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## Combined effects of exploitation and temperature on fish stocks in the Northeast Atlantic

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### Abstract :

Fish stock fluctuations are affected by two potentially confounding forces: the removal of individuals by fisheries and climatic variations affecting the productivity of fish populations. Disentangling the relative importance of these forces has thus been a question of primary importance for fisheries management and conservation. Through the analysis of long-term time-series for 27 fish stocks from the Northeast Atlantic, the present study shows that the sign and intensity of the effect of temperature on biomass are dependent on the geographical location: the stocks located at the southernmost and northernmost latitudes of our study displayed stronger associations with temperature than the stocks located in the middle range of latitudes. As a consequence, the investigation of the combined effects of exploitation and the environment revealed that the stocks at the northern/southern boundaries of the spatial extent of the species were more prone to combined effects. The interplay between geographic location, climate and exploitation thus plays a significant role in fish stock productivity, which is generally ignored during assessment, thus affecting management procedures.

**Keywords** : combined climate/exploitation effect, exploitation, fish stocks, Northeast Atlantic, response to climate, time-series

## 40 INTRODUCTION

41 In fish stock assessment, exploitation is often defined as the key driving force acting on population  
42 size. Climatic and other environmental factors are most often perceived as additional random white  
43 noise generating variations in biological processes, such as growth or recruitment. In marine ecology  
44 the general paradigm has been similar (i.e. additive processes), but somehow reversed, as  
45 environmental changes were shown to be crucial to understand the patterns of variations in exploited  
46 fish populations (Alheit and Bakun, 2010; Lehodey *et al.*, 2006). The variability of oceanographic  
47 variables is dominated by slow, long-term fluctuations (Halley, 1996; Rouyer *et al.*, 2010; Vasseur and  
48 Yodzis, 2004) and have thus been proposed as a potential force driving the long-term fluctuations in  
49 fish stocks (Lawton, 1988; Steele, 1984). Such environmental variations have been advocated to  
50 explain long-term changes in the overall level of fish populations (Alheit and Hagen, 1997; Toresen  
51 and Østvedt, 2000), while historical time series have revealed that fish populations display substantially  
52 large fluctuations even in the absence of intensive exploitation (Baumgartner *et al.*, 1992; Ravier and  
53 Fromentin, 2004). Environmental conditions impact recruitment strength during the early life stages of  
54 fishes, which affect the level of the stock as it flows through the age-classes of the population over its  
55 lifetime (Hjort, 1914, 1926). Abiotic variations also impact other key biological/ecological processes of  
56 wild fish populations, such as growth (Brander, 1995) or migratory behavior (Fromentin, 2009).  
57 Nonetheless, the response of fish stocks to environmental fluctuations depends on their life-history  
58 characteristics and on the time scales at which the environment fluctuates (Johst and Wissel, 1997;  
59 Nisbet and Gurney, 1993; Petchey, 2000). Furthermore, the sensitivity to the environment is not  
60 uniform across populations as it may depend upon their geographical location within the range of the  
61 species' repartition (Miller *et al.*, 1991). For instance, the intensity of the temperature effect on the  
62 recruitment of Atlantic cod (*Gadus morhua*) has been shown to be stronger for stocks located at the  
63 boundaries of the species geographical distribution, but also opposed between its northern and southern

64 edges (Planque and Fox, 1998; Planque and Frédou, 1999; Stige *et al.*, 2006).

65

66 Understanding the driving forces of population fluctuations has been the source of an intensive ongoing  
67 research effort that attracted a particular interest in the field of fisheries ecology for its implications in  
68 conservation and management in the context of global change. This question has remained much  
69 debated because of the complex nature of these fluctuations, originating from the interplay between  
70 deterministic and stochastic forces (Bjørnstad and Grenfell, 2001). Particularly, the effects of  
71 environmental changes and exploitation on fish populations have been seen until very recently as  
72 additive, but growing evidence shows that they have in fact important combined effects. Substantial  
73 changes in stock productivity associated to climate fluctuations can create the conditions for the over-  
74 exploitation of a given stock under an otherwise acceptable fishing effort, because the level of  
75 exploitation cannot be sustained anymore under the new environmental conditions (Perry *et al.*, 2010;  
76 Planque *et al.*, 2010). Here, we investigated a set of long-term total biomass time series for 27  
77 commercial fish stocks from the Northeast Atlantic, to test (i) whether combined effects of climate and  
78 exploitation could be detected in their long-term fluctuations and (ii) how those effects depend on the  
79 stocks geographical location.

80

81 MATERIALS AND METHODS

82 *Fisheries time series*

83 Recruitment, total biomass and fishing mortality time series were extracted from reports made available  
84 by the relevant working groups at the International Council for the exploration of the Sea (ICES,  
85 <http://www.ices.dk/indexfla.asp>). Among the 40 fish stocks comprised in the data set, we selected 27  
86 stocks (Table 1) for which long-term and age-structured data of abundance, weight, maturity and  
87 fishing mortality were available. Because the length of time series is a key parameter for the robustness  
88 of the analysis, only fish stocks displaying at least 30 years of continuous data were selected. The  
89 Baltic Sea stocks were voluntarily left out from the analyses as the environmental driver associated to  
90 the regime shift that drastically changed its composition is believed to be salinity rather than  
91 temperature (Möllmann *et al.*, 2008). However, the analyses proved robust to the inclusion of those  
92 stocks. For two stocks, the Northeast Arctic cod and the Norwegian Spring-spawning herring,  
93 exceptionally long data was available (94 and 100 years, respectively). For those, the ICES data have  
94 been extended back using the historical analyses of (Hysten, 2002) for the Northeast Arctic cod and  
95 (Toreisen and Østvedt, 2000) for the Norwegian Spring-spawning herring, further details are available  
96 in (Rouyer *et al.*, 2011). These stocks from 9 different species are located in contrasting geographical  
97 areas and display various life history traits and exploitation histories

98

99 Fishing mortality was used to quantify the effects of the variations in exploitation on total biomass,  
100 while temperature was used to assess the effect of changing environmental conditions on total biomass.  
101 Total biomass has the advantage to be more directly linked to temperature than the spawning stock  
102 biomass, for which a lag equal to the age at maturity has to be accounted for and that is plagued by two  
103 problems. First, age at maturity can substantially vary over time, which was observed for some stocks  
104 (e.g., Northeast Arctic cod). Second, introducing a lag implies removing as many data points as the

105 number of delays introduced, which reduces the length of the time series.

106 The relative latitudinal position of each stock within the species range was obtained by the following  
107 approach. The centre of the latitudinal distribution for each stock was estimated on maps of the stocks  
108 distribution. The positions were then rescaled so that a “0” indicates a stock at the southern end of the  
109 distribution and “1” a stock at the northern end, using data on fish species distribution extracted from  
110 FishBase (<http://www.fishbase.org>).

111

### 112 *Temperature time series*

113 Temperature time series were extracted over each stock distribution area from the extended  
114 reconstructed sea surface temperature (*NOAA\_ERSST\_V3*) data set provided by the *NOAA/OAR/ESRL*  
115 *PSD, Boulder, Colorado, USA*, from their web-site at <http://www.esrl.noaa.gov/psd/> and available from  
116 1880 to 2009. This data set is constructed using the International Comprehensive Ocean-Atmosphere  
117 Data Set (ICOADS) Sea Surface Temperature data and improved statistical methods. The extracted  
118 time series were averaged over space and months in order to obtain one yearly time series per stock  
119 matching the length of the fisheries data. For Arctic stocks, the temperature in the Barents Sea from  
120 1921 until 2006 from the Kola meridian transect (33°30' E, 70°30'-72°30' N) was used (34). The  
121 recruitment age in the age-structured data used to compute total biomass time series varied among  
122 stocks. Thus, the time series of temperature were lagged so that the first data point of the first age-class  
123 was aligned with the temperature the year before its birth.

124

### 125 *Extracting long-term fluctuations from temperature and fishing mortality time series*

126 Long-term fluctuations were defined in the present study as fluctuations occurring at time scales above  
127 5 years. This empirical limit was set to capture and focus on long-lasting periods of high/low values in  
128 fishing mortality/temperature, while being consistent with the length of the time series. The wavelet

129 approach was used to decompose time series and extract long-term fluctuations (Cazelles *et al.*, 2008;  
130 Lau and Weng, 1995; Torrence and Compo, 1998). The wavelet transform decomposes the variability  
131 of time series in both time and frequency and allows to cope with non-stationarity and transients. The  
132 reversibility property of the wavelet transform was used to extract the short-term fluctuations in the  
133 time series, defined as the fluctuations below 5 years. The long-term fluctuations in temperature and  
134 fishing mortality were then obtained by subtracting the filtered signal from the original time series and  
135 removing the residual variability by a low-pass filter.

136

### 137 *Testing associations*

138 On the contrary to exploitation, which is expected to be negatively associated with total biomass, the  
139 association between temperature and total biomass is likely to vary in sign and intensity across species  
140 and stock locations (Brunel and Boucher, 2007; Planque and Frédou, 1999). The association between  
141 temperature and total biomass/recruitment were thus investigated. As it is more robust to outliers and  
142 does not require Gaussian distributions, the non-parametric Spearman correlation coefficient was used  
143 to quantify the associations between the long-term fluctuations in temperature and in total  
144 biomass/recruitment. Dealing with autocorrelation for assessing the statistical significance of  
145 associations is a classical challenge in ecology (Pyper and Peterman, 1998). To cope with this, the  
146 Fourier surrogate technique was used (Rouyer *et al.*, 2008b; Royer and Fromentin, 2006; Schreiber and  
147 Schmitz, 2000). Through spectral synthesis, Fourier surrogates generate random time series that  
148 preserve the power spectrum of the original time series and therefore its autocorrelation structure.  
149 Surrogate time series thus display exactly the same fluctuations than the original series, but occurring at  
150 random times. This technique was then used to generate an empirical distribution for each Spearman  
151 correlation coefficient, allowing to test whether or not the correlation could be produced by chance  
152 while formally accounting for autocorrelation.

153

154

155 *Semi-quantitative approach*

156 Detecting the combined effects of two variables upon a third one can be done within the classical

157 generalized modelling framework (GLMs), which also provides ways to control for autocorrelation.

158 However for most of the stocks the number of events for which total biomass, temperature and fishing

159 mortality were all in a high/low period was relatively low, which was thus equivalent to an unbalanced

160 dataset. This therefore limited the ability of GLMs to quantify such events. In addition, the raw total

161 biomass time series displayed a high degree of autocorrelation and the inclusion of auto-regressive

162 terms to 'clean-up' the autocorrelation in the residuals was found inoperative. Including autoregressive

163 terms often lead to explain most of the trend in the time series, which can impair the detection of

164 external effects (Pyper and Peterman, 1998). To overcome these limitations a test was designed using a

165 semi-quantitative method, intuitive yet powerful, inspired from Brander (2005). For each time series of

166 raw total biomass and for each time series of long-term fluctuations in temperature and fishing

167 mortality, the points were ranked into categories of 'low' and 'high' levels. For each variable and each

168 time series, the years of 'low' and 'high' levels were defined as the values below the 33<sup>rd</sup> and above the

169 66<sup>th</sup> percentiles of the distribution, respectively. Attempts to use higher percentiles, such as the 20<sup>th</sup> and

170 the 80<sup>th</sup>, critically reduced the number of observations within the 'low' and 'high' levels. Since the sign

171 of the effect of temperature on recruitment varies across latitudes and stocks (Brunel and Boucher,

172 2007; Planque and Frédou, 1999), the ranking of temperature for stocks displaying a negative

173 correlation between recruitment and temperature was inverted so that 'low temperature' could be

174 interpreted as 'unfavourable conditions' and 'high temperature' could be interpreted as 'favourable

175 conditions' for all stocks (regardless of the significance of the correlation). The effect of exploitation,

176 temperature and the combined effect of both on total biomass were quantified by comparing the

177 number of years following the “expected behaviour” to the number of years following the “unexpected  
178 behaviour”. The “expected” and “unexpected” behaviours were defined as follows. For exploitation the  
179 expected behaviour corresponded to a year of low total biomass with a high fishing mortality, or to a  
180 year of high total biomass with a low fishing mortality. Conversely the unexpected behaviour  
181 corresponded to a year of low total biomass with a low fishing mortality, or to a year of high total  
182 biomass with a high fishing mortality. Similarly, the expected behaviour for temperature was a high  
183 total biomass with a favourable temperature or a low total biomass with an unfavourable temperature.  
184 The unexpected behaviour was thus defined as a low total biomass with a favourable temperature or a  
185 high biomass with an unfavourable temperature. Finally, the expected behaviour for the combined  
186 effect of temperature and fishing mortality corresponded to a year of low total biomass with a high  
187 fishing mortality and an unfavourable temperature, or to a year of high total biomass with a low fishing  
188 mortality and a favourable temperature. The unexpected behaviour was thus a year of low total biomass  
189 with a low fishing mortality and a favourable temperature or a year of high total biomass with a high  
190 fishing mortality and an unfavourable temperature. Note that this test did not directly address  
191 counteracting effects between temperature and exploitation on biomass, as such effects are more likely  
192 to range around the median of the biomass (i.e., not in the 'low'/'high' levels). For a given stock, the  
193 strength of the effect was quantified through the percentage of years conform to the expected  
194 behaviour. An effect was considered meaningful if at least 50% of the years displayed an expected  
195 behaviour. However, this proportion is largely affected by the fact that in autocorrelated time series two  
196 consecutive points will have a high probability to belong to the same category. To assess whether or not  
197 the proportions observed were higher than those obtained by chance, a test based on the surrogates  
198 approach was performed. For each stock, 1000 Fourier surrogates (see *Testing associations*) of the raw  
199 total biomass time series were produced. These allowed to compute the distribution of the unexpected  
200 and expected behaviours for each effect. For a given effect, the proportion of years displaying an

201 expected behaviour was considered significant when the observed proportion was higher than for 90%  
202 of the surrogates.

203 RESULTS

204 The association between recruitment and temperature was found similar in sign and intensity to the  
205 association between total biomass and temperature (Pearson correlation coefficient  $C=0.81$ ,  $p<0.0001$ ,  
206 Fig. 1a). Stocks for which a weak association between recruitment and temperature was found also  
207 displayed a weak association between biomass and temperature. However, significant associations with  
208 temperature were only found for a reduced number of stocks for both recruitment (11 stocks) and total  
209 biomass (6 stocks). The association between total biomass and temperature was found variable in sign  
210 and intensity over the species latitudinal range (Fig. 1b) as it was also the case for the association  
211 between recruitment and temperature (Fig. 1c), suggesting a difference in sensitivity to temperature  
212 across the range of the species distribution. Correlations were generally aligned along a latitudinal  
213 gradient, but this pattern often disappeared at the species level. The biomass/temperature and  
214 recruitment/temperature correlations were significantly associated to the mean latitude of the stock's  
215 geographical location (Pearson correlation coefficient  $C=0.44$ ,  $p=0.02$ , Fig. 1b and Pearson correlation  
216 coefficient  $C=0.68$ ,  $p<0.0001$ , Fig. 1c, respectively).

217

218 Figure 1 around here

219

220 They were generally negative for the stocks located at lower latitudes, whereas they were positive for  
221 the stocks located at higher latitudes. In the middle range of latitudes the correlations were weaker and  
222 the few significant correlations between biomass and temperature were only found in the northernmost  
223 and southernmost areas. Comparable fluctuations were also noticed between stocks with different life  
224 history traits and exploitation histories, but located in the northernmost and southernmost areas of the  
225 study. This was particularly clear for the North-east Arctic cod, the Arctic haddock and the Norwegian  
226 Spring Spawning herring that displayed common periods of low and high biomass consistent with the

227 long-term fluctuations in temperature at the Kola section (Fig. 2a). Likewise, the more southern Irish  
228 Sea cod and sole stocks both displayed common long-term fluctuations consistent with their negative  
229 relationship to temperature (Fig. 2b).

230

231 Figure 2 around here

232

233 The long-term fluctuations in fishing mortality also often displayed patterns opposite to those of  
234 biomass, such as for the Northeast Arctic cod, the Irish Sea cod, the Icelandic haddock, the Northern  
235 hake, the Arctic Saithe and the Celtic Sea sole (Fig. 3). This illustrated the difficulty to disentangle the  
236 relative effects of temperature and fishing mortality on total biomass, as both variables seemed to have  
237 substantial effects on the stocks.

238

239 Figure 3 around here

240

241 The effect of temperature and exploitation on biomass were then investigated by analysing co-  
242 occurrences of periods of low/high biomass with periods of high/low levels in the long-term  
243 fluctuations of temperature and exploitation. The long-term fluctuations of temperature and fishing  
244 mortality were obtained by filtering the time series above a 5 years threshold (Table 1). The raw time  
245 series of both fishing mortality and temperature were dominated by long-term fluctuations, which  
246 explained between 51.4% and 93.7% of the total variance for fishing mortality (77% on average) and  
247 between 52.2% and 86.7% of the total variance for temperature (73.7% on average, Table 1).

248

249 Table 1 around here

250

251 The effects of temperature and exploitation were difficult to rank as when pooling all stocks, the  
252 percentage of expected events for these factors was 67% and 68%, respectively. Furthermore,  
253 significant temperature effects were found for 8 stocks, whereas 9 stocks displayed a significant effect  
254 of exploitation (Fig. 4). In contrast, although the pool of points with a combined effect was smaller in  
255 comparison, the percentage of expected events was found higher (79%) than for exploitation or  
256 temperature alone. Among 27 stocks, 8 were found to display significant combined effects between  
257 temperature and exploitation (Fig.4). For 5 stocks the proportion of expected combined effects was  
258 found to be 100%, whereas the effect was not significant (Faroe cod, West Scotland cod, Icelandic  
259 haddock, North Sea saithe and Celtic sole, Fig. 4). This was due to the small amount of points with a  
260 combined effect (below 6) that was easily reproduced in the test by random time series displaying  
261 similar fluctuations (i.e. surrogates). As expected, significant combined effects were found for stocks  
262 that displayed a significant effect of exploitation and/or temperature. Therefore, the geographical  
263 location of stocks appeared important as stocks in the southernmost and northernmost latitudes, which  
264 displayed higher correlations with temperature, were thus more prone to undergo stronger combined  
265 effects.

266

267

Figure 4 around here

268

269 A couple of examples from the northernmost latitudes provided a clear illustration of periods with  
270 combined effects, albeit with slight differences (Fig. 5). For instance, the North-east Arctic cod has  
271 been fully to over-exploited since the late 1950s, but the total biomass remained at relatively high  
272 levels and mostly dropped during the mid-1970s and mid-1980s, when the temperature decreased and  
273 possibly drove the recruitment to low levels (Fig. 5a). The Norwegian Spring spawning herring case is  
274 even more striking, as the crash of this population in the 1970s coincided with a large peak in fishing

275 mortality that followed a prolonged period of unfavourable temperature associated to consecutive years  
276 of poor recruitment. The later rebuilding of this population corresponded to favourable temperature  
277 while fishing mortality remained at average levels (Fig. 5b).

278

279 DISCUSSION

280 Long-term effects of climate on recruitment have already been documented for a large number of  
281 commercial stocks from the North-east Atlantic (Beaugrand *et al.*, 2003; Brunel and Boucher, 2007)  
282 and are likely to be associated to substantial changes in productivity (Alheit and Hagen, 1997; Toresen  
283 and Østvedt, 2000; Vert-pre *et al.*, 2013). Here we showed that the fluctuations in temperature were  
284 associated to those in recruitment and total biomass with a comparable sign and intensity (Fig. 1a) and  
285 that those associations varied with the latitudinal location of the stocks (Fig. 1b,c). Such a spatial  
286 pattern has already been documented for the recruitment of Atlantic cod stocks (Planque and Fox,  
287 1998; Planque and Frédou, 1999; Stige *et al.*, 2006) and has been generalized to a broad range of  
288 species (Brunel and Boucher, 2006). In accordance with the species range hypothesis (Miller *et al.*,  
289 1991), these studies showed that the recruitment of stocks located at the limits of the species'  
290 geographical distribution were more prone to be affected by environmental variability than stocks in  
291 the middle range, while the sign of this effect was opposed between the northernmost and southernmost  
292 stocks. Even though the present data set did not cover the total range of the species repartition,  
293 consistently with these studies the results showed that such a pattern could also be detected in biomass  
294 fluctuations (Fig. 1b). The importance of the geographical location was further stressed by the common  
295 features in the long-term fluctuations of stocks with different life-history traits and exploitation history  
296 but located close to each other (Fig. 2). This appeared clearly for the Irish Sea cod and sole stocks, but  
297 also for three stocks from the Barents Sea area for which the environmental link has already been put  
298 forward to explain the relative synchrony of their recruitment dynamics (Ottersen and Loeng, 2000).  
299 With the current climate change, such a geographical effect could have great consequences. The raise  
300 of the sea surface temperature over the North-east Atlantic during the past 30 years could have  
301 contrasted effects on the productivity of northern and southern fish stock. Averaging the standardized  
302 biomass time series of the northernmost and southernmost stocks illustrated this (Fig. 6). Whereas the

303 trend in biomass of the northernmost stocks indicates a slight increase during this period, it indicates a  
304 clear decrease for the southernmost ones (Fig. 6). Indeed, such a general figure has to be interpreted  
305 carefully because it does not account for the variability in management efficiency across stocks and  
306 areas and also because our dataset does not cover the total species' distribution.

307

308  Figure 6 around here

309

310 In our database, even though the longest available time series were used, most of them began long after  
311 the development of the fisheries and therefore integrated long periods of overfishing. The percentage of  
312 points above the fishing mortality of the precautionary approach was on average equal to 61% when  
313 pooling the stocks together and ranged from 8% to 100%. Furthermore, 70% of the stocks displayed  
314 more than 50% of years above the fishing mortality of the precautionary approach, illustrating the  
315 general state of over-exploitation of the North-east Atlantic stocks during the studied period. The lack  
316 of contrast in the state of exploitation of the stocks, due to the relative shortness of the time series, was  
317 thus a limit that impaired the detection of combined effects. The procedure detected synchronous  
318 events based on 'high' and 'low' levels, which were defined by the distribution of the time series and not  
319 by an objective baseline. Therefore, in the case of stocks for which relatively short time series were  
320 available but that have been overexploited for a long time, 'high' biomass may in fact represent medium  
321 to low levels of biomass relatively to virgin biomass, which is difficult to estimate precisely. It was  
322 therefore not surprising that the clearest examples were found for the stocks for which the longest data  
323 were available (Fig. 5). This underlines the importance of past levels to define baselines and assess the  
324 current state of stocks. Finally, it has to be noted that our results were based upon the analysis of SST in  
325 the North-East Atlantic, but that other environmental variables in other areas are known to be important  
326 drivers of fish stock productivity (Möllmann *et al.*, 2008).

327

328 Little doubt can be cast over the crucial contribution of exploitation on the decline of commercial fish  
329 stocks from the North Atlantic (Christensen *et al.*, 2003) and our results documented clear effects of  
330 fishing mortality on the variations of total biomass of fish stocks. However, our results also show that  
331 substantial changes in fish stock productivity induced by long-term changes in the environment have  
332 probably amplified or dampened the impacts of exploitation, increasing the risk of depletion during  
333 periods of unfavourable environmental conditions, or stabilizing the biomass during periods of  
334 favourable conditions. The combined effects seemed overall stock-specific as stocks from common  
335 areas seemed to have been affected differently. For instance, the decline of the North-east Arctic cod  
336 seemed to be primarily due to a long period of overfishing, enhanced by a period of unfavourable  
337 temperatures inducing poor recruitment that drove the whole stock to its lowest level. The collapse of  
338 the Norwegian Spring Spawning herring stock in the 1970s seemed to correspond to a prolonged period  
339 of unfavourable temperatures that have probably translated into consecutive years of poor recruitment  
340 during a period of moderate exploitation. Then, the dramatic and sharp increase in fishing mortality in  
341 the early 1960s seems to have been enhanced by the constant exploitation level while the stock was  
342 declining. These two examples suggest that fishing could have amplified the effects of temperature in  
343 the case of the Norwegian Spring herring, whereas temperature could have amplified the effect of  
344 fishing in the case of the Northeast Arctic cod. Interplays between geographical location, environment  
345 and exploitation have already been noted for tuna and tuna-like species (Rouyer *et al.*, 2008a) and  
346 question the management of stocks using stable reference points. As suggested by the Norwegian  
347 Spring herring case, maintaining an otherwise acceptable level of exploitation during prolonged periods  
348 of unfavourable conditions could induce dramatic stock declines.

349

350 Our results highlight the spatial and temporal scales as key parameters to understand fish stocks

351 dynamics and emphasize that combined effects of climate and exploitation appear to have substantial  
352 effects on the long-term fluctuations of exploited fish stocks. This stresses the importance of studying  
353 the dynamics of exploited fish populations within the context of global warming, because over-  
354 exploitation has been shown to be related to poor governance, the *common-pool* nature of these  
355 resources and the globalization of sea food markets (Garcia and Grainger, 2005; Hilborn *et al.*, 2005;  
356 Sethi *et al.*, 2010).

357

358 ACKNOWLEDGEMENTS

359 Funding was provided by the Norwegian Research Council through the ARCWARM Project

360 (#178239). M.H. received support from Marie Curie Intra-European fellowship (IEF, FP7-PEOPLE-

361 IEF-2008, European Commission; project No 236549).

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- Alheit, J., and Bakun, A. 2010. Population synchronies within and between ocean basins: Apparent teleconnections and implications as to physical–biological linkage mechanisms. *Journal of Marine Systems*, 79: 267–285.
- Alheit, J., and Hagen, E. 1997. Long-term climate forcing of European herring and sardine populations. *Fisheries Oceanography*, 6: 130–139.
- Baumgartner, T., Soutar, A., and Ferreira-Bartrina, V. 1992. Reconstruction of the history of pacific sardine and northern anchovy populations over the past two millenia from sediments of the Santa-Barbara basin, California. *Calif. Coop. Oceanic Fish. Invest. Rep.*, 33: 24–40.
- Beaugrand, G., Brander, K. M., Lindley, J. A., Souissi, S., and Reid, P. C. 2003. Plankton effect on cod recruitment in the North Sea. *Nature*, 426: 661–664.
- Bjørnstad, O. N., and Grenfell, B. T. 2001. Noisy Clockwork: Time Series Analysis of Population Fluctuations in Animals. *Science*, 293: 638–643.
- Brander, K. M. 1995. The effects of temperature on growth of Atlantic cod (*Gadus morhua* L.). *ICES Journal of Marine Science*, 52: 1–10.
- Brander, K. M. 2005. Cod recruitment is strongly affected by climate when stock biomass is low. *ICES Journal of Marine Science*, 62: 339–343.
- Brunel, T., and Boucher, J. 2006. Pattern of recruitment variability in the geographical range of the exploited northeast Atlantic fish species. *Journal of Sea Research*, 55: 156–168.
- Brunel, T., and Boucher, J. 2007. Long-term trends in fish recruitment in the north-east Atlantic related to climate change. *Fisheries Oceanography*, 16: 336–349.
- Cazelles, B., Chavez, M., Berteaux, D., Ménard, F., Vik, J. O., Jenouvrier, S., and Stenseth, N. C. 2008. Wavelet analysis of ecological time series. *Oecologia*, 156: 287–304.
- Christensen, V., Guénette, S., Heymans, J. J., Walters, C. J., Watson, R., Zeller, D., and Pauly, D. 2003. Hundred-year decline of North Atlantic predatory fishes. *Fish and Fisheries*, 4: 1–24.
- Fromentin, J.-M. 2009. Lessons from the past: investigating historical data from bluefin tuna fisheries. *Fish and Fisheries*, 10: 197–216.
- Garcia, S. G., and Grainger, J. R. 2005. Gloom and doom? The future of marine capture fisheries. *Phil. Trans. R. Soc. Lond. B.*, 360: 21–46.
- Halley, J. M. 1996. Ecology, evolution and  $1/f$  noise. *Trends Ecol. Evol.*, 11: 33–38.
- Hilborn, R., Orensanz, J. M., and Parma, A. M. 2005. Institutions, incentives and the future of fisheries. *Phil. Trans. R. Soc. B.*, 360: 47–57.
- Hjort, J. 1914. Fluctuations in the great fisheries of northern Europe. *Rapp. P.-V. Cons. Int. Explor.*

- Mer, 20: 1–227.
- Hjort, J. 1926. Fluctuations in the year classes of important food fishes. *J. Cons. Int. Explor. Mer*, 1: 5–38.
- Hysten, A. 2002. Fluctuations in abundance of Northeast Arctic cod during the 20th century. *ICES Marine Science Symposia*, 215: 543–550.
- Johst, K., and Wissel, C. 1997. Extinction Risk in a Temporally Correlated Fluctuating Environment. *Theoretical Population Biology*, 52: 91–100.
- Lau, K. M., and Weng, H. 1995. Climate Signal Detection Using Wavelet Transform: How to Make a Time Series Sing. *B. Am. Meteorol. Soc.*, 76: 2391–2402.
- Lawton, J. H. 1988. More time means more variation. *Nature*, 334: 563.
- Lehodey, P., Alheit, J., Barange, M., Baumgartner, T., Beaugrand, G., Drinkwater, K., Fromentin, J.-M., *et al.* 2006. Climate Variability, Fish, and Fisheries. *Journal of Climate*, 19: 5009–5030.
- Miller, J. M., Burke, J. S., and Fitzhugh, G. R. 1991. Early life history patterns of Atlantic North American flatfish: Likely (and unlikely) factors controlling recruitment. *Netherlands Journal of Sea Research*, 27: 261–275.
- Möllmann, C., Müller-Karulis, B., Kornilovs, G., and John, M. A. S. 2008. Effects of climate and overfishing on zooplankton dynamics and ecosystem structure: regime shifts, trophic cascade, and feedback loops in a simple ecosystem. *ICES Journal of Marine Science: Journal du Conseil*, 65: 302–310.
- Nisbet, R. M., and Gurney, W. S. 1993. *Modelling Fluctuating Populations*. Books on Demand. 393 pp.
- Ottersen, G., and Loeng, H. 2000. Covariability in early growth and year-class strength of Barents Sea cod, haddock, and herring: the environmental link. *ICES Journal of Marine Science: Journal du Conseil*, 57: 339–348.
- Perry, R. I., Cury, P., Brander, K., Jennings, S., Möllmann, C., and Planque, B. 2010. Sensitivity of marine systems to climate and fishing: Concepts, issues and management responses. *Journal of Marine Systems*, 79: 427–435.
- Petchey, O. L. 2000. Environmental colour affects aspects of single-species population dynamics. *Proc. R. Soc. Lond., B*, 267: 747–754.
- Planque, B., and Fox, C. 1998. Interannual variability in temperature and the recruitment of Irish Sea cod. *Marine Ecology Progress Series*, 172: 101–105.
- Planque, B., and Frédou, T. 1999. Temperature and the recruitment of Atlantic cod (*Gadus morhua*). *Can. J. Fish. Aquat. Sci.*, 56: 2069–2077.

- Planque, B., Fromentin, J.-M., Cury, P., Drinkwater, K. F., Jennings, S., Perry, R. I., and Kifani, S. 2010. How does fishing alter marine populations and ecosystems sensitivity to climate? *Journal of Marine Systems*, 79: 403–417.
- Pyper, B., and Peterman, B. J. 1998. Comparison of methods to account for autocorrelation in correlation analyses of fish data. *Can. J. Fish. Aquat. Sci.*, 55: 2127–2140.
- Ravier, C., and Fromentin, J. M. 2004. Are the long-term fluctuations in Atlantic bluefin tuna (*Thunnus thynnus*) population related to environmental changes ? *Fish. Oceanogr.*, 13: 145–160.
- Rouyer, T., Fromentin, J.-M., Menard, F., Cazelles, B., Briand, K., Pianet, R., Planque, B., *et al.* 2008a. Complex interplays among population dynamics, environmental forcing, and exploitation in fisheries. *Proc. Nat. Ac. Sc.*, 105: 5420–5425.
- Rouyer, T., Fromentin, J.-M., and Stenseth, N. C. 2010. Environmental noise affects the fluctuations of Atlantic large pelagics. *Progress in Oceanography*, 86: 267–275.
- Rouyer, T., Fromentin, J.-M., Stenseth, N. C., and Cazelles, B. 2008b. Analysing multiple time series and extending significance testing in wavelet analysis. *Marine Ecology Progress Series.*, 359: 11–23.
- Rouyer, T., Ottersen, G., Durant, J., Hidalgo, M., Hjermann, D., Persson, J., Stige, L., *et al.* 2011. Shifting dynamic forces in fish stock fluctuations triggered by age truncation ? *Global Change Biology*, 17: 3046–3057.
- Royer, F., and Fromentin, J. M. 2006. Recurrent and density-dependent patterns in long-term fluctuations of Atlantic bluefin tuna trap catches. *Mar. Ecol. Prog. Ser.*, 319: 237–249.
- Schreiber, T., and Schmitz, A. 2000. Surrogate time series. *Physica D*, 142: 346–382.
- Sethi, S. A., Branch, T. A., and Watson, R. 2010. Global fishery development patterns are driven by profit but not trophic level. *Proceedings of the National Academy of Sciences of the United States of America.*, 107: 12163–12167.
- Steele, J. H. 1984. Modeling Long-Term Fluctuations in Fish Stocks. *Science*, 224: 985–987.
- Stige, L., Ottersen, G., Brander, K., Chan, K., and Stenseth, N. 2006. Cod and climate: effect of the North Atlantic Oscillation on recruitment in the North Atlantic. *Marine Ecology Progress Series*, 325: 227–241.
- Toresen, R., and Østvedt, O. j. 2000. Variation in abundance of Norwegian spring-spawning herring (*Clupea harengus*, Clupeidae) throughout the 20th century and the influence of climatic fluctuations. *Fish and Fisheries*, 1: 231–256.
- Torrence, C., and Compo, G. P. 1998. A practical Guide to Wavelet Analysis. *B. Am. Meteorol. Soc.*, 79: 61–78.

Vasseur, D. A., and Yodzis, P. 2004. The color of environmental noise. *Ecology*, 85: 1146–1152.

Vert-pre, K. A., Amoroso, R. O., Jensen, O. P., and Hilborn, R. 2013. Frequency and intensity of productivity regime shifts in marine fish stocks. *Proceedings of the National Academy of Sciences*. <http://www.pnas.org/content/early/2013/01/15/1214879110> (Accessed 17 January 2013).

364 FIGURE LEGENDS

365

366 Figure 1: Recruitment/temperature correlations  $r_{(Rec,Temp)}$  against the biomass/temperature correlations  
367  $r_{(TotBiom,Temp)}$  (a). Black dots indicate statistical significance of  $r_{(Rec,Temp)}$  at the 10% level. Average latitude  
368 of the stock against  $r_{(TotBiom,Temp)}$  (b). Black dots indicate statistical significance of  $r_{(TotBiom,Temp)}$  at the 10%  
369 level. Average latitude of the stock against  $r_{(Rec,Temp)}$  (c). Black dots indicate statistical significance of  
370  $r_{(TotBiom,Temp)}$  at the 10% level. All correlations were Spearman correlation coefficients computed on the  
371 raw time series and their significance was assessed using the surrogates method (see Materials and  
372 Methods).

373

374 Figure 2: Common features in the fluctuations exhibited by the raw total biomass time series and the  
375 filtered temperature series for a) the North-east Arctic cod, the Arctic haddock and the Norwegian  
376 Spring-Spawning herring stocks and for b) the Irish Sea cod and sole stocks.

377

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379

380 Figure 3: Opposite fluctuations in the raw total biomass time series (black line and black circles) and  
381 the filtered fishing mortality time series (solid line) for the North-east Arctic cod (a), the Irish Sea cod  
382 (b), the Icelandic haddock (c), the Northern hake (d), the Norwegian Spring-Spawning herring (d) and  
383 the Arctic saithe (e) stocks.

384

385

386 Figure 4: Histograms of the proportion of years displaying an expected behaviour for the exploitation  
387 (Exp.), temperature (temp.) and combined (comb.) effects. The numbers above the bars indicate the

388 proportion of years displaying an expected behaviour and a black bar indicates that this proportion is  
389 significantly higher than found by chance. For each stock this test was realised by computing the  
390 proportion of years displaying an expected behaviour for 1000 surrogates time series; a proportion was  
391 considered significant when the observed proportion was found higher than for 90% of the surrogates.

392

393

394 Figure 5: Combined effects for a) the North-east Arctic cod (top three panels), b) the Norwegian Spring  
395 Spawning herring (bottom three panels). The left panels represent the effect of exploitation, the plain  
396 line represents fishing mortality and the grey broken line represents the total biomass. The middle  
397 panels depict the effect of temperature, the plain line represents temperature and the grey broken line  
398 represents the total biomass. The right panels depict combined effects and the plain line represents the  
399 total biomass. Horizontal lines depict the limits between the levels of fishing mortality, temperature and  
400 total biomass for the left, middle and right panels, respectively. Years complying with expected cases  
401 for fishing mortality, temperature and combined effects are identified with white squares, whereas  
402 black diamonds identify unexpected cases.

403

404 Figure 6: Standardized time series of biomass for the northernmost and southernmost stocks (plain line  
405 white circles and plain line with black squares, respectively) and time series of the average temperature  
406 over the North-east Atlantic (grey line). The northernmost and southernmost stocks were selected as the  
407 stocks above the 66<sup>th</sup> percentile and below the 33<sup>rd</sup> percentile of the distribution of the mean latitude of  
408 the stocks, respectively. The time series were computed as the average of the standardized time series  
409 of total biomass over the period 1979-2006 and re-standardized. The North-east Atlantic temperature  
410 was obtained by averaging over space and years time series extracted between 30°W and 30°E, and  
411 45°N and 80°N. The resulting time series was standardized over 1900-2006.



## 413 TABLES

414 Table 1: Available period of data for the studied stocks, percentage of variance explained by the long-  
 415 term fluctuations (above 5 years) in the fishing mortality and temperature time series.

Species	Stock	Data period	Variance (%) fishing mortality	Variance (%) temperature
cod	Northeast_Arctic	1913-2006	93.7	52.2
cod	Faroe	1961-2008	61	86.2
cod	Iceland	1955-2009	83.1	81.4
cod	North_Sea	1963-2008	72.5	75.9
cod	West_scotland	1978-2007	51.4	85.7
cod	Irish_Sea	1968-2007	79.4	70.1
cod	Celtic_Sea	1971-2007	74.1	68.3
haddock	Northeast_Arctic	1950-2008	73.7	58.9
haddock	Faroe	1957-2008	86	86.7
haddock	Iceland	1979-2008	74.4	75.9
haddock	North_Sea	1963-2008	82.2	75.9
haddock	West_scotland_a	1978-2007	51.7	84.5
hake	northern	1978-2008	79.9	78.7
herring	North_Sea	1960-2008	82.4	76.2
herring	Northeast_Arctic	1907-2006	83.1	52.2
herring	West_scotland_a	1957-2008	78.7	85.3
plaice	North_Sea	1957-2008	79.8	74.4
plaice	Irish_Sea	1964-2007	83.1	70
plaice	Channel_West	1976-2007	66.4	70.5
saithe	Arctic	1960-2008	93.3	64.5
saithe	Faroe	1961-2008	79.5	86.2
saithe	Iceland	1974-2008	79.4	75.6
saithe	North_Sea	1967-2008	89.9	78.1
sole	North_Sea	1957-2008	87.8	74.4
sole	Irish_Sea	1970-2006	58.9	66.8
sole	Channel_West	1969-2007	82.3	68.3
sole	Celtic_Sea	1971-2007	83.7	67.1

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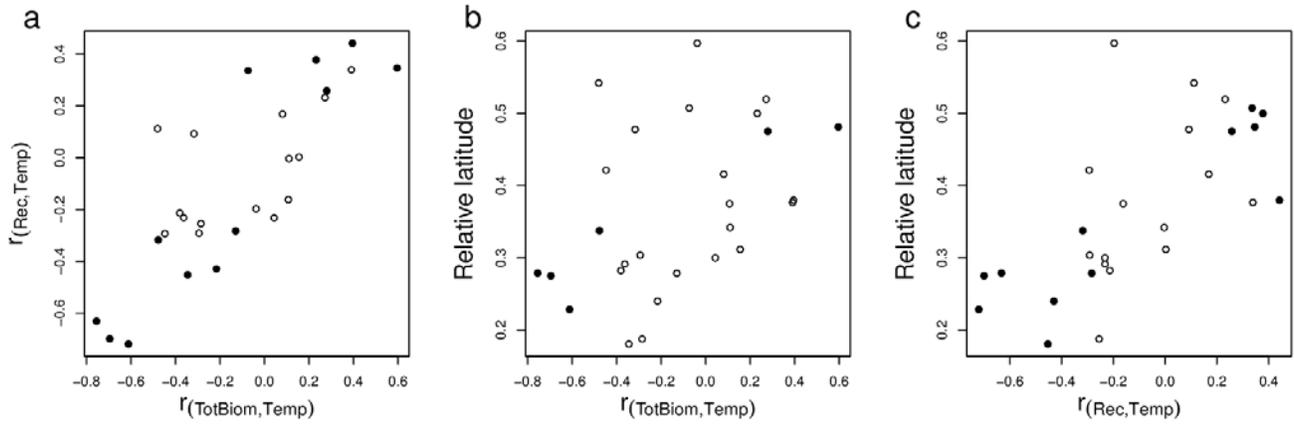
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420 FIGURES

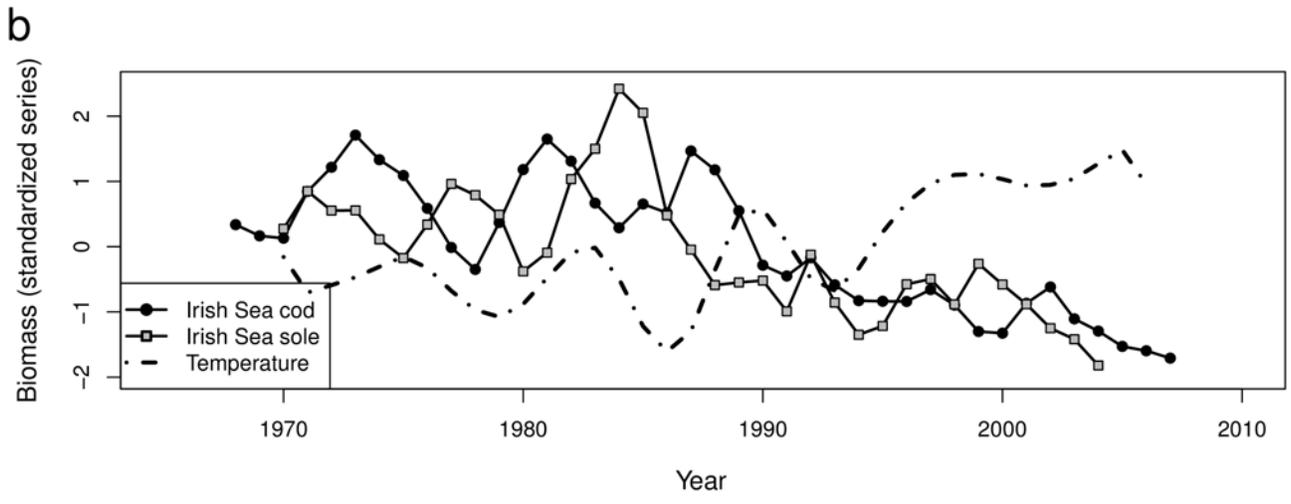
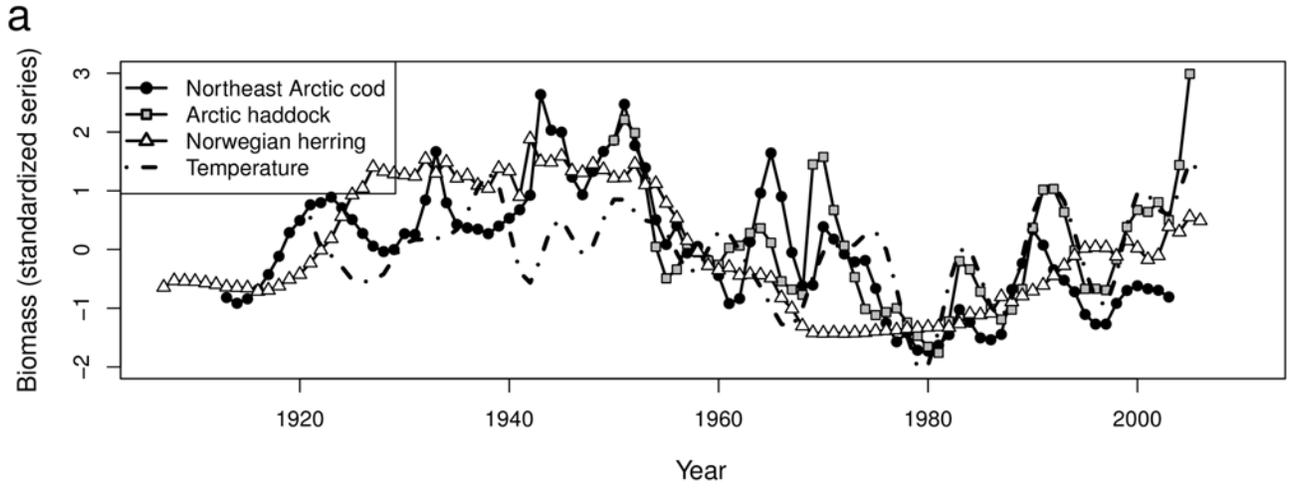
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422 FIGURE 1

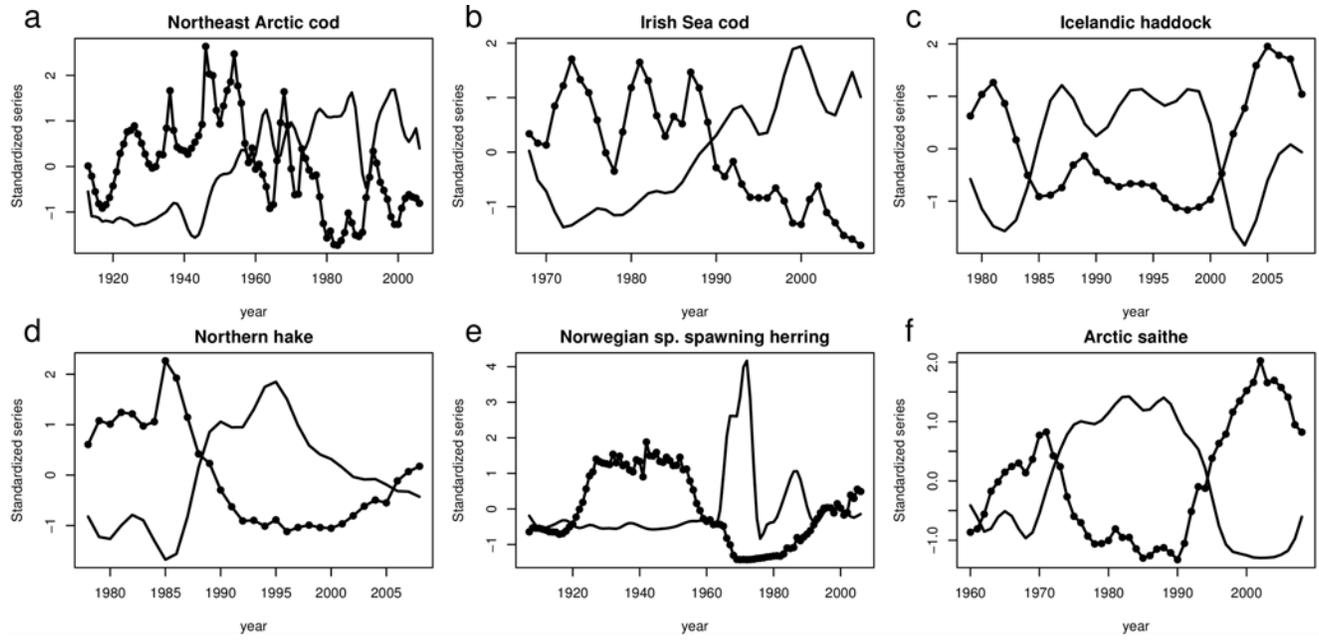


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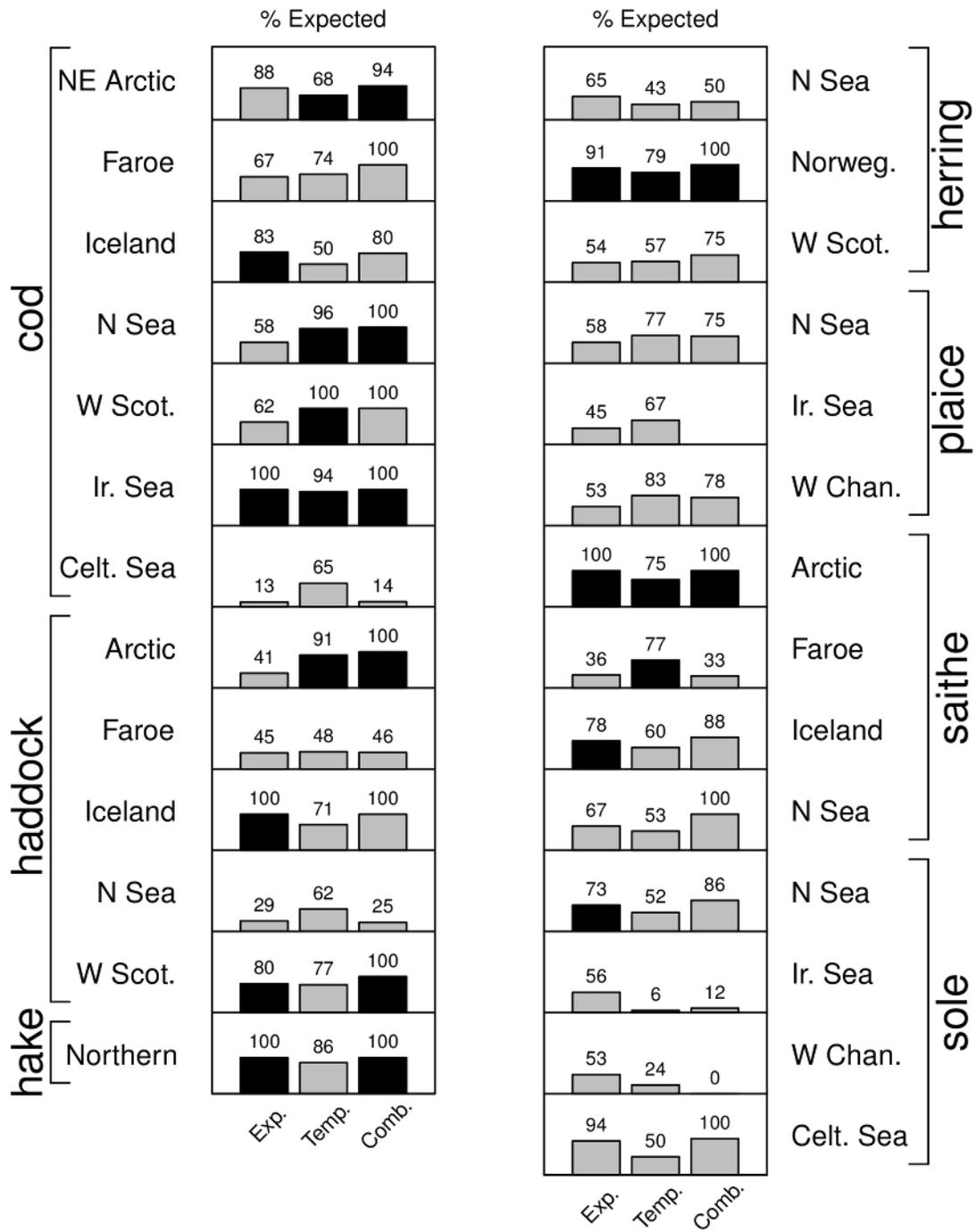
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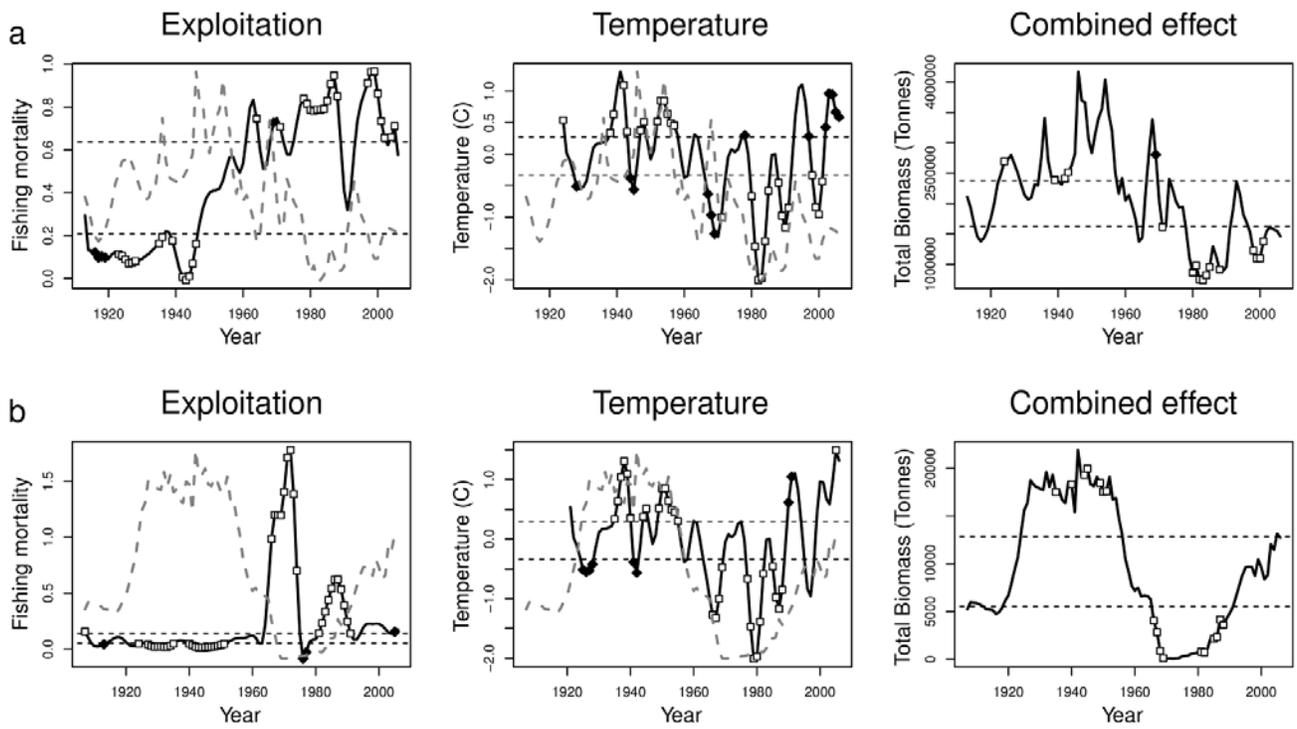
429 FIGURE 3



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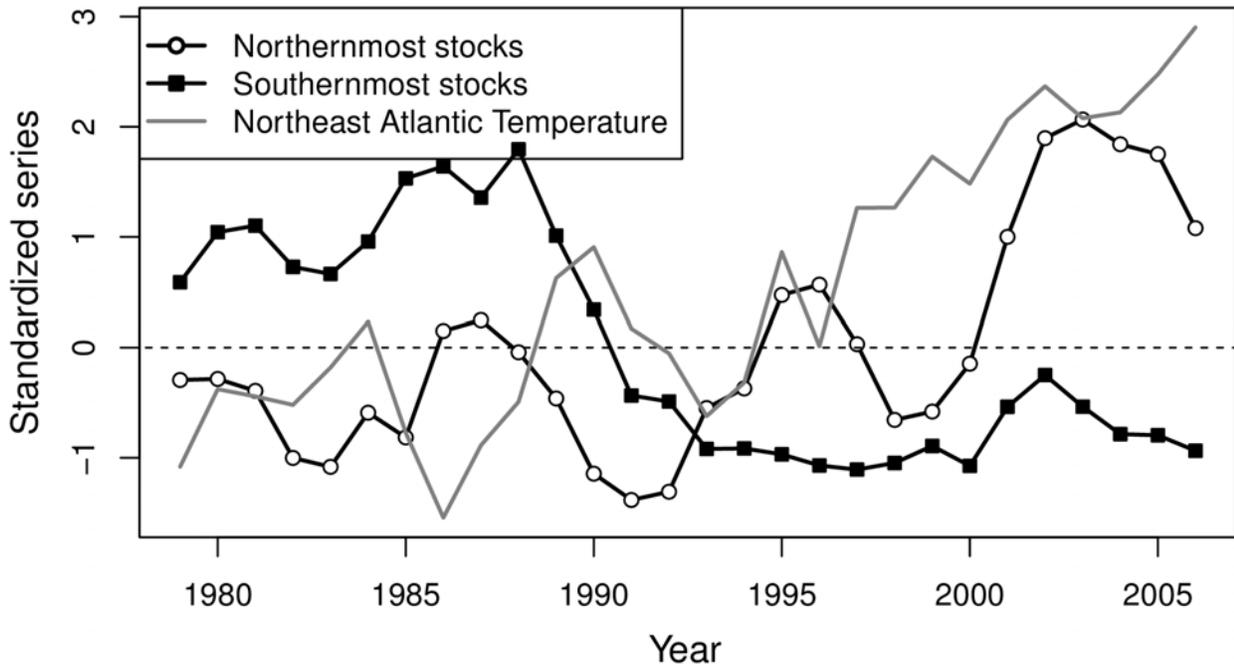


435 FIGURE 5  
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439 FIGURE 6



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