
Development of emergency response tools for accidental radiological contamination of French coastal areas

Duffa Céline^{1,*}, Bailly Du Bois Pascal², Caillaud Matthieu⁵, Charmasson Sabine¹, Couvez Céline³, Didier Damien⁴, Dumas Franck⁶, Fievet Bruno², Morillon Mehdi², Renaud Philippe³, Thébault Hervé¹

¹ Institut de Radioprotection et de Sureté Nucléaire (IRSN), PRP-ENV/SESURE/LERCM, Antenne de Radioécologie Marine, Centre Ifremer, Zone portuaire de Brégaillon, 13507 La Seyne sur Mer, France

² IRSN/PRP-ENV/SERIS/LRC, BP 10, Rue Max Pol Fouchet, 50130 Cherbourg-Octeville, France

³ IRSN/PRP-ENV/SESURE/EC, 31 rue de l'écluse, BP 40035, 78116 Le Vésinet Cedex, France

⁴ IRSN/PRP-CRI/SESUC/BMTA, 31, avenue de la Division Leclerc, BP 17, 92260 Fontenay-aux-Roses, France

⁵ ACTIMAR, 36 quai de la Douane, 29200 Brest, France

⁶ IFREMER - Centre Bretagne, ZI de la Pointe du Diable, CS 10070, 29280 Plouzané, France

* Corresponding author : Céline Duffa, email address : celine.duffa@irsn.fr

Abstract :

The Fukushima nuclear accident resulted in the largest ever accidental release of artificial radionuclides in coastal waters. This accident has shown the importance of marine assessment capabilities for emergency response and the need to develop tools for adequately predicting the evolution and potential impact of radioactive releases to the marine environment. The French Institute for Radiological Protection and Nuclear Safety (IRSN) equips its emergency response centre with operational tools to assist experts and decision makers in the event of accidental atmospheric releases and contamination of the terrestrial environment. The on-going project aims to develop tools for the management of marine contamination events in French coastal areas. This should allow us to evaluate and anticipate post-accident conditions, including potential contamination sites, contamination levels and potential consequences. In order to achieve this goal, two complementary tools are developed: site-specific marine data sheets and a dedicated simulation tool (STERNE, Simulation du Transport et du transfert d'Eléments Radioactifs dans l'environnement marin). Marine data sheets are used to summarize the marine environment characteristics of the various sites considered, and to identify vulnerable areas requiring implementation of population protection measures, such as aquaculture areas, beaches or industrial water intakes, as well as areas of major ecological interest. Local climatological data (dominant sea currents as a function of meteorological or tidal conditions) serving as the basis for an initial environmental sampling strategy is provided whenever possible, along with a list of possible local contacts for operational management purposes. The STERNE simulation tool is designed to predict radionuclide dispersion and contamination in seawater and marine species by incorporating spatio-temporal data. 3D hydrodynamic forecasts are used as input data. Direct discharge points or atmospheric deposition source terms can be taken into account. STERNE calculates Eulerian radionuclide dispersion using advection and diffusion equations established offline from hydrodynamic calculations. A radioecological model based on dynamic transfer equations is implemented to evaluate activity concentrations in aquatic organisms. Essential radioecological parameters (concentration

factors and single or multicomponent biological half-lives) have been compiled for main radionuclides and generic marine species (fish, molluscs, crustaceans and algae). Dispersion and transfer calculations are performed simultaneously on a 3D grid. Results can be plotted on maps, with possible tracking of spatio-temporal evolution. Post-processing and visualization can then be performed.

Highlights

► After Fukushima accident, it appears essential to take this risk of contamination of coastal areas into account. ► This project aims to provide IRSN with enhanced capabilities of impact assessment and management in case of marine crisis. ► Tools are in development for modelling dispersion in seawater and for assessing the potential impact on affected areas. ► STERNE tool is designed to assess the radiological impact of accidental releases affecting the marine environment.

Keywords : Decision support, Marine, Modelling, Radioecology, Nuclear accident

1. Introduction

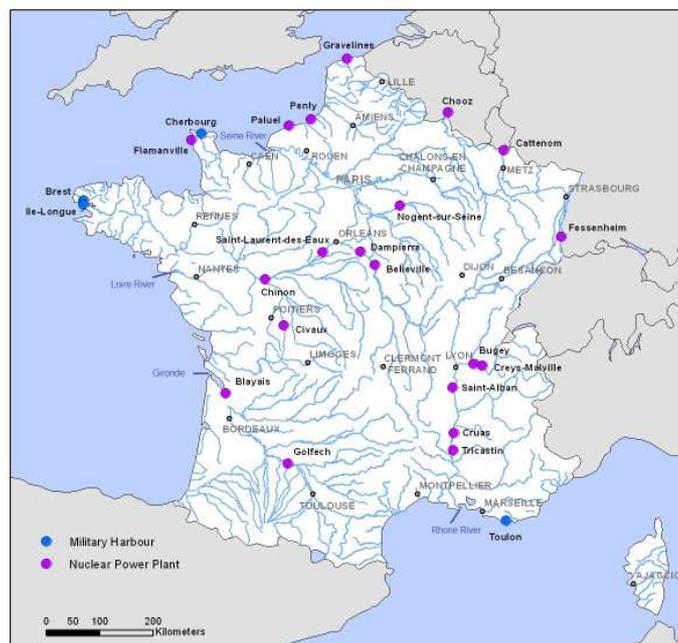
The nuclear accident at Fukushima in 2011 resulted in the largest ever accidental release of artificial radionuclides in coastal waters (UNSCEAR, 2014, Povinec et al., 2014, Science Council of Japan, 2014). The environmental and economic impact on this coastal area is enormous, particularly for fisheries (Okuda and Ohashi, 2012).

Henceforth, it appears essential to take this risk of contamination of coastal areas into account, particularly in those areas most potentially exposed to accidental releases from land based nuclear installations. To ensure optimal preparedness in the event of a nuclear emergency affecting the marine environment, it is necessary to develop and implementspecific tools for assessing the evolution and impact of radioactive marine contamination events. This information will be used to facilitate decision-making during an emergency and

49 could serve as a basis for post-accident sampling strategies leading to realistic environmental
50 impact assessment.

51 The French Institute for Radiological Protection and Nuclear Safety (IRSN) has for many
52 years now equipped its emergency response centre with operational tools to assist experts in
53 the assessment of potential risks to local populations and terrestrial environments in the event
54 of accidental release of radionuclides to the atmosphere. These tools were used in particular in
55 the case of the Fukushima accident, among other things to simulate the short and long-range
56 atmospheric dispersion of released radionuclides (Mathieu *et al.*, 2011; Korakissok *et al.*,
57 2013; Saunier *et al.*, 2013). Atmospheric dispersion computer codes are combined with
58 computational modules designed to predict exposure levels and activity concentrations in
59 different environmental compartments.

60 This project aims to equip IRSN with supplementary tools for impact assessment and
61 management in the event of accidental marine contamination of French coastal areas. Given
62 the length of its coastlines and the large number of nuclear installations in operation, it is
63 extremely important for France to implement such capabilities. In addition to nuclear power
64 plants located directly along the coast (Gravelines, Penly, Paluel and Flamanville NPPs, La
65 Hague reprocessing plant), several nuclear installations are located along rivers that flow into
66 the French Atlantic or Mediterranean coastal waters. Also to be noted is the presence of
67 nuclear-powered ships in the military ports of Brest and Toulon. The maritime transport of
68 nuclear materials must also be included in this inventory of potential source terms (see Figure
69 1).



70
71 **Figure 1: French nuclear installations and coastal areas to be considered**
72

73 This work aims to provide enhanced capabilities for predicting radionuclide dispersion in
74 seawater and for assessing the potential impact on affected areas. In particular, specific tools
75 and resources must be implemented to provide the following information:

- 76 - Initial estimates of expected activity concentrations in seawater, particularly near coastal
77 areas, to effectively ensure the protection of populations directly or indirectly exposed to
78 contaminated environments.

- 79 - Initial estimates of expected activity concentrations in aquatic organisms, particularly
80 those intended for human consumption (fishing or aquaculture products).
81 - Contamination distribution and spatio-temporal evolution maps, to provide guidance for
82 sampling strategies intended to characterise environmental impact.
83 - Detailed information regarding the site-specific environmental sensitivity and ecological,
84 economic and health-related interests of identified areas, to facilitate risk assessment and
85 decision making.

86 The approach adopted to meet these objectives is twofold: Preparation of site-specific data
87 sheets for each coastal area identified as particularly vulnerable in terms of exposure to an
88 accidental release of radionuclides (coastal nuclear installations, river mouths, military ports),
89 and development of a computer code to simulate the dispersion of radionuclides in seawater
90 and their transfer to aquatic organisms.

91

92 **2. Materials and methods**

93 2.1. Marine data sheets

94 For each site identified, a data sheet will be drawn up, listing all information required for
95 preliminary analysis of environmental impact near the release point, as well as planned
96 population protection measures. These data sheets must provide all necessary input data,
97 including the identification and characterisation of particularly vulnerable areas as a function
98 of hydrological conditions, and all information required to prepare sampling plans (sampling
99 locations and sample types) for characterization of environmental impact.

100 These data sheets should therefore include the following:

- 101 - Descriptions of dominant sea currents as a function of meteorological or tidal conditions,
102 allowing for rapid identification of vulnerable areas and assessment of corresponding time
103 frames. This information will provide guidance for the implementation of initial
104 population protection actions (prohibition of swimming, fishing or other site-specific
105 activities, suspension of water intake and port operating activities).
106 - Sampling plans consisting of maps corresponding to different dispersion conditions,
107 including identification of optimal sampling points for contamination assessment
108 purposes.
109 - Maps showing local site-specific interests, including identification of coastal occupation
110 or activity areas for effective implementation of population protection measures. These
111 maps must also show areas of economic interest (fishing, aquaculture and associated
112 activities, industrial activities requiring water intake, sea therapy) and areas of major
113 ecological interest (protected natural areas, significant ecological sites).
114 - List of contact information for local actors in the area considered (port authorities, fishing
115 committees, aquaculture operators, etc.).

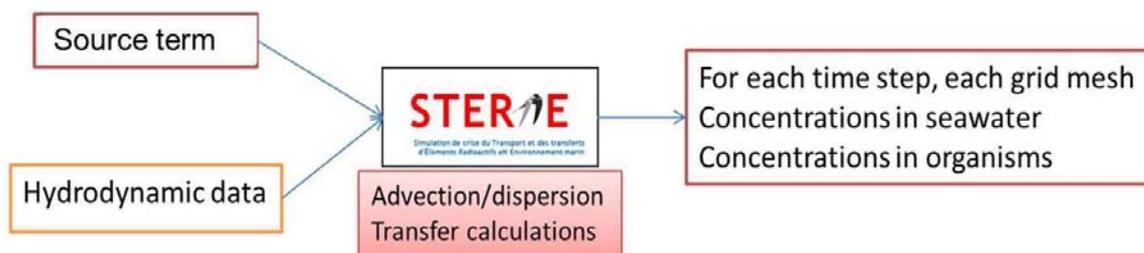
116

117

118 2.2. STERNE simulation tool

119 The STERNE simulation tool ("Simulation du Transport et du transfert d'Eléments
120 Radioactifs dans l'environnement marin", translating as "Simulation of radionuclide
121 transport and transfer in marine environments") is designed to assess the radiological impact
122 of accidental releases affecting the aquatic environment. Similarly to the atmospheric
123 dispersion simulation tools currently available, this new tool is intended to simulate

124 radionuclide dispersion in seawater and to calculate expected activity concentrations in
 125 different biological compartments. The results obtained with STERNE can be used both for
 126 predicting the evolution of contamination events and for dose assessment purposes (via post-
 127 processing tools). When cross-referenced with contextual data, these results can be used to
 128 define measures prohibiting swimming, fishing or other site-specific activities, and also to
 129 provide guidance for sampling strategies during emergency response and post-accident
 130 phases. When used in analysis mode, the STERNE simulation tool can also generate maps of
 131 contaminated areas for different site-specific release scenarios (for example, to generate
 132 information for marine data sheets).



133

134 **Figure 2: Schematic diagram of STERNE implementation principle**

135

136 The implementation principle of the STERNE simulation tool is shown schematically in
 137 Figure 2. The basic principle is the same as for atmospheric dispersion calculations currently
 138 performed at IRSN's emergency response centre, with source terms and meteorological data
 139 used as input data. For aquatic dispersion calculations, source terms and hydrodynamic data
 140 are fed directly into the simulation tool.

141

142 2.2.1. Input data

143 The hydrodynamic data serving as the basis for dispersion calculations is supplied as a single
 144 NetCDF file including data calculated on a 3D grid. Cumulative water fluxes in x, y and z
 145 directions, free surface elevation and diffusion coefficients are available for each mesh and
 146 each time step. Cumulative water fluxes are used to calculate the exact quantity of water
 147 passing through the grid meshes at each instant, thereby satisfying the continuity equation.
 148 Hydrodynamic datasets are generated using the MARS3D model (Model for Applications at
 149 Regional Scale; Lazure and Dumas, 2008) developed by IFREMER (French Research
 150 Institute for Exploitation of the Sea). The MARS3D model has been previously used and
 151 validated by IRSN in various case studies (Bailly du Bois, 2005; Bailly du Bois et al., 2012a;
 152 Bailly du Bois et al., 2014; Duffa *et al.*, 2011; Dufresne *et al.*, 2014). IFREMER generates
 153 seven-day hindcasts and forecasts of hydrodynamic conditions in French coastal areas
 154 (<http://www.previmer.fr>), with spatial resolutions of 1.2 km (Northern Mediterranean Sea,
 155 André et al., 2005) and 2.5 km (Atlantic Coast and English Channel, Lazure et al., 2009). The
 156 use of fixed-mesh models with kilometre-scale resolution ensures acceptable calculation times
 157 for such large areas. Hydrodynamic models are generated based on hindcasts and forecasts of
 158 meteorological and tidal forcing and river flow for major French rivers. Boundary conditions
 159 are forced using the Mercator global model (<http://www.mercator-ocean.fr/>, Ferry *et al.*,
 160 2007).

161 Source terms are characterised by known quantities of radionuclide releases at known
 162 locations (or predefined areas) and instants (or time series). Accidental releases of

163 radionuclides to the marine environment can occur along different pathways and in various
164 modes. Source terms described in data files include release point coordinates, activity input to
165 seawater and their temporal evolution for one or several radionuclides of interest.

166 Two main types of source terms can be considered:

167 - Punctual release of a known concentration of one or more radionuclides. Such releases
168 are instantaneous or follow a specific accidental release scenario. STERNE allows users to
169 create source term description files. The standard example is the Fukushima accident, where a
170 single release point is defined for an accidental release scenario lasting several days and
171 involving several radionuclides (UNSCEAR, 2014).

172 - Atmospheric radionuclide deposition on sea surface further to accidental release from
173 land-based nuclear installation or nuclear-powered ship. This source term may evolve both
174 spatially and temporally. NetCDF datasets generated by IRSN atmospheric dispersion models
175 can be fed directly into the STERNE simulation tool, which computes interpolated deposition
176 values at hydrodynamic model grid points from atmospheric model data.

177

178 2.2.2. Calculations

179

180 The STERNE simulation tool uses offline calculations of radionuclide dispersion and transfer.
181 A specific FORTRAN95 code has been developed for this purpose.

182 Eulerian radionuclide dispersion is calculated using a tracer advection-diffusion equation
183 (Equation 1).

$$184 \quad \frac{\partial C}{\partial t} = \text{Div}(\overline{\overline{K}}\vec{\nabla}C) - \vec{\nabla}(\vec{u}C) \quad \text{Eq. 1}$$

185 Where: C is the radionuclide concentration

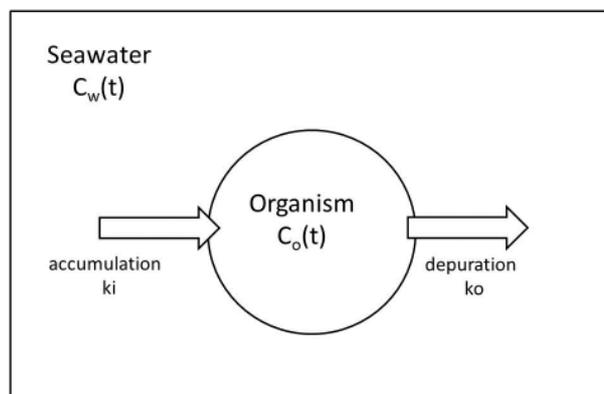
186 $\overline{\overline{K}}$ is the turbulent diffusion tensor

187 \vec{u} is the advection current

188 t is the time elapsed

189 These calculations take into account the physical decay of each radionuclide considered.
190 Activity concentrations are calculated at each grid point and each time step. Time step values
191 are user-defined based on an acceptable compromise between calculation time and numerical
192 stability. The choice of time step depends on the mesh size and maximum sea current velocity
193 for the area considered. For example, for 1.2 km resolution (Northern Mediterranean Sea), the
194 time step is set to 50 seconds.

195 The STERNE simulation tool uses a radioecological model to calculate activity
196 concentrations in aquatic organisms based on a dynamic transfer approach. Various types of
197 models are available to simulate the transfer of contaminants from seawater to aquatic
198 organisms. The strategy adopted consists of using a dynamic compartmental model as
199 proposed by Fievet and Plet (2003) and Vives i Battle *et al.* (2008). The use of single or
200 multiple compartments to represent an organism allows for a simplified operational model.
201 Actual physiological processes are far more complex (different contamination pathways, e.g.
202 feeding or breathing, different trophic levels, specific physiological parameters such as
203 physical growth or trophic factors). Nevertheless, our approach allows for implementing a
204 relatively simple biokinetic model (Figure 3) to simulate radionuclide transfers from seawater
205 to the organism of interest based on two types of parameters: concentration factors (CF, ratio
206 of radionuclide concentration in species considered vs. concentration in seawater) and
207 biological half-lives (T_b).



208
209 **Figure 3: Biokinetic model of radionuclide transfer from seawater to aquatic organisms**
210

211 Based on this model, the following equation (Equation 2) is used to calculate activity
212 concentrations in living organisms for a given concentration in seawater.
213

$$214 \quad \frac{dC_o(t)}{dt} = k_i \cdot C_w(t) - (\lambda_p + k_o) \cdot C_o(t) \quad \text{Eq. 2}$$

215 where $k_i = (k_o + \lambda_p) \cdot CF$

216 C_o is the activity concentration in the organism (Bq.kg⁻¹ fresh weight)

217 C_w is the activity concentration in seawater (Bq.l⁻¹)

218 k_i is the uptake or accumulation rate constant (d⁻¹)

219 k_o is the elimination or depuration rate constant (d⁻¹)

220 λ_p is the physical decay constant (d⁻¹)

221 CF is the concentration factor (l.kg⁻¹ f.w.)
222

223 In STERNE code, this equation is computed for a time step i as follows (see Fievet and Plet,
224 2003 for details):
225

$$226 \quad C_{o(i)} = a \times C_{o(i-1)} + FC \times (1 - a) \times C_{w(i)} \quad \text{Eq. 3}$$

227 where i is the time step.
228
229

230 In this equation $a = e^{-T(k_o + \lambda_p)}$

231 where $T = t(i) - t(i - 1)$ is the constant time step duration.

232 In order to refine this one-compartment dynamic model and adapt it to post-accident
233 conditions (for which various studies report multiple depuration rate constants), the STERNE
234 simulation tool allows users to combine two independent compartments as A.Co1+B.Co2. A
235 and B have values between 0 and 1, and A+B=1. Each compartment Co1 and Co2 has its own
236 transfer parameters (CF and Tb).

237 All required radioecological parameters (concentration factors and single or multi-component
238 biological half-lives) are compiled from literature (IAEA, 2004, Gomez *et al.*, 1991 and Vives
239 i Battle *et al.*, 2007) for main radionuclides and generic marine species (fish, molluscs,
240 crustaceans, algae). Default values can be changed by the user if necessary.

241 Activity concentrations in organisms are calculated at each 2D calculation grid point and each
242 time step used to calculate seawater activity concentrations. Activity concentrations in bottom
243 fish, molluscs, crustaceans and algae are calculated based on bottom-water concentrations.
244 For pelagic fish, two calculations are performed, i.e. based on the mean activity concentration
245 in seawater over the entire water column at the 2D calculation grid point, and based on the
246 maximum activity concentration in the water column at the same grid point.
247 One limitation of this model is that it does not consider potential contamination pathways
248 associated with bottom sediments, which are still scarcely understood. This can lead to
249 underestimates of activity concentrations in bottom fish exposed to medium or long-term
250 contamination.

251

252 3. Results

253

254 Calculations performed with the STERNE simulation tool generate values of activity
255 concentrations in seawater at each 3D grid point and at a user-defined time step. Mass
256 concentrations in selected living species (fish, algae, molluscs or crustaceans) are also
257 calculated with the chosen radioecological parameters. These values can be calculated over a
258 spatio-temporal domain smaller than or equal to that of the hydrodynamic input data.

259 Dispersion and transfer calculations are performed simultaneously on a 3D grid. Results are
260 generated in the form of a global NetCDF file or a set of time-series files for specified
261 tracking points.

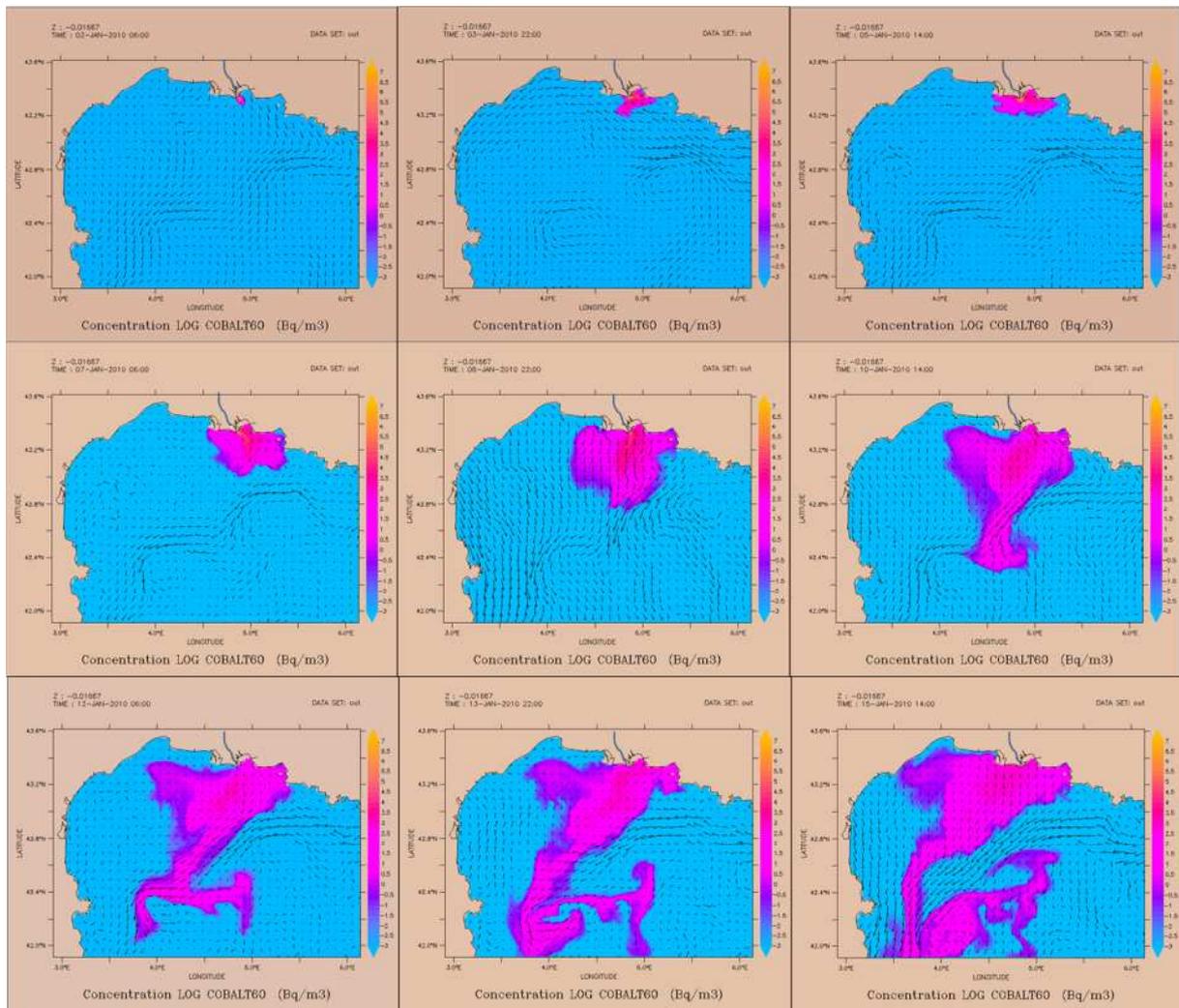
262 As an example, calculations are performed for a theoretical Rhone River accidental ^{137}Cs
263 discharge scenario where a total of 10^{14} Bq of ^{137}Cs is discharged at a constant rate for 7 days
264 starting on 2 January 2010. Realistic hydrodynamic data provided by the MARS3D model for
265 the North Mediterranean Sea (1.2 km horizontal resolution, 30 vertical layers) is used to
266 calculate radionuclide dispersion over a period of 1 month (from 2 January 2010). The area
267 considered is shown in Figure 4. Three coastal stations are defined as tracking points to export
268 calculated concentration time-series data for seawater and pelagic fish.

269 Radioecological parameters used to calculate the transfer of ^{137}Cs to fish are taken from the
270 literature (IAEA, 2004; Gomez, 1991) as follows: $CF=100$ and $Tb=58$ days, with one kinetic
271 parameter.



272

273 **Figure 4: Area considered for the simulation of ^{137}Cs dispersion along the French**
274 **Mediterranean Coast, and location of tracking stations (Rhone River Mouth, Cap**
275 **d'Agde and Banyuls sur Mer)**

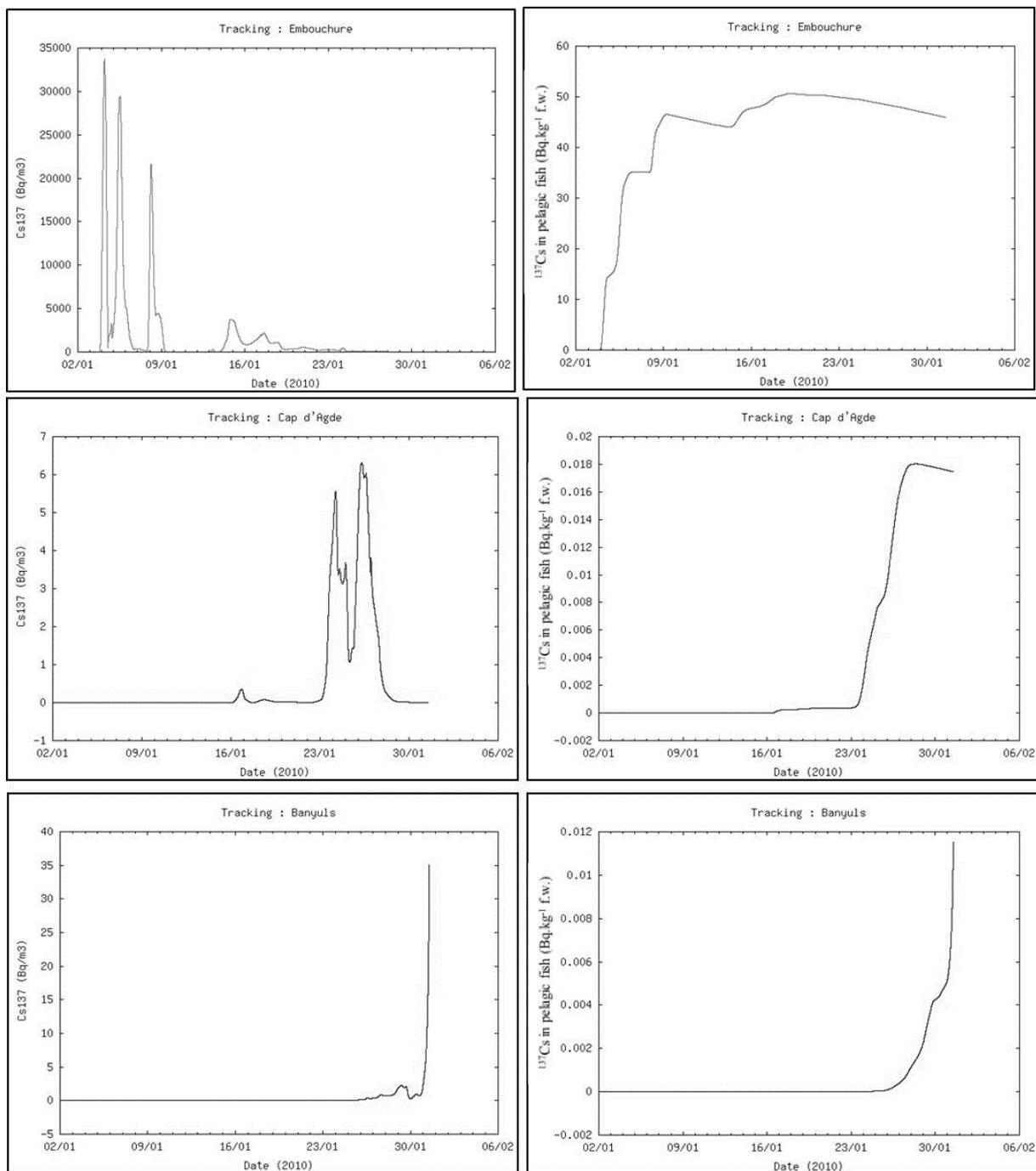


276

277 **Figure 5: Example of simulation results for surface dispersion of 10^{14} Bq of ^{137}Cs**
 278 **discharged from the Rhone River for 7 days ($\text{Bq}\cdot\text{m}^{-3}$ in surface water, logarithmic scale)**
 279 **starting on 2 January 2010, using realistic hydrodynamic data**
 280

281 Contaminated freshwater from the Rhone River spreads on the marine surface layer for the first 3
 282 days and then reaches the North Mediterranean Current. Subsequently, contaminated waters shift to
 283 the southwest and are transported toward the Spanish coast. Figure 6 shows the concentration time
 284 series in seawater and pelagic fish at the 3 tracking stations. The ^{137}Cs plume reaches the Cap d'Agde
 285 station, located 110 km from the Rhone River discharge point, approximately 22 days after the first
 286 discharge to the sea, and it reaches the Banyuls-sur-Mer station, located 165 km from the Rhone
 287 River discharge point, approximately 8 days later. In this dispersion scenario, it is interesting to note
 288 the significantly higher seawater activity concentrations at the Banyuls-sur-Mer station, where the
 289 main plume of contaminated water is directly advected by the North Mediterranean Current
 290 traveling along the continental shelf, as shown by Millot and Taupier-Letage (2005).

291 Calculated activity concentrations in pelagic fish are shown in Figure 6, with the concentration time
 292 series clearly indicating the persistence of ^{137}Cs in these organisms, due to the 58-day biological half-
 293 life used in the transfer model.



294
 295 **Figure 6: Concentration time series in seawater (Bq.m⁻³) and pelagic fish (Bq.kg⁻¹ f.w.) at**
 296 **Rhone River Mouth, Cap d'Agde and Banyuls-sur-Mer stations**

297

298 **4. Conclusion**

299
300 IRSN is currently developing two complementary tools for use in the event of a radiological
301 marine accident in French coastal areas. Site-specific marine data sheets will allow for initial
302 assessment of potential consequences and will help experts define sampling and monitoring
303 strategies.

304 The STERNE simulation tool is still in the validation stage and only available for case
305 studies. The accuracy of dispersion simulation results is directly dependent on hydrodynamic
306 forcing and source term realism. It is therefore essential to use forecasts from validated
307 hydrodynamic models, along with the best possible source term characterization.
308 Hydrodynamic conditions simulated by the Mars3D model for French coastal areas compare
309 well with measurements taken over short time intervals (hours, days) but are difficult to verify
310 for longer simulations (weeks, months). Such implementations would require an accurate,
311 validated representation of local hydrodynamics, such as for example that produced for the
312 Toulon area (Duffa *et al*, 2011; Dufresnes *et al*, 2014).

313 The modelling approach chosen to simulate transfers to marine biota still needs to be
314 validated. Additional kinetic parameters for the simulation of transfers between seawater and
315 aquatic organisms also need to be documented.

316 At this stage, considering the intended use as an emergency response tool, only dissolved-
317 phase dispersion modelling is applied. In order to take into account the fraction of
318 radionuclides attached to suspended sediment particles, it would be necessary to include a
319 sediment transport module. The integration of such a module is relatively complex and
320 computation-time consuming and, most importantly, would require adjustment to in situ
321 measurements of various sediment-specific parameters. Sediment transport modelling could
322 be subsequently implemented for more in-depth calculations during post-emergency phases or
323 for use in analysis mode. It is therefore not included in the initial development.

324 The use of marine data sheets in combination with the STERNE simulation tool for various
325 case studies should provide indications as to the strategy to be adopted.

326

327 **5. Acknowledgments**

328

329 The authors would like to thank the IFREMER modelling team for its support and use of the
330 Caparmor computational server.

331

332 **References**

333

334 André G, Garreau P, Garnier V, Fraunié P (2005) Modelling variability of the sea surface circulation in the
335 North-Western Mediterranean Sea and in the Gulf of Lions. *Ocean Dynamics* 55:294–308

336 IAEA, 2004, Technical Report Series No. 422, Sediment distribution coefficients and concentration factors for
337 biota in the marine environment.

338 Bailly du Bois P. and Dumas F., 2005. Fast hydrodynamic model for of medium- and long-term dispersion in
339 seawater in the English Channel and southern North Sea, qualitative and quantitative validation by
340 radionuclide tracers. *Ocean Modelling* 9,2 169-210.

341 Duffa C., Dufois F. Coudray S., 2011, An operational model to simulate post-accidental radionuclide transfers in
342 Toulon marine area : preliminary development, *Ocean Dynamics* DOI 10.1007/s10236-011-0429-0.

343 Dufresne C., Duffa C., Rey V. 2014, 3D circulation in the Bay of Toulon and water exchanges at the Little Bay
344 fairway, *Ocean Dynamics*, Vol.64 (2), pp 209-224.

345 Fievet B. and Plet D., 2003, Estimating biological half-life of radionuclides in marine compartments from
346 environmental time-series measurements, *Journal of Environmental Radioactivity* 65, 91-107.

347 Ferry N, Rémy E, Brasseur P, Maes C (2007) The Mercator global ocean operational analysis system:
348 Assessment and validation of an 11-year reanalysis. *Journal of Marine Systems* 65 (1–4): 540-560.

349 Gomez L.S., Marietta M.G., Jackson D.W., 1991, Compilation of selected marine radioecological data for the
350 formerly utilized sites remedial action program: summaries of available radioecological concentration
351 factors and biological half-lives, SANDIA Report, SAND89- 1585.

352 Korsakissok I., Mathieu A. and Didier D., 2013, Atmospheric dispersion and ground deposition induced by the
353 Fukushima Nuclear power plant accident: a local-scale simulation and sensitivity study,
354 *Atmospheric Environment* 70: 267-279.

355 Lazure P. and Dumas F., 2008, An external–internal mode coupling for a 3D hydrodynamical model for
356 applications at regional scale (MARS). *Advances in Water Resources* 31(2), 233-250.

357 Lazure P., Garnier V, Dumas F, Herry C., Chifflet M., 2009, Development of a hydrodynamic model of the Bay
358 of Biscay. *Validation of hydrology, Continental Shelf Research, Volume 29, Issue 8, 985-99.*

359 Mathieu A., Korsakissok I., Quélo D., Groëll J., Tombette M., Didier D., Quentric E., Saunier O., Benoît J.-P.,
360 Isnard O., 2011. Atmospheric Dispersion and Deposition of Radionuclides from the Fukushima Daiichi
361 Nuclear Power Plant Accident. *Elements* volume 8, number 3. 1811-5209/12/0008-0195\$2.50 DOI:
362 10.2113/gselements.8.3.195.

363 Millot C., and I. Taupier-Letage, 2005. *Circulation in the Mediterranean Sea. The Handbook of Environmental
364 Chemistry*”, Volume 5 Part K, Alain Saliot volume Ed., Springer Verlag, Berlin Heidelberg, 29-66.

365 Okuda K., Ohashi M., 2012, On the studies of recovery and reconstruction of fisheries hit by the Great East
366 Japan earthquake, *Procedia Technology* 5, 208-214.

367 Povinec P., Hirose K., Aoyama M., 2013, Fukushima accident: Radioactivity impact on the environment, ISBN:
368 978-0-12-408132-1, Elsevier Ed.

369 Saunier O., Mathieu A., Didier D., Tombette M., Quélo D., Winiarek V., and Bocquet M., 2013, An inverse
370 modeling method to assess the source term of the Fukushima Nuclear Power Plant accident using gamma
371 dose rate observations, *Atmos. Chem. Phys.*, 13, 11403-11421, doi:10.5194/acp-13-11403-2013.

372 Science-Council-of-Japan, 2014. A review of the model comparison of transportation and deposition of
373 radioactive materials released to the environment as a result of the Tokyo Electric Power Company’s
374 Fukushima Daiichi Nuclear Power Plant accident. Science Council of Japan, in: Sectional Committee on
375 Nuclear Accident Committee on Comprehensive Synthetic Engineering, S.C.o.J. (Ed.), 111 p

376 UNSCEAR, 2014, Sources, effects and Risks of Ionizing Radiation - UNSCEAR 2013 Report to the General
377 Assembly with Scientific Annexes, Volume1, United Nation.

378 Vives i Batlle J., Wilson R.C., McDonald P., 2007, Allometric methodology for the calculation of biokinetic
379 parameters for marine biota, *Science of the Total Environment* 388, 256-269.

380 Vives i Batlle J., Wilson R.C., Watts S.J., Jones S.R., McDonald P., Vives-Lynch S., 2008, Dynamic model for
381 the assessment of radiological exposure to marine biota, *Journal of Environmental Radioactivity*, 99,
382 1711-1730.