, *Toxicology*, 265
508 (2009) 10-14.
- 509 [52] F. Haq, M. Mahoney, J. Koropatnick, Signaling events for metallothionein induction,
510 *Mutation Research - Fundamental and Molecular Mechanisms of Mutagenesis*, 533 (2003)
511 211-226.
- 512 [53] C.I. Jones Iii, H. Zhu, S.F. Martin, Z. Han, Y. Li, B.R. Alevriadou, Regulation of
513 antioxidants and phase 2 enzymes by shear-induced reactive oxygen species in endothelial
514 cells, *Annals of Biomedical Engineering*, 35 (2007) 683-693.

- 515 [54] A. Piano, P. Valbonesi, E. Fabbri, Expression of cytoprotective proteins, heat shock
516 protein 70 and metallothioneins, in tissues of *Ostrea edulis* exposed to heat and heavy
517 metals, *Cell Stress Chaperones*, 9 (2004) 134-142.
- 518 [55] C. Urani, P. Melchiorretto, C. Canevali, G.F. Crosta, Cytotoxicity and induction of
519 protective mechanisms in HepG2 cells exposed to cadmium, *Toxicology in Vitro*, 19 (2005)
520 887-892.
- 521 [56] V. Tamasi, K. Monostory, R.A. Prough, A. Falus, Role of xenobiotic metabolism in
522 cancer: involvement of transcriptional and miRNA regulation of P450s, *Cellular and*
523 *molecular life sciences : CMLS*, 68 (2011) 1131-1146.
- 524 [57] O. Hankinson, THE ARYL-HYDROCARBON RECEPTOR COMPLEX, *Annu. Rev.*
525 *Pharmacol. Toxicol.*, 35 (1995) 307-340.
- 526 [58] C. Risso-de Faverney, M. Lafaurie, J.P. Girard, R. Rahmani, Effects of heavy metals and
527 3-methylcholanthrene on expression and induction of CYP1A1 and metallothionein levels in
528 trout (*Oncorhynchus mykiss*) hepatocyte cultures, *Environmental Toxicology and Chemistry*,
529 19 (2000) 2239-2248.
- 530 [59] D.D. Vakharia, N. Liu, R. Pause, M. Fasco, E. Bessette, Q.Y. Zhang, L.S. Kaminsky,
531 Effect of metals on polycyclic aromatic hydrocarbon induction of CYP1A1 and CYP1A2 in
532 human hepatocyte cultures, *Toxicol. Appl. Pharmacol.*, 170 (2001) 93-103.
- 533 [60] F. Regoli, G.W. Winston, S. Gorbi, G. Frenzilli, M. Nigro, I. Corsi, S. Focardi, Integrating
534 enzymatic responses to organic chemical exposure with total oxyradical absorbing capacity
535 and DNA damage in the European eel *Anguilla anguilla*, *Environmental Toxicology and*
536 *Chemistry*, 22 (2003) 2120-2129.
- 537 [61] P. Hoarau, G. Damiens, M. Romeo, M. Gnassia-Barelli, M.J. Bebianno, Cloning and
538 expression of a GST-pi gene in *Mytilus galloprovincialis*. Attempt to use the GST-pi transcript
539 as a biomarker of pollution, *Comparative Biochemistry and Physiology C-Toxicology &*
540 *Pharmacology*, 143 (2006) 196-203.
- 541 [62] S.M. Bard, Multixenobiotic resistance as a cellular defense mechanism in aquatic
542 organisms, *Aquatic Toxicology*, 48 (2000) 357-389.
- 543 [63] T. Smital, Uptake and efflux transport proteins as integral elements of the cellular
544 detoxification and environmental stress response in aquatic organisms, *Comparative*
545 *Biochemistry and Physiology a-Molecular & Integrative Physiology*, 163 (2012) S2-S2.
- 546 [64] N. Sugihara, K. Toyama, A. Michihara, K. Akasaki, H. Tsuji, K. Furuno, Effect of benzo a
547 pyrene on P-glycoprotein-mediated transport in Caco-2 cell monolayer, *Toxicology*, 223
548 (2006) 156-165.
- 549 [65] A. Prevodnik, K. Lija, T. Bollner, Benzo a pyrene up-regulates the expression of the
550 proliferating cell nuclear antigen (PCNA) and multixenobiotic resistance polyglycoprotein (P-
551 gp) in Baltic Sea blue mussels (*Mytilus edulis* L.), *Comparative Biochemistry and Physiology*
552 *C-Toxicology & Pharmacology*, 145 (2007) 265-274.
- 553 [66] L.J. Dallas, T.P. Bean, A. Turner, B.P. Lyons, A.N. Jha, Oxidative DNA damage may not
554 mediate Ni-induced genotoxicity in marine mussels: Assessment of genotoxic biomarkers
555 and transcriptional responses of key stress genes, *Mutation Research-Genetic Toxicology*
556 *and Environmental Mutagenesis*, 754 (2013) 22-31.
- 557
558
559

559

	Control	Cadmium	Cd+BaP	BaP
12 hr	15.42±0.8	24.23±1.22	13±0.61	15.56±0.85
Day 1	10.95±0.54	22.12±1.37***	16.92±0.72***	14.12±0.73**
Day 3	14.55±0.88	24.95±1.17***	19.13±0.99***	32.39±1.47***
Day 5	11.66±0.63	11.83±0.61	13.27±0.77	14.98±0.84
Day 11	12.84±0.68	14.95±0.88*	12.84±0.73	15.51±0.95

560

561

561 **Highlights**

562

563

564

- cadmium (Cd) and Benzo[a]pyrene (BaP) induced DNA damages in zebra mussels

565

- Cd is a strong inducer of Metallothionein (MT) and HSP70 genes

566

- Metallothionein might be implicated in cadmium detoxification

567

- BaP is a strong inducer of P-gP and AHR genes

568

- Cd and BaP induced in moderate changes in antioxidant enzymes mRNA

569

570

Accepted Manuscript



