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Variability in iron (II) oxidation kinetics across diverse hydrothermal sites on the northern Mid Atlantic Ridge

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Abstract :

One of the recently recognized main sources of iron to the deep ocean inventory is the hydrothermal activity associated with mid-ocean ridges. Little is known about the oxidation kinetics of iron(II) within these environments, especially the dependence on physico-chemical parameters such as temperature (T), pH, particle size-fractionation and the effect of organic matter.

Following sample collection during the GA13 section cruise, the iron(II) oxidation at six hydrothermal vent sites (Menez Gwen, Lucky Strike, Rainbow, Lost City, Broken Spur and TAG) along the Mid-Atlantic Ridge were investigated, revealing high variability. The Fe(II) oxidation rate constant analysis from multiple stations at two sites (Rainbow and TAG), revealed that factors, other than T and pH, controlled the oxidation process. Experiments on the effect of particle size-fractionation and organic matter at different pH showed that the presence of organic ligands and colloidal size particles delayed the oxidation process, while not affecting the overall pH dependency.

Extending our analysis to the broader relationship between the Fe(II) oxidation rate constants across a range of temperatures (between 2 and 25 °C) and pH (between 7 and 8) on a set of selected hydrothermal samples allowed us to derive a multiparametric equation to model the iron(II) oxidation rate constants in the ocean. This equation covers a larger range of temperatures than previous published equations, improving its applicability for global biogeochemical models.

Highlights

► Organic ligands and nanoparticles increase the resistance to oxidation. ► We present an improved multiparametric equation to calculate Fe(II) oxidation rates. ► Oxidation rates close to the vent sites are systematically faster than theorized.

Keywords : Fe(II) oxidation kinetics, Fe(II) half-life time, Kinetic equation, Hydrothermal, Mid Atlantic Ridge

1 Introduction

It is now well-established that phytoplankton growth is limited by the concentration of dissolved iron (Fe) in up to 30% of the world oceans (Moore et al., 2013). Iron in the ocean, can be found as both ferric (Fe(III)) and ferrous (Fe(II)) iron, where 99% of the dissolved Fe(III) is complexed by organic ligands (Rue and Bruland, 1995). In oxic waters, Fe(II) tends to oxidize to the most thermodynamically stable form, Fe(III), in a period of minutes to hours

(Santana-Casiano et al., 2005). Oxidation of Fe causes the solubility to decrease (Millero et al., 1995), therefore, reducing its bioavailability (Shaked et al., 2005; Morel et al., 2008).

Recently, in the context of the international GEOTRACES program (www.geotraces.org), several studies showed that Fe supplied by hydrothermal activity from mid-ocean ridges can have a substantial impact on Fe distributions at basin scales (Tagliabue et al., 2010; Conway and John, 2014; Resing et al., 2015; Fitzsimmons et al., 2017; Tagliabue et al., 2017). Hydrothermal vent supply of dFe is required to reproduce deep ocean observations (Tagliabue et al., 2010) could represent up to 11-22 % of the dissolved iron (dFe, filtered by 0.2 μm) present in the deep ocean globally (Bennett et al., 2008), accounting for a similar dFe flux as continental riverine runoff (Elderfield and Schultz, 1996). Therefore, hydrothermalism acts as an important Fe source in the deep ocean which may affect the global biological carbon pump (Tagliabue et al., 2010; Resing et al., 2015).

The concentration of dissolved Fe(II) (dFe(II)) in the open ocean is very low, between 0.02 and 2 nmol L^{-1}, due to its fast oxidation under oxic conditions following the Fenton reaction mechanism (González-Davila et al., 2005; Santana-Casiano et al., 2005). However, specific oceanic regions present dFe(II) concentrations which are orders of magnitude higher. In marine sediment interstitial waters, concentrations can reach around 300 μmol L^{-1} (de Baar and De Jong, 2001), upwelling systems can increase Fe(II) concentrations up to 50 nmol L^{-1} (Hong and Kester, 1986), while hydrothermal fluids can contain as much as 25 mmol L^{-1} of dFe(II) (German and Von Damm, 2004). Water emanating from hydrothermal systems contains high concentrations of Fe(II) amongst other elements (Mottl et al., 1979; Mandernack and Tebo, 1993). Studies conducted at hydrothermal vent sites in different ocean basins have revealed different Fe(II) oxidation rate constants, with Fe(II) half-life times varying from 0.5 to 220 min, which has important implications for the dispersion of produced

Fe-oxide phases (Massoth et al., 1998; Field and Sherrell, 2000; Statham et al., 2005; Wang et al., 2012; Santana-González et al., 2017; Rouxel et al., 2018). Recently, Gartman and Findlay (2020) compiled all the available data for Fe(II) half-life from four hydrothermal systems. One site was the TAG vent site in the Atlantic Ocean, while the others were located in the Indian and Pacific Oceans, with Fe(II) half-life estimated using large Fe(II) spikes (500 nmol L^{-1}) that may affect the role played by different reactive oxygen species (González-Dávila et al., 2006).

Temperature (T), pH, and oxygen concentration (O_2) are the primary variables known to directly affect the Fe(II) oxidation (Stumm and Lee, 1961; Millero et al., 1987; Millero and Sotolongo, 1989; Santana-Casiano et al., 2005). Salinity also affects the Fe(II) oxidation, as reported in previous laboratory experiments (Millero et al., 1987; Santana-Casiano et al., 2005). However, other factors such as changes in the inorganic complexation of Fe(II) due to the specific ionic composition of the vent fluid, or organic complexation can influence these oxidation rate constants (Rose and Waite, 2003; Statham et al., 2005). Therefore, the oxidation rate constants must be termed "apparent" due to the effect of organic complexation (Voelker and Sulzberger, 1996), as the nature and concentrations of organic ligands vary between samples. Moreover, pyrite nanoparticles formed in hydrothermal vent sites, have been estimated to oxidize more slowly than dissolved Fe(II) (Yücel et al., 2011), showing that particle composition also influences the oxidation rate constants. The first empirical equation describing the Fe(II) oxidation rate constants in seawater was reported by Millero et al. (1987), and later updated by Santana-Casiano et al. (2005) for nanomolar concentrations. This equation has been included in coupled physical-biogeochemical models in the ocean (OGCBMs) to explore the impact of future climate change on the oxidation of Fe(II) (Tagliabue and Völker, 2011). However, this equation was established for water temperatures

ranging from 10 to 25 ºC, which leads to discrepancies for the Fe(II) oxidation rate constants for cold, deep ocean water masses.

The aims of this study were to measure the oxidation kinetic constants (*k'*) of Fe(II) from different hydrothermal vents and to investigate whether the variations of pH, T and S can explain the observed variability, or if other factors such as the presence of particles, colloids or dissolved organic matter (DOM) are significant contributors. This updated version of the equation incorporates a wider range of temperatures, which lowers the error associated with the theoretical Fe(II) oxidation rate constants. Thus, it becomes applicable to the calculation of theoretical Fe(II) oxidation for the deep waters in which hydrothermal systems are typically found.

2 Methods

2.1 Sampling sites and characteristics

Seawater samples were collected during the GA13 section cruise (JC156) along the Mid Atlantic Ridge (MAR) onboard the *RRS James Cook*. This cruise departed Southampton, UK on December 20th, 2017 and arrived into Guadeloupe, France on February 1st, 2018. Alongside the large scale ocean section survey, sampling focused on six hydrothermal vent sites along the MAR: Menez Gwen (37° 50'N 31° 31'W; Charlou et al., 2000), Lucky Strike (37° 17'N 32° 16'W; Charlou et al., 2000), Rainbow (36° 14'N 33° 54'W; Douville et al., 2002) , Lost City (30° 7'N 42° 7'W; Kelley et al., 2001), Broken Spur (29 °N; James et al., 1995), and TAG (26°N; Gamo et al., 1996) (Fig. 1). These six hydrothermal vent sites represent a wide range of possible hydrothermal conditions due to different underlying geological conditions. Lost City and Menez Gwen are the shallowest sites, with the vent site depths located at 700-720 m and 840-850 m, respectively. The Lucky Strike hydrothermal

field is situated between 1560 and 1730 m depth. The Rainbow vent site is found at 2270-2280 m, within the average range (2000-3000 m) of oceanic ridge depth (Searle, 2013), while Broken Spur and TAG are deeper hydrothermal systems located at 3100-3300 m and 3670-3700 m, respectively. Chemically, Lost City is an alkaline hydrothermal vent, with pH ranging from 9.0 to 9.8. The vent fluid is at a lower temperature of 40-75 °C, compared to 185-365 °C at the other vent sites and has extremely low concentrations of most trace elements (e.g. Fe below the detection limit compared to 2-24000 μmol L^{-1} at the other vent sites) (Ludwig et al., 2006; Fouquet et al., 2013). All other hydrothermal vent sites are acidic with pH lower than 4, and act as important sources of trace metals (Charlou et al., 2000; Douville et al., 2002; Andreani et al., 2014). Geologically, Menez Gwen and Lucky Strike are similar, their basement rocks are mafic, enriched-Mid-Ocean Ridge Basalt (E-MORB, enriched with light rare earth elements, LREE), and present sulfide mounds. Rainbow also presents sulfide mounds, but its basement rock is ultramafic. Further south, the alkaline Lost City is also situated above an ultramafic base with carbonate chimneys. Finally, Broken Spur and TAG have sulfide mounds and basement rocks formed by normal-MORB (N-MORB, LREE depletion) (references within Fouquet et al., 2010).

Seawater samples were collected according to the GEOTRACES guidelines (http://www.geotraces.org/images/Cookbook.pdf). Briefly, a titanium rosette fitted with 24 x 10 L trace metal-clean Teflon-coated OTE (Ocean Test Equipment) bottles, a CTD profiler (Sea-bird Scientific), light scattering sensor (PMEL MAPR), and a redox potential sensor (PMEL MAPR), were deployed on a conducting Kevlar wire to collect samples from the water column. Sampling depths were actively chosen from the continuous T, salinity (S), light scattering sensor (https://data-search.nerc.ac.uk:443/geonetwork/srv/api/records/8ba648f2759d57fbb2246671f8065939) and redox potential readings.

Upon recovery, the OTE bottles were transported into an ISO class 6 clean air van. Unfiltered samples (UF) for Fe(II) oxidation kinetics were immediately transferred into thoroughly rinsed acid-cleaned 250 mL or 1 L low-density polyethylene bottles (LDPE; Nalgene) following the GEOTRACES Cookbook (http://www.geotraces.org/images/Cookbook.pdf) and frozen at -20 °C. Samples remained frozen in a -20 °C freezer for one year before their analysis. pH samples were collected in 125 mL borosilicate bottles and kept in the dark until the time of analysis.

2.2 pH determination

pH was measured onboard and during laboratory-based kinetic experiments in the free scale using a Tris buffer solution (Millero, 1986). Onboard pH samples were kept in the dark, inserted into a 25 °C water bath, and measured at 25 °C, within less than 1 hour after collection. Laboratory-based kinetic experiments used a Titrino 719 (Metrohm^{TM}) which automatically added 0.01M hydrochloric acid (HCl, PANREAC Hiperpur-plus) in order to keep the pH constant. The pH at *in situ* conditions were computed considering the total alkalinity for an average profile of the area (Key et al., 2015; Olsen et al., 2016). The average profile was obtained from plotting all the available alkalinity data from the studied region (20-30 °N and 40-50 °W). The associated error of the estimate was ±6 μmol kg^{-1}. Gray et al. (2011) showed that computed pH values are relatively insensitive to errors in alkalinity, and that pH measurements can be used to compute the magnitude and variability of inorganic carbon variables. A fluctuation of 400 μmol kg^{-1} in total alkalinity values (e.g. from 2325 to 2725 μmol kg^{-1}) will change the pH in +0.005, which lies within the error of determination for potentiometric pH. The set of constants from Lueker et al. (2000) were used.

All the pH data in the text and in the figures, including kinetics and profile data are expressed in the free scale.

2.3 Fe(II) measurements

In order to measure the Fe(II) oxidation rate constants, continuous Fe(II) concentrations were determined using the FeLume System (Waterville Analytical), with a data point acquisition every 1 s. This direct Flow Injection Analysis system uses luminol (5-amino-2,3-dihydro-1,4-phthalazinedione, Sigma-Aldrich) as the reagent (King et al., 1995) which reacts with Fe(II). The sample was continuously aerated with pure air (Carburos Metalicos S.A.). A 0.8 nmol L^{-1} addition of Fe(II) (ammonium iron(II) sulfate hexahydrate, Sigma-Aldrich) was introduced into each sample and the Fe(II) concentration was continuously monitored until the signal had reduced by 70 %. This method produces apparent rate constants (k_{app}) because the oxidation rate constant is measured on a Fe(II) spike rather than on the ambient Fe(II). Although, the oxidation rate constant is not affected by the initial Fe(II) in solution (Millero et al., 1987; Roy et al., 2008; Roy and Wells, 2011; Santana-González et al., 2018), the equilibrium with the organic ligands and particle surfaces could be affected. Moreover, due to the effects of organic complexation and that the concentrations of the organic ligands in each sample are not constant, the apparent oxidation rate constants are consequently not true constants.

Some previous oxidation kinetic studies have found a second-order dependency of the Fe(II) concentration against time (e.g. Statham et al., 2005). However, in oxygen saturated samples a pseudo-first order rate dependency was obtained, where the slope (ln([Fe(II)] against time) represents the pseudo-first order rate constant *k'* (Eq. 1). The linearity remained longer than the half-life time and was thus considered acceptable in all the analyses ($R^2 = 0.983 \pm 0.017$, n samples = 298). The observed linearity confirms that the equilibrium between the added Fe(II) with organic ligands and particles is achieved before 30 s (our first Fe(II) determination time limit). The *k'* error associated with these high R^2 values was 2% for *in*

situ condition experiments, and 1% for experiments carried out at 25 °C. The half-life ($t_{1/2}$) of Fe(II) in seawater can be calculated from *k'* (Eq. 1 and 2).

$$d[Fe(II)]/dt = -k_{app}[O2][Fe(II)] \quad (1)$$

In oxygen saturated concentrations, the corresponding pseudo-first order rate constant $k' = k_{app}[O_2]$.

$$t_{1/2} = \ln(2)/k' \quad (2)$$

The minimum energy required to start the oxidation reaction (activation energy, E_a, J mol^{-1}) and the corresponding preexponential or frequency factor (A) were computed from the calculated *k'* using the empirical Arrhenius equation, where R is the universal gas constant (8.314 J mol^{-1} K^{-1}):

$$\log(k') = \log(A) - \frac{E_a}{R \cdot \ln(10) \cdot T} \quad (3)$$

Iron (II) blanks were obtained from a 2 h aerated, magnetically stirred sample which was kept in the dark to achieve the oxidation of Fe(II) (Santana-Casiano et al., 2005). Iron (II) in theses blanks always below the detection limit. Calibration curves (0 to 1 nmol L^{-1}) were produced daily and confirmed linearity in the Fe(II) signal. The detection limit, defined as three times the standard deviation of the blank after being aerated for 10 minutes, was lower than 50 pmol L^{-1}. Reproducibility was checked by running the same sample three times throughout the day after an addition of 1nmol L^{-1} of Fe(II) with an agreement in concentrations that was better than ± 0.1 nmol L^{-1}. Accuracy could not be determined since reference materials currently do not exist. Results are systematically given as $t_{1/2}$ (in min), followed by the measured *k'* (in min^{-1}).

2.4 Experimental setup

Using natural seawater samples, kinetic experiments were carried out under different conditions, with pH and T set to *in situ* conditions (data shown in the supplementary Table S1), and with pH and T normalized to constant values, in unfiltered, filtered and filtered and UV-irradiated samples. These later experiments required a high sample volume, limiting the number of experiments.

The steps required for the Fe(II) oxidation kinetics determination are as follows. The day before the experiments, three to four frozen samples were taken out of the freezer and placed in a fridge set to 4 ˚C. The next morning, a temperature-controlled water bath was set to the coldest *in situ* potential temperature (supplementary Table S1). While the water bath was cooling down, each sample was subsampled into three 60 mL acid cleaned LDPE bottles and placed back into the fridge. One of the subsamples was inserted into the temperature-controlled water bath, which was set to the original *in situ* potential temperature of the sample. The sample was left in the water bath for acclimatization during 15 min and then transferred into the reaction vessel. In the reaction vessel, the sample was aerated prior to and during the study. After 10 minutes of aeration, the Fe(II) addition was made and the Fe(II) concentration and time were continuously monitored. After finishing with the coldest subsample, the temperature of the water bath was increased to the following coldest temperature, and the next sample was left to acclimate.

2.4.1 Quantification of the temperature and pH effect on the iron (II) oxidation rate constant

In order to improve the temperature range of the existing Fe(II) oxidation kinetic equation, 1 L samples of deep ocean water were collected during the JC156 cruise, at stations 13 (2339 m), 16 (2051 m), 24 (2829 m), 30 (2700 m), 31 (2796 m), and 35 (1999, 3350 and 3601 m). The samples remained frozen at -20 °C during one year until analysis. Two days prior to the

234 analysis, samples were thawed in a 4 °C fridge. This allowed a slow melting of the samples in
235 the dark, thus limiting any photoproduction of Fe(II) and also removal of any superoxide and
236 hydroxyl radicals in the sample (Voelker and Sedlak, 1995).

237 The samples were then transfered into 60 mL LDPE bottles (one for each condition).
238 Temperature effect experiments were performed at constant pH, and those for pH effect at
239 constant temperature. The temperature effect experiments were carried out at *in situ*
240 temperature, 10, 20 and 25 °C at a fixed pH of 8. The pH effect experiments were carried out
241 at *in situ* pH, and pH 7, 7.5 and 8 at a fixed temperature of 25 °C.

242 ***2.4.2 Oxidation rate constants through the water column at the six hydrothermal vent sites***

243 To assess the expected Fe(II) oxidation rate constants at *in situ* conditions, the experiments
244 were carried out at *in situ* potential temperature and pH ($T_{is}pH_{is}$). The temperature was then
245 increased to 25 °C ($T_{25}pH_{is}$) to obtain the temperature effect on the oxidation rate constants.
246 A third set of experiments were performed at 25 °C and pH 8 ($T_{25}pH_{8}$). The T and pH
247 normalization of all the samples allowed for the interpretation of the effect of other variables,
248 such as organic matter and particles, in the oxidation rate constants. This allows to investigate
249 where seawater composition is influencing the oxidation rate, a process that cannot be
250 isolated when comparing two samples with different temperatures and/or pH.

251 ***2.4.3 Experimental setup to assess the effect of colloidal and organic matter***

252 The unused unfiltered (UF) sample volumes remaining from the quantification of the
253 temperature and pH effect experiments were used to assess the impact of particles, colloids
254 and organic matter. Unfiltered samples underwent sequential filtration. Three aliquots were
255 filtered through 0.2 µm (F0.2; Whatman® Anodisc), and one of them was then filtered
256 through 0.02 µm (F0.02; Whatman® Anotop®), while the last aliquot was filtered through

0.02 µm followed by UV-irradiation (F0.02+UV). All samples were left in the dark for 6 days to limit the effect of any photochemically generated reactive oxygen species (ROS) produced during the UV-irradiation (Yuan and Shiller, 2001; Roy et al., 2008). All samples were analyzed following the same procedure as for the assessment of the pH effect (see section 2.4.2 above). In order to assess whether the generated ROS species could affect the Fe(II) oxidation rate constants determined after 0.02 µm filtration and UV-irradiation, the effect of the amount of hydrogen peroxide (H_2O_2) generated during this process was considered (Yuan and Shiller, 2001; O'Sullivan et al., 2005). Other short-lived ROS species such as hydroxyl radicals were not accounted for, since their nanosecond half-life times do not allow for their determination using our analytical method.

3 Results

3.1 Temperature and pH effects on the oxidation rate constant*s*

Experiments for the temperature effect study were carried out at a pH of 8. When plotting log *k'* vs. 1/T (K^{-1}) (Fig. 2, Eq. 3), results showed an average slope ($-\Delta E_a/R/\ln(10)$) of -4992 ± 232. Notably, the data obtained from samples collected west of TAG at 2700 m depth (Station 30) showed a slope of -3732 ± 216, which is more than three standard deviations away from the average value. This lower slope corresponds to a much lower activation energy for this sample, with a value of 71 ± 4 KJ mol^{-1}. This effect was probably caused by the interaction of organic compounds or particles with the Fe(II) species which affected the limiting Fe(II) oxidation step, and therefore a result of a different oxidation reaction mechanism. If this sample is removed, the average slope increases to -5434 ± 183, and the activation energy of all the samples considered was 104 ± 3 KJ mol^{-1}. These values are within the experimental error of previous determinations from Gulf Stream water (-5362 ± 162, and an activation energy of 103 ± 3 KJ mol^{-1} ; Santana-Casiano et al., 2005).

When experiments were carried out at a constant T of 25 ºC and at 5 different pH values in the range from 7 to 8, log *k'* and pH were linearly related, with slopes of 1.00 ± 0.06 (Fig. 3) and with no data points outside of 3standard deviations. As with the temperature relationship, the sample from the station West of TAG at 2700 m presented a slope (1.19 ± 0.08) out of the 1:1 dependence. A second sample, (TAG at 1999 m depth, over 1600 m above the vent site) showed a lower slope (0.81 ± 0.12).

The unique use of lower temperature samples in this study, allowed a better estimation of Fe(II) oxidation kinetic rate constants at temperatures lower than 10 ºC, increasing the range of applicability of the Fe(II) oxidation kinetic equation of Santana-Casiano et al. (2005). Using the equation and dataset from Santana-Casiano et al. (2005), which had a wider range in pH and salinity, together with the additional pH data and a wider range of temperature from the present study, a revised equation was obtained. The pH and salinity dependence did not change, while the T relationship was modified (Eq. 4). Thus, an updated equation for the theoretical Fe(II) oxidation appropriate for deep waters is presented:

$$\log k'(s^{-1}) = 35.627 - 6.7109 * pH + 0.5342 * pH^2 - 5434.02/T - 0.04406 * S^{1/2} - 0.002847 * S \quad (4)$$

The T factor has been modified from -5362.6/T to -5434.02/T obtained from the average slope of Figure 2. To assess the performance of the updated Fe(II) oxidation kinetic equation (Eq. 4), it was compared to three existing equations. The chosen equations were the first derived equation (Millero et al., 1987), the equation by Santana-Casiano et al. (2005), and a recent equation using Labrador Seawater (Santana-González et al., 2019). The four equations were compared to the measured oxidation rate constants at *in situ* pH and T conditions. The residual sum of squares (RSS) and root mean square error (RMSE) between predicted and observed values were calculated. Smallest values of RSS and RMSE indicate the best

equation fit to observations. Additionally, the difference in logarithms of the likelihood (Δlog(L)), based on a Gaussian error distribution with a constant variance of log-transformed data, was used to compare the equation performance (Armstrong et al., 2001 and references therein). The log(L) was calculated as in Moriceau et al. (2009), where a difference in log(L) of at least two points is considered as a criterion of a better fit to the observations. The three most recent equations present better RSS, RMSE and log(L), with the new equation presented here providing consistently better statistical results (Table 1).

	Millero et al. (1987)	Santana-Casiano et al. (2005)	Santana-González et al. (2019)	González-Santana et al. (Eq. 4 of this manuscript)
RSS	0.019	0.003	0.006	0.002
RMSE	0.019	0.008	0.011	0.007
log(L)	24.2	33.6	32.5	36.7

Table 1. Statistical comparison of the available Fe(II) oxidation kinetic equations with the measured oxidation rate constants for deep ocean waters without including the hydrothermal neutrally buoyant plumes. Low residual sum of squares (RSS) and root mean square error (RMSE) indicate the best equation fit to observations. Greater than two points of difference in the difference in logarithms of the likelihood (Δlog(L)) presents a better fit.

3.2 Iron(II) oxidation rate constant*s throughout hydrothermal plumes along the MAR*

Figures 4 and 5 present the half-life times of Fe(II) samples collected above each hydrothermal site with Figure 5 summarizing the results using boxplots. The measured $t_{1/2}$

values presented a wider range than those determined using the physico-chemical parameters in the improved theoretical equation. The Lost City vent site, was the only station where all the measured $t_{1/2}$ at $T_{is}pH_{is}$ conditions presented measured oxidation rate constants higher (shorter $t_{1/2}$) than the theoretically calculated rate constants. At $T_{25}pH_8$ conditions, all the measured oxidation rate constants were higher than the theoretical ones for all stations except at TAG.

The theoretical $t_{1/2}$ values presented in Figure 5a, show the expected high variability in the Fe(II) half-life times associated with T, pH and S. These discrepancies were further highlighted in the measured samples, where only Menez Gwen and Lucky Strike theoretical values are within the measured results (a paired t-test did not show a statistical difference between the theoretical and the measured $t_{1/2}$, with $P > 0.35$ for both).

At Menez Gwen (Fig. 4a), measured *in situ* $t_{1/2}$ varied between 8.3 and 32.0 min (*k'* between 0.084 and 0.022 min^{-1}), with short measured $t_{1/2}$ observed in the samples collected at 743, 821, and 827 m. When the temperature was increased and fixed to 25 ºC, while the pH was kept at *in situ* conditions (Supplementary Table S1), a significant T effect was observed in *k'*, with the measured $t_{1/2}$ decreasing to less than 4 min. When the pH was fixed to a constant value of 8 along the water column, the measured $t_{1/2}$ was slightly affected indicating a small effect, as the change in pH was always lower than 0.22. Overall, at T_{25}, samples presented measured $t_{1/2}$ in the same order of magnitude as the theoretical $T_{25}\,pH_8$. Samples collected within the lower part of the plume, at less than 30 m above the seafloor (hereafter *asf*) presented higher oxidation rate constants than theoretical ones ($t_{1/2}$ decreased).

Lucky Strike (Fig. 5a) presented lower oxidation rate constants than Menez Gwen related to lower temperatures in deeper waters (from about 8.5 ºC to 4.5 ºC). From 1600 m to 1680 m, measured $t_{1/2}$ averaged 46.5 ± 2.5 min (*k'* of 0.015 ± 0.001 min^{-1}, n = 4) (Fig. 4b). This $t_{1/2}$

344 average is higher than the theoretical value of 39.7 min, while at 1703 m, the Fe(II) half-life
345 decreased to 26.3 min (*k'* to 0.026 min^{-1}). Overall, at T_{25}, samples presented measured $t_{1/2}$ in
346 the same order of magnitude as the theoretical T_{25} pH_8. *In situ* pH averaged 7.89, where the
347 increase in acidity produced a lower oxidation rate constant than the measured $t_{1/2}$ at T_{25} pH_8.

348 At Rainbow, station 16, the $t_{1/2}$ measured at *in situ* conditions presented a wide range of
349 variation, from 17.3 to 65.3 min (*k'* varied from 0.040 min^{-1} to 0.011 min^{-1}). The theoretically
350 expected values were within that range, with an average value of 39 ± 3 min (n=8, Fig. 4c).
351 The shortest $t_{1/2}$ were observed between 1986 m and 2108 m (175 to 295 m *asf*) averaging
352 22.2 ± 3.6 min (*k'* = 0.032 ± 0.006 min^{-1}) while the shorter $t_{1/2}$ were found below the plume
353 and above 300 m *asf* (Fig. 4c). The light scattering data shows that the Rainbow plume
354 reached heights over 400 m *asf* (although samples were only collected within the first 330 m).
355 Temperature was an important factor controlling the oxidation rate constant. At T_{25}
356 conditions $t_{1/2}$ were similar to the theoretical value. When the pH was fixed to 8, small
357 changes were observed, and most of the values were close to the theoretical value except for
358 those around 1986 m and 2108 m which had shorter $t_{1/2}$.

359 Lost City (Fig. 4d), the only alkaline hydrothermal vent, was drastically different from the
360 other hydrothermal sites (Fig. 4d), with temperatures around 10 ºC and pH_{is} of 8.07
361 (Supplementary Table S1). The theoretical $t_{1/2}$ associated with these temperatures and pH
362 ranged between 16 and 20 min. However, most of the measured $t_{1/2}$ were shorter than 3 min,
363 reaching 0.5 min (*k'* of 1.299 min^{-1}) at 725 m with the longest $t_{1/2}$ of 7.8 min (*k'* of 0.088 min^{-1})
364 at 750 m. At 750 m depth with $T_{25}pH_{is}$ conditions, the $t_{1/2}$ decreased to 1.9 min (*k'* of 0.357
365 min^{-1}) and further decreased to 1.5 min (*k'* of 0.468 min^{-1}) at $T_{25}pH_8$ conditions. All the
366 measured $t_{1/2}$ were significantly shorter than the theoretical $t_{1/2}$ of 3.2 min (*k'* of 0.218 min^{-1})
367 at normalized conditions.

368 Broken Spur presented a wide range in measured $t_{1/2}$, with values under *in situ* conditions
369 ranging from 1.0 to 65.1 min (*k'* ranging from 0.714 min^{-1} at 2929 m to 0.011 min^{-1} at 2858
370 m; Fig. 4e). Two different behaviors can be observed. The two shallowest samples (2820 and
371 2830 m) were collected in local light scattering minima, with $t_{1/2}$ averaging 63.2 ± 2.7 min (*k'*
372 of 0.011 ± 0.001 min^{-1}) and values slightly higher than the theoretical ones. Deeper than 2830
373 m, the samples were collected in local light scattering maxima, where half-life times were
374 shorter with an average $t_{1/2}$ of 1.9 ± 1.2 min (averaged *k'* of 0.464 ± 0.220 min^{-1}).
375 Experiments at $T_{25}pH_8$ also showed that samples below 2830 m presented shorter $t_{1/2}$ (high *k'*
376 values) than what was theoretically expected.

377 Half-life times measured at *in situ* conditions from the TAG vent site samples presented the
378 widest range in $t_{1/2}$ (Figs. 4 and 5a), varying from 35.1 to 131.2 min (k' between 0.019 min^{-1}
379 and 0.005 min^{-1}). These samples were characterized by little variability of T_{is} (2.67 ± 0.01 ºC)
380 and pH_{is} (8.17 ± 0.01), producing a stable theoretical $t_{1/2}$ of 43 ± 2 min. Furthermore, the
381 measured $t_{1/2}$ were the longest obtained in these experiments. When the experiments were
382 repeated at both fixed T_{25} and pH_8 the measured values encompassed the theoretical value of
383 3.3 min, with values slightly higher at around 3325 m (4.3-4.4 min). Therefore, even when
384 temperature and pH affected the observed oxidation rate constants, the high range of values at
385 *in situ* conditions could not be explained by only these two physico-chemical variables.

386 ***3.3 Investigating the spatial variability of Fe(II) oxidation rate*** constants ***around Rainbow***
387 ***and TAG sites***

388 To investigate the spatial variability in $t_{1/2}$ for Fe(II), samples were collected from four
389 stations around the Rainbow site at 0.3º (30 km) S (St 15), E (St 12) and W (St 13) and 1º
390 (120 km) E (St 10) (Fig. 1) and from five stations around TAG at 0.3º N (St 26), S (St 27), W
391 (St 30), and E (St 31) and 1º W (St 29) (Fig. 1). At both vent sites, turbidity and light

392 scattering anomalies matched and described the shape of the plume (gray shading in Fig. 4).
393 The Rainbow plume anomaly was observed between 1930-1995 m, 2060-2090 m, and most
394 strongly between 2100 and 2210 m. At the TAG site, the signal was three times lower than
395 that observed at Rainbow, indicating the presence of a lower amount of particles, which could
396 also be due to a displacement of the sampling with respect to the vent site. The largest
397 anomalies were measured between 3160-3390 m, with several minor anomalies above (e.g.
398 2980-3030 m) and below (e.g. 3430-3520 m and 3600-3625 m; Fig. 4).

399 Turbidity and light scattering were useful for tracking the plume displacement even at 30 km
400 from the vent site. The Rainbow plume was observed to the south and west (Stn. 13 and 15)
401 at depths between 2000 and 2040 m, and 2270 and 2300 m. For TAG, only stations 27 (south
402 of TAG, but inside the MAR valley) and 31 (east) showed deeper and thinner turbidity
403 anomalies between 3400 and 3690 m and a smaller signal between 2920 and 3060 m.

404 For the Rainbow field, samples collected at the vent site in the 1980-2150 m range had
405 shorter $t_{1/2}$ at *in situ* conditions (high *k'* values) than the theoretical $t_{1/2}$ (Fig. 6a). All the other
406 samples (except for Stn 10 at 2430 m) had higher $t_{1/2}$ values than those calculated
407 theoretically, and greater $t_{1/2}$ than observed at Rainbow at depths shallower than 1980 m,
408 indicating very low oxidation rate constants. This allowed Fe(II) to stay in solution three
409 times longer than under inorganic seawater conditions. However, when T and pH were fixed
410 (Fig. 6b), the oxidation rate constants between stations were much closer to the theoretical
411 value, indicating the T and pH control over the Fe(II) oxidation rate constant. Nevertheless,
412 plume samples at the 2100-2210 m range and around 2400 m indicated $t_{1/2}$ values that were
413 still at least double the theoretical value.

414 At the TAG hydrothermal field, stations situated towards the N and W of TAG presented
415 shorter $t_{1/2}$ (high *k'* values) than those at the vent site and towards the S and E (Fig. 7a).

Station 30, situated 30 km W of TAG, exhibited both patterns, with long $t_{1/2}$ (low *k'* values) at depths shallower than 2600 m and short $t_{1/2}$ (high *k'* values) at depths below 2700 m. When all the TAG hydrothermal field station samples were normalized (i.e. $T_{25}pH_8$; Fig. 7b), most of the variability was removed, and $t_{1/2}$ averaged 3 ± 1.5 min. The same pattern as for *in situ* conditions was maintained, with stations situated towards the N and W having shorter $t_{1/2}$ than the theoretical values. These three stations were characterized by their lower turbidity in the deep waters.

3.4 The effect of particles, colloids, and organic matter on oxidation rate constants

Experiments with samples from Rainbow (Stn 16, 2051 m), West of Rainbow (Stn 13, 2339 m), Broken Spur (Stn 24, 2829 m) and TAG (Stn 35, 3350 m) were performed to improve our understanding of the impact of particulate and organic matter on the oxidation kinetics of hydrothermal vent fluids. The Rainbow sample at 2051 m, located outside the plume, did not show any difference in the log(*k'*) due to the presence of particulates, with values at each pH unit varying by ± 0.03 and an average slope of log(*k'*) against pH of 0.96 ± 0.07 for UF, F0.2 and F0.02 treatments (Fig. 8). For the other three samples (Figs. 8b-d), oxidation rate constants for UF and F0.2 did not present significant differences, with values that were reproducible within ±0.02 for the studied pH range. However, F0.02 samples had higher oxidation rate constants, while also maintaining the pH dependence (0.95 ± 0.07). Therefore, the presence of colloidal size particles delayed the oxidation process while not affecting the pH dependence, as shown in Fig. 8, where the intercept changes but the slope is unchanged. All of the UV irradiated samples in Figs. 8a-d indicate that the absence of dissolved organic matter increased the oxidation rate constants. For TAG (3350 m) and Rainbow (2051 m), the k' values were so high that it was impossible to determine them at pH = 8 under our experimental conditions.

4 Discussion

4.1. Derived oxidation rate constants as a function of pH and temperature

In the 1990s, Fe(II) oxidation kinetics variability was reported at the TAG hydrothermal field (2-3 min) (Rudnicki and Elderfield, 1993) and near the Juan de Fuca Ridge hydrothermal systems (32 h) (Chin et al., 1994; Massoth et al., 1994) using the Millero et al. (1987) equation. Nevertheless, the variability between those two sites was difficult to explain and reflected fluctuations caused by differences in O_2, T, and pH among the different sites.

When combining the six hydrothermal vent sites on the MAR (Fig. 9), the relationship between the measured oxidation rate constants under *in situ* conditions against the theoretically derived oxidation rate constants as a function of pH and temperature (in aerated water) can be obtained. Overall, equation 4 can explain 76 out of 104 data points. When color-coding the data points by stations, we observed that all the outliers corresponded to samples that were collected inside hydrothermal plumes (defined by turbidity, light scatter, and redox potential). Consequently, the results obtained from the theoretical equation (Eq. 4) and figure 9, agree well with previous results that O_2, S, T, and pH can be used to characterize the Fe(II) oxidation rate constant in open ocean waters but not from within hydrothermal plumes.

Our study clearly indicates (Figs. 8 and 9) that the presence of both colloidal particles and dissolved organic matter increases the half-life of Fe(II) (decreased k'). This agrees with previous work showing a decrease in the Fe(II) oxidation rate constants in seawater in the presence of some natural organic compounds (Santana-González et al., 2019). However, organic matter can have a positive or negative effect on the oxidation rate constants of Fe(II) depending on the type of organic matter and its degree of remineralization (Rose and Waite,

2003). The variability in the effect of total organic carbon on Fe(II) oxidation proves that total organic carbon cannot be used as a variable in an equation describing k' (Santana-Casiano et al., 2000; Santana-González et al., 2018; Santana-González et al., 2019). In this sense, a more detailed characterization of organic matter would be necessary, especially in a hydrothermal environment.

4.2. Influence of organic species on Fe(II) oxidation rate constants

The Fe(II) oxidation rate constant (Eq. 5) can be expressed as a function of each Fe(II) species (Fe^{2+}, $FeOH^{+}$, $Fe(OH)_2$, $FeHCO_3^{+}$, $Fe(CO_3)$, $Fe(CO_3)_2^{2-}$, $Fe(CO_3)OH^{-}$, $FeCl^{+}$, $FeSO_4$, FeH_3SiO_4 and any organically complexed Fe species). Species such as $Fe(CO_3)_2^{2-}$ and $Fe(OH)_2$ could become important at pH higher than 8 (Ussher et al., 2004; González-Davila et al., 2005; Santana-Casiano et al., 2005) in solution. Considering the side reaction coefficient for each inorganic (α_i) and organic (α_j) Fe(II) species in seawater and the associated second-order oxidation rate constant for the kinetically reactive Fe(II) species (k_i and k_j) (Eq. 6) (Millero and Sotolongo, 1989; King, 1998; Santana-Casiano et al., 2005), the apparent oxidation rate constant can be expressed as:

$$k_{\mathrm{app}} = k_{Fe^{2+}}\alpha_{Fe^{2+}} + k_{FeOH^{+}}\alpha_{FeOH^{+}} + k_{Fe(OH)_2}\alpha_{Fe(OH)_2} + k_{FeHCO_3^{+}}\alpha_{FeHCO_3^{+}} + k_{Fe(CO_3)}\alpha_{Fe(CO_3)} + k_{Fe(CO_3)_2^{2-}}\alpha_{Fe(CO_3)_2^{2-}} + k_{Fe(CO_3)OH^{-}}\alpha_{Fe(CO_3)OH^{-}} + k_{FeCl^{+}}\alpha_{FeCl^{+}} + k_{FeSO_4}\alpha_{FeSO_4} + k_{FeH_3Si\ _4}\alpha_{FeH_3SiO_4} + \sum_j k_{FeL_j}\alpha_{FeL_j} \quad (5)$$

$$k_{\mathrm{app}} = \sum_i \alpha_i k_i + \sum_j \alpha_j k_j \quad (6)$$

$$\alpha_{Fe^{2+}} = 1/\left(1 + \sum K_{L_i}[L_i] + \sum K_{L_j}[L_j]\right) \quad (7)$$

Where $\alpha_{Fe^{2+}}$ is the fraction of free Fe(II) including inorganic and organic speciation, $[L_i]$ is the concentration of each inorganic ligand with a conditional stability constant K_i and $[L_j]$ is

485 the concentration of each organic ligand with a conditional stability constant K_{Lj}. The
486 inorganic site reaction coefficient at pH 8 and 25 ºC was previously determined as 0.38
487 (Santana-Casiano et al., 2006). For any inorganic or organic FeL complexes:

488 $$\alpha_{FeL} = K_{FeL}[L]\alpha_{Fe^{2+}} \quad (8)$$

489 According to our results, the presence of DOM in solution generated Fe(II)-organic
490 complexes that were overall less reactive to oxidation and therefore, when removed by UV
491 irradiation the resulting oxidation rate constants were higher (Fig. 8).

492 The amount of dissolved organic carbon (DOC) in the samples ranged between 50 and 79
493 μmol kg^{-1} (data not shown) and according to O'Sullivan et al. (2005) the maximum amount
494 of H_2O_2 produced during irradiation could be 125 to 147 nmol kg^{-1}, respectively. Following
495 González-Davila et al. (2005) and assuming the extreme case of no H_2O_2 decay after 6 days,
496 the UV generated H_2O_2 could theoretically increase the Fe(II) oxidation rate constants in
497 oxygen saturated conditions. The theoretical $t_{1/2}$ associated with the H_2O_2 oxidation was
498 between 48 % and 53 % of the $t_{1/2}$ measured for 0.02 μm filtered samples. Nevertheless, the
499 measured $t_{1/2}$ for UV-irradiated samples (Supplementary Table S1 and Fig. 8) were shorter
500 (higher *k'*), indicating that any H_2O_2 remaining in solution after 6 days was not the main
501 factor that increased the oxidation rate constants and that the presence of dissolved organic
502 matter was stabilizing Fe(II) in solution. Furthermore, laboratory experiments have shown
503 that the $t_{1/2}$ of H_2O_2 in oceanic waters is shorter than 5.5 days (Yuan and Shiller, 2001),
504 therefore the impact of H_2O_2 in the experiments would be further reduced.

505 The role of organically complexed Fe(II) could be calculated from the following
506 assumptions. Firstly, the Fe(II)-L complex was assumed to not oxidize in the pH range of 7.5
507 to 8. Secondly, the observed reduction in Fe(II) oxidation rate constants (Fig. 8 and

508 Supplementary Table S1) at 25 ºC between the 0.02 µm filtered and the filtered and UV
509 irradiated samples was due to organic complexation. The computed role of organic
510 complexation produced a 78 ± 6% decrease in the concentration of reactive inorganic Fe(II)
511 species. The effect of organic Fe(II) species was highest at Broken Spur and TAG sites (81 ±
512 5 %) and lowest at Rainbow (74.1 ± 4 %). Roy et al. (2008) found a value that was slightly
513 lower for the western subarctic Pacific water, with a 61 ± 8% decrease. As both the organic
514 ligand concentrations and the conditional constants are unknown, only the product $K_{Lj} \cdot [L_j]$
515 (Eq. 7) can be inferred. Knowing the inorganic distribution coefficient of Fe(II), which is pH
516 dependent but can be calculated from Santana-Casiano et al., (2005), and considering a
517 unique class of ligand (L), the product $K_{Lj} \cdot [L_j]$ should be 9±2 (higher values at Broken Spur
518 and TAG, while lower values at Rainbow) while $1 + \sum K_{L_i}[L_i]$ in Eq. 7 is 2.6 (2.6 = 1/0.38,
519 Santana-Casiano et al., (2006)).

520 If the concentration of Fe(II) organic ligands in seawater is on the same order as Fe(III)
521 ligands (1 to 10 nmol L^{-1}), as assumed by Roy et al. (2008), the corresponding conditional
522 constant should range between 10^9 to 10^{10}, (respectively for 10 and 1 nmol L^{-1} of ligands).
523 The strength of these Fe(II) complexing ligands compared to those for Fe(III) is in the low
524 affinity range (Gledhill and Buck, 2012; Bundy et al., 2014). However, the formation of these
525 complexes would decrease the Fe(II) oxidation rate constant through both the effects on the
526 side reaction coefficients (Eqs. 7-8) and because the complexes can be non-reactive to
527 oxidation or may oxidize slower than other inorganic species. Moreover, it should be
528 considered that organic matter might even increase the Fe(II) $t_{1/2}$ due to the reduction of
529 organically complexed Fe(III), which would increase the Fe(II) in solution, and therefore the
530 $t_{1/2}$ (Santana-Casiano et al., 2010).

The UV-irradiation treatment (Fig. 8) removes the organic matter effect on the Fe-complexation, which for the studied samples resulted in an increase in the oxidation rate constants. Consequently, in these analyzed natural samples, organic matter ultimately lowered the oxidation rate constants (black data points below the straight line in Fig. 9).

The DOM is prone to natural and thermal degradation in hydrothermal systems (Hawkes et al., 2016). Accordingly, the variety of organic-complexing ligands and their effect on oxidation could change with plume transport and distance to the vent site as observed in figures 6 and 7. In the proximities to the vent sites, samples presented greater than theoretical oxidation rate constants (above the straight line in Fig. 9). This could be due to the presence of organic matter with ligands complexing Fe that oxidizes faster than the inorganic Fe(II) species. At these locations, the organic ligand effect would dominate over processes that lower oxidation rate constants such as nanoparticle concentrations. Similar experiments considering organic matter degradation just above the vent site should be carried out to confirm this hypothesis. Nevertheless, the complexity of the organic matter pool and the extent and nature of Fe-DOM interactions make the resulting effect on the oxidation rate constants difficult to predict.

4.3. Effects of physical speciation

Lough et al. (2019) have shown that there is an Fe exchange between the different particle size fractions within the hydrothermal plume. This exchange could affect Fe(II) oxidation as found in this study for particles smaller than 0.2 µm and colloidal sized particles (0.2-0.02 µm) at different locations in TAG and in Rainbow (Fig. 8). The results shown in Figure 8 seem to indicate that the size-fractionation can significantly affect Fe(II) oxidation and thus, is a factor that should be considered in future studies. In this sense, previous work has shown that particulate Fe(II), such as pyrite nanoparticles are more resistant to oxidation, with Fe(II)

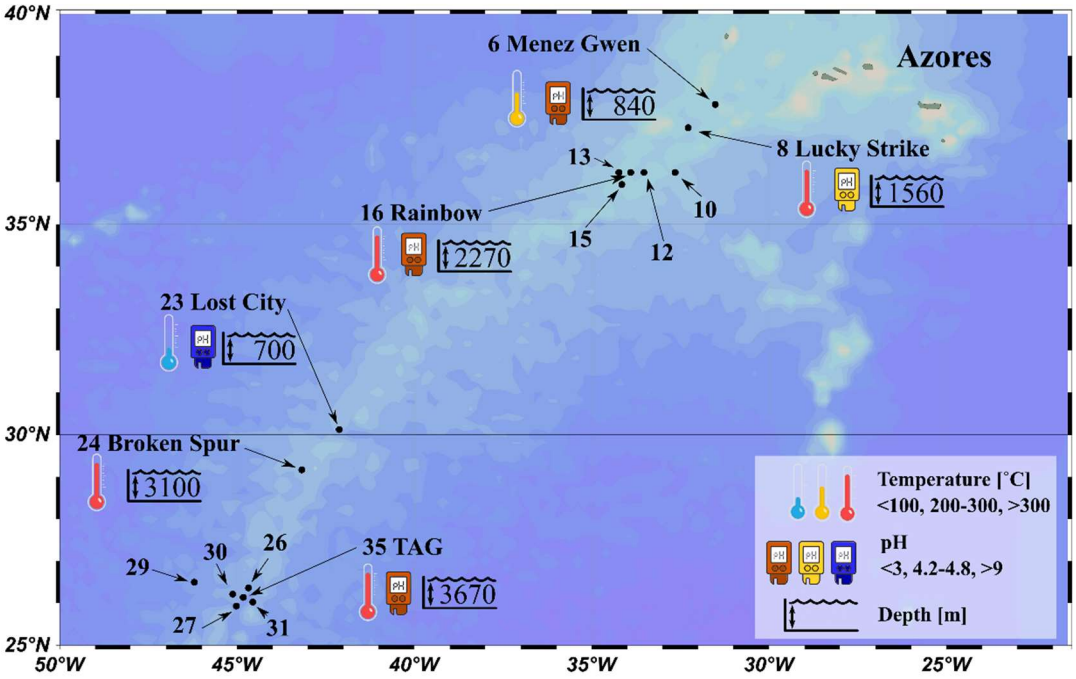
half-life between 4 and 48 months at 2 °C (Yücel et al., 2011). The presence of colloidal particles also enhances the decay of ROS allowing for longer $t_{1/2}$ (Yuan and Shiller, 2001). Therefore, nanoparticles could account for a substantial fraction of the decrease in the oxidation rate constants (Fig. 8). Another consequence of filtration is the reduction in larger size organic matter, which could also affect the stabilization of Fe(II) (Buck et al., 2015).

The variability observed in the oxidation rate constants inside the plume for the six studied hydrothermal vents indicates that changes in the composition of both the DOM and colloidal particles, both between sites and at different distances from the vent, can strongly affect the oxidation rate constants, thus, their role requires further study.

5 Conclusions

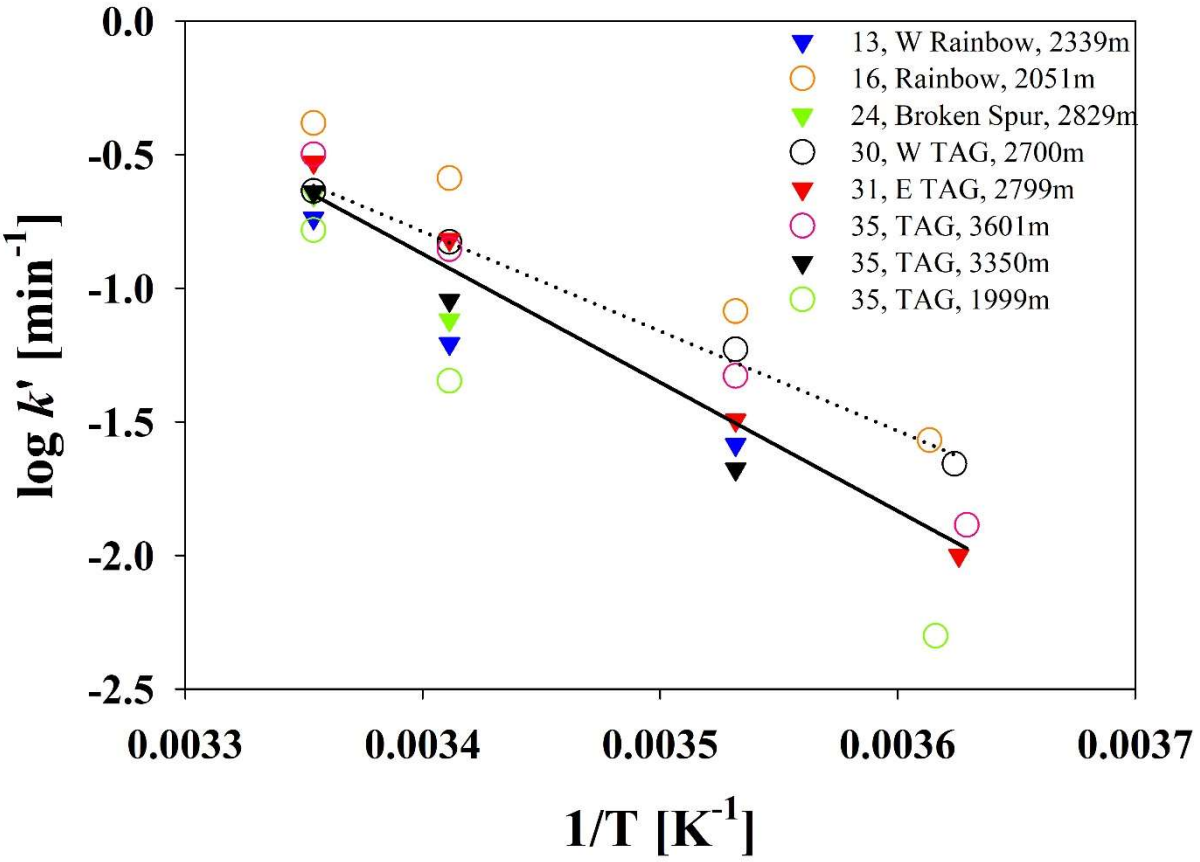
This study highlights the complexity of Fe(II) chemistry in the ocean, with a focus on hydrothermal vent settings. Fe(II) oxidation rate constants, and therefore persistence of Fe(II) in the environment, is shown to not only depend on the physico-chemical conditions (T, S, pH, O_2 content) but also the physical state (soluble, colloidal and particulate) of the inorganic and organic species that are present. Our results indicate that organic ligands can stabilize Fe(II) and that nanoparticles present in hydrothermal plumes could increase the resistance to oxidation. Both processes contribute to the observed variability in the oxidation rate constants in the proximities of vents. We were able to revise the theoretical equation for the estimation of k', and to further increase the existing temperature range to include seawater temperatures as low as 2 ºC. Future studies on iron oxidation kinetics in the marine environment should be accompanied by a characterization of the organic matter present in the medium, a size-fractionation of the samples, and a detailed description of the conventional physico-chemical factors (pH, T, and S).

578 **Figures**



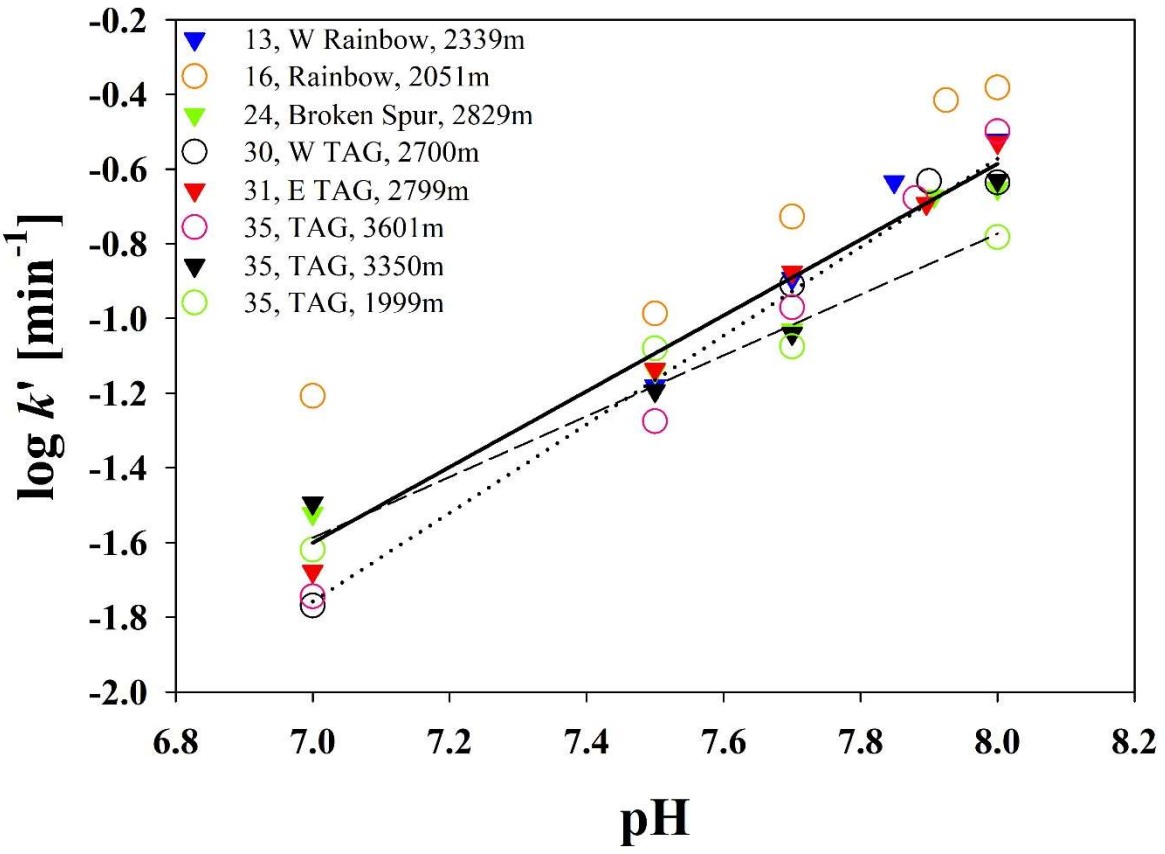
579

580 Figure 1. Map of the investigated hydrothermal vent sites along the MAR, showing the
581 station numbers and hydrothermal vent site names for the sampling sites. Symbols represent
582 endmember characteristics. The pH of the Broken Spur endmember was not available.



583

584 Figure 2. Arrhenius relationship between the oxidation rate constant (log *k'* (min^{-1})) and
585 temperature (1/T (K^{-1})) in the range 2 to 25 ºC for eight selected samples. Note that the
586 sample from station 30 (W TAG, 2700m, open squares, dotted line) shows a different
587 behavior than samples from other stations.



588

589 Figure 3. pH effect (from 7.0 to 8.0) on the Fe(II) oxidation rate constants (log *k'* (min^{-1})) for
590 eight selected samples collected along the MAR. Note that the samples from 2700 m depth to
591 the west of TAG (Stn 30, open squares) and from 1999 m depth at the TAG site (Stn 35, open
592 diamonds) have different slopes than samples from other stations (data from these two
593 stations are plotted with dotted lines).

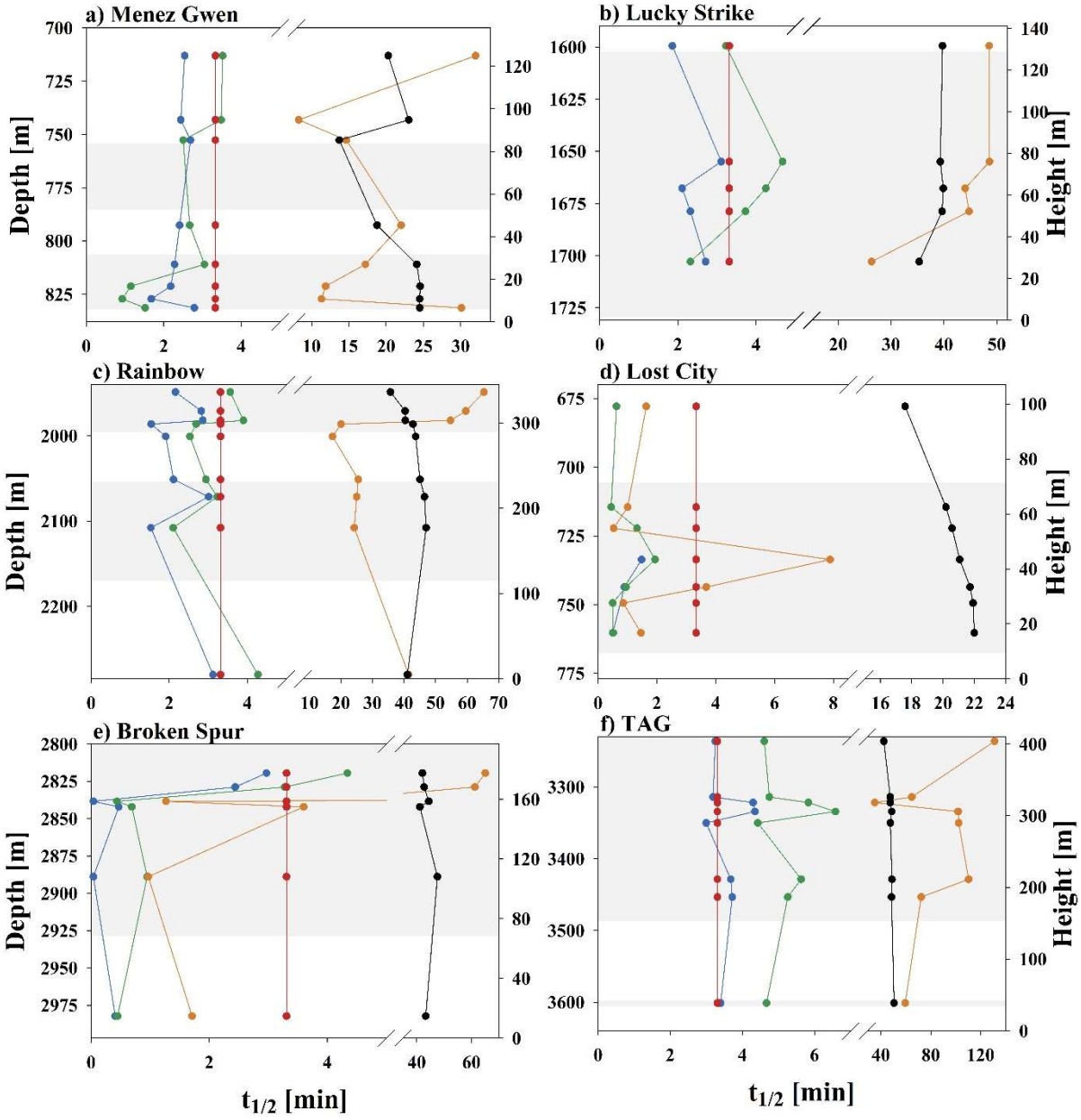
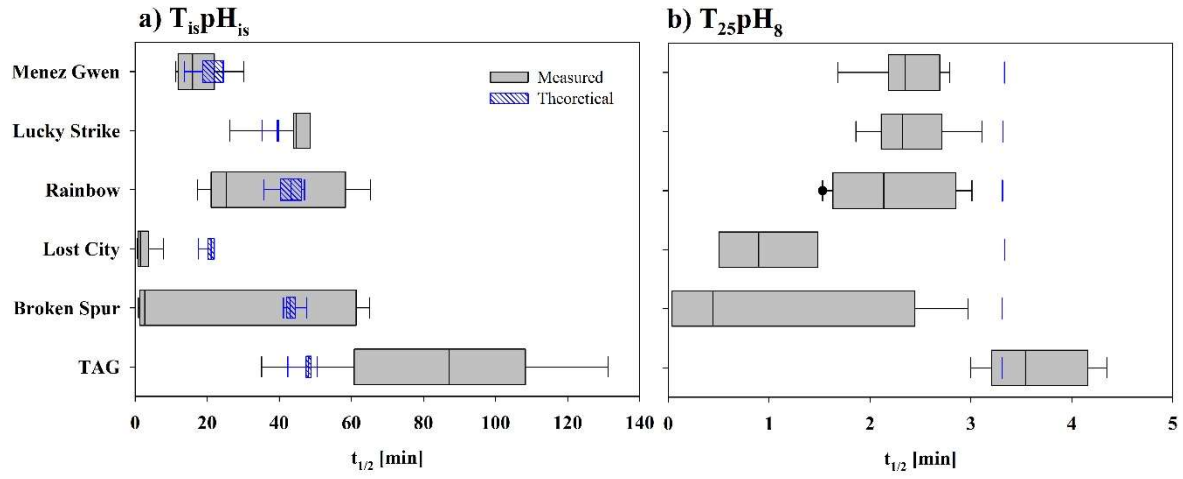
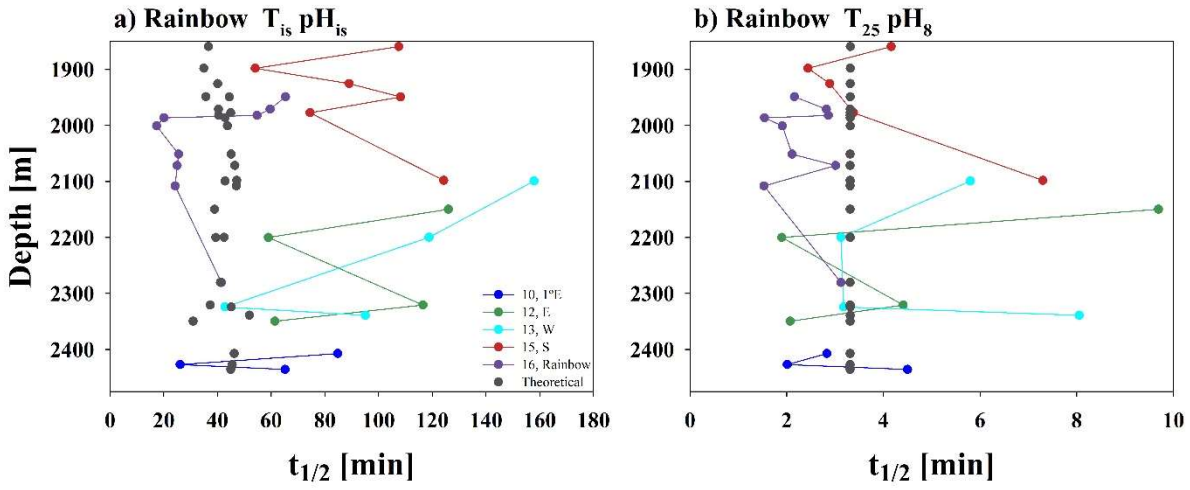


Figure 4. The measured Fe(II) half-life ($t_{1/2}$, min) at $T_{is}pH_{is}$ (orange), $T_{25}pH_{is}$ (green), and $T_{25}pH_8$ (blue), and the theoretical (Theo) values for $T_{is}pH_{is}$ (black), and $T_{25}pH_8$ (red), at the six vents. The shaded area shows the anomaly depths determined using light scattering data . Both, the depth (left y-axis) and the height above the seafloor (right y-axis) of the sampling are indicated.



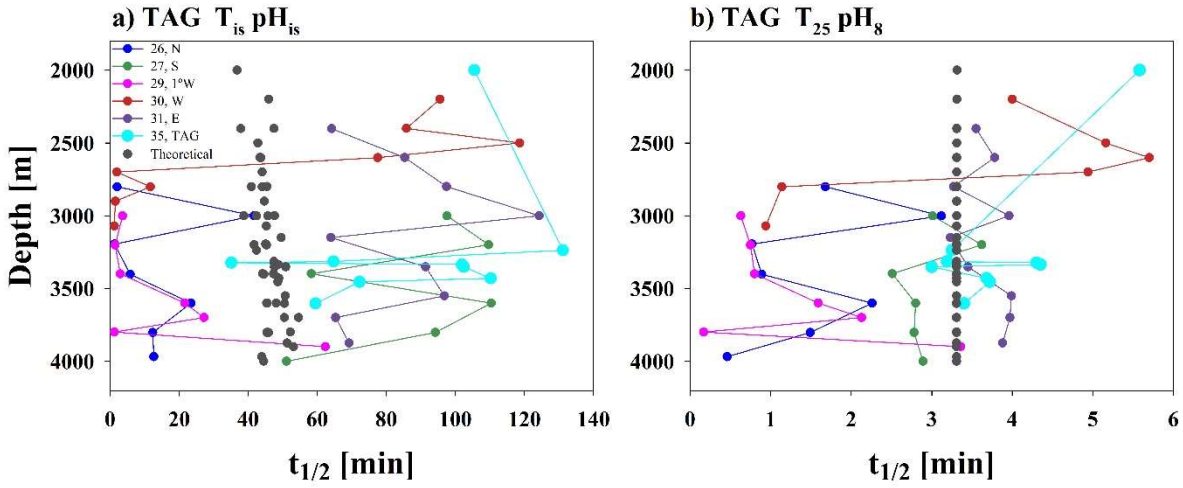
600

601 Figure 5. Boxplots of the Fe(II) half-life ($t_{1/2}$, min) within the six hydrothermal plumes. a)
602 The measured $t_{1/2}$ at *in situ* conditions (grey) and the theoretical $t_{1/2}$ (blue; obtained using Eq.
603 4). b) The measured $t_{1/2}$ at constant temperature (25 ºC) and pH 8 (grey) and the
604 corresponding theoretical $t_{1/2}$, the slight variances are caused by S changes (blue).



605

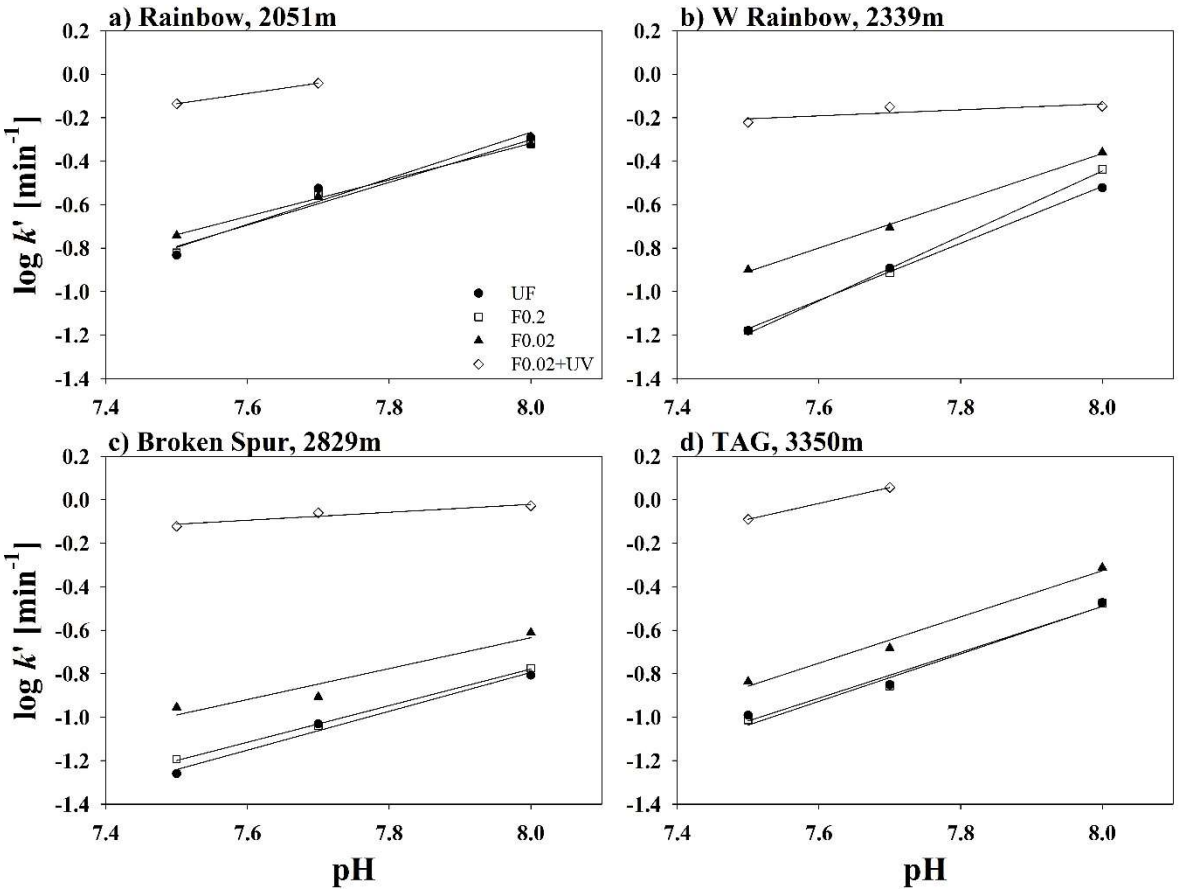
606 Figure 6. The Fe(II) half-life ($t_{1/2}$, min) at the Rainbow field a) under *in situ* conditions (see
607 also Supplementary Table S1) and b) normalized to constant T of 25 ºC and a pH of 8.



608

609 Figure 7. The Fe(II) half-life ($t_{1/2}$, min) at the TAG field a) under *in situ* conditions (see also

610 Supplementary Table S1) and b) normalized to constant T of 25 ºC and a pH of 8.

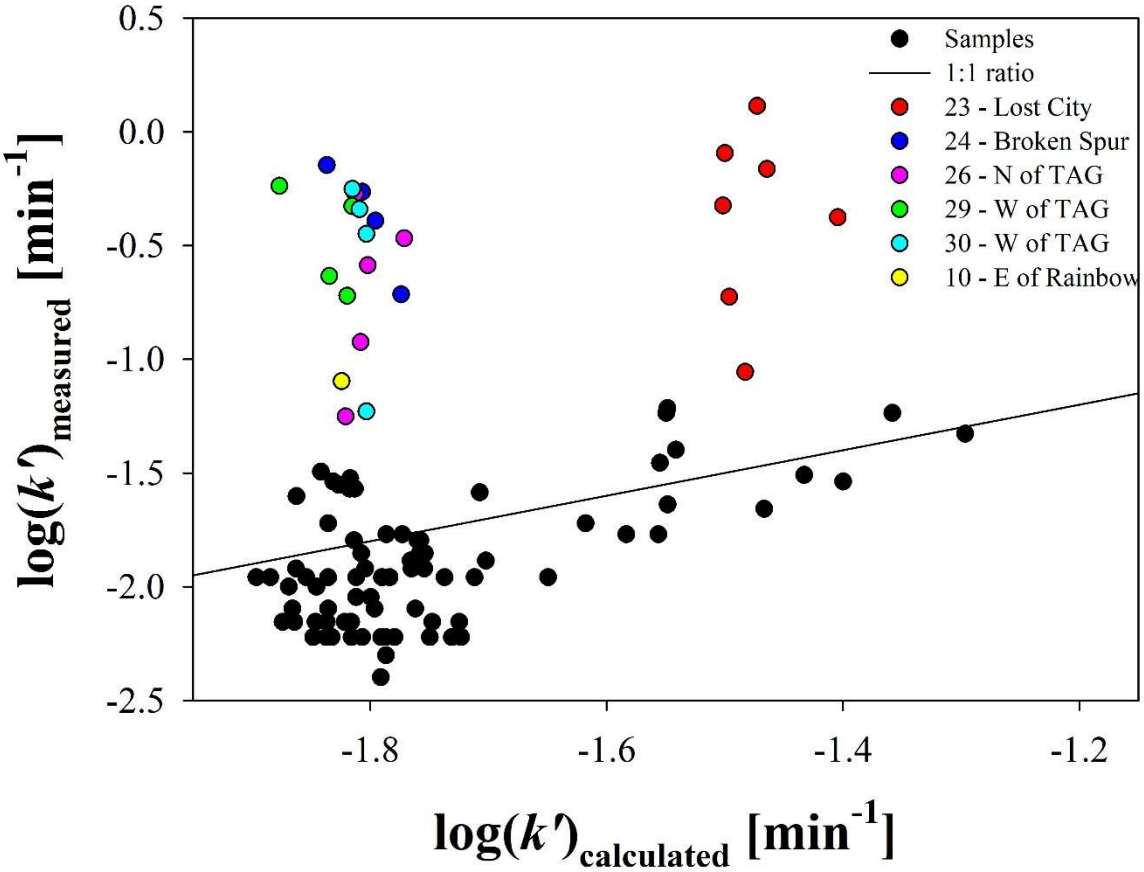


611

612 Figure 8. Oxidation rate constants at different pH (7.5, 7.7 and 8), for selected unfiltered (UF,

613 filled circles), filtered through 0.2 µm (F0.2, open squares), filtered through 0.02 µm (F0.02,

614 filled triangles) and filtered through 0.02 µm followed by UV irradiation (F0.02+UV, open
615 diamonds) samples at a) Rainbow (Stn 16, 2051 m), b) West of Rainbow (Stn 13, 2339 m), c)
616 Broken Spur (Stn 24, 2829 m) and d) TAG (Stn 35, 3350 m).



617

618 Figure 9. Fe(II) oxidation rate constants (log(*k'*)) determined under *in situ* conditions against
619 the anticipated expected theoretical value. All data points are shown in black. The black line
620 represents a 1:1 ratio between both oxidation rate constants. Color coded data points
621 correspond to samples with measured oxidation rate constants greater than 20% of the
622 theoretically derived rate constants.

623 **Author contributions**

DGS wrote the manuscript with significant contributions from all coauthors. DGS was invited to the research cruise by MCL and AT, where DGS, MCL, and LA collected the samples. DGS analyzed the samples with support from MSC and MGD. Figures for the results were designed by HP and GS. Fieldwork logistics and costs were overseen by MCL and AT, while analysis costs were handled by MSC, MGD, GS, and HP.

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Competing interests

The authors declare no competing interests.

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