

Biological interactions: The overlooked aspects of marine climate change refugia

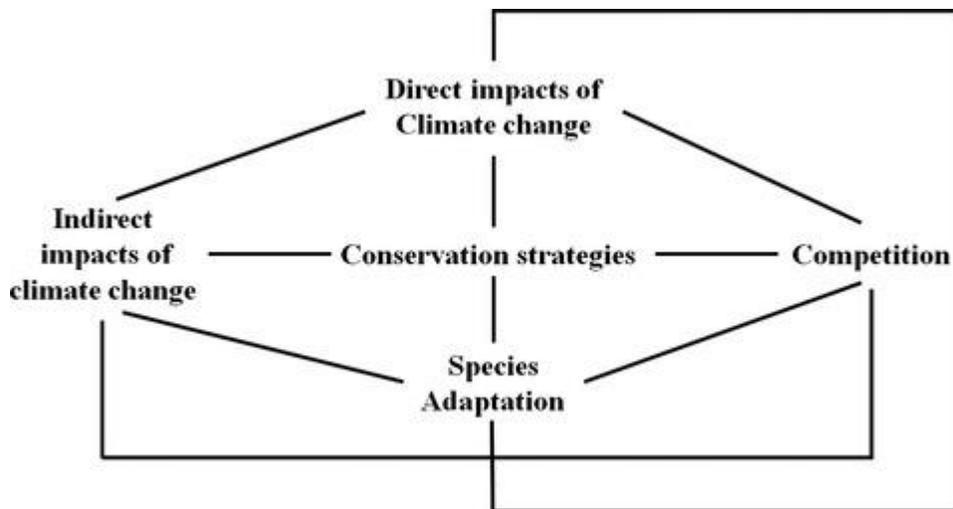
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Abstract :

This article emphasizes on biological interactions as an important overlooked criterion to better assess the chance of target marine species in potential refugia to survive climate change. It proposes future climate change refugia studies must consider the reciprocal interactions among climate change, biological factor, and conservation strategies.



10 Climate change refugia are currently considered an inseparable part of biological conservation in
11 both marine and terrestrial ecosystems. However, our understanding of marine refugia
12 effectiveness remains limited despite many valuable efforts made in recent years. Certain studies
13 have suggested criteria to assess potential marine climate change refugia (e.g., Kavousi & Keppel,
14 2018), while others proposed specific refugia to conserve (e.g. Cacciapaglia & van Woesik, 2016
15 and references therein). While I acknowledge that determining the most effective marine climate
16 change refugia is an urgent necessary action to protect marine biota, I propose that ruling out
17 biological interactions may result in far less accurate estimations of the effectiveness of potential
18 long-term refugia and short-term refuges (defined by Keppel *et al.*, 2012) for marine species,
19 which could subsequently lead to overly optimistic ecosystem management strategies.

20 The high level of uncertainty associated with climate change impacts on marine ecosystems
21 remains an obstacle to establishing long-term conservation plans (Beyer *et al.*, 2018) and is a
22 hindrance to identifying effective marine climate change refugia. While refugia can provide
23 protection from some climatic changes for target species, there remains a lack of data directly
24 assessing how potential refugia and their species respond to the *indirect* impacts of climate change,

25 such as climate change-induced and -intensified biological phenomena (e.g., the outbreak of
26 infectious diseases, harmful algal blooms, invasive species, or intensified predation). For instance,
27 global projections suggest that warming-induced coral diseases will be a major driver of coral reef
28 declines, equal to that of coral bleaching (Maynard *et al.*, 2015). The increased temperatures
29 causing the former either have already started or will start in the coming decades, primarily before
30 the onset of annual coral bleaching for the majority of coral reefs (Maynard *et al.*, 2015), which
31 includes nearly all proposed large-scale coral reef global warming refugia (Fig. 1). Since marine
32 refugia are not strongly decoupled from their surroundings (Kavousi & Keppel, 2018), biological
33 threats in refugia and refuges can sometimes appear with catastrophic impacts (see e.g., Bongaerts
34 *et al.*, 2010). Notably, there does not appear to be a practical barrier to maintaining refugial
35 capacities—the ability of refugia to mitigate the magnitude, duration, and increased rate of
36 stressors—under the indirect impacts of climate change.

37 Like all other environments, refugia are prone to competition over space, light, and other factors.
38 Small size, novel biotic interactions (e.g., dispersal of species into an *ex situ* refugium that would
39 be outside a species' range, as defined by Keppel *et al.*, 2012), and size shrinkage due to climate
40 change (e.g., Assis *et al.*, 2018) can intensify competition in refugia, which may lead to
41 competitive exclusion under persistent climate change (Mclaughlin *et al.*, 2017). Evidently, any
42 species that is unable to cope with the negative effects of climate change and must abandon their
43 original unsuitable habitats will strive to benefit from any potential refugium that they manage to
44 reach. Those species can be historical competitors (e.g., reef corals and macroalgae) or novel
45 competitors (endemic and invasive species). Therefore, refugia must either be available for
46 colonization or support the persistence of existing communities (Mclaughlin *et al.*, 2017).

47 To date, three major type refugia for marine species—including higher latitudes, deep areas, and
48 upwelling regions—are suggested to act as global warming refugia for many species, including
49 reef corals, macroalgae, and kelp. However, all of these regions are generally predicted to
50 experience higher rates of intensified ocean acidification in the coming decades (Gattuso &
51 Hansson, 2011). Thus, the direct negative effects of ocean acidification may favor some species,
52 such as mat-forming algae, in competition with ecosystem engineers such as reef-building coral
53 and kelp (Connell *et al.*, 2013). As a result, competition within refugia would become even more
54 complicated. Therefore, how competition will shape communities within marine refugia may not
55 be only determined by the refugium limitations and competition level, but also by the rate in which
56 climatic change (and other non-climate factors) can alter the competition capabilities of the species
57 involved.

58 Overall, changes in biotic interactions under climate change represent a known cause of species
59 extinction (Blois *et al.*, 2013), thus they cannot be ignored in climate change refugia studies. Such
60 biotic interactions are even more important than the direct abiotic impacts of climate change
61 (Ockendon *et al.*, 2014). Therefore, the identification of effective climate change refugia/refuges
62 without considering biological interactions—particularly those induced and/or intensified by
63 climate change such as competition, marine diseases, and phytoplankton blooms—seems to be
64 simplistic. Although, such interactions will undoubtedly increase the complexity and practicality
65 of projections, their implications are, in my opinion, inevitable. To support my claim, I exemplify
66 the Persian Gulf, which has been suggested to act as a coral reef refugium against global warming
67 by the year 2100 (Fig. 1; Cacciapaglia & van Woesik, 2016). However, along with multiple mass
68 bleaching events, the Persian Gulf has already experienced mass mortalities of corals due to
69 diseases and phytoplankton blooms (Riegl *et al.*, 2012; 2018 and references therein). Therefore, I

70 suggest that simply pinpointing potential marine climate change refugia based on their capacities
71 to buffer global warming or ocean acidification (i.e., the direct impacts of climate change) should
72 not be incorporated into modern conservation frameworks and marine spatial planning initiatives.
73 Consequently, I propose that to be realistic and reliable, future climate change refugia studies must
74 consider the reciprocal interactions among climate change, biological factor and conservation
75 strategies (Fig. 2). Without effective management plans, even the least vulnerable ecosystems may
76 be at risk (see Beyer *et al.*, 2018). Such strategies can also reduce non-climate change stressors,
77 which should not be ignored while identifying marine refugia (Kavousi & Keppel, 2018). I
78 emphasize my understanding of how the suggested study structure can be complicated, time-
79 consuming, and costly to perform. However, the painful truth is that conserving our oceans against
80 climate change is complicated by multilevel impacts that require the application of complex,
81 multilevel research methodologies and management strategies.

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83 **References**

- 84 Assis J., Araújo M.B., Serrão E.A. (2018). Projected climate changes threaten ancient refugia of
85 kelp forests in the North Atlantic. *Global Change Biology*, 24(1), e55-e66.
- 86 Beyer H.L., Kennedy E.V., Beger M., Chen C.A., Cinner J.E., Darling E.S., Eakin C.M., Gates
87 R.D., Heron S.F., Knowlton N., Obura D.O. (2018). Risk-sensitive planning for conserving
88 coral reefs under rapid climate change. *Conservation Letters*, 11(6), p.e12587.
- 89 Blois J.L., Zarnetske P.L., Fitzpatrick M.C., Finnegan S. (2013). Climate change and the past,
90 present, and future of biotic interactions. *Science*, 341(6145), 499-504.
- 91 Bongaerts P., Ridgway T., Sampayo E.M., Hoegh-Guldberg O. (2010). Assessing the ‘deep reef
92 refugia’ hypothesis: focus on Caribbean reefs. *Coral Reefs*, 29(2), 309-327.
- 93 Cacciapaglia C., Van Woesik R. (2016). Climate-change refugia: shading reef corals by turbidity.
94 *Global Change Biology*. 22(3), 1145-1154.
- 95 Connell S.D., Kroeker K.J., Fabricius K.E., Kline D.I., Russell B.D. (2013). The other ocean
96 acidification problem: CO₂ as a resource among competitors for ecosystem dominance.
97 *Philosophical Transactions of the Royal Society B: Biological Sciences*, 368(1627),
98 20120442.

99 Gattuso, J.P., Hansson, L. (2011). Ocean acidification. New York: Oxford University Press.
100 Kavousi J., Keppel G. (2018). Clarifying the concept of climate change refugia for coral reefs.
101 ICES Journal of Marine Science. 75(1), 43-49.
102 Keppel G., Van Niel K.P., Wardell-Johnson G.W., Yates C.J., Byrne M., Mucina L., Schut A.G.,
103 Hopper S.D., Franklin S.E. (2012). Refugia: identifying and understanding safe havens for
104 biodiversity under climate change. *Global Ecology and Biogeography*, 21(4), 393-404.
105 Maynard J., Van Hooidek R., Eakin C.M., Puotinen M., Garren M., Williams G., Heron S.F.,
106 Lamb J., Weil E., Willis B., Harvell C.D. (2015) Projections of climate conditions that
107 increase coral disease susceptibility and pathogen abundance and virulence. *Nature*
108 *Climate Change*. 5(7), 688-694.
109 McLaughlin B.C., Ackerly D.D., Klos P.Z., Natali J., Dawson T.E., Thompson S.E. (2017).
110 Hydrologic refugia, plants, and climate change. *Global Change Biology*, 23(8), 2941-2961.
111 Ockendon N., Baker D.J., Carr J.A., White E.C., Almond R.E., Amano T., Bertram E., Bradbury
112 R.B., Bradley C., Butchart S.H.M., Doswald N., Foden W., Gill D.J.C., Green R.E.,
113 Sutherland W.J., Tanner E.V.J., Pearce-Higgins J.W. (2014). Mechanisms underpinning
114 climatic impacts on natural populations: altered species interactions are more important
115 than direct effects. *Global Change Biology*, 20(7), 2221-2229.
116 Riegl B., Bruckner A.W., Samimi-Namin K., Purkis S.J. (2012) Diseases, harmful algae blooms
117 (HABs) and their effects on Gulf coral populations and communities. In: B. Riegl & S.J.
118 Purkis (Eds.), *Coral Reefs of the Gulf* (pp. 107-125). Dordrecht: Springer.
119 Riegl B., Johnston M., Purkis S., Howells E., Burt J., Steiner S.C., Sheppard C.R., Bauman A.
120 (2018). Population collapse dynamics in *Acropora downingi*, an Arabian/Persian Gulf
121 ecosystem-engineering coral, linked to rising temperature. *Global Change Biology*, 24(6),
122 2447-2462.

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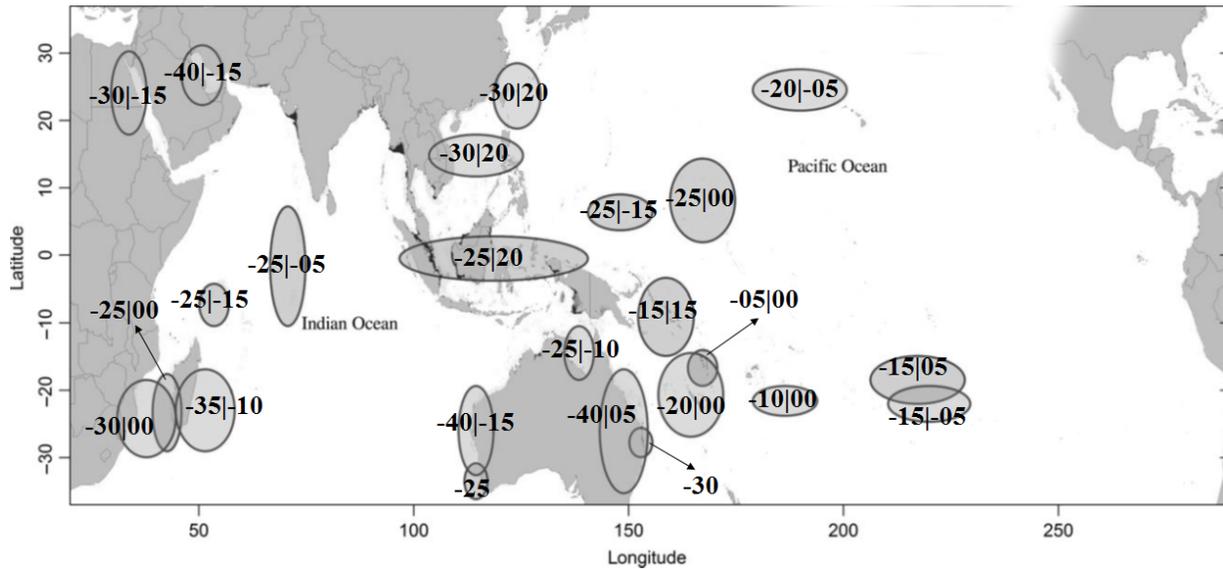
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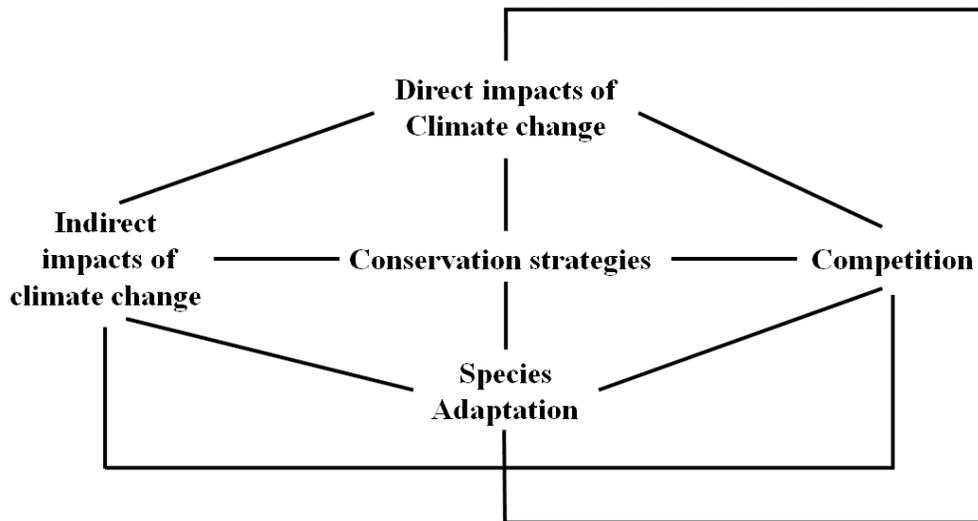
130 **Figures**



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132 **Fig. 1** The difference in timing between when at least two of the three temperature conditions for
133 disease —i.e. coral susceptibility to disease, pathogen abundance, and pathogen virulence — will
134 be reached and the onset of severe annual bleaching (extracted from Maynard *et al.*, 2015) in 24
135 suggested coral reef refugia (Cacciapaglia & van Woesik, 2016). These values reflect the earliest
136 and final year for each refugium in which the conditions can occur before bleaching. Negative
137 values imply that at least two of the three temperature conditions for disease are projected to occur
138 before annual severe bleaching conditions.

139



140

141 **Fig. 2** Simplified suggested structure of reciprocal interactions that should be considered in future
 142 management studies to better assess the chance of target marine species in potential refugia to
 143 survive climate change.