Fatty acid ester metabolites of gymnodimine in shellfish collected from China and in mussels (*Mytilus* galloprovincialis) exposed to Karenia selliformis

Ji Ying ^{1, 2}, Che Yijia ^{1, 2}, Wright Elliott J. ³, McCarron Pearse ³, Hess Philipp ⁴, Li Aifeng ^{1, 2, *}

¹ College of Environmental Science and Engineering, Ocean University of China, Qingdao 266100, China

² Key Laboratory of Marine Environment and Ecology, Ocean University of China, Ministry of Education, Qingdao 266100, China

³ Biotoxin Metrology, National Research Council Canada, 1411 Oxford Street, Halifax, Nova Scotia B3H 3Z1, Canada

⁴ Ifremer, Phycotoxins Laboratory, Nantes, France

* Corresponding author : Aifeng Li, email address : lafouc@ouc.edu.cn

Abstract :

Marine shellfish exposed to the microalgae Karenia selliformis can accumulate gymnodimines (GYM). Shellfish samples collected from Beihai City in Guangxi Autonomous Region, and Ningde City in Fujian Province, in the South China Sea, as well as mussels Mytilus galloprovincialis fed on K. selliformis under laboratory conditions were analyzed. Gymnodimines and various fatty acid ester metabolites were detected in the clam Antigona lamellaris and pen shell Atrina pectinata, while no esters were found in the oyster Crassostrea sp. and the gastropod Batillaria zonalis despite positive detection of free GYM in both species. When present, the predominant acyl esters observed were 18:0-GYM-A and 20:1-GYM-A. Under laboratory conditions GYM-A was accumulated and metabolized to fatty acid esters in mussels exposed to K. selliformis, with 16:0-GYM-A and 20:1-GYM-A as the major variants. A novel compound with the same accurate mass as GYM-A and its 16:0 fatty acid ester were observed in the experimental mussels but was not present in the microalgal strain to which mussels were exposed. No significant differences of reactive oxygen species (ROS) levels and antioxidant enzymes were found between mussels fed on K. selliformis or GYM-free microalgae Isochrysis galbana. This suggests the accumulation of GYM and its metabolites does not significantly impact the physiological status of mussels. While it is currently not proven that GYM affects human health, risk assessments should consider the presence of GYM esters in naturally contaminated shellfish as part of exposure analysis.

Graphical abstract



Highlights

► GYM-A was detected in four species of field marine mollusks from Chinese coast. ► Diverse fatty acid esters of GYM-A were found in clam and pen shell, but not in oyster and gastropod. ► GYM-A and its different fatty acid esters accumulated in mussels fed with *Karenia selliformis*. ► The esters 16:0-GYM-A and 20:1-GYM-A were the major components in the experimental mussels. ► A novel compound with the same accurate mass as GYM-A and its esters were also observed in mussels.

Keywords : Gymnodimines (GYM), Fatty acid ester, *Karenia selliformis*, Toxin profiles, Antioxidant enzymes

56 **1. Introduction**

Marine bivalves have the capacity to accumulate phycotoxins produced by 57 harmful algae and generate a diverse range of secondary metabolites, collectively 58 known as shellfish toxins (Salas and Clarke, 2019). These toxins can be divided into 59 eight groups, including saxitoxins, azaspiracids, brevetoxins, cyclic imines, domoic 60 acids, okadaic acids, pectenotoxins, and yessotoxins (Toyofuku, 2006). Gymnodimines 61 (GYM) (Fig. 1) are a group of fast-acting cyclic imine toxins initially identified in 62 oysters collected in New Zealand (Seki et al., 1995; Stewart et al., 1997). GYM-A, 63 64 GYM-B and GYM-C are produced by the dinoflagellate Karenia selliformis (Miles et al., 2000, 2003; Seki et al., 1995). 12-methyl-gymnodimine was subsequently verified 65 as natural product in the dinoflagellate Alexandrium ostenfeldii from estuaries on the 66 U.S. East Coast (Van Wagoner et al., 2011) and the Netherlands (Van de Waal et al., 67 68 2015). Additionally, A. ostenfeldii from the northern Baltic Sea was shown to produce a new analogue, GYM-D (Harju et al., 2016). Gymnodimines have also been 69 documented in shellfish from coastal regions of Tunisia (Biré et al., 2002; Marrouchi 70 71 et al., 2010), Australia (Takahashi et al., 2007), Europe & North America (Kharrat et al., 2008), South Africa (Krock et al., 2009) and China (Jiang et al., 2017; Li et al., 2015; 72 73 Liu et al., 2011a). A recent study demonstrated that fatty acid ester metabolites of GYM-A constitute the majority (>90%) of GYM toxins present in clam samples (Ruditapes 74 decussatus) from Tunisia (de la Iglesia et al., 2013). Several unknown GYM-like 75 compounds have been detected in shellfish (McCarron et al., 2014) and in A. ostenfeldii 76 by liquid chromatography-mass spectrometry (LC-MS) (Harju et al., 2016; Qiu et al., 77 2018). Considering the high intraperitoneal toxicity (LD₅₀ of 80-96 mg kg⁻¹) of GYM-78 79 A to mice, and albeit lower relative oral toxicity (Munday et al., 2004), these 80 metabolites should be considered when evaluating the risks of GYMs in shellfish for 81 human consumption.

Antioxidant enzyme activity in bivalves is influenced by reactive oxygen species (ROS) and may be an indication of organism health and oxidative stress (Box et al., 2007; Fernández et al., 2012; Gillis et al., 2014; Hu et al., 2015a; Regoli and Principato,

3

1995). Antioxidant responses are enhanced when the production of ROS increases, 85 resulting in oxidative stress (Hu et al., 2015b; Pan et al., 2006). The antioxidant and 86 87 immune systems of shellfish, including superoxide dismutase (SOD), catalase (CAT), glutathione peroxidase (GSH-Px), glutathione reductase (GR), glutathione (GSH) and 88 lysozyme (LZM), are responsible for eliminating ROS and foreign particles. Paralytic 89 90 shellfish toxins accumulated from toxic microalgae induced oxidative stress in mussels 91 (Mytilus galloprovincialis) and scallops (Patinopecten vessoensis), where ROS were 92 reduced by antioxidant enzymes under laboratory conditions (Qiu et al., 2013).

The objective of this work was to analyze the profiles of GYM and its fatty acid esters in various species of shellfish relevant to the coastal regions of China, with a specific focus on the South China Sea. Laboratory studies were conducted on mussels (*M. galloprovincialis*) fed with the GYM-producing microalgae *K. selliformis* to study the uptake and metabolism of GYM-A, and to monitor the response of antioxidant enzymes (SOD, CAT, GSH).

99 2. Materials and Methods

100 2.1 Reagents

Ammonium formate and formic acid were from Fisher Scientific (Fair Lawn, NJ,
USA) or Honeywell-Fluka (Oakville, ON, Canada). Acetonitrile, methanol,
monopotassium phosphate (KH₂PO₄) and disodium hydrogen phosphate (Na₂HPO₄)
were from Merck Ltd. (White-house Station, NJ, USA). A certified reference material
(CRM) of GYM-A was obtained from the National Research Council of Canada
(Halifax, NS, Canada). Milli-Q water was supplied by a Milli-Q water purification
system from Millipore Ltd. (Bedford, MA, USA) to 18.2 MΩ quality or better.

108 2.2 Culture of microalgae and sample preparation

109 *K. selliformis* (strain GM94GAB) used in the feeding experiment was isolated 110 from the Gulf of Gabes, Tunisia (Medhioub et al., 2009). The strain was maintained in 111 sterilized seawater filtered with 0.45 μ m mixed fiber membrane (Jinjing Ltd., China) in 112 photo-bioreactors (120 L) without aeration. The culture was enriched with f/2 medium 113 without silicate (Guillard and Ryther, 1962) at 18±2°C under 6000 lx light intensity with a 12 h: 12 h illumination cycle. Algal cells were counted by optical microscopy at
a magnification of 100 on a Sedgewick Rafter Counting Chamber and collected at the
stable growth stage to feed mussels.

117 *Isochrysis galbana* 3011 kept at Ocean University of China was used as a control 118 diet in the feeding experiment. The culture conditions of *I. galbana* were as above for 119 *K. selliformis*, without aeration. The batch of *I. galbana* was collected when the algal 120 density approached 10^8 cells mL⁻¹.

121 A freeze-dried pellet $(2.6 \times 10^6 \text{ cells})$ of the cultured *K. selliformis* was extracted 122 using a 1/16" microtip ultrasonicator probe with coupler (QSonica LLC (CT, USA) in 123 1 mL methanol using a cycle with 40% amplitude pulse mode for a total of 3 min. The 124 sample was held in an ice bath to prevent the solution from heating during extraction. 125 The sample was then filtered using a 0.45 µm PVDF centrifugal filter (Merck Millipore 126 Ltd.) prior to LC-MS analysis.

127 2.3 Shellfish samples

128 Multiple species of commercially grown shellfish were collected at shellfish 129 harvesting regions in the South China Sea in March 2016. This included clams 130 (Antigona lamellaris), pen shells (Atrina pectinata) and oysters (Crassostrea sp.) from Beihai City, Guangxi Autonomous Region, and gastropod (Batillaria zonalis) from 131 Ningde City, Fujian Province. Whole tissue samples (~1 g) were weighed and extracted 132 133 with 3 mL methanol on an ice using a superfine homogenizer (F6/10, Fluko) at speeds ranging from 5000 to 32000 rpm in three 30 s cycles. The supernatant was carefully 134 135 transferred to a 10 mL volumetric flask after centrifugation (2400 $\times g$ for 10 min at 4°C). The remaining pellet was re-extracted twice more as above and the combined 136 137 supernatants were adjusted to a final volume of 10 mL and passed through 0.22 µm polyamide nylon filters into HPLC vials and stored at -20°C until LC-MS analysis. 138

139 2.4 Feeding experiment and sample preparation

Mussels (*M. galloprovincialis*) were obtained from a marine aquaculture zone along the coast of Qingdao, China, in October 2016. They were transported alive to the laboratory and washed to remove epibionts. Mussels were randomly divided into two tanks (60 L) and acclimated in filtered seawater with continuous aeration at $16 \pm 2^{\circ}$ C for 3 days. During this period, seawater (without any feed) was renewed twice daily at 08:00 and 20:00.

After acclimation, mussels in the first tank (treatment group) were fed with 146 K. selliformis for four days and then starved for three days. They were maintained in a 147 60 L tank with fresh seawater containing the toxic microalgae at an initial density of 148 1000 cells mL⁻¹. The K. selliformis culture and fresh seawater were continuously 149 150 supplied for four days using a peristaltic pump to maintain the initial algal density. The volume of seawater in the tank was kept constant. Fresh seawater alone was added for 151 the three days of starvation. Acclimated mussels in the second tank were used as a blank 152 control. Experimental settings for the control group were the same as for the treatment 153 group, with the exception of the diet. The non-toxic microalga I. galbana was pumped 154 and maintained at a density of 10 000 cells mL⁻¹ for the first four days. Both tanks were 155 continuously aerated during the entire experiment. Mussels (n=10) were selected 156 randomly and dissected at various time points (0, 12, 24, 48, 72, 96, 120, 144, and 168 157 158 h). The muscle (containing foot and adductor muscle) and digestive gland of mussels were separated. Tissue samples $(1 \pm 0.1 \text{ g})$ of both treatment groups were prepared as 159 described in section 2.3. 160

For ROS determinations, tissue samples (0.5 g) were homogenized in 10 mL phosphate buffer (0.1 M, pH=7.0) containing 2.6 mM KH₂PO₄ and 4.1 mM Na₂HPO₄ for 3 mins on an ice-water bath. The supernatant from centrifugation (2400 ×g, 4°C, 20 min) was collected. To analyze activity of antioxidant enzymes (CAT, SOD, GSH) additional tissue samples (1 \pm 0.1 g) were homogenized and centrifuged in 9 mL physiological saline (0.9%) following the conditions as described above. The supernatant was collected.

168 2.5 LC-MS analysis for GYM and esters

The LC-MS method was modified from a previous study (Li et al., 2016). An
HPLC (Agilent 1290, Palo Alta, CA, USA) was coupled with a triple quadrupole mass
spectrometer (Agilent 6430, Palo Alta, CA, USA) equipped with an ESI interface. A

Luna C₁₈ column (50 \times 2.1 mm i.d., 3 µm, Phenomenex) was used at 35°C. Mobile 172 phases A (water) and B (95% acetonitrile) contained with 50 mM formic acid and 2 173 mM ammonium formate. A gradient at 300 µL min⁻¹ was run from 25% to 100% B over 174 7 min, held for 3 min and returned to 25%B over 1 min before re-equilibration for 3 175 min. Injection volume was set at 5 µL. Selected reaction monitoring (SRM) transitions 176 for all the acyl ester derivatives of GYM-A were described previously (de la Iglesia et 177 al., 2013). Product ion m/z 490.3, fragmentor energy 150 V, collision energy 30 V, and 178 179 dwell time 15 ms were used to scan fatty acid esters of GYM-A. Concentrations of GYM and metabolites of GYM-A were roughly estimated using CRM-GYM-A as a 180 reference material assuming equal molar responses. 181

182 2.6 LC-HRMS analysis for GYM and esters

183 An HPLC (Agilent 1200, Palo Alto, CA, USA), was coupled with high resolution mass spectrometer (Q-Exactive HF, Thermo Fisher Scientific, Bremen, Germany) 184 equipped with a heated electrospray ionization probe (HESI-II). A Luna C₁₈ column (50 185 \times 2 mm i.d., 2.5 μ m, Phenomenex) was used with mobile phase, flow rate, and 186 187 temperature the same as described in section 2.5. The gradient ran from 15% to 100% 188 B over 10 min, held for 15 min, and then returned to 15%B over 0.1 min before reequilibration for 7 min. Injection volumes were 1 µL. Mobile phase was diverted to 189 waste for the first 1.5 minutes of the run. Source conditions were spray voltage 3000, 190 191 capillary temperature 350°C, sheath gas and auxillary gas at 35 and 10 (arbitrary units) respectively with heat set to 300°C, and S-Lens RF Level at 50. Full scan data was 192 acquired in positive mode from m/z 400-900 using the 60 k resolution setting with an 193 194 automatic gain control (AGC) target 1e6 ions, 100 ms maximum injection time. MS/MS 195 spectra were collected through 'TopN' experiments. These TopN experiments used data-dependent acquisition with the top 10 most intense ions (not on an exclusion list 196 197 of background ions) selected for fragmentation from the full scan survey, with preferential fragmentation given to ions above an intensity threshold of 1.6e5 ions from 198 199 an inclusion list of known GYM and acyl ester variants. Apex trigger was set from 2-6 s with dynamic exclusion for 4 s in efforts to capture spectra near the apex of the 200

chromatographic peak. Spectra were collected at the 15 k resolution setting, 50 ms maximum injection time, an isolation window of 1 m/z about the parent mass with a collision energy spread of 30, 65 eV. Focused collection of free GYM spectra was done using parallel reaction monitoring (PRM) mode at the 30 k resolution setting, 2e5 AGC target, 100 ms maximum injection time, isolation window 0.4 m/z, and collision energy spread set at 40, 75 eV.

207 2.7 Analysis of antioxidant enzymes

The level of ROS and activities of antioxidant enzymes (SOD, CAT, GSH) in the digestive gland and muscle of mussels from both groups were analyzed using test kits for antioxidant enzyme biomarkers acquired from Nanjing Jiancheng Bioengineering Institute Ltd. (Nanjing, China) as described by (Liu, et al., 2011b; Ji et al., 2018a). The measuring principles are in accordance with specifications in kits.

213 **3. Results**

214 3.1 Profiles of GYM and esters in field shellfish collected from China

Varying concentrations of GYM-A was present in the four shellfish species tested. 215 Concentrations of GYM-A as well as the relative molar percentages of predominant 216 217 acyl esters in field mollusk samples are shown in Table 1. Acyl esters of GYM-A were found in the clam (Antigona lamellaris) and the pen shell (Atrina pectinata), while no 218 fatty acid esters were found in the oyster (Crassostrea sp.) and the gastropod (Batillaria 219 zonalis). The profile of GYM-A acyl esters observed varied between clams and pen 220 221 shell samples. Octadecanoic acid (18:0) ester was the most abundant (50%), followed by the eicosenoic acid (20:1) (23%) in the clam. While 20:1-GYM ester (23%), 18:0-222 223 GYM ester (20%), and the docosadienoic acid (22:2) (18%) ester dominated the profile in the pen shell sample. The extent of GYM-A esterification was estimated for clam 224 225 and pen shell samples assuming equimolar response for the esters against GYM-A standard, and were approximately 58% and 87% respectively. 226

227 3.2 GYM and esters in mussels from feeding experiment

228 LC-MS/MS SRM chromatograms for esters of GYM-A in digestive gland tissue of

mussels (*M. galloprovincialis*) fed with *K. selliformis* after 96 h are shown in Fig. 2. 229 Twenty-eight fatty acid esters (odd- and even-chains) were identified in the 230 experimental mussels. The K. selliformis used as feed microalga for mussels in this 231 study primarily produces only free GYM-A and a trace amount of GYM-B (Fig. S1A). 232 The mussels were free of GYM toxins before feeding on K. selliformis (Fig. S1C and 233 D). LC-HRMS full scan chromatograms for GYM-A CRM, 16:0 GYM-A positive 234 control tissue which was previously confirmed in using synthetic 16:0 GYM-A (Iglesia 235 236 et al., 2013), K. selliformis culture and experimental mussels are shown in Fig. 3. Product-ion spectra (MS/MS) (Fig. 4) of GYM-A in the extract of digestive gland 237 sample (treatment group-96 h), were acquired to confirm the accumulation and 238 esterification of GYM-A in mussels fed with K. selliformis in laboratory. An additional 239 GYM peak (4.66 min), not present in the microalgae, was also observed in the shellfish 240 with the same measured accurate mass as GYM (Fig. 3 and 4). This new GYM analogue 241 also formed esters (14.84 min) in the mussels (Fig. 3). 242

The concentrations of GYM-A and the relative amounts of acyl ester metabolites 243 244 varied in the muscle and digestive gland tissues of mussels (Fig. 5). The acyl esters accounted for an estimated 98% of the GYM total in the muscle and digestive glands. 245 Toxin concentrations increased during the algae-feeding period and stabilized during 246 the starvation period. The highest concentration in the muscle tissues was at 120 h and 247 96 h for the digestive gland tissues. The sum concentration of GYM and GYM esters 248 reached a maximum of \sim 760 nmol kg⁻¹ in the muscle tissues, while it was four times 249 higher in the digestive gland samples. The relative proportion of GYM-A and the major 250 ester metabolites (16:0, 18:0, 18:1, 20:1, 20:2, 22:2) differed in both tissues (Fig. 5). In 251 252 the muscle samples, the relative abundance of 16:0-GYM-A was dominant (33%) in the first 48 h but gradually declined to 14%, while the relative abundance of 20:1-253 GYM-A increased from 12% to 34%, and reached the highest level among ester 254 metabolites after 48 h. The relative proportions of acyl esters in the digestive gland 255 samples was consistent over the duration of the experiment with 16:0-GYM-A 256 constituting approximately 30% esters detected. 257

258

Molar concentrations of GYM-A in the muscle and digestive gland tissues are

shown in Fig. S2A. The esters 22:2-GYM-A, 20:1-GYM-A, 18:0-GYM-A and 16:0-GYM-A were most abundant for their respective carbon chain lengths (Fig. S2B-E). The nutritionally important ω -3-fatty acids, EPA (C20:5) and DHA (C22:6) abundant in mussels (Fernández-Reiriz et al., 1996), formed much less acyl esters compared to the four derivatives above. Trends of each metabolite over time in these four groups were similar to the variations of GYM-A in the digestive gland tissues.

265 3.3 Responses of Antioxidant System

The ROS levels from the control and treatment groups followed similar trends (Fig. S3). ROS levels reduced slightly when feeding on toxic or non-toxic microalgae at 12 h, and did not change significantly until 144 h. Variations of CAT activity, SOD activity and GSH level in muscle and digestive glands are also shown in Fig. S3. These enzyme activity levels exhibited similar trends, showing minor fluctuation without significant changes over the course of the study.

272 4. Discussion

273 Accumulation of marine phycotoxins in shellfish is an important seafood safety 274 issue due to frequent occurrence of toxic algal blooms in coastal areas worldwide (James et al., 2010). To properly assess the toxicity of shellfish exposed to harmful algal 275 species, esterification of lipophilic toxins such as okadaic acid, pectenotoxins, 276 277 spirolides and pinnatoxins has been studied (JØrgensen et al., 2005; Aasen et al., 2006; McCarron et al., 2012). The occurrence of GYM esters in shellfish highlights the need 278 for methods to quantitate total profiles of GYM. Gymnodimines degrade under alkaline 279 hydrolysis conditions and the enzymatic hydrolysis method still needs modification to 280 281 be applied to cyclic imines (de la Iglesia et al., 2013; Doucet et al., 2007). Currently direct analysis of acyl esters by LC-MS is the most practical option for GYM ester 282 283 measurement, however this is limited by availability of standards for GYM esters. In 284 this work, individual ester concentrations were estimated assuming equal response of 285 the esters and GYM-A. While total GYM esters measured were significant, these numbers can only be considered indicative. Monitoring lipophilic toxins in cultured 286 287 mussels (Mytilus coruscus) during a previous study (Li et al., 2015) showed low

amounts of GYM-A are common throughout the year in the mariculture zone of Gougi 288 Island in the East China Sea. A subsequent study on lipophilic shellfish toxins showed 289 higher concentrations of GYM-A were frequently detected in shellfish collected from 290 the South China Sea (Ji et al., 2018b). GYM-A was present in the four field mollusk 291 samples tested here. GYM-B, an analog of GYM-A, was first isolated and elucidated 292 from K. selliformis (Miles et al., 2000). No GYM-B was detected in field samples. 293 Trace amounts of GYM-B were present in the K. selliformis, and both the experimental 294 295 mussels and seawater collected at 96 h during the feeding experiment (Fig. S1A and B). 296 A positive detection of GYM-B in clams also occurred in a previous study (Naila et al., 297 2012).

Acyl ester metabolites have been observed for a variety of marine phycotoxins in 298 a range of shellfish species (Comesaña Losada et al., 1999; García et al., 2004; 299 JØrgensen et al., 2005; Torgersen et al., 2008a; 2008b; Wilkins et al., 2006). Based on 300 previous reports of acyl ester metabolites for other classes of cyclic imines including 301 spirolides (Aasen et al., 2006) and pinnatoxins (McCarron et al., 2012), and because of 302 303 the previous study showing GYM esters in Tunisian clams (Ruditapes decussatus) (de al Iglesia et al., 2013), the presence of GYM esters in Chinese shellfish is not 304 unexpected. In this study, concentration and profile of ester metabolites varied in clam 305 (Antigona lamellaris) and pen shell (Atrina pectinata) samples, but weren't detected in 306 oyster (Crassostrea sp.) or gastropod (B. zonalis) samples (Table 1), indicating that 307 GYM esterification in mollusks may be species-specific. Possibly some key enzymes 308 govern the metabolism of GYM esters in shellfish. The species specific profiles 309 observed in GYM esterification should be studied further. 310

The feeding experiment demonstrated the acylation process under controlled laboratory conditions. GYM-A and GYM derived esters analyzed in this work showed parent masses < 5 ppm, and 490-fragment masses < 3.5 ppm. The 392-fragment observed for GYM-A and its esters, was not present for GYM-like species (*m/z* 508) (Fig. 4). Various ester metabolites were confirmed in the experimental mussels based on characterized transitions and relative retention times of esters published before (Fig. 2). The high relative abundance (90%-98%) of esterified GYM at each sampling-time

point (Fig. 5), including the first time point at 12 h, indicates that these Phase-II 318 metabolites of GYM-A are formed very rapidly and that esterification constitutes a 319 major metabolic process for these toxins. Relatively static distribution of acyl esters 320 and GYM-A in digestive gland samples indicates GYM-A and its metabolites are 321 readily retained in shellfish, possibly explaining the persistence of GYM-A in oysters 322 (Seki et al., 1995) and clams (Naila et al., 2012). Considering the relative abundance of 323 different GYM-A esters in the digestive gland tissue, percentages of free GYM-A and 324 325 16:0 ester slightly decreased in the depuration period after 96 h, which demonstrated that they were excreted or transformed faster than 20:1 ester, 22:2 ester, and 18:0 ester. 326 This tendency is also observed for GYM-A and the 16:0 ester in muscle tissue in the 327 depuration period (data not shown). 328

It is interesting to note that the unidentified GYM analog observed in the mussels 329 330 after feeding, with the same accurate mass as GYM-A, also formed ester metabolites (Fig. 3 and 4). The origin of this GYM analog was not confirmed in this study, but it 331 was not observed in direct analysis of the toxin-producing K. selliformis by LC-HRMS. 332 333 The typical GYM-A product ions m/z 392 and 446 were absent in the MS/MS spectra of this new compound (Fig. 4), indicating a possible structural variation between C1-4 334 (Fig. 1). This compound may have been a product of metabolism in the mussels, or may 335 be an artefact from sample preparation or processing. LC-HRMS experiments did not 336 reveal the presence of any other significant GYM related compounds in the microalgae 337 (data not shown). The formation of this compound will be considered in further studies. 338 339 All mussels were alive during the feeding experiment. Similar to the response of scallops (Chlamys farreri) and mussels (M. galloprovincialis) exposed to azaspiracids 340 341 in our previous study (Ji et al., 2018), no significant difference of ROS levels and 342 antioxidant enzymes were found in mussels exposed to the GYM-producing microalgae. This indicates there is no measurable impact of GYM accumulation on the antioxidant 343 enzyme system of mussels based on this study. 344

345 **5.** Conclusions

346

GYM-A and ester metabolites were confirmed in the clam (Antigona lamellaris)

and pen shell (Atrina pectinata) samples collected from the South China Sea and in 347 experimental mussels (M. galloprovincialis) fed with K. selliformis under laboratory 348 conditions. While GYM-A was detected in the oyster (Crassostrea sp.) and gastropod 349 (B. zonalis) samples, no GYM-A esters were detected. Various acyl esters were detected 350 by SRM analysis and the predominant esters were confirmed by LC-HRMS. 351 Esterification and ester profiles appear to be species-specific for mollusks. In mussels, 352 ester metabolites constituted the majority of total GYMs (>90% for laboratory fed 353 354 samples), with the 18:0-GYM-A and 20:1-GYM-A formed as the most abundant ester derivatives in the field samples analyzed. K. selliformis feeding demonstrated no toxic 355 effects on mussels based on responses of antioxidant enzymes. A novel GYM analogue 356 with the same accurate mass as GYM-A, which also formed fatty acid ester, was found 357 in the experimental mussels. Risks posed by GYMs and their acyl esters should be 358 assessed in future toxicological studies and risk assessments. 359

360

361 **Conflict of interest**

362 All authors declare that there are no conflicts of interest.

363

364 Acknowledgement

This work was funded by the National Natural Science Foundation of China (41876112) and the Fundamental Research Funds for the Central Universities (201841003). The authors are grateful for the support of Florent Malo (Ifremer) for the provision of cell pellets of cultures of *K. selliformis*.

- 369
- 370
- 371
- 372
- 373
- 374
- 375

376 **References:**

- Aasen, J.A.B., Hardstaff, W., Aune, T., Quilliam, M.A. (2006). Discovery of fatty acid ester
 metabolites of spirolide toxins in mussels from Norway using liquid chromatography/tandem
 mass spectrometry. *Rapid Communications in Mass Spectrometry*, 20(10), 1531-1537. doi:
 10.1002/rcm.2501
- Biré, R., Krys, S., Frémy, J.M, Dragacci, S., Stirling, D., Kharrat, R. (2002). First evidence on
 occurrence of gymnodimine in clams from Tunisia. *Journal of Nature Toxins*, 4(11), 269-275
- Blanco, J., Mariño, C., Martín, H., Acosta, C.P. (2007). Anatomical distribution of diarrhetic
 shellfish poisoning (DSP) toxins in the mussel *Mytilus galloprovincialis*. *Toxicon*, 50(8), 10111018. doi: 10.1016/j.toxicon.2007.09.002
- Box, A., Sureda, A., Galgani, F., Pons, A., Deudero, S. (2007). Assessment of environmental
 pollution at Balearic Islands applying oxidative stress biomarkers in the mussel *Mytilus galloprovincialis. Comparative Biochemistry and Physiology Part C: Toxicology & Pharmacology*, 146(4), 531-539. doi: 10.1016/j.cbpc.2007.06.006
- Comesaña Losada, M., Leão, J.M., Gago-Martínez, A., Rodríguez Vázquez, J.A., Quilliam, M.A.
 (1999). Further Studies on the Analysis of DSP Toxin Profiles in Galician Mussels. *Journal of Agricultural and Food Chemistry*, 47(2), 618-621. doi: 10.1021/jf971043a
- de la Iglesia, P., McCarron, P., Diogène, J., Quilliam, M.A. (2013). Discovery of gymnodimine fatty
 acid ester metabolites in shellfish using liquid chromatography/mass spectrometry. *Rapid Communications in Mass Spectrometry*, 27(5), 643-653. doi: 10.1002/rcm.6491
- Doucet, E., Ross, N.N., Quilliam, M.A. (2007). Enzymatic hydrolysis of esterified diarrhetic
 shellfish poisoning toxins and pectenotoxins. *Analytical and Bioanalytical Chemistry*, 389(1),
 335-342. doi: 10.1007/s00216-007-1489-3
- Fernández, B., Campillo, J.A., Martínez-Gómez, C., Benedicto, J. (2012). Assessment of the
 mechanisms of detoxification of chemical compounds and antioxidant enzymes in the digestive
 gland of mussels, *Mytilus galloprovincialis*, from Mediterranean coastal sites. *Chemosphere*,
 87(11), 1235-1245. doi: 10.1016/j.chemosphere.2012.01.024
- 403 Fernández-Reiriz, M.J., Labarta, U., Babarro, J.M.F. (1996). Comparative allometries in growth and 404 chemical composition of mussel (Mytilus galloprovincialis Lmk) cultured in two zones in the Ria 405 sada (Galicia, NW Spain). Journal of Shellfish Research, 15, 349-353. 406 http://hdl.handle.net/10261/61211
- García, C., González, V., Cornejo, C., Palma-Fleming, H., Lagos, N. (2004). First evidence of
 Dinophysistoxin-1 ester and carcinogenic polycyclic aromatic hydrocarbons in smoked bivalves
 collected in the Patagonia fjords. *Toxicon*, 43(2), 121-131. doi: 10.1016/j.toxicon.2003.10.028
- Gillis, P.L., Higgins, S.K., Jorge, M.B. (2014). Evidence of oxidative stress in wild freshwater
 mussels (*Lasmigona costata*) exposed to urban-derived contaminants. *Ecotoxicology and Environmental Safety*, 102, 62-69. doi: 10.1016/j.ecoenv.2013.12.026
- 413 Guillard, R.R., Ryther, J.H. (1962). Studies of marine planktonic diatoms: I. Cyclotella Nana

- Hustedt, and Detonula Confervacea (CLEVE) Gran. *Canadian Journal of Microbiology*, 8(2),
 229-239. doi: 10.1139/m62-029
- Harju, K., Koskela, H., Kremp, A., Suikkanen, S., de la Iglesia, P., Miles, C.O., Krock, B., Vanninen,
 P. (2016). Identification of gymnodimine D and presence of gymnodimine variants in the
 dinoflagellate *Alexandrium ostenfeldii* from the Baltic Sea. *Toxicon*, 112, 68-76. doi:
 10.1016/j.toxicon.2016.01.064
- Hu, M., Li, L., Sui, Y., Li, J., Wang, Y., Lu, W., Dupont, S. (2015b). Effect of pH and temperature
 on antioxidant responses of the thick shell mussel *Mytilus coruscus*. *Fish & Shellfish Immunology*, 46(2), 573-583. doi: 10.1016/j.fsi.2015.07.025
- Hu, M., Wu, F., Yuan, M., Li, Q., Gu, Y., Wang, Y., Liu, Q. (2015a). Antioxidant responses of
 triangle sail mussel *Hyriopsis cumingii* exposed to harmful algae *Microcystis aeruginosa* and
 hypoxia. *Chemosphere*, 139, 541-549. doi: 10.1016/j.chemosphere.2015.07.074
- James, K.J., Carey, B., O'Halloran J., van Pelt F.N.A.M., Škrabáková Z. (2010). Shellfish toxicity:
 human health implications of marine algal toxins. *Epidemiology & Infection*, 138(7), 927-940.
 doi: 10.1017/S0950268810000853
- Ji, Y., Hu, Y., Song, J.L., Chen, H.D., Li, A.F. (2018b). Characteristics of components and regional
 distribution of lipophilic shellfish toxins in bivalves cultured along the Chinese coast in Spring.
 Chinese Fishery Quality and Standards, 8(4), 15-24. (in Chinese)
- Ji, Y., Qiu, J., Xie, T., McCarron, P., Li, A. (2018a). Accumulation and transformation of
 azaspiracids in scallops (*Chlamys farreri*) and mussels (*Mytilus galloprovincialis*) fed with *Azadinium poporum*, and response of antioxidant enzymes. *Toxicon*, 143, 20-28. doi:
 10.1016/j.toxicon.2017.12.040
- Jiang, T., Liu, L., Li, Y., Zhang, J., Tan, Z., Wu, H., Jiang, T., Lu, S. (2017). Occurrence of marine
 algal toxins in oyster and phytoplankton samples in Daya Bay, South China Sea. *Chemosphere*,
 183, 80-88. doi: 10.1016/j.chemosphere.2017.05.067
- JØrgensen, K., Scanlon, S., Jensen, L.B. (2005). Diarrhetic shellfish poisoning toxin esters in
 Danish blue mussels and surf clams. *Food Additives and Contaminants*, 22(8), 743-751. doi:
 10.1080/02652030500136928
- Kharrat, R., Servent, D., Girard, E., Ouanounou, G., Amar, M., Marrouchi, R., Benoit, E., Molgó,
 J. (2008). The marine phycotoxin gymnodimine targets muscular and neuronal nicotinic
 acetylcholine receptor subtypes with high affinity. *Journal of Neurochemistry*. doi:
 10.1111/j.1471-4159.2008.05677.x
- Krock, B., Pitcher, G., Ntuli, J., Cembella, A. (2009). Confirmed identification of gymnodimine in
 oysters from the west coast of South Africa by liquid chromatography tandem mass
 spectrometry, 31(1), 113-118. doi: 10.2989/AJMS.2009.31.1.12.783
- Li, A., Jiang, B., Chen, H., Gu, H. (2016). Growth and toxin production of *Azadinium poporum*strains in batch cultures under different nutrient conditions. *Ecotoxicology and Environmental Safety*, 127, 117-126. doi: 10.1016/j.ecoenv.2016.01.017

- Li, A., Sun, G., Qiu, J., Fan, L. (2015). Lipophilic shellfish toxins in *Dinophysis caudata* picked
 cells and in shellfish from the East China Sea. *Environmental Science and Pollution Research*,
 22(4), 3116-3126. doi: 10.1007/s11356-014-3595-z
- Liu, R., Liang, Y., Wu, X., Xu, D., Liu, Y., Liu, L. (2011a). First report on the detection of
 pectenotoxin groups in Chinese shellfish by LC–MS/MS. *Toxicon*, 57(7-8), 1000-1007. doi:
 10.1016/j.toxicon.2011.04.002
- Liu, X., Xi, Q., Yang, L., Li, H., Jiang, Q., Shu, G., Wang, S., Gao, P., Zhu, X., Zhang, Y. (2011b).
 The effect of dietary *Panax ginseng* polysaccharide extract on the immune responses in white
 shrimp, *Litopenaeus vannamei*. *Fish & Shellfish Immunology*, 30(2), 495-500. doi:
 10.1016/j.fsi.2010.11.018
- Marrouchi, R., Dziri, F., Belayouni, N., Hamza, A., Benoit, E., Molgó, J., Kharrat, R. (2010).
 Quantitative Determination of Gymnodimine-A by High Performance Liquid Chromatography
 in Contaminated Clams from Tunisia Coastline. *Marine Biotechnology*, 12(5), 579-585. doi:
 10.1007/s10126-009-9245-7
- McCarron, P., Rourke, W.A., Hardstaff, W., Pooley, B., Quilliam, M.A. (2012). Identification of
 Pinnatoxins and Discovery of Their Fatty Acid Ester Metabolites in Mussels (*Mytilus edulis*)
 from Eastern Canada. *Journal of Agricultural and Food Chemistry*, 60(6), 1437-1446. doi:
 10.1021/jf204824s
- McCarron, P., Wright, E., Quilliam, M.A. (2014). Liquid Chromatography/Mass Spectrometry of
 Domoic Acid and Lipophilic Shellfish Toxins with Selected Reaction Monitoring and Optional
 Confirmation by Library Searching of Product Ion Spectra. *Journal of AOAC International*,
 97(2), 316-324. doi: 10.5740/jaoacint.SGEMcCarron
- Medhioub, A., Medhioub, W., Amzil, Z., Sibat, M., Bardouil, M., Ben Neila, I., Mezghani, S.,
 Hamza, A., Lassus, P. (2009). Influence of environmental parameters on *Karenia selliformis*toxin content in culture. *CBM Cahiers de Biologie Marine*, 50(4), 333-342
- Miles, C.O., Wilkins, A.L., Stirling, D.J., MacKenzie, A.L. (2000). New Analogue of Gymnodimine
 from a Gymnodinium Species. *Journal of Agricultural and Food Chemistry*, 48(4), 1373-1376.
 doi: 10.1021/jf991031k
- Miles, C.O., Wilkins, A.L., Stirling, D.J., MacKenzie, A.L. (2003). Gymnodimine C, an Isomer of
 Gymnodimine B, from *Karenia selliformis. Journal of Agricultural and Food Chemistry*, 51(16),
 482 4838-4840. doi: 10.1021/jf030101r
- Munday, R., Towers, N.R., Mackenzie, L., Beuzenberg, V., Holland, P.T., Miles, C.O. (2004).
 Acute toxicity of gymnodimine to mice. *Toxicon*, 44(2), 173-178. doi: 10.1016/j.toxicon.2004.05.017
- Naila, I.B., Hamza, A., Gdoura, R., Diogène, J., de la Iglesia, P. (2012). Prevalence and persistence
 of gymnodimines in clams from the Gulf of Gabes (Tunisia) studied by mouse bioassay and LC–
 MS/MS. *Harmful Algae*, 18, 56-64. doi: 10.1016/j.hal.2012.04.004
- Pan, L.Q., Ren, J., Liu, J. (2006). Responses of antioxidant systems and LPO level to
 benzo(a)pyrene and benzo(k)fluoranthene in the haemolymph of the scallop *Chlamys ferrari*.

- 491 Environmental Pollution, 141(3), 443-451. doi: 10.1016/j.envpol.2005.08.069
- 492 Qiu J., Rafuse C., Lewis N.I., Li A., Meng F., Beach D.G., McCarron P. (2018) Screening of cyclic
 493 imine and paralytic shellfish toxins in isolates of the genus *Alexandrium* (Dinophyceae) from
 494 Atlantic Canada. *Harmful Algae*, 77, 108-118.
- Qiu, J., Ma, F., Fan, H., Li, A. (2013). Effects of feeding *Alexandrium tamarense*, a paralytic
 shellfish toxin producer, on antioxidant enzymes in scallops (*Patinopecten yessoensis*) and
 mussels (*Mytilus galloprovincialis*). *Aquaculture*, 396-399, 76-81. doi:
 10.1016/j.aquaculture.2013.02.040
- Regoli, F., Principato, G. (1995). Glutathione, glutathione-dependent and antioxidant enzymes in mussel, Mytilus galloprovincialis, exposed to metals under field and laboratory conditions: implications for the use of biochemical biomarkers. *Aquatic Toxicology*, 31(2), 143-164. doi: 10.1016/0166-445X(94)00064-W
- Salas R., and Clarke D. (2019). Review of DSP toxicity in Ireland: Long-term trend impacts,
 biodiversity and toxin profiles from a monitoring perspective. Toxins, 11(61), 1-19.
 doi:10.3390/toxins11020061
- Seki, T., Satake, M., Mackenzie, L., Kaspar, H.F., Yasumoto, T. (1995). Gymnodimine, a New
 Marine Toxin of Unprecedented Structure Isolated from New Zealand Oysters and the
 Dinoflagellate, *Gymnodinium* sp. *Tetrahedron Letters*, 36(39), 7093-7096. doi: 10.1016/00404039(95)01434-J
- Stewart, M., Blunt, J.W., Munro, M.H.G., Robinson, W.T., Hannah, D.J. (1997). The absolute
 stereochemistry of the New Zealand shellfish toxin gymnodimine. *Tetrahedron Letters*, 38(27),
 4889-4890. doi: 10.1016/S0040-4039(97)01050-2
- Takahashi, E., Yu, Q., Eaglesham, G., Connell, D.W., McBroom, J., Costanzo, S., Shaw, G.R.
 (2007). Occurrence and seasonal variations of algal toxins in water, phytoplankton and shellfish
 from North Stradbroke Island, Queensland, Australia. *Marine Environmental Research*, 64(4),
 429-442. doi: 10.1016/j.marenvres.2007.03.005
- 517 Torgersen, T., Lindegarth, S., Ungfors, A., Sandvik, M. (2008a). Profiles and levels of fatty acid
 518 esters of okadaic acid group toxins and pectenotoxins during toxin depuration, Part I: Brown
 519 crab (*Cancer pagurus*). *Toxicon*, 52(3), 407-417. doi: 10.1016/j.toxicon.2008.06.010
- Torgersen, T., Sandvik, M., Lundve, B., Lindegarth, S. (2008b). Profiles and levels of fatty acid
 esters of okadaic acid group toxins and pectenotoxins during toxin depuration. Part II: Blue
 mussels (*Mytilus edulis*) and flat oyster (*Ostrea edulis*). *Toxicon*, 52(3), 418-427. doi:
 10.1016/j.toxicon.2008.06.011
- Toyofuku, H. (2006). Joint FAO/WHO/IOC activities to provide scientific advice on marine
 biotoxins (research report). *Marine Pollution Bulletin*, 52(12), 1735-1745. doi:
 10.1016/j.marpolbul.2006.07.007
- Van de Waal, D.B., Tillmann, U., Martens, H., Krock, B., van Scheppingen, Y., John, U. (2015).
 Characterization of multiple isolates from an *Alexandrium ostenfeldii* bloom in The Netherlands. *Harmful Algae*, 49, 94-104. doi: 10.1016/j.hal.2015.08.002

Van Wagoner, R.M., Misner, I., Tomas, C.R., Wright, J.L.C. (2011). Occurrence of 12methylgymnodimine in a spirolide-producing dinoflagellate *Alexandrium peruvianum* and the
biogenetic implications. *Tetrahedron Letters*, 52(33), 4243-4246. doi:
10.1016/j.tetlet.2011.05.137

- Wilkins, A.L., Rehmann, N., Torgersen, T., Rundberget, T., Keogh, M., Petersen, D., Hess, P., Rise,
 F., Miles, C.O. (2006). Identification of Fatty Acid Esters of Pectenotoxin-2 Seco Acid in Blue
 Mussels (*Mytilus edulis*) from Ireland. *Journal of Agricultural and Food Chemistry*, 54(15),
- 537 5672-5678. doi: 10.1021/jf060396j

538

539

Tables

Table 1. Concentrations of GYM-A (nmol/kg) and molar percentages of predominantGYM-A esters (to the sum esters in each sample) in the field samples collected fromChinese coasts.

Sample	GYM-A	percentage of predominant esters (%)						ER*
	(nmol/kg)	C16:0	C18:0	C18:1	C20:1	C20:2	C22:2	(%)
clam (Antigono lamellaris)	382	4.9	50	4.3	23	5.9	4.2	58
pen shell (Atrina pectinata)	99	11	20	7.6	23	7.1	18	87
oyster (Crassostrea sp.)	42	-	-	-	-	-	-	-
gastropod (Batillaria zonalis)	55	-	-	-	-	-	-	-

Esters of GYM-A were estimated using GYM-A as calibration standards.

* ER means the esterification rate of GYM-A roughly estimated by GYM-A reference material; ER(%) = esters/(esters = esters)/(esters = esters)/(esters)/(esters = esters)/(e

+ GYM-A) ×100.

- means not detected in samples.

Legends of Figures

Figure 1. Chemical structures of select gymnodimine analogs.

Figure 2. LC-MS/MS chromatograms for fatty acid esters of GYM-A in digestive gland tissue of mussels (*M. galloprovincialis*) fed with *K. selliformis* after 96 h.

Figure 3. LC-HRMS full scan chromatograms for GYM-A (left) and 16:0 GYM-A (right) using a 1.5 ppm mass tolerance, normalized to 100% intensity for *m/z* 508.3421 in each sample. Samples shown are NRC-CRM gymnodimine (A), NRC positive control tissue (B), microalgal feed culture (*K. selliformis*) (C), and muscle tissues (D) and digestive gland tissues (E) of mussels (*M. galloprovincialis*) after 96 h feeding time point of this study.

Figure 4. High resolution MS/MS spectra for GYM-A in certified reference material, and GYM-A and a second peak bearing GYM-A mass in mussel digestive gland (MDG) sample (LEFT). Spectra for 16:0 GYM-A acyl ester in positive control tissue, and the two 16:0 acyl esters derived from the GYM-A and the second GYM-like compound in MDG sample (RIGHT).

Figure 5. Fatty acid esters of GYM-A in muscle (A) and digestive gland (B) tissues of mussels (*M. galloprovincialis*) collected at each sampling time point. The bars show the relative distribution of GYM-A including free form, selected individual ester and other esters (total free form and esters of GYM-A=100). The lines show the sum of the relative amounts of esters to the total toxin burden in % (filled triangle), the level of esters (filled squares) and esters + toxin (filled circles), both in nmol kg⁻¹ tissue (right hand side axis).



Fig. 1



Fig.2



Fig. 3



Fig. 4



Fig. 5