Resupply of mesopelagic dissolved iron controlled by particulate iron composition

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Abstract :

The dissolved iron supply controls half of the oceans' primary productivity. Resupply by the remineralization of sinking particles, and subsequent vertical mixing, largely sustains this productivity. However, our understanding of the drivers of dissolved iron resupply, and their influence on its vertical distribution across the oceans, is still limited due to sparse observations. There is a lack of empirical evidence as to what controls the subsurface iron remineralization due to difficulties in studying mesopelagic biogeochemistry. Here we present estimates of particulate transformations to dissolved iron, concurrent oxygen consumption and iron-binding ligand replenishment based on in situ mesopelagic experiments. Dissolved iron regeneration efficiencies (that is, replenishment over oxygen consumption) were 10- to 100-fold higher in low-dust subantarctic waters relative to higher-dust Mediterranean sites. Regeneration efficiencies are heavily influenced by particle composition. Their make-up dictates ligand release, controls scavenging, modulates ballasting and may lead to the differential remineralization of biogenic versus lithogenic iron. At high-dust sites, these processes together increase the iron remineralization length scale. Modelling reveals that in oceanic regions near deserts, enhanced lithogenic fluxes deepen the ferricline, which alter the vertical patterns of dissolved iron replenishment, and set its redistribution at the global scale. Such wide-ranging regeneration efficiencies drive different vertical patterns in dissolved iron replenishment across oceanic provinces.

Globally, the productivity of major phytoplankton groups, including diatoms and diazotrophs, is set by dissolved iron (DFe) supply[1]. Twenty years of research has revealed diverse modes of DFe supply, from dust to hydrothermal vents, and their regional influences[2,3]. Iron biogeochemistry is a rapidly evolving field, driving the development of global modelling initiatives[4]. However, iron cycling in the oceans' interior, a fundamental component of iron biogeochemistry[5,6], represents a major unknown, and critically is hindering model development[7].

51

Deep winter mixing is a key vector in the annual resupply of upper ocean DFe stocks[8]. At depth, DFe is replenished via biotic (e.g., microbial solubilization) and abiotic (e.g., dissolution) transformations (also termed remineralization or replenishment here) of particulate Fe (PFe)[7]. To date, internal iron cycling has been investigated using three distinctive approaches, each of which has improved it's representation in biogeochemical models[9-11]. Firstly, studies deploying trace metal-clean multi-depth sediment traps, in regions dominated by biogenic PFe, provided initial evidence of subsurface decoupling in PFe remineralization relative to phosphorus (P), carbon (C) or nitrogen[12-15]. Second, bacterially-mediated PFe remineralization – suppressed in sediment trap studies due to preservatives – was investigated in shipboard time-series incubations of resuspended mesopelagic particles in which the release of DFe and Fe-binding ligands resulted from particle degradation[9,16]. Third, dust addition experiments within nearshore 15 m deep mesocosms demonstrated that lithogenic Fe, conventionally viewed as a major source of external Fe to surface waters[1,2,17], can either release (dissolution) or remove (scavenging) DFe depending on the initial biogeochemical conditions[18-20].

67

These three approaches to internal iron cycling targeted different processes, such as ligand and DFe release[9,16], patterns in PFe flux attenuation[12-15], or scavenging and dissolution[18-20]. However, it is difficult to compare the findings of these studies as they each sampled particle assemblages with differing contributions from biogenic and lithogenic iron. Hence, no study so far has examined the relative role of the different processes associated with biogenic versus lithogenic Fe on internal iron cycling. The hypothesis[9] that the composition of particles will largely determine mesopelagic patterns of DFe replenishment remains untested due to difficulties in studying this stratum[21] and in discriminating between biogenic and lithogenic

76 PFe within sinking particles[7]. Here, we overcame both limitations by targeting regions

77 dominated by biogenic, a mix of biogenic/lithogenic, or lithogenic sinking PFe, and applying

78 a novel *in situ* particle interceptor/incubator – RESPIRE[22] – to iron biogeochemistry to

79 concurrently elucidate the roles of biogenic and lithogenic PFe and their fate within the upper

80 mesopelagic zone (~100-200 m depth).

81

82 **Contrasting biogeochemical areas** – Mesopelagic Fe cycling was studied using a trace

83 metal-clean version of RESPIRE (TM-RESPIRE; Supplementary-Fig. 1) on a surface-

84 tethered free-drifting mooring to non-intrusively intercept settling particles, and then

85 immediately (i.e., at the end of the 1-2 d collection period) incubate them within this device at

86 *in situ* pressure and temperature. RESPIRE[22] provides rates of remineralization by particle-

87 attached bacteria based on an oxygen consumption time-series (Fig. 1a-b). TM-RESPIRE

88 provides the PFe/P/C sinking fluxes and associated DFe/P/C replenishment rates along with

89 concurrent release of iron-binding ligands. In doing so, the TM-RESPIRE approach (along

90 with subsequent analysis to estimate scavenging) combines all three previous approaches and

91 permits the investigation *in situ* of the different processes driving the Fe remineralization and

92 their interplay.

93

94 RESPIRE and TM-RESPIRE were deployed (vertically separated by <20 m; Methods) at 1 or

95 2 depths in the upper mesopelagic (110-200 m depth) during GEOTRACES process studies in

96 the Subantarctic Zone (SAZ) and Mediterranean Sea (Supplementary-Table 1; Methods).

97 These sites span wide-ranging dust deposition and productivity regimes (Fig. 1c). At SAZ,

98 High-Nitrate-Low-Chlorophyll (HNLC) surface waters are characterized by low dust

99 deposition and biologically-limiting DFe levels[24]. In contrast, Fe-rich oligotrophic waters of

100 the Eastern Mediterranean (ION site) encounter intense Saharan dust deposition event[23]. In

101   addition, a site within the Algerian basin (Western Mediterranean, ALG site) was selected

102   since it had intermediate characteristics between these end-members (Fig. 1c).

103

104   These site-specific characteristics were reflected in the widely-differing particulate organic

105   carbon (POC) and PFe sinking fluxes (Fig. 2; Supplementary-Table 1). SAZ had low POC

106   fluxes (0.9-1.2 mmol m^{-2} d^{-1}), but in contrast to the seasonal biological trends evident in Fig.

107   1c, ~4-fold higher POC fluxes were measured at ION and ALG relative to SAZ, suggesting

108   that a higher proportion of productivity was exported at these higher dust sites. This trend was

109   probably due to lithogenic ballasting which is often associated with a high proportion of the

110   POC export in the Mediterranean Sea[25,26]. PFe fluxes at the Mediterranean sites were also

111   ~15-fold higher (8.0-12.5 µmol m^{-2} d^{-1} at 115 m depth) relative to the Subantarctic. This

112   order-of-magnitude difference was driven by dust-derived lithogenic material

113   (Supplementary-Fig. 2b and f), a major constituent of Mediterranean sinking fluxes[26] but

114   negligible at SAZ[27]. Indeed, the Fe/C ratio of the sinking particles, a proxy of the

115   biogenic:lithogenic PFe (Methods), ranged >30-fold (~460-13970 µmol/mol) confirming the

116   SAZ site as the low lithogenic Fe end member (Fig. 3b-c). At ALG and ION, contrasting POC

117   and PFe flux attenuation patterns were evident, with PFe increasing with depth while a 3-fold

118   attenuation of POC flux occurred between 115-195 m depth (Fig. 2). Consequently, a

119   decrease in the relative proportion of biogenic PFe occurred over this stratum, evident by

120   increased Fe/C ratios (Fig. 3b-c).

121

122   **Mesopelagic bacterial remineralization** – Oxygen-based remineralization rates were

123   correlated linearly to the POC concentrations within the RESPIRE (i.e., the intercepted POC

124   flux; Fig. 3a). The absence of significant relationship(s) with other flux characteristics

125   (Supplementary-Fig. 2a-d), reveals that POC exerted a first-order control on remineralization

126 rates. To assess whether DFe replenishment is largely driven by microbial solubilization of

127 biogenic Fe in settling particles, we investigated trends in Fe regeneration efficiency ($R_{Fe/O2}$).

128 Both site-specific and water-column processes contributed to the wide range of $R_{Fe/O2}$ (Fig.

129 3b). High $R_{Fe/O2}$ efficiencies were observed at the SAZ site (148-421 µmol/mol) but were one-

130 to two-orders of magnitude lower at both Mediterranean sites. These trends in $R_{Fe/O2}$ are

131 consistent with vertical DFe stocks increasing with depth at SAZ, and in contrast decreasing

132 at the Mediterranean sites (Supplementary-Fig. 3). Similarly, replenishment rates of P and C

133 were highest at SAZ (Supplementary-Table 2). Although the present study offers a snapshot

134 of the annual cycle of mesopelagic remineralization, our observations are consistent with

135 high-latitude studies characterized as sites with relatively labile particles prone to microbially-

136 mediated remineralization[28,29].

137

138 Despite the complex transformations that characterize the internal cycle of Fe, a strong

139 inverse relationship is observed between $R_{Fe/O2}$ and the composition of the PFe flux estimated

140 from the particulate Fe/C ratio (Fig. 3b). Critically, this inverse relationship reveals that $R_{Fe/O2}$

141 is not determined by the magnitude of the PFe flux but rather by its particle composition

142 (Supplementary-Fig. 2g-h). Furthermore, this negative relationship strongly suggests that

143 biogenic PFe is efficiently regenerated while the dissolution of lithogenic PFe (predominant

144 at the Mediterranean sites; Fig. 2) takes place at much lower rates, corroborating 1D model

145 simulations showing that biogenic and lithogenic PFe fluxes exhibit distinctly different

146 vertical attenuation[9].

147

148 **Drivers of mesopelagic iron remineralization** – To develop a better understanding of the

149 drivers of mesopelagic iron biogeochemistry, $R_{Fe/O2}$ was converted into a Fe/C regeneration

150 ratio ($R_{Fe/C}$; Methods) and compared with the Fe/C and biogenic Fe/C (Fe_{bio}/C)

151   stoichiometries of the intercepted particles (Fig. 3c). A positive linear relationship (i.e.,

152   similar biogenic flux and regenerative stoichiometries) should be observed if bacterial

153   solubilization exerts a first-order control on DFe resupply. Here, the absence of such a

154   relationship, along with systematically lower DFe replenishment rates relative to P and C

155   (Supplementary-Table 2), confirm that DFe resupply results from a combination of biotic and

156   abiotic transformations of sinking PFe[15,17]. As highlighted in Fig. 3c, dissolution of lithogenic

157   Fe (i.e., DFe release without O_2 consumption) increases $R_{Fe/O2}$ and hence $R_{Fe/C}$, whereas Fe

158   scavenging (i.e., DFe removal without O_2 consumption) has the opposite effect. As expected

159   for areas dominated by biogenic PFe such as the subantarctic, bacterial solubilization

160   explained most of the $R_{Fe/C}$ relative to that of scavenging (Fig. 3c). The relatively large excess

161   in post-incubation Fe-binding ligands (L*) observed at SAZ (Fig. 4), driven by bacterial

162   degradation of biogenic-dominated sinking particles[9,16], is consistent with these low

163   scavenging rates pointing to complex interplay between processes associated with biogenic

164   and lithogenic PFe.

165

166   In contrast, a pronounced mismatch was observed between $R_{Fe/C}$ and the biogenic flux

167   stoichiometry at ALG at the deeper depth (195 m) and both depths at ION (Fig. 3c). Although

168   this trend results from the dominance of particle scavenging over solubilization/dissolution

169   (Fig. 4), the increase of this mismatch with depth may be explained by increasing scavenging

170   and/or decreasing solubilization/dissolution. Saharan dust-derived Fe dissolution rates are

171   reported to remain constant for several days in lab-studies[30], arguing for a relatively constant

172   dissolution rate over this 115-195 m stratum that particles will sink through on this

173   timescale[7,9]. Although changes in the bacterial solubilization rate of PFe, over this depth

174   range, cannot be directly assessed, the increasing proportion of C respired with depth

175   (Supplementary-Table 2) is not consistent with decreased bacterial solubilization rate of PFe

176   (by assuming constant or increasing biogenic Fe/C ratio[15]). Thus, increased scavenging (Fig.

177   4), is the most likely mechanism to account for the trend in $R_{Fe/C}$ with depth.

178

179   Measurements of size and concentration of particles collected at the Mediterranean sites

180   revealed decreasing cumulative particle volume concentrations with depth (Supplementary-

181   Table 4), excluding these two parameters as possible drivers of increased scavenging rate.

182   Therefore, the increase with depth in the relative proportion of lithogenic material being

183   exported (Supplementary-Fig. 2b and f), and the resulting decrease in the release of Fe-

184   binding ligands (Supplementary-Table 3), are most likely jointly responsible for this shift

185   within the upper mesopelagic (115-195 m; ALG/ION) stratum in the balance between

186   remineralization of biogenic PFe (DFe and ligand release), dissolution of lithogenic PFe (DFe

187   release) and scavenging processes. These findings demonstrate the confounding role played

188   by dust-derived lithogenic particles, conventionally viewed as a major pelagic DFe source[2,5],

189   but shown here in the upper 100-200 m stratum to act primarily as a scavenging-modulated

190   sink for DFe, and as a ballasting agent[25,31,32], each influencing where in the water column DFe

191   is replenished.

192

193   **Vertical resetting of mesopelagic biogeochemical conditions** – Figure 4 summarises how

194   the interplay of biogenic and lithogenic processes establish iron biogeochemical conditions at

195   each site, and importantly reset conditions with depth. Biogenic PFe is the main source of

196   DFe replenishment via bacterial solubilization, and this is also the case for the concurrent

197   release of iron-binding ligands. By combining the observations from TM-RESPIRE with

198   dissolution rate and scavenging ratios derived from prior dust-addition studies[18,30], it is

199   evident that lithogenic PFe contributes ~10 times more to Fe scavenging relative to Fe

200   dissolution in the upper mesopelagic (Fig. 4) supporting our observations in Fig. 3c. It is

201 important to note that high scavenging rates were reported even when L^* was (slightly)

202 positive (Fig. 3c and 4). It is probable that the higher the value of L^* the less likelihood of

203 pronounced scavenging, however, this trend, of scavenging when L^* is positive, points to an

204 unknown or unexplored mechanism (e.g., sorption of complexed Fe, colloidal aggregation).

205 These contrasting biogenic and lithogenic roles point to the importance of the composition of

206 the PFe flux in setting the Fe remineralization length-scale, with high PFe attenuation at the

207 biogenic end-member site, and in contrast an increase in PFe sinking flux at the lithogenic

208 end-member site.

209

210 Conceptually, depth-dependant changes in the relative proportion of biogenic vs. lithogenic

211 PFe, driven by their different attenuation length-scales, result in an ongoing resetting of the

212 biogeochemical conditions (e.g., L^*, ballasting via changes to the specific gravity of the

213 particle) while particles are settling in the water column (Fig. 4). The vertical trend observed

214 at the ALG site – characterized by intermediate surface characteristics relative to the two

215 other sites (Fig. 1c) – illustrates this dynamic situation. At 115 m depth, bacterial degradation

216 of heterogenous (i.e., biogenic/lithogenic) particles resulted in a relatively high post-

217 incubation L^* (Fig. 4) and a low scavenging rate (Fig. 3c and 4). However, 80 m deeper,

218 alteration of the composition of the PFe flux resulted in a decrease in post-incubation L^* (Fig.

219 4; Supplementary-Table 3) and a marked increase in the scavenging rate (Fig. 3c and 4). A

220 similar increase in scavenging with depth was observed at the ION site (Fig. 4). The dynamic

221 interplay between biogenically- and lithogenically-modulated mechanisms explain the

222 unexpected high spatial (and by analogy, temporal) and vertical variability in the PFe

223 remineralization reported in this study, but not so far captured in biogeochemical models[7].

224

225 We report a ~thousand-fold range in the DFe replenishment rate, while only modest changes

226 in P and C remineralization rates occurred at all sites (Supplementary-Table 2). At the SAZ,

227 the decoupling between Fe and C remineralization was relatively low and comparable to that

228 reported for sinking diatoms in subtropical waters[15]. In contrast, a pronounced decoupling

229 between Fe and both C and P remineralization was observed at ALG and ION. Critically, the

230 lithogenic component of the sinking flux, by having virtually no influence on C and P

231 remineralization, amplifies the decoupling between the Fe and both the C and P

232 remineralization length-scales. This finding highlights that the multi-faceted effects of particle

233 composition and dynamics on mesopelagic iron recycling needs to be considered in global

234 biogeochemical models, to better explain the spatial variability in the decoupling of nutrient

235 recycling.

236

237 **Controls on the global iron distribution** – The most straightforward route to examining the

238 broader role of biogenic:lithogenic particle composition on iron distributions was to focus on

239 the links to lithogenic iron, as dust supply is a well-established component of iron

240 biogeochemical models[4]. The modelling simulations exploited the observed Fe/O$_2$

241 relationship with dust (Fig. 3b-c) to develop a simple first-order parameterization that

242 captures observed links between Fe/O$_2$ and dust (Fig. 5a). Thus, an additional modulator of Fe

243 remineralization, based on the atmospherically-derived lithogenic particle concentration, was

244 added to the PISCES model parameterization, which enabled the sensitivity of iron

245 biogeochemistry to the impact of lithogenic particles on mesopelagic DFe cycling to be

246 addressed (Methods).

247

248 This simulation was employed to assess the wider implications of the multi-faceted roles of

249 lithogenic Fe on ocean iron cycling. Projections of the upper mesopelagic (100-250 m) DFe

250 inventory, in this amended simulation, decreased by 23.7% relative to the control run (Fig.

251 5b) and as expected, this trend was especially marked in areas influenced by high dust

252 deposition (i.e., North Atlantic, North Pacific, and Indian Oceans). In parallel, the DFe

253 inventory between 1000-1250 m depth increased by 6.4% across the global ocean (Fig. 5b).

254 Ultimately, the effect of lithogenic particles on PFe remineralization is to deepen the vertical

255 profile of DFe and therefore drive the replenishment of DFe deeper in the water column (Fig.

256 5c). This significant redistribution of DFe over the upper 1000 m has important ramifications

257 when considering that this depth stratum is heavily influenced by mode and intermediate

258 waters lateral transport[33], which can then alter DFe supply to different ocean regions and the

259 associated primary production (Supplementary-Fig. 4b). Therefore, the vertical redistribution

260 of iron in regions dominated by settling lithogenic particles may be pronounced when changes

261 in dust delivery to the ocean[34] may be accompanied by altered ocean circulation in the coming

262 decades[5].

263

264 Our findings suggest that predicted changes in dust inputs[34], by altering the

265 biogenic:lithogenic composition of the sinking particle assemblage, may impact the

266 replenishment of the subsurface DFe inventory and vertical supply of DFe not only in dusty

267 regions, but across the global ocean. This alteration, across many oceanic provinces, of DFe

268 resupply may in return have profound effects on the carbon sequestration efficiency of the

269 biological pump[35]. Our study enhances the wider understanding of the role of dust in altering

270 particle composition which in turn influences the replenishment of the subsurface DFe

271 inventory. This research points to the need for further studies on the internal cycling of trace

272 metals if we are to fully understand how they are returned to surface waters via their

273 biogeochemical cycles.

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References

1. Moore, C. M. *et al*. Processes and patterns of oceanic nutrient limitation. *Nat. Geosci.* **6**, 701-710 (2013).

2. Jickells, T. D. *et al*. Global iron connections between desert dust, ocean biogeochemistry, and climate. *Science* **308**, 67-71 (2005).

3. Guieu, C., *et al*. Iron from a submarine source impacts the productive layer of the Western Tropical South Pacific (WTSP). *Sci. Rep.* **8**, 9075 (2018).

4. Tagliabue, A. *et al.* How well do global ocean biogeochemistry models simulate dissolved iron distributions? *Glob. Biogeochem. Cycles* **30**, 149-174 (2016).

5. Tagliabue, A. *et al*. The integral role of iron in ocean biogeochemistry. *Nature* **543**, 51-59 (2017).

6. GEOTRACES Planning Group. GEOTRACES Science Plan. Baltimore, Maryland: Scientific Committee on Oceanic Research (2006).

7. Boyd, P. W., Ellwood, M. J., Tagliabue, A., & Twining, B. S. Biotic and abiotic retention, recycling and remineralization of metals in the ocean. *Nat. Geosci.* **10**, 167-173 (2017).

8. Tagliabue, A. *et al.* Surface-water iron supplies in the Southern Ocean sustained by deep winter mixing. *Nat. Geosci.* **7**, 314-320 (2014).

9. Boyd, P. W., Ibisanmi, E., Sander, S. G., Hunter, K. A., & Jackson, G. A. Remineralization of upper ocean particles: Implications for iron biogeochemistry. *Limnol. Oceanogr.* **55**, 1271-1288 (2010).

10. Ye, Y., Wagener, T., Völker, C., Guieu, C., & Wolf-Gladrow, D. A.. Dust deposition: iron source or sink? A case study. *Biogeosciences* **8**, 2107-2124 (2011).

11. Völker, C., & Tagliabue, A. Modeling organic iron-binding ligands in a three-dimensional biogeochemical ocean model. *Mar. Chem.* **173**, 67-77 (2015).

299   12. Frew, R. D. *et al.* Particulate iron dynamics during FeCycle in subantarctic waters

300   southeast of New Zealand. *Glob. Biogeochem. Cycles* **20**, GB1S93 (2006).

301   13. Lamborg, C. H., Buesseler, K. O., & Lam, P. J. Sinking fluxes of minor and trace

302   elements in the North Pacific Ocean measured during the VERTIGO program. *Deep Sea*

303   *Res. Part II* **55**, 1564-1577 (2008).

304   14. Ellwood, M. J. *et al.* Pelagic iron cycling during the subtropical spring bloom, east of

305   New Zealand. *Mar. Chem.* **160**, 18-33 (2014).

306   15. Twining, B. S. *et al.* Differential remineralization of major and trace elements in sinking

307   diatoms. *Limnol. Oceanogr.* **59**, 689-704 (2014).

308   16. Velasquez, I. B. *et al*. Ferrioxamine siderophores detected amongst iron binding ligands

309   produced during the remineralization of marine particles. *Front. Mar. Sci.* **3**, 172 (2016).

310   17. Boyd, P. W., & Ellwood, M. J. The biogeochemical cycle of iron in the ocean. *Nat.*

311   *Geosci.* **3**, 675-682 (2010).

312   18. Wagener, T., Guieu, C., & Leblond, N. Effects of dust deposition on iron cycle in the

313   surface Mediterranean Sea: results from a mesocosm seeding experiment. *Biogeosciences*

314   **7**, 3769-3781 (2010).

315   19. Bressac, M., & Guieu, C. Post-depositional processes: What really happens to new

316   atmospheric iron in the ocean's surface? *Glob. Biogeochem. Cycles* **27**, 859-870 (2013).

317   20. Wuttig, K. *et al*. Impacts of dust deposition on dissolved trace metal concentrations (Mn,

318   Al and Fe) during a mesocosm experiment. *Biogeosciences* **10**, 2583-2600 (2013).

319   21. Burd, A. B. *et al*. Assessing the apparent imbalance between geochemical and

320   biochemical indicators of meso-and bathypelagic biological activity: What the @$#! is

321   wrong with present calculations of carbon budgets? *Deep Sea Res. Part II* **57**, 1557-1571

322   (2010).

323 22. Boyd, P. W., McDonnell, A., Valdez, J., Lefèvre, D., & Gall, M. P. RESPIRE: An in situ

324   particle interceptor to conduct particle remineralization and microbial dynamics studies in

325   the oceans' Twilight Zone. *Limnol. Oceanogr. Methods* **13**, 494-508 (2015).

326 23. Albani, S. *et al.* Improved dust representation in the Community Atmosphere Model. *J.*

327   *Adv. Model Earth Sy.* **6**, 541-570 (2014).

328 24. Sedwick, P. N., *et al.* Limitation of algal growth by iron deficiency in the Australian

329   Subantarctic region. *Geophys. Res. Lett.* **26**, 2865-2868 (1999).

330 25. Bressac, M. *et al*. Quantification of the lithogenic carbon pump following a simulated

331   dust-deposition event in large mesocosms. *Biogeosciences* **11**, 1007-1020 (2014).

332 26. Ternon, E. *et al.* The impact of Saharan dust on the particulate export in the water column

333   of the North Western Mediterranean Sea. *Biogeosciences* **7**, 809-826 (2010).

334 27. Trull, T. W., Bray, S. G., Manganini, S. J., Honjo, S., & Francois, R. Moored sediment

335   trap measurements of carbon export in the Subantarctic and Polar Frontal Zones of the

336   Southern Ocean, south of Australia. *J. Geophys. Res.* **106**, 31489-31509 (2001).

337 28. Francois, R., Honjo, S., Krishfield, R., & Manganini, S. Factors controlling the flux of

338   organic carbon to the bathypelagic zone of the ocean. *Glob. Biogeochem. Cycles* **16**, 34-1

339   (2002).

340 29. Henson, S. A., Sanders, R., & Madsen, E. Global patterns in efficiency of particulate

341   organic carbon export and transfer to the deep ocean. *Glob. Biogeochem. Cycles* **26**,

342   GB1028 (2012).

343 30. Wagener, T., Pulido-Villena, E., & Guieu, C. Dust iron dissolution in seawater: Results

344   from a one-year time-series in the Mediterranean Sea. *Geophys. Res. Lett.* **35**, L16601

345   (2008).

346   31. Louis, J., Pedrotti, M. L., Gazeau, F., & Guieu, C. Experimental evidence of formation of

347   Transparent Exopolymer Particles (TEP) and POC export provoked by dust addition

348   under current and high pCO$_2$ conditions. *PloS one* **12**, e0171980 (2017).

349   32. Pabortsava, K. *et al.* Carbon sequestration in the deep Atlantic enhanced by Saharan dust.

350   *Nat. Geosci.* **10**, 189 (2017).

351   33. Hanawa, K. & Talley, L. D. in *Ocean Circulation and Climate*. Eds Siedler, G. & Church,

352   J., Academic, San Diego, 373–386 (2001).

353   34. Mahowald, N. M., & Luo, C. A less dusty future?. *Geophys. Res. Lett.* **30**, 1903 (2003).

354   35. Blain, S. *et al.* Effect of natural iron fertilization on carbon sequestration in the Southern

355   Ocean. *Nature* **446**, 1070-1074 (2007).

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**Figure 1. *In situ* particle remineralization measurements at contrasting biogeochemical
sites. a**, Conceptual alteration of particles within the (TM-)RESPIRE. Sinking particles are
intercepted for ~36h and regularly transferred into the inner chamber. Next, particles are
immediately incubated for >24h at *in situ* pressure and temperature during which bacterial
solubilization (i.e., diminished green particles) releases elements. See Methods for chemical
assays, and corrections applied to obtain pre-incubation particulate concentrations. **b**, Typical
oxygen optode time-series measured within RESPIRE. Bacterially-mediated remineralization
is derived from ΔO_{2max} and Δt (i.e., the time-period where the slope of the linear regression is
maximum). **c**, Weekly chlorophyll climatology from Modis-Aqua (green; 2003–2017; 1°x1°

396   resolution) and monthly climatology of simulated dust deposition (brown; 0.9°x1.25°

397   resolution)[23] in the subantarctic (SAZ: upper-panel), and Mediterranean Sea (ALG: middle;

398   ION: lower-panel). Grey vertical bars are (TM-)RESPIRE sampling periods.

399

400   **Figure 2. Summary of downward particle fluxes and composition from (TM-)RESPIRE**

401   **at SAZ (green), ALG (red) and ION (blue).** Coloured solid circles denote POC fluxes

402   (mmol m^{-2} d^{-1}; averaged (TM-)RESPIRE fluxes) at each depth (left-hand downward arrow,

403   corresponding to the centre of each circle). PFe fluxes (μmol m^{-2} d^{-1}) are represented by

404   coloured circle rims, and the partitioning of the PFe flux by the proportion of grey

405   (lithogenic) and white (biogenic) within each circle (Methods). All fluxes were corrected for

406   bacterial remineralization during the incubation (Methods). The circle size represents the flux

407   magnitude (Supplementary-Table 1) and are enlarged 10-fold (inside the dashed rectangle) at

408   the low-flux SAZ site. At SAZ, fluxes are the mean of two successive deployments (same

409   depth).

410

411   **Figure 3. Bacterial particle remineralization and iron regeneration efficiency at SAZ**

412   **(green), ALG (red) and ION (blue). a**, Remineralization versus POC concentrations

413   (corrected for remineralization during the incubation) measured within RESPIRE at 115 m

414   (triangles), 160 m (circles), and 195 m (diamonds). The best-fit of the linear model is plotted.

415   **b**, Regeneration efficiencies ($R_{Fe/O2}$) versus the Fe/C molar ratios of intercepted particles (a

416   proxy for PFe flux composition, higher ratios have more PFe_{litho}; Methods). The best-fit of the

417   power law model is presented. **c**, Fe/C regeneration ratios ($R_{Fe/C}$), obtained from $R_{Fe/O2}$ and a

418     $C:O_2$ conversion factor (Methods), versus the bulk (open symbols) and biogenic (Fe_{bio}/C;

419   Methods; closed symbols) Fe/C molar ratios. To aid interpretation, Fe-specific processes

420   (vertical arrows) in relation to $R_{Fe/C}$ and Fe_{bio}/C are displayed. The grey triangle illustrates the

17

421   increasing proportion of lithogenic PFe across sites and with depth. Error bars were derived

422   using uncertainty-propagation laws (Methods). Fluxes were expressed as concentrations to

423   permit cross-comparison between sites (different collection/incubation times employed).

424

425   **Figure 4. Synthesis of key processes that together set the PFe remineralization length-**

426   **scale expressed as a function of the relative proportion of sinking biogenic:lithogenic**

427   **PFe.** Sites and depths are assigned into each of three idealized categories: biogenic-dominated

428   (SAZ), heterogeneous (biogenic/lithogenic; ALG 115 m), and lithogenic-dominated (ALG

429   195 m, ION 115-195 m) PFe fluxes. Note, how intercepted sinking particles at ALG shift

430   categories with depth. For PFe attenuation, + and − denote the magnitude of the decrease or

431   increase in the flux with depth, respectively. Note, the dynamic nature of concurrent ligand

432   release and scavenging as particles settle, means that L*>0 may not impede scavenging (see

433   Fig. 3c). Nevertheless, the higher the value of L* the less likelihood of pronounced

434   scavenging.

435

436   **Figure 5. Results of model simulations using lithogenic particle-dependent modulation of**

437   **iron remineralization. a**, Relationship between $R_{Fe/O2}$ and lithogenic PFe concentrations in

438   offshore regions (>3000 m depth; 30°S-30°N) employing different lithogenic particle-

439   dependent modulators of iron remineralization (kd; Methods). **b**, Change in DFe inventory

440   vertically integrated over 100-250 m (upper-panel) and 1000-1250 m (lower-panel) depth

441   strata relative to the control run. This simulation is based on a 500-year simulation employing

442   a kd of 0.1 µg m^{-3} (i.e., which reproduced the inverse relationship observed between $R_{Fe/O2}$

443   and the lithogenic PFe concentration in panel **a**). **c**, Alteration of the global mean vertical DFe

444   profile from a simulation employing a kd of 0.1 µg m^{-3} relative to the control run.

445

METHODS

Site selection – Datasets were acquired during two GEOTRACES process studies, in the Subantarctic Zone (SAZ) southwest of Tasmania, aboard the *RV Investigator* (March 2017; SOTS project), and in the central (Ionian Sea, ION) and western (Algerian Basin, ALG) Mediterranean Sea aboard the *RV Pourquoi Pas?* (May/June 2017; Peacetime project). Sites were selected for their contrasting magnitude and composition of the downward particle flux. In particular, ~30-fold higher lithogenic fluxes have been reported at ~1000 m depth at ALG (12.7 g m^{-2} yr^{-1})[36] and ION (13.9 g m^{-2} yr^{-1})[37], relative to SAZ (0.4 g m^{-2} yr^{-1})[27]. In contrast, POC fluxes at ~1000 m depth are relatively similar at SAZ (1.1-1.4 g m^{-2} yr^{-1})[38], ALG (1.4-1.7 g m^{-2} yr^{-1})[36], and ION (0.7-0.9 g m^{-2} yr^{-1})[37].

The SAZ represents >50% of the areal extent of the ice-free Southern Ocean. This HNLC area has both low silicate and DFe concentrations year round[24] along with moderate phytoplankton biomass[39]. The low dust flux to this area originates primarily from Australia, with the highest fluxes between October and March[40]. POC (0.91-1.23 mmol m^{-2} d^{-1}) and PFe (0.48-0.67 μmol m^{-2} d^{-1}) fluxes measured in this study compared well with POC (3.34 ± 1.81 mmol m^{-2} d^{-1})[41] and PFe (0.17 ± 0.09 μmol m^{-2} d^{-1})[42] fluxes measured at the same site and depth, but in January/February.

The Mediterranean Sea has a west-to-east gradient of increasing oligotrophy, with a deficiency of phosphorus and nitrogen. Relatively weak winter convection[43] prevents efficient uplift of nutrients to the surface waters. Deposition of Saharan desert dust, characterized by strong variability and dominated by extreme events[44], constitutes the main source of new nutrients. ION is an ultra-oligotrophic area, while the ALG is one of the most productive areas in the Mediterranean Sea[45]. At ALG/ION, POC fluxes measured at 195 m depth (1.14-1.67 mmol m^{-2} d^{-1}) are consistent with previous measurements (0.4-3.0 mmol m^{-2} d^{-1}, 1st/3rd quartiles)[46], but ~4-fold higher than fluxes simultaneously measured at 250 m depth with a

471   PPS5 sediment trap (0.32-0.37 mmol m^{-2} d^{-1}; N. Leblond, pers. comm.). Similarly, 2-fold

472   higher PFe fluxes were collected at 195 m depth with the TM-RESPIRE (13.4-16.1 µmol m^{-2}

473   d^{-1}) compared to PFe fluxes at 250 m depth (PPS5; 4.5-8.5 µmol m^{-2} d^{-1}; N. Leblond, pers.

474   comm.).

475

476   **(TM-)RESPIRE** – The conceptual view, functioning and potential artefacts of the RESPIRE

477   particle interceptor are detailed in ref.[22]. TM-RESPIRE, a trace metal-clean version of

478   RESPIRE, was developed to quantify PFe remineralization rates. TM-RESPIRE is

479   constructed from polycarbonate (PC) except for the PVC indented rotating sphere (IRS), and

480   has identical dimensions to RESPIRE (Supplementary-Fig. 1). The IRS excludes

481   mesozooplankton from the incubation chamber, and delivers particles every ~10 min into this

482   chamber. When the IRS is not rotating, the chamber beneath it is completely closed, avoiding

483   any exchange with the upper part of TM-RESPIRE. Trace metal cleanliness and optode-based

484   oxygen measurements are not reconcilable since plastic material essential for trace elements

485   studies often have high oxygen permeability, and optodes require a metal window. To

486   circumvent these issues, TM-RESPIRE was systematically deployed concurrently with the

487   RESPIRE fitted with an Aanderaa 3830 oxygen optode (resolution < 1 µM; accuracy < 5

488   µM).

489   Optodes were post-calibrated using the Winkler method, and oxygen time-series were

490   corrected for pressure and salinity (Aanderaa operating manual). A pressure sensor (RBR,

491   Canada) was deployed alongside RESPIRE to determine the deployment depth. The vertical

492   distance between the TM-RESPIRE and RESPIRE (10-20 m; Supplementary-Table 1) was

493   constrained by the ship's dimensions (i.e., distance above the waterline) and the necessity to

494   keep a minimum distance between the two traps to avoid contamination through contact with

495   the vessels' hull or propeller wash during deployment/recovery.

496   Remineralization rates measured within the RESPIRE were assumed to be comparable in the

497   TM-RESPIRE. This assumption is supported by similar rates obtained at SAZ after two

498   successive deployments several days apart (5.1 ± 0.2 mmol O_2 m^{-3} d^{-1}; Fig. 3a), and by

499   previous replicate RESPIRE measurements exhibiting small variability in the remineralization

500   rates[47]. In the present study, POC fluxes from RESPIRE and TM-RESPIRE (vertically

501   separated by 10-20 m) varied on average by a factor of 1.8 ± 0.6 (n = 6). Similar POC flux

502   variabilities were observed at ALG/ION (250 m depth) and SAZ (150 m depth) across

503   individual 12-24 h samples obtained by conventional sediment traps which varied on average

504   by 1.6 ± 0.8 (N. Leblond, pers. comm.) and 3[41], respectively. This relatively high variability

505   in successive upper mesopelagic POC flux measurements suggests that the differences

506   observed in the present study is driven largely by inherent variability in the POC flux in this

507   stratum[48], rather than by the vertical spacing between the RESPIRE and TM-RESPIRE.

508

509   **(TM-)RESPIRE procedure** – RESPIRE were cleaned following ref.[22]. Before each voyage,

510   TM-RESPIRE were soaked in 2% Neutracon (7d), 2 M HCl (reagent grade; 30d), 1.2 M HCl

511   (TM grade; 7d), and copiously rinsed with Ultrapure water. Before deployment, each TM-

512   RESPIRE was filled overnight with 0.12 M HCl (TM grade), rinsed with Ultrapure water and

513   pre-conditioned with low-Fe filtered seawater to remove all trace of acid. Several hours

514   before deployment, RESPIRE and TM-RESPIRE were filled with filtered seawater collected

515   at the deployment depth. Clean polyethylene (PE) bags covering the traps were removed just

516   prior to deployment. Upon recovery, RESPIRE and TM-RESPIRE were immediately covered

517   with PE bags. TM-RESPIRE's were transferred into a class-100 clean laboratory. Seawater

518   above the IRS was siphoned off using clean Teflon PFA tubing, and the incubation chamber

519   was sampled via a Teflon PFA stopcock valve. The absence of mesozooplankton in the

520   chamber was confirmed by visual inspection. Samples were split into equal fractions and

521   subsamples for DFe, Fe-binding ligands, and nutrients were filtered through acid-cleaned 0.2-

522   µm PC membranes. DFe samples were stored in low density PE (LDPE) bottles and acidified

523   to pH 1.8 (quartz-distilled HCl). Fe-binding ligand samples were transferred to high density

524   PE   bottles   (cleaned   following   GEOTRACES   protocols,

525   http://www.geotraces.org/science/science-highlight/intercalibration/222-sampling-and-

526   sample-handling-protocols-for-geotraces-cruises), double-bagged and stored at -20°C. 0.2-µm

527   PC membranes were dried under a laminar flow hood and used for particulate trace element

528   analysis. Subsamples for POC were obtained by filtration onto pre-combusted 13-mm QMA

529   or GF/F filters. RESPIRE was sampled for POC and nutrients. All steps were performed

530   within 2-3 hours of recovery of (TM-)RESPIRE.

531   TM-RESPIRE procedural blank measurements were performed onboard during voyages.

532   Blanks comprised incubating <0.2-µm seawater from 150 m depth within the TM-RESPIRE.

533   After 72 h, the blank samples were subjected to identical processing protocols (i.e.,

534   subsampling, filtration) as for deployments. The average DFe blank (DFe_{blank}; 0.38 ± 0.03

535   nM) was used to correct DFe release (see Calculation of metrics). However, we acknowledge

536   that deployment and recovery are two steps with high risk of contamination. Concentrations

537   in trace elements other than Fe (such as Zn) measured within the TM-RESPIRE were used to

538   assess possible contamination (not necessarily conspicuous with DFe) that could have

539   occurred during deployment/recovery. This approach allowed us to reject one contaminated

540   deployment (Subantarctic voyage, April 2016).

541

542   **Water column sampling** – Samples were collected using a Titanium Rosette mounted with

543   Teflon-coated 12 L Niskin (SAZ) or Go-Flo bottles (ALG/ION) deployed on a Kevlar cable.

544   After recovery, bottles were transferred inside a class-100 clean laboratory container.

545   Seawater samples were directly filtered from the bottles through acid-cleaned 0.2-µm capsule

546   filters (Sartorius Sartobran-P-capsule 0.45/0.2-µm). DFe samples were stored in LDPE bottles

547   and acidified to pH 1.8 (quartz-distilled HCl), while nutrient samples were analyzed at sea.

548   Suspended particulate trace elements were sampled using in situ pumps (McLane; acid-

549   cleaned 1-µm PC membranes) at SAZ and pressurized Go-Flo bottles at ALG/ION (acid-

550   cleaned 25-mm diameter Supor 0.45-µm polyethersulfone filters; 4.8 L on average) following

551   GEOTRACES recommendations.

552

553   **Analytical methods** – DFe concentrations were measured by flow injection with online

554   preconcentration and chemiluminescence detection[3,49]. An internal acidified seawater

555   standard was measured daily to control the stability of the analysis. During the analysis of

556   TM-RESPIRE and water-column samples, the detection limit was 15 pM and the accuracy of

557   the method was controlled by analyzing the SAFe S (0.086 ± 0.010 nmol/kg (n = 3);

558   consensus value 0.093 ± 0.008 nmol/kg), SAFe D1 (0.64 ± 0.13 nmol/kg (n = 19); consensus

559   value 0.67 ± 0.04 nmol/kg), GD (1.04 ± 0.10 nmol/kg (n = 10); consensus value 1.00 ± 0.10

560   nmol/kg), and GSC (1.37 ± 0.16 nmol/kg (n = 4); consensus value not available) seawater

561   standards.

562   Dissolved nutrients were analysed onboard with a segmented flow analyser (AAIII HR Seal

563   Analytical; detection limits were 0.02 µM for P, 0.05 µM for N, and 0.08 µM for Si) [50,51].

564   Iron organic speciation measurements were performed using CLE-CSV with 2-(2-

565   Thiazolylazo)-p-cresol (TAC) as the competing ligand[52]. Reagent blanks for Fe were

566   undetectable and the detection limit for ligand concentrations was calculated as 3 times the

567   standard deviation of the concentrations (ranging from 0.09-0.99 nM and always lower than

568   the concentration).

569   Particulate trace element samples were digested (10% HF/50% HNO_3 (v/v)) following the

570   protocol described in the "GEOTRACES Cookbook" and ref.[53]. Procedural blanks consisted

571 of unused acid-cleaned filters. Analyses were performed on a high resolution ICP-MS

572 (Element XR, Thermo-Fisher Scientific). The accuracy of the measurements was established

573 using a range of Certified Reference Materials, including MESS-4. The recoveries in these

574 reference materials were 80-130% for iron.

575 POC samples were acidified overnight with 2 M HCl to remove inorganic C, and then dried at

576 60°C for 2 d. Samples were analyzed on a CHN analyser (Thermo Finnigan EA 1112 Series

577 Flash Elemental Analyser).

578

579 **Calculations of metrics** – Mesozooplankton, free-living and particle-attached heterotrophic

580 bacteria drive mesopelagic remineralization[54]. By excluding mesozooplankton (using the IRS)

581 and boosting particle-attached bacterial abundances relative to free-living bacteria (through

582 particle interception), the measured remineralization rates were dominated by the particle-

583 attached microbial assemblage[22].

584 **Particle remineralization** was calculated as follows: the pre-incubation O_2 concentration

585 (Fig. 1b) was subtracted from each data-point obtained during the incubation, and the sign

586 reversed. The slope of the linear regression between this remineralization signature and the

587 time elapsed corresponds to the particle remineralization rate. Note that the decrease in O_2

588 was not systematically linear, and a plateau can be attained toward the end of the incubation

589 (Fig. 1b). The explanation for this trend remains unclear, but may be related to particle

590 containment, a shift in the microbial community[55], and/or altered organic matter lability[22].

591 When such a plateau was evident, remineralization was calculated over the time period where

592 the decrease in O_2 was maximum (Fig. 1b). Since remineralization rates were assumed to be

593 comparable within the RESPIRE and TM-RESPIRE, the error in remineralization rates was

594 calculated by propagating the uncertainty from the slope of the linear regression and the

595 relative standard deviation of the POC fluxes collected by the RESPIRE/TM-RESPIRE. The

596   error in ΔO_2 (see below) was calculated in the same way.

597   **The iron regeneration efficiency** ($R_{Fe/O2}$; μmol/mol; Fig. 3b) was calculated as: $R_{Fe/O2} =$

598   $\Delta DFe / \Delta O_2$, where $\Delta DFe = DFe_{post-incubation} - (DFe_{initial} + DFe_{blank})$, and $\Delta O_2 = O_{2\ pre-incubation} -$

599   $O_{2\ post-incubation}$ (initial, pre-incubation, and post-incubation terms are illustrated in Fig. 1a). The

600   error in $R_{Fe/O2}$ was calculated by propagating the uncertainties from ΔDFe and ΔO_2.

601   **The Fe/C regeneration ratio** ($R_{Fe/C}$; μmol/mol; Fig. 3c) was calculated as: $R_{Fe/C} = R_{Fe/O2} /$

602   0.69, where 0.69 is a $C:O_2$ conversion factor (i.e., $117/170$)[56]. Note that different conversion

603   factors can be used to convert the oxygen-based remineralization rate to carbon (discussed in

604   detail in ref.[22]).

605   **Particulate fluxes were expressed as concentrations** to permit cross-comparison between

606   sites in which different collection times were employed (Supplementary-Table 1). **Pre-**

607   **incubation particulate Fe, P, and OC concentrations** correspond to the sum of the post-

608   incubation particulate concentration, plus the concentration of the respective element released

609   into the dissolved phase (i.e., ΔDFe, ΔPO_4 (= $PO_{4\ post-incubation} - PO_{4\ initial}$), and ΔOC (= $\Delta O_2 \times$

610   0.69), respectively).

611   **The replenishment rate of Fe** (% d^{-1}; Supplementary-Table 2) was calculated as: ($\Delta DFe /$

612   $PFe_{pre-incubation}$) $\times$ 100 / Δt, (Δt corresponds to the incubation time). The error was calculated

613   by propagating the uncertainties from ΔDFe and $PFe_{pre-incubation}$. P and C replenishment rates

614   were calculated in the same way.

615   **Quantification of the lithogenic and biogenic fractions of sinking PFe** has large

616   uncertainties. Twining et al.[15] reported a 2.3 and 4.4-fold increase in the Fe/P and Fe/S (proxy

617   of Fe/C) ratios of sinking diatom by 200 m depth, respectively, highlighting difficulties in

618   estimating the biogenic fraction of sinking PFe from surface cell quotas. An alternative

619   approach is to estimate the lithogenic fraction of PFe by using the Fe/Al ratio. However, the

620   present study encompasses different regions where lithogenic material has different Fe and Al

621 compositions. In addition, dissolution/scavenging of Fe and Al differ during particle

622 settling[60], altering their pre-depositional Fe/Al ratio. Thus, lithogenic PFe estimated at

623 ALG/ION using a Saharan dust end-member Fe/Al ratio[59] is systematically higher than total

624 PFe. In contrast, suspended particles collected at ALG/ION had a Fe/Al ratio comparable to

625 the crustal Fe/Al ratio[58] (Supplementary-Fig. 3). Thus, the crustal Fe/Al molar ratio was used

626 to estimate biogenic and lithogenic PFe components in Fig. 2 and 3c. For the remainder of the

627 study, the PFe/POC molar ratio (hereafter termed Fe/C) of the intercepted particles (Fig. 3)

628 was used as a proxy for the composition of the PFe flux. In using this approach, we consider

629 POC and PFe as proxies of biogenic (algal/detrital) and lithogenic PFe, respectively[9]. This

630 ratio increases when the relative proportion of lithogenic Fe increases, and vice-versa. We

631 acknowledge that the biogenic Fe/C ratio differs between Fe-limited and Fe-replete areas,

632 however, we believe that this approach is relatively robust when considering such contrasting

633 sites.

634

635 **Ancillary biogeochemical information –** Surface chlorophyll-*a* was derived from MODIS-

636 Aqua. The 4 km resolution eight-day composite images were averaged over the 2003-2017

637 period (due to extensive subantarctic cloud-cover) for a 1°x1° box centered at each site. At

638 SAZ, MODIS chlorophyll-*a* concentrations were corrected using an improved regional

639 algorithm[61].

640 Monthly estimates of total (wet + dry) dust deposition annually averaged were obtained using

641 an atmospheric model (CAM4-BAM, case C4fn)[23] run for 30 years, of which we considered

642 the last 10 years, with a spatial resolution of 0.9°x1.25°.

643

644 **The PISCES biogeochemical model experiment –** PISCES[62,63] is a relatively complex

645 general ocean circulation and biogeochemistry model with two PFe pools (large and small)

26

646   characterized by different sinking rates, and two analogous POC size classes sourced from the

647   'mortality' of organic Fe and C pools. Uncomplexed DFe is subjected to scavenging, while

648   colloidal iron undergoes coagulation losses, both of which augment two PFe size classes.

649   Scavenging rate depends on the particle abundance, including the lithogenic PFe pool (based

650   on the dust input at the surface and a simple sinking speed). Subsurface dissolution of

651   lithogenic PFe occurs with a ~500 m length-scale, a sinking rate of 2 m d^{-1}, and a reduced

652   solubility. PFe remineralization takes into account changes in particle size and lability due to

653   bacterial solubilization via reactivity continuum[64].

654   We conducted a range of different simulations with PISCES aimed at addressing the first-

655   order influence of lithogenic particles (i.e., dust) on $R_{Fe/O2}$. Building this parameterization on

656   dust has many advantages. Indeed, biogenic Fe is a complex pool (detritus/algal)

657   characterized by different cell quotas[65], while dust supply is a well-established component[4]

658   that does not require explicit modelling of extra pools and can be incorporated more widely

659   into models. In the model, $R_{Fe/O2}$ is derived by dividing the annually integrated

660   remineralization flux of iron from PFe by the O_2 consumption during remineralization at each

661   model grid cell. We conducted a set of sensitivity experiments where PFe remineralization

662   was modulated by the lithogenic particle concentration at each model grid cell. This

663   modulator (M, an unitless quantity) is a function of the lithogenic particle concentration: M =

664   1 – [lithogenic particles / (lithogenic particles + kd)]. M has a range of different sensitivities

665   (kd = 0.1, 1, 5 and 10 µg m^{-3}). To avoid double accounting, Fe scavenging by dust included in

666   PISCES[62] was switched off. We then ran for 500 years experiments as well as a 500-year

667   model run in which PFe remineralization was left unchanged (Supplementary-Fig. 4a). We

668   then compared the range of $R_{Fe/O2}$ outputs from each experiment with the local lithogenic PFe

669   concentrations (Fig. 5a), which showed that kd = 0.1 µg m^{-3} was the most realistic as it

670   reproduced the inverse relationship observed between $R_{Fe/O2}$ and the lithogenic PFe

671   concentrations (Fig. 3b). To minimize the impact of high iron inputs (e.g., near shelves), only

672   offshore regions (>3000 m depth; 30°S-30°N) were considered. Finally, the impact of this

673   new parameterization on DFe inventories, vertical DFe distribution, and surface iron-driven

674   processes was investigated.

675

676   **Exploration of caveats** – Our approach comes with some caveats, most of which were

677   identified and discussed in ref.[22]. Here, we discuss potential biases that may have affected Fe

678   cycling within the TM-RESPIRE. Specifically, we acknowledge that the pre-concentration of

679   particles in a 1.6 L chamber, along with the widely-differing downward fluxes (Fig. 2), may

680   have influenced the DFe replenishment rates.

681   The potential effects of this experimental bias, related to the need of concentrating particles

682   and to the contrasting sites, were tested via *in vitro* incubation experiments (Supplementary-

683   Fig. 5). The rationale, method, and conclusions are described in the Figure caption. Results

684   from these experiments suggest that during the <48 h incubation within the TM-RESPIRE,

685   the limited loss of DFe by adsorption onto walls/particles was not significantly influenced by

686   the particle concentration, and by the surface-area-to-volume (SA:Vol) ratio of the incubation

687   bottles (0.31-0.59 cm^{-1}). This later finding allows us to conclude that a different SA:Vol ratio

688   of the TM-RESPIRE incubation chamber (0.57 cm^{-1}) would not have changed our

689   conclusions.

690   In our study, the concomitant decreases in DFe replenishment and particle volume

691   concentrations observed with depth at ALG/ION (Supplementary-Tables 2 and 4) do not

692   support a particle concentration effect on the replenishment of Fe. This experimental bias, that

693   would act to lower DFe concentration, has mainly been observed during dust dissolution

694   experiments[66]. This effect is likely offset in incubations with mixed biogenic/lithogenic

695   particles (along with the associated bacterial communities) by the concurrent release of Fe-

696 binding ligands (which increased on average by a factor >2 in our study; Supplementary-

697 Table 3). Consistent with the conclusions of this study, the particle composition and/or

698 bacterial communities, rather than experimental parameters, appear to be the primary control

699 of the DFe replenishment rate in these *in vitro* experiments.

700

701 **Data availability** – Modis Chl-a concentrations (ALG and ION sites) were obtained with the

702 Giovanni online data system, developed and maintained by the NASA GES DISC. MODIS

703 Chl-a concentrations corrected using an improved regional algorithm for the Southern Ocean

704 (SAZ site) are publicly available via the Australian Integrated Marine Observing System

705 (IMOS) Ocean Portal (www.imos.org.au). Following publication, the dataset generated and

706 analysed during the current study (mostly available in the Supplementary Information section)

707 will be made available (i.e. open access) through the IMAS/UTAS data portal

708 (http://www.imas.utas.edu.au/data).

709

710 **Code availability** – The NEMO-PISCES model we use in this work is freely available

711 (http://www.nemo-ocean.eu/)   under   the   CeCILL   free   software   licence

712 (http://www.cecill.info/index.en.html). We used a modified version of the PISCES

713 biogeochemical model. These modifications concern the representation of the particulate iron

714 remineralization and this is not yet present in the freely available NEMO release but will be

715 provided upon contacting A.T.
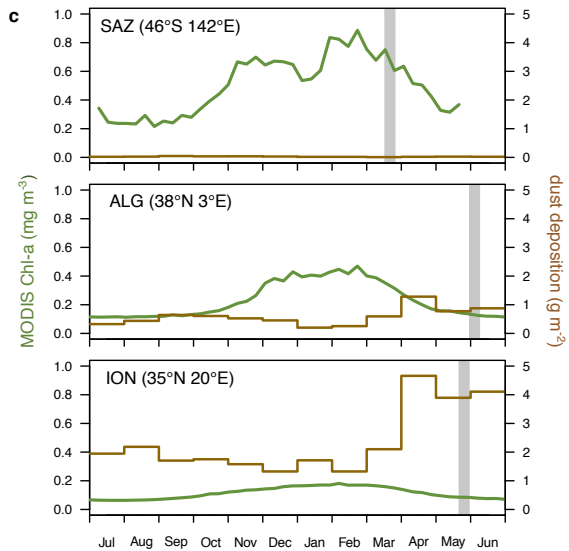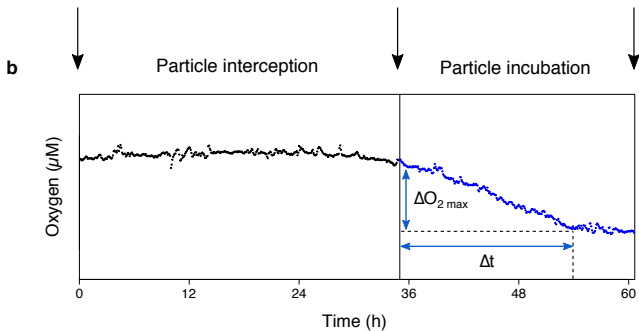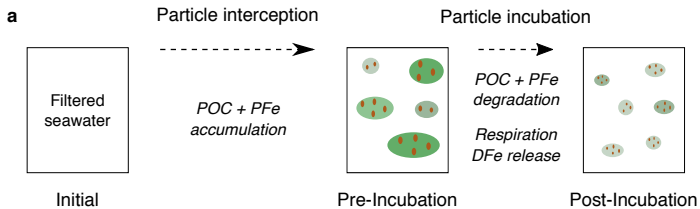
716

717 **References**

718 36. Zúñiga, D., et al. Particulate organic carbon budget in the open Algero-Balearic Basin

719   (Western Mediterranean): Assessment from a one-year sediment trap experiment. *Deep

720   Sea Res. Part I* **54**, 1530-1548 (2007).

721 37. Stavrakakis, S., *et al*. Downward fluxes of sinking particulate matter in the deep Ionian

722 Sea (NESTOR site), eastern Mediterranean: seasonal and interannual variability.

723 *Biogeosciences* **10**, 7235 (2013).

724 38. Rigual-Hernández, A. S., Trull, T. W., Bray, S. G., & Armand, L. K. The fate of diatom

725 valves in the Subantarctic and Polar Frontal Zones of the Southern Ocean: Sediment trap

726 versus surface sediment assemblages. *Palaeogeogr. Palaeoclimatol. Palaeoecol.* **457**,

727 129-143 (2016).

728 39. Eriksen, R., *et al*. Seasonal succession of phytoplankton community structure from

729 autonomous sampling at the Australian Southern Ocean Time Series (SOTS) observatory.

730 *Mar. Ecol. Prog. Ser.* **589**, 13-31 (2018).

731 40. Li, F., Ginoux, P., & Ramaswamy, V. Distribution, transport, and deposition of mineral

732 dust in the Southern Ocean and Antarctica: Contribution of major sources. *J. Geophys.*

733 *Res.* **113**, D10207 (2008).

734 41. Ebersbach, F., Trull, T. W., Davies, D. M., & Bray, S. G. Controls on mesopelagic

735 particle fluxes in the Sub-Antarctic and Polar Frontal Zones in the Southern Ocean south

736 of Australia in summer—Perspectives from free-drifting sediment traps. *Deep Sea Res.*

737 *Part II* **58**, 2260-2276 (2011).

738 42. Bowie, A. R. *et al*. Biogeochemical iron budgets of the Southern Ocean south of

739 Australia: Decoupling of iron and nutrient cycles in the subantarctic zone by the

740 summertime supply. *Glob. Biogeochem. Cycles*, **23**, GB4034 (2009).

741 43. D'Ortenzio, F., *et al.* Seasonal variability of the mixed layer depth in the Mediterranean

742 Sea as derived from in situ profiles. *Geophys. Res. Lett.* **32**, L12605 (2005).

743 44. Guerzoni, S., *et al*. The role of atmospheric deposition in the biogeochemistry of the

744 Mediterranean Sea. *Progr. Oceanogr.* **44**, 147-190 (1999).

745   45. D'Ortenzio, F., & Ribera d'Alcalà, M. On the trophic regimes of the Mediterranean Sea: a

746   satellite analysis. *Biogeosciences* **6**, 139-148 (2009).

747   46. Ramondenc, S., et al. An initial carbon export assessment in the Mediterranean Sea based

748   on drifting sediment traps and the Underwater Vision Profiler data sets. *Deep Sea Res.*

749   *Part I* **117**, 107-119 (2016).

750   47. McDonnell, A. M., Boyd, P. W., & Buesseler, K. O. Effects of sinking velocities and

751   microbial respiration rates on the attenuation of particulate carbon fluxes through the

752   mesopelagic zone. *Glob. Biogeochem. Cycles* **29**, 175-193 (2015).

753   48. Lamborg, C. H., *et al*. The flux of bio-and lithogenic material associated with sinking

754   particles in the mesopelagic "twilight zone" of the northwest and North Central Pacific

755   Ocean. *Deep Sea Res. Part II* **55**, 1540-1563 (2008).

756   49. Bonnet, S., & Guieu, C. Atmospheric forcing on the annual iron cycle in the western

757   Mediterranean Sea: A 1-year survey. *J. Geophys. Res. Oceans* **111**, C09010 (2006).

758   50. Rees, C., *et al*. Methods for reproducible shipboard SFA nutrient measurement using

759   RMNS and automated data processing. *Limnol. Oceanogr. Methods* **17**, 25-41 (2019).

760   51. Aminot, A., & Kérouel, R. Dosage automatique des nutriments dans les eaux marines:

761   méthodes en flux continu, Ed. Ifremer, Méthodes d'analyse en milieu marin (2007).

762   52. Croot, P. L. & Johansson, M. Determination of Iron Speciation by Cathodic Stripping

763   Voltammetry in Seawater Using the Competing Ligand 2-(2-Thiazolylazo)-p-cresol

764   (TAC). *Electroanalysis* **12**, 565-576 (2000).

765   53. Planquette, H., & Sherrell, R. M. Sampling for particulate trace element determination

766   using water sampling bottles: methodology and comparison to in situ pumps. *Limnol.*

767   *Oceanogr. Methods* **10**, 367-388 (2012).

768   54. Steinberg, D. K., *et al*. Bacterial vs. zooplankton control of sinking particle flux in the

769   ocean's twilight zone. *Limnol. Oceanogr.* **53**, 1327-1338 (2008).

770   55. LeCleir, G. R., DeBruyn, J. M., Maas, E. W., Boyd, P. W., & Wilhelm, S. W. Temporal

771   changes in particle-associated microbial communities after interception by nonlethal

772   sediment traps. *FEMS microbiol. Ecol.* **87**, 153-163 (2014).

773   56. Anderson, L. A., & Sarmiento, J. L. Redfield ratios of remineralization determined by

774   nutrient data analysis. *Glob. Biogeochem. Cycles* **8**, 65-80 (1994).

775   57. Reinthaler, T., *et al*. Prokaryotic respiration and production in the meso-and bathypelagic

776   realm of the eastern and western North Atlantic basin. *Limnol. Oceanogr.* **51**, 1262-1273

777   (2006).

778   58. Wedepohl, K. H. The composition of the continental crust. *Geochim. Cosmochim. Acta*

779   **59**, 1217-1232 (1995).

780   59. Guieu, C., Loÿe-Pilot, M. D., Ridame, C., & Thomas, C. Chemical characterization of the

781   Saharan dust end-member: Some biogeochemical implications for the western

782   Mediterranean Sea. *J. Geophys. Res.* **107**, ACH-5 (2002).

783   60. Baker, A. R., Adams, C., Bell, T. G., Jickells, T. D., & Ganzeveld, L. Estimation of

784   atmospheric nutrient inputs to the Atlantic Ocean from 50° N to 50° S based on large-

785   scale field sampling: Iron and other dust-associated elements. *Glob. Biogeoch. Cycles* **27**,

786   755-767 (2013).

787   61. Johnson, R., Strutton, P. G., Wright, S. W., McMinn, A., & Meiners, K. M. Three

788   improved satellite chlorophyll algorithms for the Southern Ocean. *J. Geophys. Res.* **118**,

789   3694-3703 (2013).

790   62. Aumont, O., Ethé, C., Tagliabue, A., Bopp, L., & Gehlen M. PISCES-v2: an ocean

791   biogeochemical model for carbon and ecosystem studies, *Geoscientific Model*

792   *Development*, **8**, 2465-2513 (2015).

793   63. Völker, C., & Tagliabue, A. Modeling organic iron-binding ligands in a three-dimensional

794   biogeochemical ocean model, *Mar. Chem.*, **173**, 67-77 (2015).
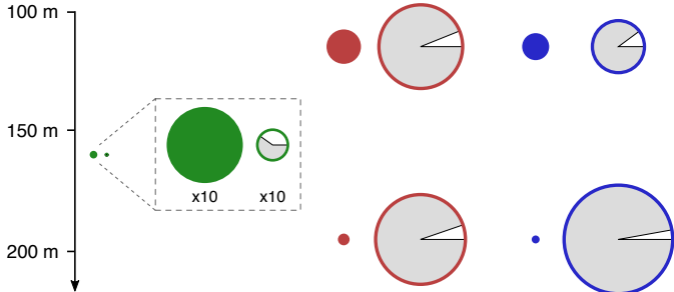
795   64. Aumont, O., *et al*. Variable reactivity of particulate organic matter in a global ocean

796   biogeochemical model, *Biogeosciences*, **14**, 2321-2341 (2017).

797   65. Boyd, P. W. *et al*. Why are biotic iron pools uniform across high-and low-iron pelagic

798   ecosystems?. *Glob. Biogeoch. Cycles* **29**, 1028-1043 (2015).

799   66. Bonnet, S., & Guieu, C. Dissolution of atmospheric iron in seawater. *Geophys. Res. Lett.*

800   **31**, L03303 (2004).

a Particle interception → Particle incubation

Filtered seawater — *POC + PFe accumulation* → — *POC + PFe degradation* / *Respiration DFe release* →

Initial — Pre-Incubation — Post-Incubation

b Particle interception | Particle incubation

Oxygen (μM) vs Time (h)

$\Delta O_{2\,max}$

Δt

c
SAZ (46°S 142°E)

ALG (38°N 3°E)

ION (35°N 20°E)

MODIS Chl-a (mg m^{-3})

dust deposition (g m^{-2})

Jul Aug Sep Oct Nov Dec Jan Feb Mar Apr May Jun

a

$y = 0.13x + 5.05$
$R^2 = 0.93; p < 0.01$

Remineralization rate (mmol O_2 m^{-3} d^{-1})

POC (μM)

b

$y = 5.39x^{-1.1}$
RSE = 0.29; p = 0.007

$R_{Fe/O2}$ (μmol Fe/mol O_2)

Fe/C ($\mu mol/mol$)

c

microbial solubilisation

abiotic dissolution

scavenging

$R_{Fe/C}$ ($\mu mol/mol$)

Fe/C and Fe_{bio}/C ($\mu mol/mol$)

	SAZ	ALG 115 m	→	ALG 195 m	ION 115-195 m
DFe release (μmol Fe/mol O$_2$)	150-420	68	→	6	22-10
DFe$_{litho}$ [1] (% of ΔDFe)	0.01%	0.5%	→	17%	2-9%
Post-incubation L* [2]	0.43	0.25	→	0.18	0.21-0.12
Potential lithogenic scavenging [3] (% of ΔDFe)	0.3%	5%	→	170%	30-130%
PFe attenuation [4]	(+++)		(-)		(- - -)

(1) Estimates based on a dissolution rate of 0.018 nmol mg^{-1} dust d^{-1}[30]
(2) L* corresponds to the excess of ligands over DFe after the incubation phase
(3) Estimates based on a scavenging ratio of DFe on dust of 0.37 nmol mg^{-1}[18]
(4) At SAZ, estimate based on the balance between DFe/L release and scavenging

a

R_{FeO2} (µmol/mol)

Control
kd = 5.0 µg m⁻³
kd = 1.0 µg m⁻³
kd = 0.1 µg m⁻³

Ocean lithogenic PFe (mg m⁻³)

b Change in DFe inventory (µmol m⁻²)

100–250 m

1000–1250 m

c

Depth (m)

Relative change in DFe