
Risk and benefit assessment of seafood consumption harvested from the Pertuis Charentais region of France

Noger-Huet Élise ¹, Vagner Marie ^{1,2}, Le Grand Fabienne ², Graziano Nicolas ², Bideau Antoine ², Brault-Favrou Maud ¹, Churlaud Carine ¹, Bustamante Paco ³, Lacoue-Labarthe Thomas ^{1,*}

¹ UMR LIENSs, CNRS-La Rochelle Université, 2 rue Olympe de Gouges, 17 000, La Rochelle, France

² UMR LEMAR, Univ Brest, IRD, CNRS, Ifremer, Place Nicolas Copernic, Plouzané, 29 280, France

³ Institut Universitaire de France (IUF), 1 rue Descartes 75005, Paris, France

* Corresponding author : Thomas Lacoue-Labarthe, email address : tlacouel@univ-lr.fr

Abstract :

Seafood is well recognized as a major source of Long Chain n-3 Polyunsaturated Fatty Acids (LC n-3 PUFA, especially eicosapentaenoic acid, i.e. EPA and docosahexaenoic acid, i.e. DHA) and essential trace elements (As, Cu, Fe, Mn, Se, and Zn). It is also a source of non-essential trace elements (Ag, Cd, Hg, Pb) that can be deleterious for health even at low concentrations. Edible parts of sixteen species (fish, cephalopods, crustaceans and bivalves) of great importance in the Pertuis Charentais region, one of the main shellfish farming and fishing areas, were sampled in winter and analyzed to determine their fatty acid (FA) composition and trace element concentrations. Based on these analyses, a suite of indices was calculated to estimate risk and benefit of seafood consumption: the n-6/n-3 ratio, the atherogenic index, the thrombogenic index, the EPA + DHA daily recommended portion, as well as the maximum safe consumption. The results showed that fish contributed the most to LC n-3 PUFA supply, while bivalves and crustaceans were more beneficial in essential trace elements. Whatever the species, the concentrations of non-essential elements were not limiting for seafood consumption, as important amounts of the analyzed species can be eaten daily or weekly before becoming deleterious to consumers. Yet, concentrations of Hg in dogfish and seabass can become a concern for frequent seafood consumers (>three meals a week), confirming that varying seafood items is a key point for consumers to optimize the benefits of diverse seafood resources. Considering FA composition, whiting and pilchard are the most beneficial fish species for human diet, while surmullet was the least beneficial one. However, using an index integrating the relative risk due to Hg content, the surmullet appears as one of the most beneficial. This study provides a temporal shot of the quality of marine resources consumed in winter period in the studied area and highlights the complexity of a quantitative risk and benefit assessment with respect to the biochemical attributes of selected seafood.

Keywords : LC n-3 PUFA, Essential and non-essential trace elements, Mercury, Net risk/benefit index

1. Introduction

Seafood (i.e. fish, crustaceans and mollusks) is currently a significant component of food sources for humans worldwide, especially for those that live in coastal areas (60% of the world's population). The average annual intake per capita had increased from 9.0 kg in 1961, to 20.5 kg in 2017 (FAO, 2018). Seafood is part of a well-balanced human diet with several recognized benefits because, e.g. they are rich in proteins, vitamins, omega-3 fatty acids (FAs) and essential elements such as copper (Cu), iron (Fe), or zinc (Zn).

Seafood is particularly recognized as the main source of long chain highly unsaturated fatty acids of the n-3 series (LC n-3 PUFA, i.e. fatty acids with at least 4 double bonds and 20 carbon atoms), also known as LC omega-3 PUFA (e.g. Afonso *et al.*, 2013). LC n-3 PUFA, and especially the highly unsaturated ones, are major components of cell membranes but are poorly synthesized *de novo* by vertebrates including humans and thus must be supplied by food. Among these, eicosapentenoic acid (EPA, 20:5n-3) and docosahexanoic acid (DHA, 22:6n-3) in human diet are sourced mostly from seafood (Astorg *et al.*, 2004). EPA and DHA are the most beneficial LC n-3 PUFA. Their benefits on human health, and in particular in cerebral, cardiovascular, and immune functions, are now well-recognized (Gil & Gil, 2015; Mozaffarian & Rimm, 2006; Pike, 1999; Ruxton *et al.*, 2004; Simopoulos, 1991). For example, regular intakes of LC n-3 PUFA help in reducing of risk of death from a coronary heart disease through the reduction of atherogenic and thrombogenic risks (i.e. reduction of the platelet aggregation and subsequent thrombus and atheroma formation in the cardiovascular system; Valfré *et al.*, 2003). LC n-3 PUFA regular intake also induce a drop of dementia disorders and Alzheimer symptoms in elderly people (Hu *et al.*, 2002; Morris *et al.*, 2003; Oomen *et al.*, 2000; Ruxton *et al.*, 2004). The improvement of neuronal development and visual acuity of breast-fed infants by women who regularly consume marine products has also been demonstrated (Fleith & Clandinin, 2005; Koletzko *et al.*, 2008; Simopoulos, 1991). As well as a regular LC n-3 PUFA intake, a balanced dietary n-6/n-3 PUFA ratio between 1:1 and 4:1 has been shown to prevent the onset of coronary heart disease, as n-3 contribute to anti-inflammatory and anti-thrombus process, while n-6 contribute to inflammatory process (Simopoulos *et al.*, 2000; Simopoulos, 2002; 2003).

Seafood intake is, as well, one of the main sources of essential trace elements such as arsenic (As), Cu, Fe, manganese (Mn), selenium (Se) and Zn (e.g. Guerin *et al.*, 2011). In organisms, these elements are also cofactors of enzymes involved in antioxidant systems, in the DNA

80 metabolism, and in the oxygen transport (Ansari *et al.*, 2004; Olmedo *et al.*, 2013a; Uthus,
81 1992). However, these essential elements may be at risk when too high amounts are ingested
82 (Ansari *et al.*, 2004; Ersoy & Çelik, 2009; Frieden, 1985; Goldhaber, 2003; Olmedo *et al.*,
83 2013a; Storelli, 2009), as they become toxic like other contaminants usually found in seafood.
84 Indeed, mainly fisheries and aquaculture activities occur in coastal waters, where anthropogenic
85 activities (*e.g.* agriculture, industry, urbanization, oil extraction) contribute to the enrichment
86 of waters in essential elements such as As, Cu, Fe, Mn, Se and Zn but also in non-essential
87 elements like silver (Ag), cadmium (Cd), mercury (Hg) and lead (Pb), which are highly toxic
88 even at low concentrations (Leblanc *et al.*, 2006; Maher & Butler, 1988; Naser, 2013; Olmedo
89 *et al.*, 2013b). Some non-essential elements (*e.g.* Hg or Pb) can also biomagnify along the food
90 chain, reaching elevated concentrations in marine predators regardless of ambient
91 contaminations (Anual *et al.*, 2018; Langston & Bebianno, 1998; Maher & Butler, 1988;
92 Olmedo *et al.*, 2013b, Eagles-Smith *et al.*, 2018). The consumption of seafood contaminated
93 by these trace elements could thus put humans at risk, potentially leading to neurotoxic,
94 carcinogenic or cardiovascular issues (Ansari *et al.*, 2004; Ersoy & Çelik, 2009; Goldhaber,
95 2003; Leblanc *et al.*, 2006).

96 The coast of Charente-Maritime and its adjacent Pertuis Charentais area hosts the largest
97 network of intertidal bare mudflats in France, conferring to the area a high primary productivity
98 of the littoral zone that are advantageous for fisheries and shellfish farming (Blanchard *et al.*,
99 2001). The Marennes-Oléron bay is the first European basin for oyster farming (Miossec *et al.*,
100 2009). The Pertuis Charentais area and the onshore Bay of Biscay support artisanal fisheries
101 targeting local species such as *Merluccius merluccius*, *Merlangius merlangus*, *Lophius*
102 *piscatorius*, *Sardina pilchardus*, *Scomber scombrus* and *Sepia officinalis* among others
103 (FranceAgriMer, 2017). This highly touristic region is also known for its historical Cd, Cu, and
104 Zn contamination originating from the discharge of a mine treatment wastes upstream from the
105 Gironde River mouth (Grousset *et al.*, 1999; Miquel, 2001; Miramand *et al.*, 2002). Industrial
106 wastes are also an important source of Hg in the Charente River which emerges directly in the
107 Pertuis Charentais (Gagnaire *et al.*, 2003). More recently, an increasing contamination pressure
108 by Ag used as nanoparticles with antimicrobial properties and as hail clouds dispersive agent
109 to protect regional wines also has occurred in coastal waters of this area (Salles *et al.*, 2013).
110 Despite this recurrent and historic contamination, only few studies consider the nutritional
111 quality of the seafood products from this worldwide important shellfish farming and fishing
112 area (Guérin *et al.*, 2011). In this context, this study assessed the quality of wild or extensive

113 farmed seafood from a unique geographical area, both in terms of FAs and trace elements, and
114 estimated concomitantly the risk and the benefice of their consumption. A total of sixteen
115 marine highly consumed species including fish, crustaceans and mollusks were collected within
116 the Pertuis Charentais area in winter. Their compositions in fatty acids (FA) and trace elements
117 were determined. Based on these data, the exposure of local seafood consumers to beneficial
118 FAs and essential elements, and to potentially non-essential metallic contaminants was
119 assessed. Hazards and benefits related to seafood consumption were characterized using
120 national and international recommendations and by applying composite metrics.

121

122 **2. Materials and methods**

123

124 *2.1. Ethics statement*

125 The species sampled are not protected or endangered species in the fishing area of the Pertuis
126 Charentais. No field permits or ethical approvals were required for this study, as all species
127 originated from commercial fisheries and were already dead when provided to us. Fish were
128 sacrificed by the commercial fishers at sea using standard fisheries practices (all fish were dead
129 when landed).

130

131 *2.2. Sample collection and preparation*

132 Sixteen marine species were purchased from a local fishmonger between November 2018 and
133 February 2019, including fish (11 species including 1 cartilaginous and 10 teleost species),
134 crustaceans (1 species), cephalopods (2 species) and bivalves (2 species; Table 1). Animals
135 were fished maximum 2 days before being purchased and conserved on ice since fishing. Two
136 brands (“Spéciale”, *i.e.* cultured in coastal waters, and “Fine de Claire”, *i.e.* refined for
137 minimum 4 weeks in saltmarsh clay ponds) of cupped oyster species coming from a shellfish
138 farmer of Oléron Island were purchased. All the specimens were weighed and measured (Table
139 1). On each individual, the edible part (*i.e.* muscle for fish, cephalopods, crustaceans, and
140 muscle as well as gonad for the great Atlantic scallops) was collected in duplicate. One replicate
141 was used for FAs analyses and the other one for trace element analyses. For oysters, the entire
142 soft edible tissues of two oyster individuals were pooled and split in two subsamples for FAs
143 and trace element analyses. All samples dedicated to further FA analyses were directly dropped
144 into liquid nitrogen and then stored at -80°C. Samples dedicated to trace element analyses were
145 wet weighed (ww) and stored at -20°C.

146

147

2.3. FAs analysis

148

149

150

151

152

153

154

155

156

157

158

159

160

161

162

163

164

165

166

167

168

169

170

171

172

173

174

175

176

177

178

The platform “LIPIDOCEAN” (UMR 6539 - Laboratory of Environmental Marine Sciences, Plouzané, France) fulfilled the determination of FAs qualitative and quantitative compositions in specimens, according to the protocol described in Mathieu-Resuge et al. (2020), except that we performed analyses on the total lipid fraction without separating neutral and polar lipids. Briefly, frozen muscles and soft tissues were firstly homogenized by ball mill in liquid nitrogen. Total lipids from approximately 250 mg of tissue powder were extracted in 6 mL of chloroform-methanol (2:1, v/v). The total lipid extract was then sonicated 5 min at 4°C and stored at -20°C under N₂ gas. An aliquot of the total lipid extract (1 out of 6 mL) was transmethylated for 10 min at 100°C, after evaporation to dryness and addition of 2.3 µg of an internal standard (tricosanoic acid C23:0) and 800 µL of methanol/H₂SO₄ (3.4%; v/v). Resulting FA methyl esters (FAME) were recovered with 800 µL of hexane and washed 3 times with 1.5 mL of hexane-saturated distilled water. FAME were then analyzed by gas chromatography coupled to a flame-ionization detector (GC-FID) on a Varian CP8400 gas chromatograph equipped with splitless injectors. FAME were separated simultaneously on two columns, one polar (ZBWAX: 30 m × 0.25 mm ID × 0.2 µm, Phenomenex) and one apolar (ZB5HT: 30 m × 0.25 mm ID × 0.2 µm, Phenomenex). FAME were identified by comparison of their retention time on both columns with those of commercial standards or lab-made standards mixtures (chromatograms are presented in supplementary materials). FA were named as C:Yn-Z where C is the number of carbon of the aliphatic chain, Y, the number of unsaturation and Z the position of the 1st unsaturation from the terminal carbon.

The FAs analysis procedure was assessed by comparing the quantities measured with C23:0 with theoretical amount of each FA present in a standard mixture of different FA included into different lipid classes in different proportions to attest to the eventual impact of potential bias on the different FA, different class of lipid, proportion considered as well as their initial quantity (50 µg vs 100 µg vs 150 µg). The repeatability was estimated with 4.2% for the GC and 11.9% of variation for the whole FA analysis (Sardenne *et al.*, 2021). A blank was realized for each sample series (1 blank every 14 samples). Blanks follow exactly the same analytical process as samples, from lipid extraction to GC analysis. Blank subtraction was performed as they contained 16:0 and 18:0 traces. The calibration was performed using a FA mixture of known theoretical mass composition of Supelco® 37 Component FAME Mix. The mass percentage values calculated for this certified standard was compared to its known certified values and

179 remained within the 5% of error for the polar column and 6% of error for the apolar column.
180 The GC-FID analyses conducted during our experiment can then be considered as suitable to
181 obtain correct semi-quantitative values for all the fatty acids.

182 The FAME quantification was made using the standard C23:0 added to each sample before
183 transesterification. The equation used was the following:

$$184 \quad \text{Quantity FA (g)} = (\text{Area FA} \times \text{Quantity C23:0}) / \text{Area C23:0}$$

185 Concentration of each FA was estimated by considering the mass of tissue extracted, the total
186 volume of lipid extract and the aliquote volume of the lipid extract used for lipid analyses.

187 The results of each individual FA were expressed as mass percent of the total FA content (%
188 TFA) and were also given as concentrations (in mg of FA per g of wet sample, mg g^{-1} ww).

189

190 *2.4. Trace elements analysis*

191 Frozen samples were freeze-dried for 36 to 48 hours (Chris® BETA 1-8 LDplus). Then, they
192 were weighed for dry weight (dw) before being homogenized in a porcelain mortar and pestle.

193 Aliquots ~ 200 mg dw tissues were microwave digested with a mixture of 3ml of suprapure
194 nitric acid (VWR/Merck) and 1ml of suprapure chlorhydric acid (VWR/Merck). Trace elements

195 (Ag, As, Cd, Cu, Fe, Mn, Pb, Se, and Zn) were analyzed using an Inductively Coupled Plasma
196 Mass Spectrometry (ICP-MS II Series Thermo Fisher Scientific) and an Inductively Coupled

197 Plasma Atomic Emission Spectrometry (Varian Vista-Pro ICP-AES), as described by
198 Kojadinovic *et al.* (2011). The limits of detection (LOD) ranged from $0.01 \mu\text{g g}^{-1}$ dw (e.g. Ag,

199 Cd, Pb) to $5 \mu\text{g g}^{-1}$ dw (Fe). Accuracy and reproducibility were assessed by analyzing
200 procedural blanks and Certified Reference Material (CRM) (DOLT-5 dogfish liver from

201 National Research Council, Canada and IAEA-436 tuna fish flesh homogenate from the
202 International Atomic Energy Agency IAEA). Recovery rates were $96 \pm 11 \%$ for DOLT-5 (from

203 74% to 118%) and $103 \pm 8 \%$ for IAEA-436 (from 93% to 120%).

204 For Hg analysis, aliquots ranging from 5 to 10 mg dw were analyzed with an Advanced Mercury
205 Analyser spectrophotometer (Altec AMA 254, LOD of 0.05 ng). The accuracy was checked

206 using the CRM DOLT-5 with certified Hg concentration: $0.350 \pm 0.005 \mu\text{g g}^{-1}$ dw. Analyses
207 were repeated, for each individual, twice or three times until getting a relative standard

208 deviation (SD) $< 10 \%$. All the trace elements concentrations were obtained in $\mu\text{g g}^{-1}$ dw and
209 then transformed and expressed in $\mu\text{g g}^{-1}$ ww throughout the manuscript.

210

211 *2.5. Data treatment and statistics*

212 Prior to data analysis, values of trace element concentrations below the limit of detection (LOD)
213 were replaced by the lowest measured value of the corresponding element multiplied by 0.5
214 (Guérin *et al.*, 2011; Olmedo *et al.*, 2013ab).

215 A Principal Component Analysis (PCA) was fulfilled to study relationships between FA and
216 trace element concentrations in species, using ‘FactoMineR’ (Le *et al.*, 2008) and ‘Factoextra’
217 (Kassambara *et al.*, 2017) packages on R. Individuals with non-available value (NA) of trace
218 elements concentrations were removed from the PCA. PCA was based on correlation matrix
219 and normalized data, centered and divided by the standard deviation, for each variable.

220 The comparisons of FA and trace element concentrations between the species were tested using
221 means comparison tests, using ‘ggrepel’ (Slowikowski, 2019), ‘tidyverse’ (Wickham, 2017),
222 ‘cowplot’ (Wilke, 2019), and ‘multcompView’ (Graves *et al.*, 2015) packages on R. One-way
223 ANOVA and Tukey’s post-hoc tests were also performed. The conditions of application of
224 parametric tests were determined *a posteriori* by checking the normality and homoscedasticity
225 of data residuals using a Shapiro-Wilk test and a Bartlett test respectively. If these conditions
226 were not respected, non-parametric tests (Kruskal-Wallis and Wilcoxon signed-rank test
227 without adjustment method) were used. The significant level of statistical analyses was set at α
228 = 0.05. Species with individuals having trace elements concentrations below the LOD were not
229 considered for the inter-specific comparison. Results presented with boxplots were arranged in
230 descending order for each taxon, based on the means.

231 The nutritional quantity and quality in terms of FA of the studied species was compared using
232 a hierarchical clustering analyses focusing on (i) two nutritional quantity indices (real DHA and
233 EPA available quantity in biomass), and (ii) six quality indices (LC n-3 PUFA/TFA, n-6/n-3,
234 EPA/TFA, DHA/TFA, the atherogenic index (AI) and the thrombogenic index (TI) – see below)
235 in the different species using ‘Pretty Heatmaps’ (Kolde, 2019) and ‘ColorBrewer Palettes’
236 (Neuwirth, 2014) packages on R.

237 All data analyses and graphical representations were performed with R version 3.5.0 (R Core
238 Team, 2018).

239

240 2.6. Risk and benefit assessment

241 Considering essential and non-essential element concentrations, as well as FA composition,
242 some indices were calculated to estimate risk and benefit of seafood intake.

243

244 2.6.1. n-6/n-3 PUFA ratio

245 The ratio between n-6 and n-3 series concentrations, named as n-6/n-3, was determined for each
 246 species. The FA considered were: 16:2n-6, 16:3n-6, 18:2n-6, 18:3n-6, 18:4n-6, 20:2n-6, 20:3n-
 247 6, 20:4n-6, 22:2n-6, 22:4n-6, 22:5n-6 for n-6 series; and: 16:3n-3, 16:4n-3, 18:3n-3, 18:4n-3,
 248 18:5n-3, 20:3n-3, 20:4n-3, 20:5n-3, 21:5n-3, 22:5n-3, 22:6n-3 for n-3 series. A n-6/n-3 ratio in
 249 seafood between 1:1 and 4:1 indicates an improvement in the balance of the contribution of
 250 FAs in the human diet and to prevent the onset of coronary heart disease (Simopoulos, 2002).

251

252 2.6.2. *Atherogenic and thrombogenic indices*

253 The dietary factors involved in the onset of coronary heart disease are directly correlated to
 254 qualitative aspects of the lipid fraction. The MUFA and the PUFA of the series n-3 and n-6
 255 seem to have equal role in the prevention of thrombus development, while the n-3 PUFA seem
 256 to be more important in the limitation set atheroma. Long-chain SFA (14:0, 16:0, 18:0)
 257 accelerate thrombus formation, by reducing the production of arterial prostacylin, a strong
 258 antagonist of platelet aggregation, while many studies indicate that long-chain unsaturated fatty
 259 acids slow down intra-arterial occlusion and platelet aggregation. Ulbricht and Southgate
 260 (1991) summed up these numerous effects through equations for the calculations of the index
 261 of thrombogenicity index (TI) and the atherogenicity index (AI). In these indices different
 262 weights are attributed to these categories of FA in relation to their different contribution to the
 263 prevention or promotion of intra-arterial occlusion and platelet aggregation. The atherogenic
 264 (AI) and thrombogenic (TI) potentials of a resource were thus evaluating according to Ulbricht
 265 & Southgate (1991) formula:

266

267 $[AI = (12:0 + 4*14:0 + 16:0) / ((n-6 \text{ PUFA} + n-3 \text{ PUFA}) + 18:1n-9 + \text{other MUFA})]$ (in mg g⁻¹),
 268 ¹,

269 where 12:0, 14:0, 16:0 are saturated FAs; n-6 PUFA and n-3 PUFA are, respectively, the sum
 270 of the polyunsaturated FAs from n-6 series (16:2n-6, 16:3n-6, 18:2n-6, 18:3n-6, 18:4n-6, 20:2n-
 271 6, 20:3n-6, 20:4n-6, 22:2n-6, 22:4n-6, 22:5n-6) and n-3 series (16:3n-3, 16:4n-3, 18:3n-3,
 272 18:4n-3, 18:5n-3, 20:3n-3, 20:4n-3, 20:5n-3, 21:5n-3, 22:5n-3, 22:6n-3); 18:1n-9 is a
 273 monounsaturated FA and MUFA is the sum of all other monounsaturated FAs (14:1n-5, 16:1n-
 274 11, 16:1n-9, 16:1n-7, 16:1n-5, 18:1n-11, 18:1n-7, 18:1n-5, 20:1n-11, 20:1n-9, 20:1n-7, 22:1n-
 275 11, 22:1n-9, 22:1n-7, 24:1n-9). But the 12:0 was not detected in the present study, so it was not
 276 considered in the formulation.

277

278 $[TI = (14:0 + 16:0 + 18:0) / (0.5*18:1n-9 + 0.5*other\ MUFA + 0.5*n-6\ PUFA + 3*n-3\ PUFA$
 279 $+ (n-3\ PUFA / n-6\ PUFA))]$ (in mg g⁻¹),

280

281 where 14:0, 16:0, 18:0 are saturated FAs; 18:1n-9 is a monounsaturated FA; MUFA is the sum
 282 of monounsaturated FAs; n-6 PUFA and n-3 PUFA are the sum polyunsaturated FAs from n-6
 283 and n-3 series respectively, as detailed for the AI.

284

285 2.6.3. *Daily recommended portion and maximum safe consumption*

286 A daily recommended portion was determined for each species and corresponded to the intake
 287 (expressed in g per day) needed to achieve the 250 mg EPA + DHA daily dietary requirement
 288 for an adult (FAO/WHO, 2010). It was thus a function of the sum of EPA and DHA
 289 concentrations measured per species. Then, results were compared to the European value of a
 290 serving, i.e. 150 g (Roth & Knai, 2003).

291 The maximum safe consumption of each species considering their supply in each trace elements
 292 was estimated through the Maximum Safe Consumption calculation (MSC) (Metian *et al.*,
 293 2013), for a trace element A:

294

295 $[MSC_A = (W_{ind} * JL_A) / X_A]$ (in g ww per time unit),

296

297 where W_{ind} is the mean human body weight (bw, average of 70 kg); JL_A is the Provisional
 298 tolerable monthly intake (PTMI, in $\mu\text{g kg}^{-1}$ ww bw) or the Provisional Tolerable Weekly Intake
 299 (PTWI, in $\mu\text{g kg}^{-1}$ ww bw) or Provisional Maximum Tolerable Daily Intake (PMTDI, in $\mu\text{g kg}^{-1}$
 300 ww bw) of A; X_A is the mean concentration of A (in $\mu\text{g g}^{-1}$ ww) in seafood. Data under the
 301 LOD were not considered in the calculations of indices.

302

303 3. Results

304

305 3.1. *FA and trace elements and concentrations in seafood*

306 FAs (TFA, LC n-3 PUFA, DHA, and EPA) and trace elements (Ag, As, Cd, Cu, Fe, Hg, Mn,
 307 Pb, Se, and Zn) concentrations in edible tissues of seafood species studied were included in a
 308 PCA (Fig. 1). The first and the second principal components, with respectively 43% and 26%,
 309 accounted for 69% of the total variation in the analysis. Concentrations of Ag, Cd, Cu, Fe, Mn,
 310 Pb, and Zn contributed the most to the first dimension, while TFA, LC n-3 PUFA, EPA, and

311 DHA concentrations contributed the most to the second principal component. The As, Hg, and
312 Se concentrations were not explained by the two first components according to the correlation
313 circle (Fig. 1A). No correlation was observed between FAs and trace elements concentrations.
314 The projection of individuals showed that the majority of the species were gathered at the origin
315 of the PCA, meaning they were not discriminated by their FA nor trace element contents (Fig.
316 1B). Nevertheless, the PCA strongly discriminated the cupped oysters “Fine de Claire” (OFC)
317 and “Spéciale” (OSP), the gonad of great Atlantic scallop (GSG) from the other species
318 according to the first component because of their high trace element concentrations. In contrast,
319 the Atlantic mackerel (MKR), the surmullet (SRM), and the European pilchard (PIL) were
320 discriminated according to their high FA concentrations, explained by the second component
321 (Fig. 1B).

322

323 3.2. Comparison of fatty acids profile between seafood species

324 The species with the highest TFA concentrations had the highest LC n-3 PUFA and others FA
325 concentrations (Fig. 2; for more detailed see Table S1). The mackerel (MKR), the surmullet
326 (SRM), and the pilchard (PIL) had the highest LC n-3 PUFA concentrations ($29.8 \pm 14.0 \text{ mg g}^{-1}$,
327 $21.0 \pm 8.2 \text{ mg g}^{-1}$, and $10.6 \pm 2.7 \text{ mg g}^{-1}$, respectively). In descending order, these species
328 were followed by the meagre (MGR), the seabass (SBS), and the hake (HAK), for which
329 concentrations were around 4.5 mg g^{-1} (Table 1). The black seabream (SBR), the whiting
330 (WTG), the common sole (SOL), the John Dory (JDO), and the dogfish (DOG) were the fish
331 with the lowest LC n-3 PUFA concentration (below $1.9 \pm 0.3 \text{ mg g}^{-1}$ and up to $0.8 \pm 0.1 \text{ mg g}^{-1}$).
332 Concerning cephalopods, the squid (SQD) displayed a higher LC n-3 PUFA content ($3.3 \pm$
333 0.6 mg g^{-1}) than the cuttlefish (CTF; $1.8 \pm 0.1 \text{ mg g}^{-1}$). The LC n-3 PUFA concentration of
334 spider crab (SPI) was $1.0 \pm 0.1 \text{ mg g}^{-1}$, similar to those found in the great scallop (GSM) muscle
335 ($1.1 \pm 0.1 \text{ mg g}^{-1}$). Finally, the two oysters (OSP and OFC) and the scallop gonad (GSG)
336 displayed concentrations of $3.2 \pm 1.2 \text{ mg g}^{-1}$, $2.7 \pm 1.1 \text{ mg g}^{-1}$, and $4.7 \pm 1.2 \text{ mg g}^{-1}$ respectively.
337 Proportionally to their TFA content, the muscles of whiting (WTG), cuttlefish (CTF), squid
338 (SQD), and scallop (GSM) were the items with the most important fraction of LC n-3 PUFA
339 (upper than 45 % of TFA, Fig. 3, Table S2). The fraction of LC n-3 PUFA in the other species
340 varied between $30.4 \pm 5.3 \%$ of TFA in mackerel (MKR) and $38.3 \pm 2 \%$ of TFA in John Dory
341 (JDO). The only exception was surmullet (SRM), which contained the lowest LC n-3 PUFA
342 relative proportion with only $21.4 \pm 2.7 \%$ of TFA.

343 EPA and DHA proportions, which represented an important part of the LC n-3 PUFA, varied
344 with the proportion of LC n-3 PUFA on TFA (Fig. 3, Table S2). Globally, DHA concentrations
345 were higher than EPA concentrations, except for oysters “Spéciale” (OSP), “Fine de Claire”
346 (OFC) and spider crab (SPI). The DHA fraction ranged from 9.6 ± 1.3 % of TFA in surmullet
347 (SRM), to 39.1 ± 2.8 % of TFA in whiting (WTG), while the EPA fraction varied from $4.2 \pm$
348 0.6 % of TFA in dogfish (DOG) to 23.7 ± 1.6 % of TFA in spider crab (SPI).

349 For the cupped oysters, the fraction of LC n-3 PUFA was higher in “Fine de Claire” (OFC; 35
350 ± 2.1 % of TFA) than in “Spéciale” (OSP; 32.7 ± 1.8 % of TFA). The EPA fraction was not
351 significantly different between these two brands, but the DHA fraction was higher in the oyster
352 “Fine de Claire” (OFC; 15.2 ± 1.7 % of TFA) than in the “Spéciale” one (OSP; 13.2 ± 1.2 % of
353 TFA).

354

355 *3.3. Risk and benefit assessment of seafood consumption, regarding fatty acid* 356 *content*

357 The daily recommended portion of seafood varied logically according to the concentrations of
358 EPA and DHA found in edible part of consumed species (Table 2). Among the seafood species
359 analyzed, only five required more than 240 g of a portion to reach the 250 mg EPA + DHA
360 daily dietary requirement (i.e. the common sole - SOL, the John Dory - JDO, the dogfish -
361 DOG, the spider crab - SPI and the scallop muscle - GSM). Contrasting this, only 15.8 ± 19.6
362 g of the mackerel (MKR), 19.1 ± 16.7 g of the surmullet (SRM), and 28.1 ± 10.0 g of the
363 pilchard (PIL) were necessary to achieve the recommendation.

364 The benefit of seafood consumption was assessed through the n-6/n-3 ratio, the AI and the TI
365 calculations (Fig. 3, Table S3). The spider crab (SPI), the seabream (SBR), the common sole
366 (SOL), and the surmullet (SRM) displayed the highest n-6/n-3 ratio (ranging from 0.28 ± 0.04
367 to 0.33 ± 0.15), whereas the values for the other species ranged from 0.11 ± 0.01 (for the
368 whiting; WTG) to 0.21 ± 0.08 (for the dogfish; DOG). Nonetheless, the pilchard (PIL; $0.10 \pm$
369 0.02) and the two cephalopods (SQD; 0.04 ± 0.01 and CTF; 0.08 ± 0.02) had a n-6/n-3 ratio \leq
370 0.1 . These three species also showed the highest AIs with 0.53 ± 0.07 , 0.62 ± 0.03 and $0.45 \pm$
371 0.03 for the pilchard, the squid and the cuttlefish, respectively. The lowest AIs were reported
372 for the spider crab (SPI; 0.12 ± 0.01) and the whiting (WTG; 0.26 ± 0.01), while for the other
373 species the AI ranged from 0.029 ± 0.05 (OFC) to 0.40 ± 0.04 (SBR). Concerning the TI, the
374 spider crab (SPI), the whiting (WTG), the cuttlefish (CTF) and the muscle of the scallop (GSM)

375 displayed the lowest values (all equal to 0.06 ± 0.01). The other species had TI comprised
376 between 0.07 ± 0.01 for the squid and 0.25 ± 0.02 for the surmullet.

377

378 3.4. Trace element concentrations in seafood

379 The trace element concentrations in edible tissues varied in seafood species from less than 0.001
380 $\mu\text{g g}^{-1}$ for Ag, Cd, and Pb to more than $200 \mu\text{g g}^{-1}$ for Zn (Fig. 4A and 4B, Table S4). For all
381 species, the concentrations of non-essential trace elements were lower than the essential ones,
382 with concentrations ranging such as $\text{Zn} > \text{Fe} > \text{As} > \text{Cu} > \text{Mn} > \text{Se} > \text{Ag} > \text{Hg} > \text{Cd}$, and Pb,
383 based on the overall mean.

384 The concentrations of non-essential trace elements, Ag and Cd, in the muscles of all fish and
385 cephalopods, were under the limit of detection, and Pb was only significantly measured in the
386 squid (SQD, $0.004 \pm 0.001 \mu\text{g g}^{-1}$). The highest concentration of Ag was measured in the cupped
387 oysters “Spéciale” (OSP; $0.66 \pm 0.52 \mu\text{g g}^{-1}$) and “Fine de Claire” (OFC; $0.58 \pm 0.68 \mu\text{g g}^{-1}$)
388 but was also found in both tissues of the great scallop (GSG; $0.19 \pm 0.08 \mu\text{g g}^{-1}$ and GSM; 0.006
389 $\pm 0.002 \mu\text{g g}^{-1}$), and in the spider crab (SPI; $0.07 \pm 0.02 \mu\text{g g}^{-1}$). Cd was detected in the spider
390 crab (SPI; $0.013 \pm 0.003 \mu\text{g g}^{-1}$) but oysters and scallops had the highest concentrations (OSP;
391 $0.27 \pm 0.04 \mu\text{g g}^{-1} > \text{GSG}; 0.19 \pm 0.16 \mu\text{g g}^{-1} > \text{OFC}; 0.18 \pm 0.06 \mu\text{g g}^{-1} > \text{GSM}; 0.16 \pm 0.03$
392 $\mu\text{g g}^{-1}$). The gonad of the great scallop displayed the highest concentration of Pb (GSG; $0.15 \pm$
393 $0.17 \mu\text{g g}^{-1}$), followed by the oysters, both “Fine de Claire” (OFC; $0.12 \pm 0.03 \mu\text{g g}^{-1}$) and
394 “Spéciale” (OSP; $0.07 \pm 0.02 \mu\text{g g}^{-1}$). In a lesser extent, Pb was found in the spider crab (SPI;
395 $0.02 \pm 0.003 \mu\text{g g}^{-1}$) and in the muscle of the scallop (GSM; $0.009 \pm 0.012 \mu\text{g g}^{-1}$).

396 The highest Hg concentrations were measured in the dogfish (DOG; $0.40 \pm 0.12 \mu\text{g g}^{-1}$) and the
397 seabass (SBS; $0.29 \pm 0.07 \mu\text{g g}^{-1}$). The flesh of the other fish and cephalopod showed
398 concentrations of Hg comprised between $0.03 \pm 0.01 \mu\text{g g}^{-1}$ in the pilchard (PIL) and $0.09 \pm$
399 $0.02 \mu\text{g g}^{-1}$ in the seabream (SBR). The lowest Hg concentrations were measured in the bivalves
400 (the both oyster brands and the both scallop tissues) with a maximum of $0.015 \pm 0.003 \mu\text{g g}^{-1}$
401 found in the “Fine de Claire” oyster (OFC).

402 Among the six essential trace elements studied, Cu, Mn and Zn were the most present in the
403 bivalves, especially the oyster “Spéciale” (OSP) and the “Fine de Claire” (OFC) with $18.13 \pm$
404 $9.19 \mu\text{g g}^{-1}$ and $9.68 \pm 2.99 \mu\text{g g}^{-1}$ of Cu, respectively, with $7.16 \pm 1.87 \mu\text{g g}^{-1}$ and 4.28 ± 1.83
405 $\mu\text{g g}^{-1}$ of Mn, respectively, and with $211.7 \pm 69.8 \mu\text{g g}^{-1}$ and $208.6 \pm 78.6 \mu\text{g g}^{-1}$ of Zn,
406 respectively. The results showed that the concentrations of Fe in the hake, the sole, the whiting,

407 and the John Dory were under the LOD, while the highest concentrations of this essential
408 element were found in the gonad of the great scallop (GSG; $32.5 \pm 17.8 \mu\text{g g}^{-1}$) and in the
409 oysters (OFC; $29.9 \pm 6.6 \mu\text{g g}^{-1}$ and OSP; $24.5 \pm 3.4 \mu\text{g g}^{-1}$).

410 It is noteworthy that flesh of the dogfish (DOG) and the spider crab (SPI) displayed the highest
411 concentrations of As, with $37.00 \pm 9.31 \mu\text{g g}^{-1}$ and $21.10 \pm 3.05 \mu\text{g g}^{-1}$, respectively. Finally,
412 the spider crab (SPI) and the great scallop (GSG) displayed the most important concentration
413 of Se, respectively $1.26 \pm 0.16 \mu\text{g g}^{-1}$ and $1.00 \pm 0.24 \mu\text{g g}^{-1}$.

414

415 3.5. *Risk and benefit assessment of seafood consumption regarding the trace* 416 *element concentration*

417 The Maximum Safe Consumption (MSC) showed that the consumption of some species was
418 narrowed by the concentrations of some trace elements, when others did not seem to be
419 restrictive (Table 2). For all the seafood species studied, the MSCs indicated a safe intake of
420 more than (i) $2.0 \pm 0.4 \text{ kg}$ per day with respect to the Cu and Fe concentrations in the edible
421 tissues, (ii) $6.7 \pm 1.0 \text{ kg}$ per month regarding to the Cd, and (iii) $15.0 \pm 3.0 \text{ kg}$ per week regarding
422 to the Pb. The concentrations of Hg were those limiting the most the weekly consumption of
423 the John Dory at $0.87 \pm 1.49 \text{ kg}$, the European seabass at $0.40 \pm 0.10 \text{ kg}$, and the dogfish at 0.29
424 $\pm 0.07 \text{ kg}$. Considering the As concentration, the daily MSCs varied between a consumption of
425 $0.03 \pm 0.01 \text{ kg}$ of dogfish, and $0.05 \pm 0.01 \text{ kg}$ of spider crab, to $1.15 \pm 0.50 \text{ kg}$ of Atlantic
426 mackerel. Regarding to the Zn concentration, the daily MSCs of the fish, the cephalopods, and
427 the muscle of the great scallop were upper than $2.75 \pm 0.40 \text{ kg}$, when MSCs were 0.67 ± 0.05
428 kg for the spider crab, $0.80 \pm 0.30 \text{ kg}$ for the gonad of the great scallop, $0.25 \pm 0.09 \text{ kg}$ for the
429 oyster “Fine de Claire”, and $0.24 \pm 0.09 \text{ kg}$ for the oyster “Spéciale”.

430

431 4. Discussion

432

433 The consumption of seafood leads to the intake of beneficial and detrimental molecules or
434 elements. This study demonstrates that the concentrations of essential and/or non-essential trace
435 elements, as well as of fatty acids, including LC n-3 PUFA, strongly differed among the
436 different species sampled in the Pertuis Charentais for this study, and were not correlated (Fig.
437 1). In response to a lack of local data on the seafood quality in one of the most productive and
438 touristic European coastal area, a non-exhaustive baseline and discussion of the risks and
439 benefit for local consumers is presented.

440

441 4.1. *Variations in fatty acid contents among seafood species*

442 For human diet, the highest fish quality in terms of FAs is reflected by a low TFA content, a
443 high quantity of LC n-3 PUFAs, combined with a low content of undesirable FAs, especially
444 saturated FAs, like 14:0 and 16:0, that are considered highly atherogenic (Abrami *et al.*, 1992).
445 Considering the TFA content, this study shows that, not surprisingly, the Atlantic mackerel and
446 the surmullet can be considered as fat fish, with TFA concentrations comprised between 80 and
447 100 mg g⁻¹ (Ackman, 1990; Médale, 2009; Sirot *et al.*, 2008). Conversely, the other fish species
448 as well as crustaceans, cephalopods, bivalves are considered as intermediaries (25 < TFA < 80
449 mg g⁻¹) or lean species (TFA < 25 mg g⁻¹). Surprisingly, the European pilchard which is usually
450 considered a fat fish, has in this study a TFA concentration of 28.3 ± 7.2 mg g⁻¹, placing it in
451 the group of intermediaries. The lipid content of muscle tissue of fat species can fluctuate
452 according to age, sexual cycle, trophic ecology, or environmental factors, such as temperature
453 (Médale *et al.*, 2009). As an illustration, pilchard caught in November in the Bay of Biscay
454 displayed a lower energy density than fish from the English Channel. This may result from
455 contrasted regional zooplankton productivity and delay in the spawning period (Gatti *et al.*,
456 2018) that occurs later in autumn (October-November) in the Biscay (Coombs *et al.*, 2006).
457 More generally, the unique sampling season in winter showed a temporal shot of the lipid
458 composition of marine resources, hiding seasonal variations, including the effect of
459 environmental factors on life history traits, that could be significant in fat species (Sirot *et al.*,
460 2008). Thus, designation of the pilchard of the present study as intermediary fish may be partly
461 attributed to the sampling of post-spawning individuals depleted in TFA (Aidos *et al.*, 2002).

462 The TFA quantity determined in the edible parts of the species does not necessarily inform
463 about the quality of the FAs profile, including the LC n-3 PUFA relative TFA proportion (Fig.
464 3). The consumption of fat fish such as mackerel and surmullet implies the intake of a high LC
465 n-3 PUFA content, but also an enhanced intake of other PUFAs, monounsaturated FAs
466 (MUFAs), and saturated FAs (SFAs). Contrasting to this, the consumption of lean species such
467 as cuttlefish, squid, and whiting brings an optimal LC n-3 PUFA intake relative to TFA content,
468 and in turn, a lower intake of SFA, making these species qualitatively and relatively more
469 beneficial (Abrami *et al.* 1992). A PUFA/SFA dietary ratio below 0.45 have been often
470 considered undesirable for the human diet because of their potential to increase cholesterol
471 concentrations in the blood (Zhang *et al.*, 2020, Ospina *et al.*, 2012). In the present study, this
472 ratio ranged from 0.88 to 3.07 (mean ± SD: 1.67 ± 0.40, results not shown), indicating that all

473 species seems beneficial in terms of PUFA proportion supply, while they provide a great
474 disparity of quality with respect to their SFA relative proportion.

475 In this study, LC n-3 PUFAs are predominantly composed of EPA and DHA. The higher
476 EPA+DHA proportion with respect to the TFA content, the greater the quality of the dietary
477 lipid source for human diet (Abrami *et al.*, 1992). These FAs are both the main structural
478 components of human cell membranes, but they do not play the same physiological role: EPA
479 is mainly involved in immune function and response to stress, while DHA is mostly involved
480 in the development and function of nervous system and brain, as well as in cardiovascular
481 function (Narayan *et al.*, 2006). The present study shows that bivalves, crustaceans, and to a
482 lesser extent, cephalopods, are a greater source of EPA than fish, while DHA is supplied
483 similarly by all taxonomic groups, except by the spider crab, the oysters and the surmullet,
484 which showed the lowest DHA proportion. While it has recently been observed that the
485 EPA/DHA ratio can vary spatially and temporally within a species, and even be reversed,
486 depending on many factors (trophic in particular, in *Sardina pilchardus*, F. Le Grand, pers.
487 com), these results demonstrate the necessity to diversify the consumption of seafood items to
488 allow complete and beneficial intakes, and to optimize the lipid profile for human consumption.

489

490 4.2. Risk and benefit assessment regarding to the fatty acid composition

491 Quantitatively, the daily requirement of EPA + DHA has been established at 250 mg, by the
492 JECFA (Joint FAO/WHO Expert Committee on Food Additives). Considering this, a serving
493 of 150 g (usual portion size) of the richest EPA and DHA species (*i.e.* Atlantic mackerel,
494 surmullet, pilchard, meagre, hake, seabass, squid, cuttlefish, oysters, and even the whiting) is
495 enough to reach this recommendation. Noteworthy, a serving of 4-5 scallops (~ 130 g including
496 ~100 g of muscles and ~30 g of gonads) contributes to 94% of the daily requirement of EPA +
497 DHA. However, more than one serving of 150 g of the poorest ones (*i.e.* spider crab, John Dory,
498 common sole and dogfish) are necessary (*i.e.* 275 ± 39 g, 270 ± 120 g, 355 ± 205 g and $391 \pm$
499 62 g for these 4 species respectively). Obviously, the recommendations may vary depending on
500 the target population: for example, the requirements for pregnant, breastfeeding women, or
501 elderly would be higher (daily requirement of 300 mg of EPA + DHA; FAO/WHO, 2010), due
502 to the beneficial effects of EPA and DHA on the development of infants or on the decline in
503 the onset of cardiovascular disease (Hellberg *et al.*, 2012; Ruxton *et al.*, 2004). Therefore, these
504 populations should preferably eat mackerel, surmullet, pilchard, meagre, great scallop with

505 gonads, hake, seabass, squid, or even cupped oyster “Spéciale” because a serving of these
506 species is sufficient to achieve their daily requirement.

507 It is noteworthy that the daily recommended portion for mackerel, sardine and surmullet showed
508 a high variability directly linked to the high inter-individual variability in their FA content. The
509 sampling winter season is a period of spawning for sardine and mackerel in the Bay of Biscay
510 (Alheit *et al.*, 2010). Reproductive status (i.e. gonad maturation stage, pre- or post-spawning
511 status) and sex (and thus the reproductive investment of energy and nutrient in gametes
512 production) could directly modulate the FA content, including EPA and DHA, among
513 individuals (Garrido *et al.*, 2007; Caponio *et al.*, 2004). These results raise again the question
514 of seasonal variability of the resources quality and its associated benefit for human consumers
515 with respect to the biology of targeted species.

516

517 Western diets are known to be deficient in n-3, and have excessive amount of n-6 (ratio n-6/n-
518 3 between 15:1 and 40:1), while the healthy ratio of n-6/n-3 in the human diet is recommended
519 to be between 1:1 and 4:1 (Simopoulos *et al.*, 2000; Simopoulos, 2003). This imbalance is due
520 to a reduced intake of fish, combined with an excessive amount of vegetable oils rich in linoleic
521 acid (LA; 18:2n-6) (Simopoulos, 2002, 2006). The n-3 and n-6 PUFAs compete for the same
522 enzymes for eicosanoid synthesis, but do not have the same role: n-3 PUFA-derived metabolites
523 have an anti-inflammatory effect, while n-6 PUFA-derived metabolites may have an
524 inflammatory effect (for a review, see Stupin *et al.* 2019). If the eicosanoid metabolic products
525 from n-6 PUFAs, such as arachidonic acid (20:4n-6) are formed in larger quantities than those
526 formed from EPA (20:5n-3), they will contribute to the formation of thrombus and atheromas,
527 allergic and inflammatory disorders, and cell proliferation. The higher the n-6 /n-3 ratio, the
528 higher the death rate from cardiovascular disease (Simopoulos, 1991, 2006). In this study, the
529 ratios were all much lower than the maximum 4:1 recommended, indicating that the sampled
530 seafood can therefore help to reduce the gap in the total diet by the intake of n-3 PUFA and
531 thus, help to stave off the cardiovascular or inflammatory diseases (Simopoulos, 2002).
532 However, this ratio might lead to simplistic dietary advice, as it does not consider the intake of
533 specific FAs. As mentioned earlier, the highest fish quality is reflected by a simultaneous low
534 TFA content, a high quantity of LC n-3 PUFAs (especially EPA and DHA), as well as a low
535 content of undesirable FAs (*i.e.* SFA, MUFA and n-6 FA; Abrami *et al.*, 1992). For that reason,
536 the AI and TI indices based on functional effects of FAs were also employed in this study to
537 conduct a comprehensive evaluation of the nutritional quality of the studied species.

538 The AI and TI indices are related to the atherogenicity and thrombogenicity of saturated FAs,
539 such as 16:0 or 18:0. Higher the AI and TI values, higher the platelet aggregation and
540 subsequent thrombus and atheroma formation in the cardiovascular system (Valfré *et al.*, 2003).
541 Among the fish sampled in this study, the whiting and the pilchard can be considered as the two
542 most beneficial fish species for the human diet: they both have the lowest n-6/n-3 ratio.
543 Moreover, the whiting, which is a lean fish with high LC n-3 PUFA fraction, presented the
544 lowest AI and TI indices and covers the EPA+DHA supply in one single meal. In contrast, the
545 pilchard, considered as a fat fish with a lower LC n-3 PUFA proportion is a 5-fold higher source
546 of DHA and EPA, but it presented the highest AI among fish, and one of the highest TI among
547 the studied species. The surmullet, despite providing a high EPA+DHA supply, considered as
548 fat with the lowest LC n-3 PUFA and DHA fraction on TFA, has the highest n-6 /n-3 ratio and
549 the highest TI value, indicating a lower beneficial intake of this species. Concerning other taxa
550 studied here, the squid and the cuttlefish could be considered as beneficial, as they are lean
551 species presenting a high proportion of LC n-3 PUFA (including DHA and EPA), the lowest n-
552 6/n-3 ratio, a low TI, while they also stand out with the highest AI.

553

554 4.3. Variation factors of trace element concentrations

555 The concentration of trace elements varied widely between species, and usually depend on (i)
556 the accumulation pathways (accumulation through contaminated prey consumption versus
557 seawater dissolved contaminant bioconcentration), (ii) the individual characteristics (*e.g.* sex,
558 age), as well as (iii) environmental conditions (*i.e.* seawater temperature) (Rainbow, 1997;
559 Sokolova & Lannig, 2008). The high trace element concentrations found in bivalves is linked
560 to the presence of gills, digestive gland, and gonads, that are known to efficiently concentrate
561 trace elements (Bustamante & Miramand, 2004; Metian *et al.*, 2008) in the edible tissues (El
562 Moshely *et al.*, 2014; Ersoy & Çelik, 2009; Geffard *et al.*, 2001). Also, the gonads of scallop
563 are 1.2-fold more concentrated in Cd to 36-fold more concentrated in Ag than their muscle
564 tissue. Thus, seafood consumed as a whole organism contributes globally much more than
565 muscular flesh of fish, cephalopods and crustaceans to trace elements (except Hg) intake for
566 consumers. Secondly, the filter-feeders such as cupped oysters tend to accumulate more trace
567 elements than other species because they directly filter large volumes of water that can be rich
568 in trace elements from suspended particulate matter (El-Moselhy *et al.*, 2014). Interestingly,
569 the concentration of Cd, Cu, Fe, Mn, and Pb vary significantly among the two brands of cupped
570 oysters and may be attributed to different farming methods. While the oyster “Spéciale” grows

571 up in the open ocean, the “Fine de Claire” finishes its growth in shallow clay ponds in marshes
572 dependent on the arrival of freshwater from the watershed with the presence of navicular
573 microalgae greening the oysters from Marennes-Oléron area. Our results raised the question of
574 the influence of these contrasting environmental conditions on the trace element concentrations
575 in cultured oysters.

576 It is noteworthy that muscle of carnivorous fish, cephalopods and crustaceans displayed higher
577 concentrations in Hg than bivalves, consistent with Hg biomagnification along the trophic webs
578 (*e.g.* Storelli *et al.*, 2007; Coelho *et al.*, 2010; Ersoy & Çelik, 2009), and to the tropism of
579 methylmercury (MeHg) that binds tightly to the sulfhydryl groups of muscular proteins (Bloom,
580 1992). Thus, seabass showed the second highest Hg concentrations ($0.289 \pm 0.071 \mu\text{g g}^{-1}$), due
581 to its high trophic position (*e.g.* Chouvelon *et al.*, 2012). Surprisingly, the Hg concentrations in
582 the meagre, *i.e.* a predator in the same trophic position, remained relatively low in comparison
583 ($0.065 \pm 0.007 \mu\text{g g}^{-1}$). Although seabass and meagre individuals were of similar size (475 ± 12
584 mm and 431 ± 11 mm, respectively), the meagre specimens were considered younger than
585 seabass ones (1 yr *vs.* 5-6 yr old, respectively). Hg concentrations tend to increase with fish age,
586 generally proxied by size (Storelli *et al.*, 2007; Abreu *et al.*, 2000; Chouvelon *et al.*, 2012), as
587 a result of a longer dietary exposure and a poor excretion of assimilated Hg. In addition, young
588 meagre display a trophic regime based on crustaceans, *i.e.* poor Hg prey (Hubans *et al.*, 2017)
589 before switching towards a piscivorous diet (*i.e.* Hg enriched fish prey), limiting again the Hg
590 intake in these individuals. These results raise the question of maximum Hg concentrations
591 recorded in bigger seabass and meagre that could be usually found in seafood markets. In
592 addition, the dogfish showed the highest Hg concentrations, as already observed in the literature
593 (Storelli *et al.*, 2005a; Chouvelon *et al.*, 2012). This species, such as other benthic species living
594 in close association with sediment in which they bury and from where they mainly feed, is more
595 exposed to Hg and MeHg sediment-associated contamination than pelagic species (Storelli *et*
596 *al.*, 2003c; 2006a). Nevertheless, the general higher Hg concentrations found in
597 Chondrichthyan in comparison with Actinopterygian suggest the influence of metabolic factors,
598 such as specific detoxication mechanisms (Chouvelon *et al.*, 2012).

599 Finally, it is worth noting that the highest As concentrations were found in benthic and
600 nektobenthic species, *i.e.* the dogfish, the spider crab, the common sole, the seabream, the
601 surmullet and the cuttlefish. These values are comparable to those reported in previous studies
602 for marine benthic species (Storelli *et al.*, 2005a; Sirot *et al.*, 2009). Such concentrations likely
603 result from their diet based on bottom living invertebrates (Storelli *et al.*, 2005a; Wu *et al.*,

604 2014), which are enriched by the As trapped in sediment. They also confirm that seafood is an
605 important source of As in human diet. Indeed, in France, seafood contributes to more than 60%
606 of the total dietary As supply (Leblanc *et al.*, 2006).

607

608 4.4. *Risk and benefit regarding to the presence of trace elements*

609 The International and European Regulation publishes the maximum concentrations (*i.e.*
610 maximum permissible levels) of Cd, Hg and Pb that regulates the commercialization and
611 consumption of seafood (EFSA, 2014). These limits established differ among taxa and species.
612 Regarding Cd, the concentrations could not exceed $0.5 \mu\text{g g}^{-1}$ in crustaceans and $1.0 \mu\text{g g}^{-1}$ in
613 bivalves and cephalopods. In fish muscle, Cd is permitted at a maximum of $0.25 \mu\text{g g}^{-1}$ in
614 pilchard, $0.1 \mu\text{g g}^{-1}$ in mackerel and $0.05 \mu\text{g g}^{-1}$ in the other species (EC, 2014). None of the
615 individual samples analyzed in this study exceed these limits, with the exception of one scallop
616 gonad that reached $0.52 \mu\text{g g}^{-1}$. Concerning Hg, the highest concentration found in dogfish (*i.e.*
617 $0.60 \mu\text{g g}^{-1}$) is well below the maximum levels of $1 \mu\text{g g}^{-1}$ permissible for sharks. Likewise, the
618 seabass individual displaying a Hg concentration of $0.39 \mu\text{g g}^{-1}$ in muscle did not exceed the
619 $0.5 \mu\text{g g}^{-1}$ threshold fixed for fish (EC, 2008; EC, 2011). Finally, the value of $0.61 \mu\text{g g}^{-1}$ Pb
620 recorded in scallop gonad is also below the maximum levels of $1.5 \mu\text{g g}^{-1}$ authorized for
621 bivalves (EC, 2015).

622 While seafood is a well-recognized source of proteins and FAs for humans, it also
623 contributes to the chronic intake of potentially harmful trace elements leading the JECFA to
624 establish endpoints representing the permissible human daily, weekly or monthly exposure to
625 both essential and non-essential elements, *i.e.* As, Cd, Cu, Fe, Hg, Pb, and Zn. Based on these
626 recommendations, the calculations of the maximum safe consumption (MSC) indicate a safety
627 intake with respect to essential Cu, Fe and Zn allowing fish and cephalopod meals of more than
628 3 kg per day until reaching the established limits. It is noteworthy that frequent seafood
629 consumers could still reach the MSC for Zn when eating ~ 240 g of oyster flesh that corresponds
630 approximatively to two dozen of oysters. In addition, the very high MSCs for Cd and Pb
631 highlighted that local seafood is safe for consumers with respect to these both non-essential
632 metals.

633 Although it is assumed that As plays an essential role for human health (Mayer *et al.*, 1993),
634 the JECFA established a PTWI at $15 \mu\text{g kg}^{-1}$ bw, limiting at first glance the consumption of
635 seabream, common sole, dogfish, surmullet, cuttlefish, and spider crab to less than one portion
636 per week (< 150 g). However, this recommendation refers to the toxicity of the inorganic As

637 (Ansari *et al.*, 2004; Hughes, 2002; Neff, 1997), whereas As found in marine organisms are
638 predominantly organic arsenical compounds (*i.e.* arsenobetaine) known to be far less toxic
639 (Borak & Hosgood, 2007; EFSA, 2009; Olmedo *et al.*, 2013b). Seafood should be thus
640 considered safe for consumers, whereas there is still a lack of data on toxicity of some organic
641 compounds, e.g. As-sugars found in seaweeds, bivalves and crustaceans, and their metabolites
642 produced during the digestive process (Taylor *et al.*, 2017).

643 This previous point highlights that regulation and recommendation with respect to
644 contaminated seafood is based on the total concentration of contaminant and does not consider
645 its metabolic bioavailability in tissue. Indeed, the subcellular distribution of metals in cytosolic
646 (*i.e.* metal free in the cytosol) and organelles fractions (e.g. metal bound to metalloproteins, or
647 entrapped in metal rich-granules) drives the proportion of elements available for absorption at
648 the intestinal level, defined as bioaccessibility for consumers (Wallace and Luoma, 2003).
649 Experimental work demonstrated that the bioaccessible fraction rarely exceed 80% of the total
650 concentration for Zn and ranged from 50 to 90% for Cd in raw mussels and oysters (Metian *et*
651 *al.*, 2009; Gao and Wang, 2014). Consequently, this oral bioaccessibility could help unravel
652 risk associated with consumption of contaminated seafood (Gao and Wang, 2014).

653 Finally, Hg remains the element of most concern for consumers, and more particularly the
654 MeHg which is the dominant and toxic form in seafood (Andersen & Depledge, 1997; Storelli
655 *et al.*, 2005b). The MSCs calculated on the basis of the PTWI of $1.6 \mu\text{g kg}^{-1}$ bw of MeHg
656 (JECFA) highlighted that less than 400 g (*i.e.* lower than three portions, ~ 450 g) per week for
657 the European seabass, as well as for the dogfish is enough to exceed a safety intake. Thus, the
658 consumption of these two species may be at risk for high seafood consumers, also considering
659 that the Hg intake might be enhanced when bigger seabass (> 1 kg) are eaten. The essential
660 metal Se is known as a protective antagonist against Hg toxicity (*e.g.* Burger & Gochfeld.,
661 2011), implying that the Se:Hg molar ratios exceeding 1 are protective for adverse Hg effects
662 (Ralston, 2008). The flesh of dogfish and seabass displayed Se in excess in relation to Hg with
663 Se:Hg ratios of 2.6 ± 0.9 and 4.1 ± 1.3 , respectively (data not shown), but these values are the
664 lowest compared to those of the other foodstuffs that have a ratio between 13 and 206 (*i.e.*
665 mackerel and scallop gonad, respectively). Even if the mechanisms of Hg toxicity neutralization
666 by Se are clearly known for consumers, these results might indicate the nutritional importance
667 of seafood that would provide enough Se benefit to balance the Hg harm (Ralston, 2008).

668

669 4.5. *Application to a concrete case*

670 Literature reports that MeHg can counteract the cardioprotective effects of fish consumption
671 (Guallar *et al.*, 2002; Salonen *et al.*, 1995) mainly brought by the LC n-3 PUFAs. Considering
672 simultaneously the LC n-3 PUFA and Hg concentrations, two constituents that have a
673 mechanistic basis for influencing cardiovascular outcome, the present analysis makes one to
674 use an index integrating these two parameters in the same calculation to illustrates both the risk
675 and benefit of seafood consumption. In this line, the index of net risk/benefit for cardiovascular
676 endpoints was calculated on a species-specific basis according to Ginsberg & Toal (2009). The
677 method subtracts risk of adult cardiovascular heart disease due to MeHg (23% higher risk/1
678 ppm hair Hg) from the benefit thanks to PUFA (14.6% lower risk/100 mg EPA+DHA). The
679 calculated index for each species (Fig. 5) showed that the relative risk of consumption increases
680 for species with the highest Hg concentrations, *i.e.* the lesser-spotted dogfish and the European
681 seabass. Surprisingly, the consumption of seabass provides very little benefit, even if consumed
682 as a single portion per week. Fat species, like the Atlantic mackerel or the surmullet seem to
683 provide the best benefit because of their high LC n-3 PUFA concentrations and their low Hg
684 concentrations. However, this index must be considered carefully, as it does not consider the
685 proportion of LC n-3 PUFAs, nor the specific functional role of FAs, making it contradictory
686 to the n-6/n-3 ratio or AI and TI index that could count down the benefit. This is particularly
687 evident concerning the surmullet that we have previously considered not to be beneficial
688 considering only their FA composition and the calculation derived therefrom.

689

690 **5. Conclusion**

691

692 The present study highlights the benefit and risks of consuming different seafood varieties from
693 the “Pertuis charentais” area of France, a well-known region for seafood production in Europe.
694 In terms of the FA profile, all species presented a PUFA/SFA ratio, as well as a n-6/n-3 ratio
695 much lower than the threshold from which the rate of cholesterol and cardiovascular disease
696 increase. Considering all of the FA indicators measured in this study, whiting and pilchard
697 appear as the most beneficial fish species for the human diet, while the surmullet is least
698 advantageous. However, using an index that integrates the relative risk due to Hg content, the
699 surmullet and mackerel appear as the seafood with the best effects on the prevention of
700 cardiovascular disease in adults. Overall, the concentration of trace elements with respect to
701 seafood safety recommendations, are such that significant amounts of seafood can be safely
702 eaten on a daily or weekly basis. Yet, levels of Hg in dogfish and seabass can become a concern

703 for frequent seafood consumers (> three meals a week), confirming that diversity in seafood is
704 key for consumers to optimize the benefit of seafood resources.

705 It is important to note that the risk/benefit assessment for the seafood consumers is strongly
706 believed to vary with season as the lipid content and trace element concentrations may depend
707 on the nutrient availability, the physiological and the reproductive status of marine organisms
708 (Aidos *et al.*, 2002; Lozano-Bilbao *et al.*, 2020). It is also necessary to consider food
709 preparation, as cooking is known to i) damage the LC n-3 PUFAs (Türkkan *et al.*, 2008,
710 Gladyshev *et al.*, 2006, Sardenne *et al.*, 2021), and ii) decrease the bioaccessibility of trace
711 elements (He & Wang, 2011; Houlbrèque *et al.*, 2011), including the levels of other organic
712 contaminants (*i.e.* PCBs, PAHs). In addition, heightened anthropogenic activities may be
713 responsible for increased trace elements concentrations including other contaminants such as
714 persistent organic pollutant (*e.g.* DDT or PCBs) in the environment (Storelli, 2008). Also,
715 global change, through warming, acidification or deoxygenation alter the assemblages and
716 physiology of marine microalgae, leading to an overall reduction in the production of LC n-3
717 PUFAs at the base of the marine food web (Galloway, 2015, Hixson and Arts 2016). This may
718 have consequences on upper trophic organisms, including seafood species and humans, that
719 may be unable to synthesize these molecules (Hixson & Arts, 2016; Pethybridge *et al.*, 2015).

720

721 **Acknowledgments**

722 This work is a contribution to the project ECONAT Axe 1 - Ressources Marines Littorales:
723 qualité et éco-valorisation, funded by the Contrat de Plan Etat-Région and the CNRS and the
724 European Regional Development Fund through the project QUALIDRIS (Ressources Marines
725 et Littorales: Qualité et Eco-valorisation). Authors thank Lucas Weppe for its assistance in
726 sample preparation. Thanks to the CPER (Contrat de Projet Etat-Région) and the FEDER
727 (Fonds Européen de Développement Régional) for funding the AMA and the IRMS of LIENSs
728 laboratory. This work benefitted from the French GDR "Aquatic Ecotoxicology" framework
729 which aims at fostering stimulating scientific discussions and collaborations for more
730 integrative approaches. The IUF (Institut Universitaire de France) is acknowledged for its
731 support to PB as a Senior Member.

732

733

734 **References**

- 735 Abreu, S.N., Pereira, E., Vale, C., Duarte, A.C., 2000. Accumulation of mercury in sea bass
736 from a contaminated lagoon (Ria de Aveiro, Portugal). *Marine Pollution Bulletin*. 40, 293-
737 297.
- 738 Ackman, R.G., 1990. Seafood lipids and fatty acids. *Food Reviews International*. 6, 617-646.
- 739 Abrami, G., Natiello, F., Bronzi, P., McKenzie, D., Bolis, L., Agradi, E., 1992. A comparison
740 of highly unsaturated fatty acid levels in wild and farmed eels (*Anguilla anguilla*).
741 *Comparative Biochemistry and Physiology Part B: Comparative Biochemistry*. 101, 79-81.
- 742 Afonso, C., Cardoso, C., Lourenço, H.M., Anacleto, P., Bandarra, N.M., Carvalho, M.L.,
743 Castro, M., Nunes M.L., 2013. Evaluation of hazards and benefits associated with the
744 consumption of six fish species from the Portuguese coast. *Journal of Food Composition*
745 *and Analysis*. 32, 59-67.
- 746 Aidos, I., Van Der Padt, A., Luten, J.B., Boom R.M., 2002. Seasonal Changes in Crude and
747 Lipid Composition of Herring Fillets, Byproducts, and Respective Produced Oils. *Journal*
748 *of Agricultural and Food Chemistry*. 50, 4589-4599.
- 749 Alheit, J., Beare, D., Bernal, M., Casini, M., Clarke, M., Cotano, U., Dickey-Collas, M.,
750 Dransfeld, L., Harma, C., Heino, M., Massé, J., Möllmann, C., Nogueira, E., Petitgas, P.,
751 Reid, D., Silva, A., Skaret, G., Slotte, A., Stratoudakis, Y., Uriarte, A., Voss, R., 2010. Life-
752 cycle spatial patterns of small pelagic fish in the Northeast Atlantic, in: Petitgas, P. (Ed.),
753 ICES Cooperative Research Report n°396, 98 pp.
- 754 Andersen, J.L., Depledge, M.H., 1997. A survey of total mercury and methylmercury in edible
755 fish and invertebrates from Azorean waters. *Marine Environmental Research*. 44, 331-350.
- 756 Ansari, T.M., Marr, I.L., Tariq, N., 2004. Heavy metals in marine pollution perspective - A
757 mini review. *Journal of Applied Sciences*. 4, 1-20.
- 758 Anual, Z.F., Maher, W., Krikowa, F., Hakim, L., Ahmad, N.I., Foster, S., 2018. Mercury and
759 risk assessment from consumption of crustaceans, cephalopods and fish from West
760 Peninsular Malaysia. *Microchemical Journal*. 140, 214-221.
- 761 Astorg, P., Arnault, N., Czernichow, S., Noisette, N., Galan, P., Hercberg, S., 2004. Dietary
762 intakes and food sources of n-6 and n-3 PUFA in French adult men and women. *Lipids*. 39,
763 527-535.
- 764 Blanchard G.F., Guarini J.M., Orvain F., Sauriau P.G., 2001. Dynamic behaviour of benthic
765 microalgal biomass in intertidal mudflats. *Journal of Experimental Marine Biology and*
766 *Ecology*. 264, 85-100.

- 767 Bloom, N.S., 1992. On the chemical form of mercury in edible fish and marine invertebrate
768 tissue. *Canadian journal of fisheries and aquatic sciences*. 49, 1010-1017.
- 769 Borak, J., Hosgood, H.D., 2007. Seafood arsenic: implications for human risk
770 assessment. *Regulatory Toxicology and Pharmacology*. 47, 204-212.
- 771 Burger, J., Gochfeld, M., 2011. Mercury and selenium levels in 19 species of saltwater fish
772 from New Jersey as a function of species, size, and season. *Science of the Total Environment*.
773 409, 1418-1429.
- 774 Bustamante, P., Miramand, P., 2004. Interspecific and geographical variations of trace element
775 concentrations in Pectinidae from European waters. *Chemosphere*. 57, 1355-1362.
- 776 Caponio, F., Lestingi, A., Summo, C., Bilancia, M.T., Laudadio, V., 2004. Chemical
777 characteristics and lipid fraction quality of sardines (*Sardina pilchardus* W.): influence of
778 sex and length. *Journal of Applied Ichthyology*. 20, 530-535.
- 779 Chouvelon, T., Spitz, J., Caurant, F., Mèndez-Fernandez, P., Autier, J., Lassus-Débat, A.,
780 Chappuis, A., Bustamante, P., 2012. Enhanced bioaccumulation of mercury in deep-sea
781 fauna from the Bay of Biscay (north-east Atlantic) in relation to trophic positions identified
782 by analysis of carbon and nitrogen stable isotopes. *Deep Sea Research Part I:
783 Oceanographic Research Papers*. 65, 113-124.
- 784 Coelho, J.P., Santos, H., Reis, A.T., Falcão, J., Rodrigues, E.T., Pereira, M.E., Duarte, A.C.,
785 Pardal, M.A., 2010. Mercury bioaccumulation in the spotted dogfish (*Scyliorhinus canicula*)
786 from the Atlantic Ocean. *Marine Pollution Bulletin*. 60, 1372-1375.
- 787 Coombs, S.H., Smyth, T.J., Conway, D.V.P., Halliday, N.C., Bernal, M., Stratoudakis, Y.,
788 Alvarez, P., 2006. Spawning season and temperature relationships for sardine (*Sardina
789 pilchardus*) in the eastern North Atlantic. *Journal of the Marine Biological Association of
790 the United Kingdom* 86, 1245-1252.
- 791 EC, 2008. Commission Regulation (EC) No 629/2008 of 2 July 2008 amending Regulation
792 (EC) No 1881/2006 setting maximum levels for certain contaminants in foodstuffs. *Official
793 Journal of the European Union*. 173, 6-9.
- 794 EC, 2011. Commission Regulation (EU) No 420/2011 of 29 April 2011 amending Regulation
795 (EC) No 1881/2006 setting maximum levels for certain contaminants in foodstuffs. *Official
796 Journal of the European Union*. 111, 3-6.
- 797 EC, 2014. Commission Regulation (EU) No 488/2014 of 12 May 2014 amending Regulation
798 (EC) No 1881/2006 as regards maximum levels of cadmium in foodstuffs. *Official Journal
799 of the European Union*. 138, 75-79.

- 800 EC, 2015. Commission Regulation (EU) No 2015/1005 of 25 June 2015 amending Regulation
801 (EC) No 1881/2006 as regards maximum levels of lead in certain foodstuffs. *Office*
802 *Journal of the European Union*. 161, 9-13.
- 803 EFSA, 2009. Scientific opinion of the panel on dietetic products, nutrition and allergies on a
804 request from the European Commission related to labelling reference intake values for n-3
805 and n-6 polyunsaturated fatty acids. *The EFSA Journal*. 1176, 1-11.
- 806 EFSA, 2014. Annual Report of the EFSA Journal 2013. *EFSA Supporting publication*. 11,
807 721E.
- 808 Eagles-Smith, C.A., Silbergeld, E.K., Basu, N., Bustamante, P., Diaz-Barriga, F., Hopkins,
809 W.A., Kidd, K.A., Nyland, J.F., 2018. Modulators to mercury risk to wildlife and humans
810 in the context of rapid global change. *Ambio*, 47: 170–197.
- 811 El-Moselhy, K.M., Othman, A.I., Abd El-Azem, H., El-Metwally, M.E.A., 2014.
812 Bioaccumulation of heavy metals in some tissues of fish in the Red Sea, Egypt. *Egyptian*
813 *Journal of Basic and Applied Sciences*. 1, 97-105.
- 814 Ersoy, B., Çelik, M., 2009. Essential elements and contaminants in tissues of commercial
815 pelagic fish from the Eastern Mediterranean Sea. *Journal of the Science of Food and*
816 *Agriculture*. 89, 1615–1621.
- 817 FAO, 2018. The State of World Fisheries and Aquaculture 2018 – Meeting the sustainable
818 development goals, Rome.
- 819 FAO/WHO, 2010. Fats and fatty acids in human nutrition. Report of an expert consultation. 1-
820 166.
- 821 Fleith, M., Clandinin, M.T., 2005. Dietary PUFA for preterm and term infants: Review of
822 clinical studies. *Critical reviews in food science and nutrition*. 45, 205-229.
- 823 FranceAgriMer, 2017. Données de ventes déclarées en halles à marée en 2016. *Pêche et*
824 *aquaculture, données et bilans*. 1-94.
- 825 Frieden, E., 1985. New perspectives of the Essential trace elements. *Journal of Chemical*
826 *Education*. 62, 917-923.
- 827 Gagnaire, B., Renault, T., Thomas-Guyon, H., 2003. In vitro and in vivo effects of mercury on
828 haemocytes of pacific oyster, *Crassostrea gigas* (Thunberg): Development of techniques
829 evaluating estuarine pollution. *Journal de recherche océanographique*. 28, 34-38.
- 830 Gao, S., Wang, W.X., 2014. Oral bioaccessibility of toxic metals in contaminated oysters and
831 relationships with metal internal sequestration. *Ecotoxicology and Environmental Safety*
832 110, 261-268.

- 833 Garrido, S., Rosa, R., Ben-Hamadou, R., Cunha, M.E., Chicharo, M.A., van der Lingen, C.D.,
834 2007. Effect of maternal fat reserves on the fatty acid composition of sardine (*Sardina*
835 *pilchardus*) oocytes. *Comparative Biochemistry and Physiology*. 148 B, 398-409.
- 836 Gatti, P., Cominassi, L., Duhamel, E., Grellier, P., Le Delliou, H., Le Mestre, S., Petitgas, P.,
837 Rabiller, M., Spitz, J., Huret, M., 2018. Bioenergetic condition of anchovy and sardine in
838 the Bay of Biscay and English Channel. *Progress in Oceanography* 166, 129-138.
- 839 Geffard, A., Amiard-Triquet, C., Amiard, J.C., Mouneyrac, C., 2001. Temporal variations of
840 metallothionein and metal concentrations in the digestive gland of oysters (*Crassostrea*
841 *gigas*) from a clean and a metal-rich site. *Biomarkers*. 6, 91-107.
- 842 Gil, A., Gil, F., 2015. Fish, a Mediterranean source of n-3 PUFA: benefits do not justify limiting
843 consumption. *British Journal of Nutrition*. 113, 58-67.
- 844 Ginsberg, G.L., Toal, B.F., 2009. Quantitative approach for incorporating methylmercury risks
845 and omega-3 fatty acid benefits in developing species-specific fish consumption advice.
846 *Environmental Health Perspectives*. 117, 267-275.
- 847 Gladyshev, M., Sushchik, N., Gubanenko, G., Demirchieva, S., and Kalachova, G., 2006. Effect
848 of way of cooking on content of essential polyunsaturated fatty acids in muscle tissue of
849 humpback salmon (*Oncorhynchus gorbusha*). *Food Chemistry*. 96, 446-451.
- 850 Golden C., 2016. Fall in fish catch threatens human health. *Nature*. 538, 171-171.
- 851 Goldhaber, S.B., 2003. Trace element risk assessment: essentiality vs. toxicity. *Regulatory*
852 *Toxicology and Pharmacology*. 38, 232-242.
- 853 Grousset, F.E., Jouanneau, J.M., Castaing, P., Lavaux, G., Latouche, C., 1999. A 70 year record
854 of contamination from industrial activity along the Garonne River and its tributaries (SW
855 France). *Estuarine, Coastal and Shelf Science*. 48, 401-414.
- 856 Guallar E., Sanz-Gallardo M.I., van't Veer P., Bode P., Aro A., Gomez-Aracena J., Kark, J.D.,
857 Rimersma, R.A., Martin-Moreno, J.M., Kok, F.J., 2002. Mercury, fish oils, and the risk of
858 myocardial infarction. *The New England Journal of Medicine*. 347, 1747-54.
- 859 Guérin, T., Chekri, R., Vastel, C., Siro, V., Volatier, J.L., Leblanc, J.C., Noël, L., 2011.
860 Determination of 20 trace elements in fish and other seafood from the French market. *Food*
861 *Chemistry*. 127, 934-942.
- 862 He, M., Wang, W.X., 2011. Factors affecting the bioaccessibility of methylmercury in several
863 marine fish species. *Journal of Agricultural and Food Chemistry*. 59, 7155-7162.

- 864 Hellberg, R.S., Mireles Dewitt, C.A., Morrissey, M.T., 2012. Risk-benefit analysis of seafood
865 consumption: A review. *Comprehensive Reviews in Food Science and Food Safety*. 11, 490-
866 517.
- 867 Hixson S.M, Arts M.T, 2016. Climate warming is predicted to reduce omega-3, long-chain,
868 polyunsaturated fatty acid production in phytoplankton. *Global Change Biology*. 22, 2744-
869 2755.
- 870 Houlbrèque, F., Hervé-Fernandez, P., Teyssié, J.L., Oberhänsli, F., Boisson, F., Jeffree, R.,
871 2011. Cooking makes cadmium contained in Chilean mussels less bioaccessible to humans.
872 *Food Chemistry*. 126, 917-921.
- 873 Hu, F.B., Bronner, L., Willett, W.C., Stampfer, M.J., Rexrode, K.M., Albert, C.M., Hunter, D.,
874 Manson, J. E., 2002. Fish and omega-3 fatty acid intake and risk of coronary heart disease
875 in women. *Jama*. 287, 1815-1821.
- 876 Hubans, B., Chouvelon, T., Begout, M.-L., Biaï, G., Bustamante, P., Ducci, L., Mornet, F.o.,
877 Boiron, A., Coupeau, Y., Spitz, J., 2017. Trophic ecology of commercial-size meagre,
878 *Argyrosomus regius*, in the Bay of Biscay (NE Atlantic). *Aquatic Living Resources*. 30, 9.
- 879 Hughes, M.F., 2002. Arsenic toxicity and mechanisms of action. *Toxicology Letters*. 133, 1-16.
- 880 Koletzko, B., Lien, E., Agostoni, C., Böhles, H., Campoy, C., Cetin, I., Decsi, T., Dudenhausen,
881 J.W., Dupont, C., Forsyth, S., Hoesli, I., Holzgreve, W., Lapillonne, A., Putet, G., Secher,
882 N.J., Symonds, M., Szajewska, H., Willatts, P., UauyR., 2008. The roles of long-chain
883 polyunsaturated fatty acids in pregnancy, lactation and infancy: Review of current
884 knowledge and consensus recommendations. *Journal of perinatal medicine*. 36, 5-14.
- 885 Langston, W.J., Bebianno, M.J., 1998. *Metal metabolism in aquatic environments*. Springer
886 Science & Business Media.
- 887 Leblanc, J.C., Sirot, V., Volatier, J.L., Bemrah-Aouachria, N., 2006. Fish and seafood
888 consumption study and biomarker of exposure to trace elements, pollutants and omega 3.
889 *CALIPSO report. French Agency for Food, Environmental and Occupation Health & Safety,*
890 *AFSSA*. 1-162.
- 891 Lozano-Bilbao, E., Jurado-Ruzafa, A., Lozano, G., Jimenez, S., Hardisson, A., Rubio, C.,
892 Weller, D.G., Paz, S., Gutierrez, A.J., 2020. Development stage and season influence in the
893 metal content of small pelagic fish in the North-West Africa. *Chemosphere*. 261, 127692.
- 894 Maher, W., Butler, E., 1988. Arsenic in the marine environment. *Applied Organometallic*
895 *Chemistry*. 2, 191-214.

- 896 Mathieu-Resuge, M., Le Grand, F., Schaal, G., Lluch-Cota, S.R., Racotta, I.S., Kraffe, E.,
897 2020. Specific regulations of gill membrane fatty acids in response to environmental
898 variability reveal fitness differences between two suspension-feeding bivalves (*Nodipecten*
899 *subnodosus* and *Spondylus crassisquama*). *Conservation Physiology*, 8, coaa079
- 900 Mayer, D.R., Kosmus, W., Pgglichtsch, H., Mayer, D., Beyer, W., 1993. Essential trace elements
901 in humans : Serum arsenic concentrations in hemodialysis patients in comparison to healthy
902 controls. *Biological Trace Elements Research*. 37, 27-38.
- 903 Médale, F., 2009. Teneur en lipides et composition en acides gras de la chair de poissons issus
904 de la pêche et de l'élevage. *Cahiers de Nutrition et de Diététique*. 44, 173-181.
- 905 Metian, M., Bustamante, P., Cosson, R.P., Hédouin, L., Warnau, M., 2008. Investigation of Ag
906 in the king scallop *Pecten maximus* using field and laboratory approaches. *Journal of*
907 *Experimental Marine Biology and Ecology*. 367, 53-60.
- 908 Metian, M., Charbonnier, L., Oberhänsli, F., Bustamante, P., Jeffree, R., Amiard, J.C., Warnau,
909 M., 2009. Assessment of metal, metalloid, and radionuclide bioaccessibility from mussels
910 to human consumers, using centrifugation and simulated digestion methods coupled with
911 radiotracer techniques. *Ecotoxicology and Environmental Safety* 72, 1499-1502.
- 912 Metian, M., Warnau, M., Chouvelon, T., Pedraza, F., Rodriguezy y Baena, A.M., Bustamante,
913 P., 2013. Trace element bioaccumulation in reef fish from New Caledonia: Influence of
914 trophic groups and risk assessment for consumers. *Marine Environmental Research*. 87-88,
915 26-36.
- 916 Miossec, L., Le Deuff, R.M., Gouletquer, P., 2009. Alien species alert: *Crassostrea gigas*
917 (Pacific oyster). *ICES Cooperative Research Report*. 299, 1-42.
- 918 Miquel, G., 2001. Rapport sur les effets des métaux lourds sur l'environnement et la santé.
919 <https://www.senat.fr/rap/100-261/100-2611.pdf>.
- 920 Miramand P, Ferchaud R, Pigeot J, Caurant F, Bustamante P, Guyot T, 2002. Estimation of the
921 Cd intake in the human dietary from the shellfish caught in the seashore of Charente-
922 Maritime (France). *Revue de Médecine Vétérinaire*. 153, 741-746.
- 923 Morris, M.C., Evans, D.A., Bienias, J.L., Tangney, C.C., Bennett, D.A., Wilson, R.S.,
924 Aggarwal, Neelum., Schneider, J., 2003. Consumption of fish and n-3 fatty acids and risk of
925 incident Alzheimer disease. *Archives of neurology*. 60, 940-946.
- 926 Mozaffarian, D., Rimm, E.B., 2006. Fish intake, contaminants, and human health: Evaluating
927 the risks and the benefits. *Journal of the American Medical Association*. 296, 1885-1900.

- 928 Narayan, B., Miyashita, K., Hosakawa, M., 2006. Physiological effects of eicosapentaenoic
929 acid (EPA) and docosahexaenoic acid (DHA) - A review. *Food Reviews International*. 22,
930 291-307.
- 931 Naser, H.A., 2013. Assessment and management of heavy metal pollution in the marine
932 environment of the Arabian Gulf: A review. *Marine Pollution Bulletin*. 72, 6–13.
- 933 Neff, J.M., 1997. Ecotoxicology of arsenic in marine environment. *Environmental Toxicology*
934 *and Chemistry*. 16, 917-927.
- 935 Olmedo, P., Hernández, A.F., Pla, A., Femia, P., Navas-Acien, A., Gil, F., 2013a.
936 Determination of essential elements (copper, manganese, selenium and zinc) in fish and
937 shellfish samples. Risk and nutritional assessment and mercury – selenium balance. *Food*
938 *and Chemical Toxicology*. 62, 299-307.
- 939 Olmedo, P., Pla, A., Hernández, A.F., Barbier, F., Ayouni, L., Gil, F., 2013b. Determination of
940 toxic elements (mercury, cadmium, lead, tin and arsenic) in fish and shellfish samples. Risk
941 assessment for the consumers. *Environment International*. 59, 63-72.
- 942 Oomen, C.M., Feskens, E.J.M., Räsänen, L., Fidanza, F., Nissinen, A.M., Menotti, A., Kok,
943 F.J., Kromhout, D., 2000. Fish consumption and coronary heart disease mortality in Finland,
944 Italy, and the Netherlands. *American Journal of Epidemiology*. 151, 999-1006.
- 945 Ospina-E, J. C., Sierra-C, A., Ochoa, O., Pérez-Álvarez, J. A., Fernández-López, J., 2012.
946 Substitution of saturated fat in processed meat products: A review. *Critical Reviews in Food*
947 *Science and Nutrition*. 52, 113-122.
- 948 Pethybridge, H.R., Parrish, C.C., Morrongiello, J., Young, J.W., Farley, J.H., Gunasekera,
949 R.M., Nichols, P.D., 2015. Spatial patterns and temperature predictions of tuna fatty acids:
950 Tracing essential nutrients and changes in primary producers. *PLoS ONE*. 10, 1-17.
- 951 Pike, I.H., 1999. Health benefits from feeding fish oil and fish meal. *IFOMA, Herts, UK*.
- 952 Rainbow, P.S., 1997. Trace metal accumulation in marine invertebrates: Marine biology or
953 marine chemistry? *Journal of the Marine Biology Association of the United Kingdom*. 77,
954 195-210.
- 955 Ralston, N.V.C., 2008. Selenium health benefit values as seafood safety criteria. *EcoHealth*. 5,
956 442-455.
- 957 Roth, N., Knai, C., 2003. Food based dietary guidelines in the WHO European Region. *WHO:*
958 *Geneva, Switzerland*.

- 959 Ruxton, C.H.S., Reed, S.C., Simpson, M.J.A., Millington, K.J., 2004. The health benefits of
960 omega-3 polyunsaturated fatty acids: A review of the evidence. *Journal of Human Nutrition*
961 *and Dietetics*. 17, 449-459.
- 962 Salles, D., Roumezi, A., Lancelleur, L., Schäfer, J., Chiffolleau, J.F., Auger, D., Blanc, G., Petit,
963 J., Coynel, A., 2013. L'argent (Ag, Nanoag) comme contaminant émergent dans l'estuaire
964 de la Gironde : Évaluations scientifiques et gouvernance des risques. *Environnement,*
965 *Risques et Sante*. 12, 317-323.
- 966 Salonen J.T., Seppanen K., Nyyssonen K., Korpela H., Kauhanen J., Kantola M., Tuomilehto
967 J., Esterbauer H., Tatzber F., Salonen R., 1995. Intake of mercury from fish, lipid
968 peroxidation, and the risk of myocardial infarction and coronary, cardiovascular, and any
969 death in eastern Finnish men. *Circulation*. 91, 645-655.
- 970 Sardenne, F., Puccinelli, E., Vagner, M., Pecquerie, L., Bideau, A., Le Grand, F., Soudant, P.,
971 2021. Post-mortem storage conditions and cooking methods affect long-chain omega-3 fatty
972 acid content in Atlantic mackerel (*Scomber scombrus*). *Food Chemistry*. 359, 129828.
- 973 Simopoulos, A.P., 1991. Omega-3 fatty acids in health and disease and in growth and
974 development. *The American Journal of Clinical Nutrition*. 54, 438-463.
- 975 Simopoulos, A.P., Leaf, A., Salem, N.J., 2000. Workshop statement on the essentiality of and
976 recommended dietary intakes for omega-6 and omega-3 fatty acids. *Prostaglandins,*
977 *Leukotriens, and Essential Fatty Acids*. 63, 119-121.
- 978 Simopoulos, A.P., 2002. The importance of the ratio of omega-6/omega-3 essential fatty acids.
979 *Biomedicine & Pharmacotherapy*. 56, 365-379.
- 980 Simopoulos, A.P., 2003. Importance of the ratio of omega 6/omega 3 essential fatty acids:
981 Evolutionary aspects. *World Review of Nutrition and Dietetics*. 92, 1-174.
- 982 Simopoulos, A.P., 2006. Evolutionary aspects of diet, the omega-6/omega-3 ratio and genetic
983 variation: nutritional implications for chronic diseases. *Biomedicine & Pharmacotherapy*.
984 60, 502-507.
- 985 Sirot, V., Oseredczuk, M., Bemrah-Aouachria, N., Volatier, J.L., Leblanc, J.C., 2008. Lipid and
986 fatty acid composition of fish and seafood consumed in France: CALIPSO study. *Journal of*
987 *Food Composition and Analysis*. 21, 8-16.
- 988 Sirot, V., Guérin, T., Volatier, J. L., & Leblanc, J. C., 2009. Dietary exposure and biomarkers
989 of arsenic in consumers of fish and shellfish from France. *Science of the Total*
990 *Environment*. 407, 1875-1885.

- 991 Sokolova, I.M., Lannig, G., 2008. Interactive effects of metal pollution and temperature on
992 metabolism in aquatic ectotherms: implications of global climate change. *Climate Research*.
993 37, 181-201.
- 994 Storelli, M.M., Giacomini-Stuffler, R., Storelli, A., Marcotrigiano, G.O., 2003. Total
995 mercury and methylmercury content in edible fish from the Mediterranean Sea. *Journal of*
996 *Food Protection*. 66, 300-303.
- 997 Storelli, M.M., Giacomini-Stuffler, R., Storelli, A., Marcotrigiano, G.O., 2005a.
998 Accumulation of mercury, cadmium, lead and arsenic in swordfish and bluefin tuna from the
999 Mediterranean Sea: A comparative study. *Marine Pollution Bulletin*. 50, 1004-1007.
- 1000 Storelli, M.M., Storelli, A., Giacomini-Stuffler, R., Marcotrigiano, G.O., 2005b. Mercury
1001 speciation in the muscle of two commercially important fish, hake (*Merluccius merluccius*)
1002 and striped mullet (*Mullus barbatus*) from the Mediterranean Sea: Estimated weekly intake.
1003 *Food Chemistry*. 89, 295-300.
- 1004 Storelli, M.M., Giacomini-Stuffler, R., Storelli, A., Marcotrigiano, G.O., 2006. Cadmium
1005 and mercury in cephalopod molluscs: Estimated weekly intake. *Food Additives and*
1006 *Contaminants*. 23, 25-30.
- 1007 Storelli, M.M., Barone, G., Piscitelli, G., Marcotrigiano, G.O., 2007. Mercury in fish:
1008 Concentration vs. fish size and estimates of mercury intake. *Food Additives and*
1009 *Contaminant*. 24, 1353-1357.
- 1010 Storelli, M.M., 2008. Potential human health risks from metals (Hg, Cd, and Pb) and
1011 polychlorinated biphenyls (PCBs) via seafood consumption: Estimation of target hazard
1012 quotients (THQs) and toxic equivalents (TEQs). *Food and Chemical Toxicology*. 46, 2782-
1013 2788.
- 1014 Storelli, M.M., 2009. Intake of Essential Minerals and Metals via Consumption of Seafood from
1015 the Mediterranean Sea. *Journal of Food Protection*. 72, 1116-1120.
- 1016 Stupin, M., Kibel, A., Stupin, A., Selthofer-Relatić, K., Matic, A., Mihalj, M., Mihaljević, Z.,
1017 Jukić, I., Drenjančević, I., 2019. The Physiological Effect of n-3 Polyunsaturated Fatty Acids
1018 (n-3 PUFAs) Intake and Exercise on Hemorheology, Microvascular Function, and Physical
1019 Performance in Health and Cardiovascular Diseases; Is There an Interaction of Exercise and
1020 Dietary n-3 PUFA Intake? *Frontiers in Physiology*. 10, 1129.
- 1021 Taylor, V., Goodale, B., Raab, A., Schwerdtle, T., Reimer, K., Conklin, S., Karagas, M.R.,
1022 Francesconi, K.A., 2017. Human exposure to organic arsenic species from seafood. *Science*
1023 *of the Total Environment*. 580, 266-282.

- 1024 Türkkan, A.U., Cakli, S., Kilinc, B., 2008. Effects of cooking methods on the proximate
1025 composition and fatty acid composition of seabass (*Dicentrarchus labrax*, Linnaeus, 1758).
1026 *Food and Bioproducts Processing*. 86, 163-166.
- 1027 Ulbricht, T.L.V., Southgate, D.A.T., 1991. Coronary heart disease: Seven dietary factors. *The*
1028 *Lancet*. 338, 985-992.
- 1029 Uthus, E.O., 1992. Evidence for arsenic essentiality. *Environmental Geochemistry and Health*.
1030 14, 55-58.
- 1031 Valfré F., Caprino F., Turchini G.M., 2003. The Health Benefit of Seafood. *Veterinary*
1032 *Research Communications*. 27, 507-512.
- 1033 Wu, X., Gao, M., Wang, L., Luo, Y., Bi, R., Li, L., Xie, L., 2014. The arsenic content in
1034 marketed seafood and associated health risks for the residents of Shandong, China.
1035 *Ecotoxicology and Environmental Safety*. 102, 168-173.
- 1036 Zhang, X., Ning, X., He, X., Sun, X., Yu, X., Cheng, Y., Yu, R.Q., Wu, Y., 2020. Fatty acid
1037 composition analyses of commercially important fish species from the Pearl River Estuary,
1038 China. *PLoS ONE*. 15, 1-15.
- 1039
- 1040 **R Packages:**
- 1041 Graves, S., Piepho H.P., Selzer, L., with help from Dorai-Raj, S., 2015. multcompView:
1042 Visualizations of Paired Comparisons. R package version 0.1-7. [https://CRAN.R-](https://CRAN.R-project.org/package=multcompView)
1043 [project.org/package=multcompView](https://CRAN.R-project.org/package=multcompView).
- 1044 Kassambara, A., Mundt, F., 2017. factoextra: Extract and Visualize the Results of Multivariate
1045 Data Analyses. R package version 1.0.5. <https://CRAN.R-project.org/package=factoextra>.
- 1046 Kolde, R., 2019. pheatmap: Pretty Heatmaps. R package version 1.0.12. [https://CRAN.R-](https://CRAN.R-project.org/package=pheatmap)
1047 [project.org/package=pheatmap](https://CRAN.R-project.org/package=pheatmap).
- 1048 Le, S., Josse, J., Husson, F., 2008. FactoMineR: An R Package for Multivariate Analysis.
1049 *Journal of Statistical Software*. 25, 1-18. 10.18637/jss.v025.i01.
- 1050 Neuwirth, E., 2014. RColorBrewer: ColorBrewer Palettes. R package version 1.1.
1051 <https://CRAN.R-project.org/package=RColorBrewer>.
- 1052 Slowikowski, K., 2019. ggrepel: Automatically Position Non-Overlapping Text Labels with
1053 'ggplot2'. R package version 0.8.1. <https://CRAN.R-project.org/package=ggrepel>.
- 1054 Wickham, H., 2016. ggplot2: Elegant Graphics for Data Analysis. *Springer-Verlag New York*.
- 1055 Wickham, H., 2017. tidyverse: Easily Install and Load the 'Tidyverse'. R package version 1.2.1.
1056 <https://CRAN.R-project.org/package=tidyverse>.

1057 Wilke, C.O., 2019. cowplot: Streamlined Plot Theme and Plot Annotations for 'ggplot2'. R
1058 package version 0.9.4. <https://CRAN.R-project.org/package=cowplot>.

Journal Pre-proof

Table 1. Scientific and common names, acronym, number (n), length (mm, mean \pm SD) and weight (g, mean \pm SD), date of procurement of the studied organisms: fish, crustaceans, cephalopods, bivalves from fisheries (white lines) and bivalves from aquaculture (grey lines).

Taxa	Order	Scientific name	Common name	Acronym	n	Length (mm)	Weight (g)	Date of procurement	Comment
Fish	<i>Clupeiformes</i>	<i>Sardina pilchardus</i>	European pilchard	PIL	10	170 \pm 9 ^a	53 \pm 10	11/2018	not gutted ^g
	<i>Gadiformes</i>	<i>Merluccius merluccius</i>	European hake	HAK	9	434 \pm 29 ^a	538 \pm 116	11/2018	gutted ^f
	<i>Gadiformes</i>	<i>Merlangius merlangus</i>	Whiting	WTG	10	316 \pm 16 ^a	240 \pm 32	11/2018	gutted ^f
	<i>Perciformes</i>	<i>Scomber scombrus</i>	Atlantic mackerel	MKR	10	245 \pm 20 ^a	142 \pm 33	11/2018	not gutted ^g
	<i>Perciformes</i>	<i>Spondyliosoma cantharus</i>	Black seabream	SBR	10	220 \pm 15 ^a	280 \pm 24	02/2019	not gutted ^g
	<i>Perciformes</i>	<i>Dicentrarchus labrax</i>	European seabass	SBS	10	475 \pm 12 ^a	1223 \pm 128	11/2018	not gutted ^g
	<i>Perciformes</i>	<i>Argyrosomus regius</i>	Meagre	MGR	10	431 \pm 11 ^a	820 \pm 55	11/2018	not gutted ^g
	<i>Perciformes</i>	<i>Mullus surmuletus</i>	Surmullet	SRM	10	227 \pm 10 ^a	206 \pm 37	11/2018	not gutted ^g
	<i>Pleuronectiformes</i>	<i>Solea solea</i>	Common sole	SOL	10	302 \pm 8 ^a	223 \pm 39	11/2018	not gutted ^g
	<i>Zeiformes</i>	<i>Zeus faber</i>	John Dory	JDO	10	362 \pm 21 ^a	620 \pm 159	11/2018	gutted ^f
	<i>Carcharhiniformes</i>	<i>Scyliorhinus canicula</i>	Lesser-spotted dogfish	DOG	5	641 \pm 199 ^a	1149 \pm 117	02/2019	not gutted ^g
Crustacean	<i>Decapoda</i>	<i>Maja brachydactyla</i>	Atlantic spinous spider crab	SPI	5	139 \pm 4 ^c	693 \pm 66	11/2018	not gutted ^g
Cephalopod	<i>Myopsida</i>	<i>Loligo vulgaris</i>	European squid	SQD	10	214 \pm 17 ^b	241 \pm 34	11/2018	not gutted ^g
	<i>Sepiida</i>	<i>Sepia officinalis</i>	Common cuttlefish	CTF	10	106 \pm 7 ^b	159 \pm 38	11/2018	not gutted ^g
Bivalve	<i>Ostreida</i>	<i>Crassostrea gigas</i> * "Fine de Claire" (green)	Cupped oyster	OFC	20	/	8 \pm 2 ^e	02/2019	full ^h
	<i>Ostreida</i>	<i>Crassostrea gigas</i> * "Spéciale"	Cupped oyster	OSP	20	/	8 \pm 1 ^e	02/2019	full ^h
	<i>Pectinida</i>	<i>Pecten maximus</i> (gonad)	Great Atlantic scallop	GSG	10	111 \pm 4 ^d	3 \pm 1	02/2019	full ^h
	<i>Pectinida</i>	<i>Pecten maximus</i> (muscle)	Great Atlantic scallop	GSM	7	111 \pm 5 ^d	9 \pm 2	02/2019	full ^h

^afork length, ^bmantle length, ^ccarapace length, ^dshell length, ^esoft tissue weight (without the shell), ^f individuals were bought gutted, ^g individuals were bought not gutted, ^hentire individuals was bought. * The Genus *Crassostrea* was preferred to *Magallana* in these study because of the current controversy about the designation of the species (Bayne *et al.*, 2017).

Table 2. Daily recommended species consumption (in g wet weight per day, mean \pm SD) based on the eicosapentaenoic acid (EPA) + the docosahexanoic acid (DHA) requirement (250 mg per day) by the Joint FAO/WHO Expert Committee on Food Additives (JECFA) (FAO/WHO, 2008), considering EPA and DHA concentrations of each studied species (see Table S1 for the EPA and DHA content of each species) and Maximum Safe Consumption (MSC) of trace elements concentrations (expressed in kg per month for Cd; in kg per week for As, Hg, and Pb; in kg per day for Cu, Fe, and Zn), based on provisional tolerable daily, weekly or monthly intake of studied species recommended for an adult of 70 kg body weight.

Taxa	Species	EPA + DHA	Non-essential trace elements			Essential trace elements			
		Daily recommended consumption	MSC _{Cd}	MSC _{Hg}	MSC _{Pb}	MSC _{As}	MSC _{Cu}	MSC _{Fe}	MSC _{Zn}
Fish	Atlantic mackerel	15.8 \pm 19.6	1224	1.6	1152	1.150	46	12	10
	Black seabream	156.7 \pm 29.0	1576	1.3	1172	0.141	81	23	9.1
	Common sole	354.5 \pm 204.5	1540	1.6	1540	0.106	68	22	11
	European hake	68.7 \pm 21.8	1758	3.4	1758	0.548	105	64	12
	European pilchard	28.1 \pm 10.0	1235	4.8	1096	0.416	33	5.3	8.7
	European seabass	76.1 \pm 45.4	1594	0.402	1594	1.340	77	26	11
	John Dory	269.5 \pm 119.7	1571	0.874	1571	1.240	72	24	11
	Lesser-spotted dogfish	390.8 \pm 62.2	1137	0.294	1351	0.030	114	24	4.4
	Meagre	61.4 \pm 8.8	1601	1.8	1601	0.838	81	29	11
	Surmullet	19.1 \pm 16.7	1310	1.6	1168	0.163	65	21	11
	Whiting	152.0 \pm 20.3	1336	39	1197	0.331	42	9.7	9.4
Crustaceans	Atlantic spinous spider crab	274.8 \pm 39.1	142	2.4	98	0.051	6.8	21	0.673
Cephalopods	Common cuttlefish	145.8 \pm 10.9	1574	4.2	1574	0.116	25	33	3.7
	European squid	81.0 \pm 21.5	1450	2.6	458	0.245	27	21	3.6
Bivalves	Cupped oyster « Fine de Claire »	131.9 \pm 92.8	11	8.0	15	0.310	3.9	2.0	0.246
	Cupped oyster « Spéciale »	106.8 \pm 68.8	6.7	8.4	28	0.353	2.5	2.3	0.239
	Great Atlantic scallop (gonad)	62.0 \pm 19.0	17.8	9.8	26	0.408	14	2.4	0.799
	Great Atlantic scallop (muscle)	242.8 \pm 22.3	11.2	11	394	0.653	78	14	2.75

Figure Captions

Figure 1. PCA-derived projection of variables and individuals. Variables were defined as the concentrations of Ag, As, Cd, Cu, Fe, Hg, Mn, Pb, Se, and Zn (all expressed of $\mu\text{g g}^{-1}$ wet weight) and the concentrations of total fatty acids (TFA), long chain n-3 polyunsaturated fatty acids (LC n-3 PUFA), EPA, and DHA (all expressed in mg g^{-1} wet weight) in edible parts of fishes, crustacean, cephalopods and bivalves. A) Correlation circle showing the distribution of each variables on the first two components and B) Grouping of all individuals by species (acronym indicated in the white rectangle) on the first two components. Studied species were: for fishes: DOG (lesser-spotted dogfish, $n = 5$), HAK (European hake, $n = 9$), JDO (John Dory, $n = 10$), MGR (meagre, $n = 10$), MKR (Atlantic mackerel, $n = 10$), PIL (European pilchard, $n = 10$), SBR (black seabream, $n = 10$), SBS (European seabass, $n = 10$), SOL (common sole, $n = 10$), SRM (surmullet, $n = 10$), WTG (whiting, $n = 10$); for crustacean: SPI (Atlantic spinous spider crab, $n = 5$); for cephalopods: CTF (common cuttlefish, $n = 10$), SQD (European squid, $n = 10$); for bivalves: GSG (great Atlantic scallop gonad, $n = 7$), GSM (great Atlantic scallop muscle, $n = 6$), OFC (cupped oyster « Fine de Claire », $n = 10$), OSP (cupped oyster « Spéciale », $n = 10$).

Figure 2. Total fatty acid (TFA, total histogram bars) concentrations (mg g^{-1} wet weight) composed of LC n-3 PUFA (in light grey), others PUFAs (in grey) and others fatty acids (in black) comprising saturated fatty acids, monounsaturated fatty acids, branched fatty acids, and dimethyl acetal fatty acids in edible parts of fishes (blue), crustacean (purple), cephalopods (yellow) and bivalves (red). Studied species were: for fishes: DOG (lesser-spotted dogfish, $n = 5$), HAK (European hake, $n = 9$), JDO (John Dory, $n = 10$), MGR (meagre, $n = 10$), MKR (Atlantic mackerel, $n = 10$), PIL (European pilchard, $n = 10$), SBR (black seabream, $n = 10$), SBS (European seabass, $n = 10$), SOL (common sole, $n = 10$), SRM (surmullet, $n = 10$), WTG (whiting, $n = 10$); for crustacean: SPI (Atlantic spinous spider crab, $n = 5$); for cephalopods: CTF (common cuttlefish, $n = 10$), SQD (European squid, $n = 10$); for bivalves: GSG (great Atlantic scallop gonad, $n = 7$), GSM (great Atlantic scallop muscle, $n = 6$), OFC (cupped oyster « Fine de Claire », $n = 10$), OSP (cupped oyster « Spéciale », $n = 10$).

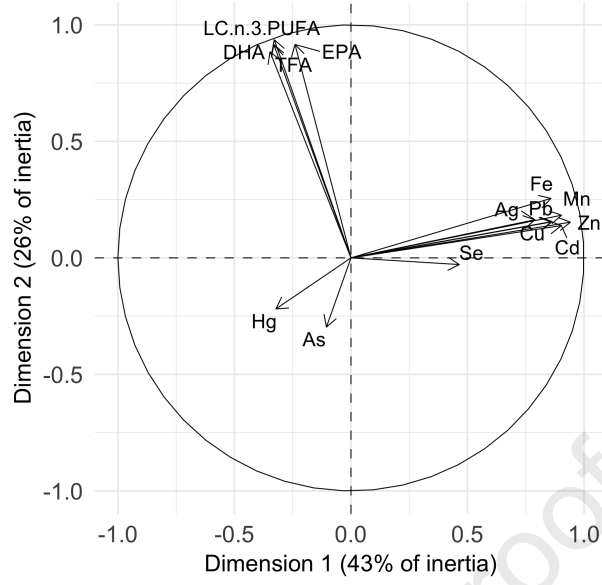
Figure 3. Clustered heatmap showing LC n-3 PUFA/TFA, EPA/TFA, DHA/TFA, n-6/n-3 ratio, atherogenic (AI) and thrombogenic (TI) indices in edible parts of fishes, cephalopods, crustacean and bivalves. Studied species were: for fishes: DOG (lesser-spotted dogfish, $n = 5$), HAK (European hake, $n = 9$), JDO (John Dory, $n = 10$), MGR (meagre, $n = 10$), MKR (Atlantic mackerel, $n = 10$), PIL (European pilchard, $n = 10$), SBR (black seabream, $n = 10$), SBS (European seabass, $n = 10$), SOL (common sole, $n = 10$), SRM (surmullet, $n = 10$), WTG (whiting, $n = 10$); for cephalopods: CTF (common cuttlefish, $n = 10$), SQD (European squid, $n = 10$); for crustacean: SPI (Atlantic spinous spider crab, $n = 5$); for bivalves: GSG (great Atlantic scallop gonad, $n = 7$), GSM (great Atlantic scallop muscle, $n = 6$), OFC (cupped oyster « Fine de Claire », $n = 10$), OSP (cupped oyster « Spéciale », $n = 10$). For all indices, the higher the value, the darker the green.

Figure 4. Boxplots showing A) non-essential (Ag, Cd, Hg, and Pb) and B) essential (As, Cu, Fe, Mn, Se, and Zn) trace element concentrations (expressed in $\mu\text{g g}^{-1}$ wet weight) in edible parts of fishes (blue), crustaceans (purple), cephalopods (yellow) and bivalves (red). Studied species were: for fishes: DOG (lesser-spotted dogfish, $n = 5$), HAK (European hake, $n = 9$), JDO (John Dory, $n = 10$), MGR (meagre, $n = 10$), MKR (Atlantic mackerel, $n = 10$), PIL (European pilchard, $n = 10$), SBR (black seabream, $n = 10$), SBS (European seabass, $n = 10$), SOL (common sole, $n = 10$), SRM (surmullet, $n = 10$), WTG (whiting, $n = 10$); for crustaceans: SPI (Atlantic spinous spider crab, $n = 5$); for cephalopods: CTF (common cuttlefish, $n = 10$), SQD (European squid, $n = 10$); for bivalves: GSG (great Atlantic scallop gonad, $n = 10$ except for Hg $n = 9$ and for Mn, and Zn $n = 8$), GSM (great Atlantic scallop muscle, $n = 7$ except for Zn $n = 6$), OFC (cupped oyster « Fine de Claire », $n = 10$), OSP (cupped oyster « Spéciale », $n = 10$). Non-parametric

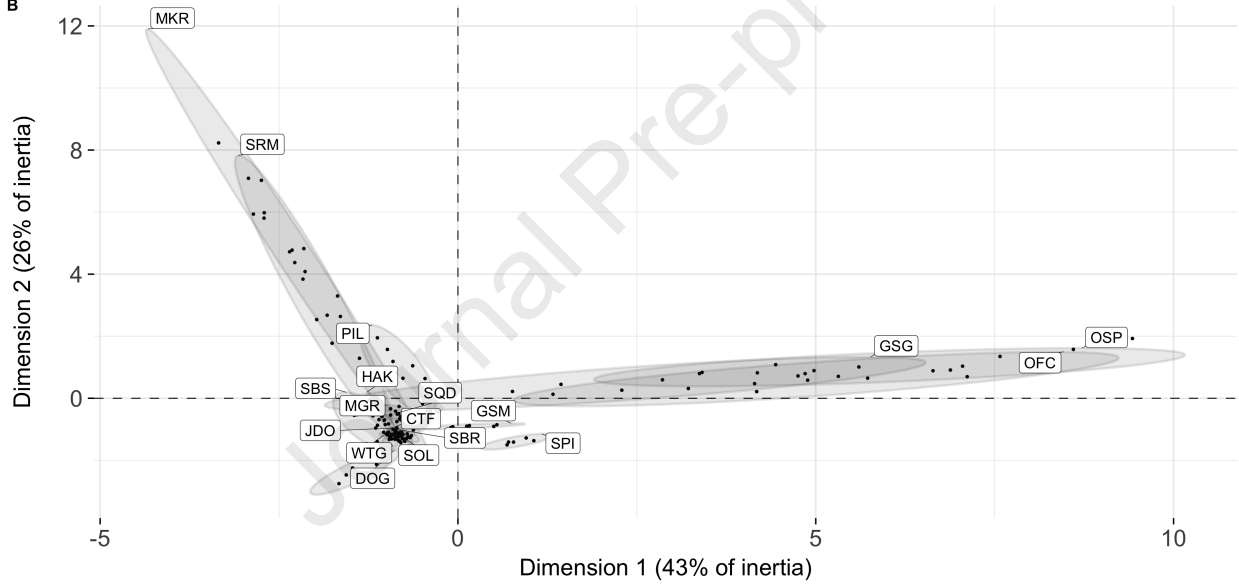
Kruskal-Wallis tests showed for all elements p -values < 0.05 (Ag: $\chi^2 = 33.749$, $df = 4$; Cd: $\chi^2 = 20.791$, $df = 4$; Hg: $\chi^2 = 143.54$, $df = 17$; Pb: $\chi^2 = 41.48$, $df = 5$; As: $\chi^2 = 150.75$, $df = 17$; Cu: $\chi^2 = 155.09$, $df = 17$; Fe: $\chi^2 = 103.5$, $df = 13$; Mn: $\chi^2 = 145.31$, $df = 17$; Se: $\chi^2 = 118.5$, $df = 17$; Zn: $\chi^2 = 143.34$, $df = 17$). Different letters denote significant differences in trace element concentrations between species (Wilcoxon signed rank tests, without adjust method, p -value < 0.05). Blanks correspond to element contents under the limit of detection.

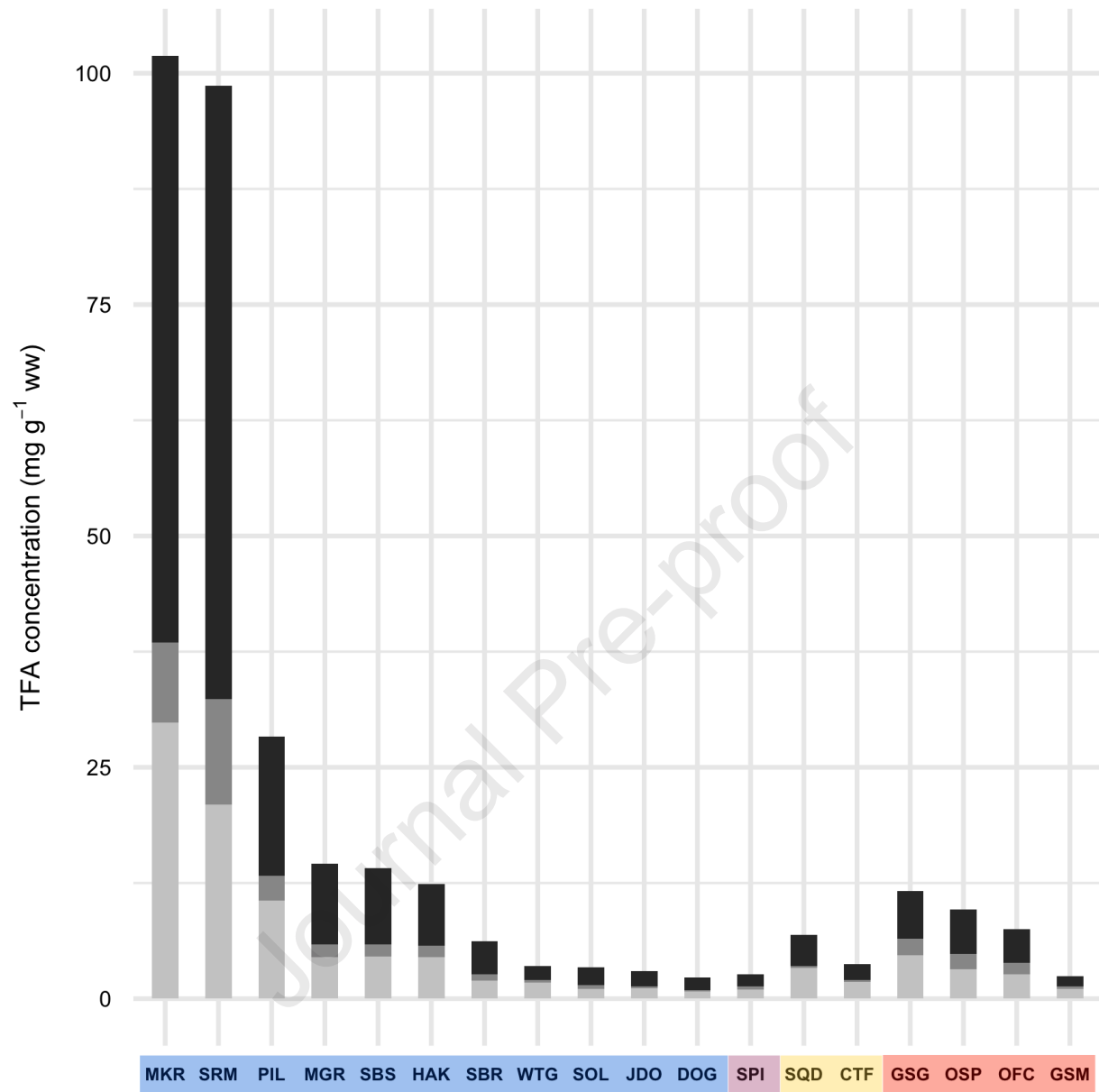
Figure 5. Estimated effect (in %) of Hg and EPA+DHA on cardiovascular heart disease risk, considering one (in black) or two (in grey) 150 g seafood servings per week of the edible parts of fish (blue), crustacean (purple), cephalopods (yellow), and bivalves (red). Studied species were: for fish: DOG (lesser-spotted dogfish, $n = 5$), HAK (European hake, $n = 9$), JDO (John Dory, $n = 10$), MGR (meagre, $n = 10$), MKR (Atlantic mackerel, $n = 10$), PIL (European pilchard, $n = 10$), SBR (black seabream, $n = 10$), SBS (European seabass, $n = 10$), SOL (common sole, $n = 10$), SRM (surmullet, $n = 10$), WTG (whiting, $n = 10$); for crustacean: SPI (Atlantic spinous spider crab, $n = 5$); for cephalopods: CTF (common cuttlefish, $n = 10$), SQD (European squid, $n = 10$); for bivalves: GSG (great Atlantic scallop gonad, $n = 7$), GSM (great Atlantic scallop muscle, $n = 6$), OFC (cupped oyster « Fine de Claire », $n = 10$), OSP (cupped oyster « Spéciale », $n = 10$).

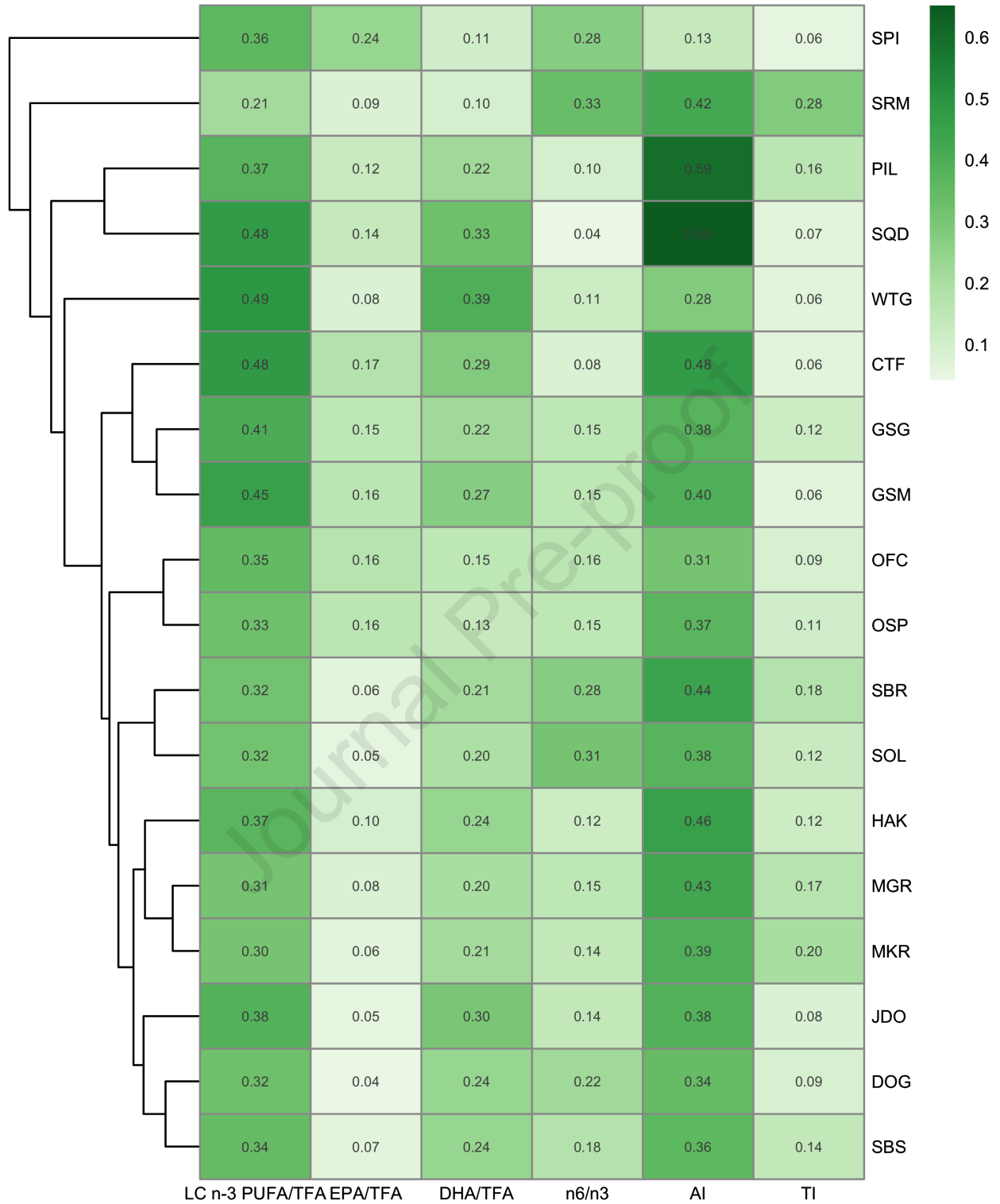
A

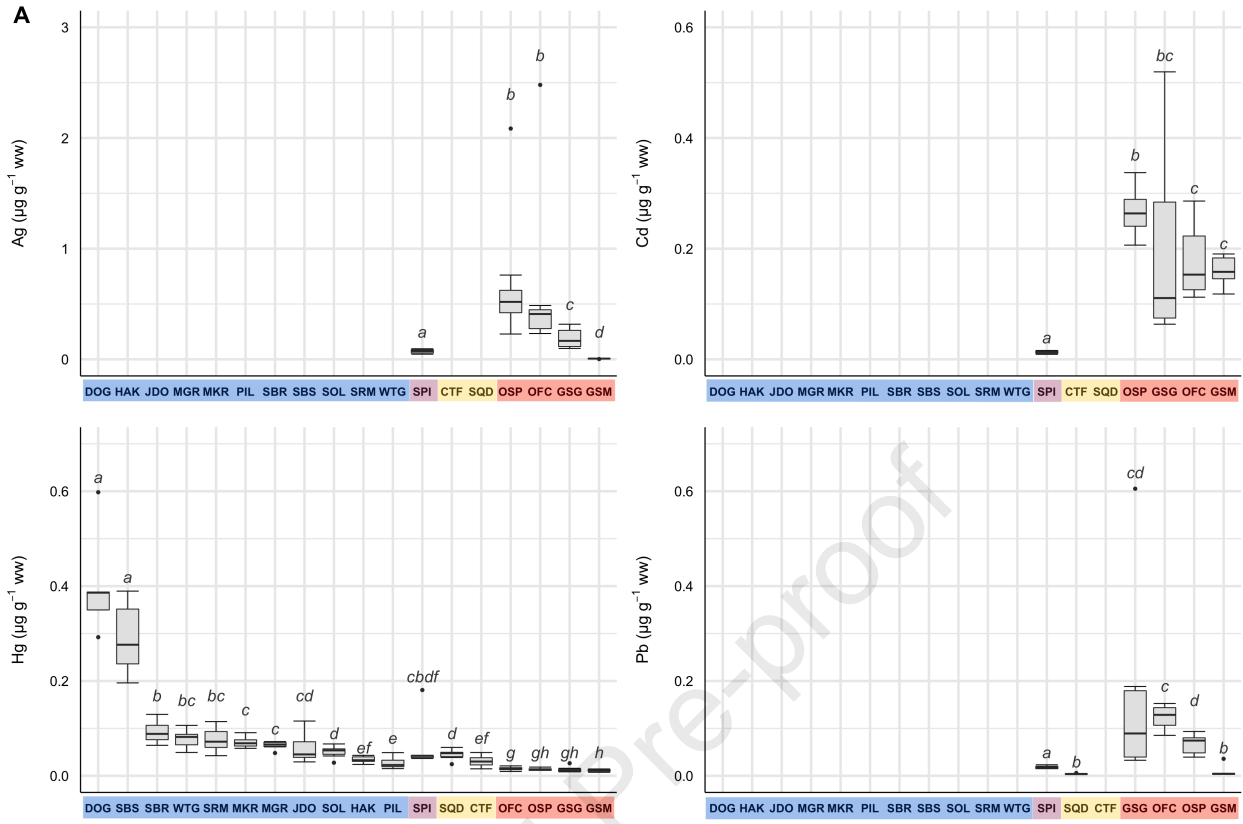


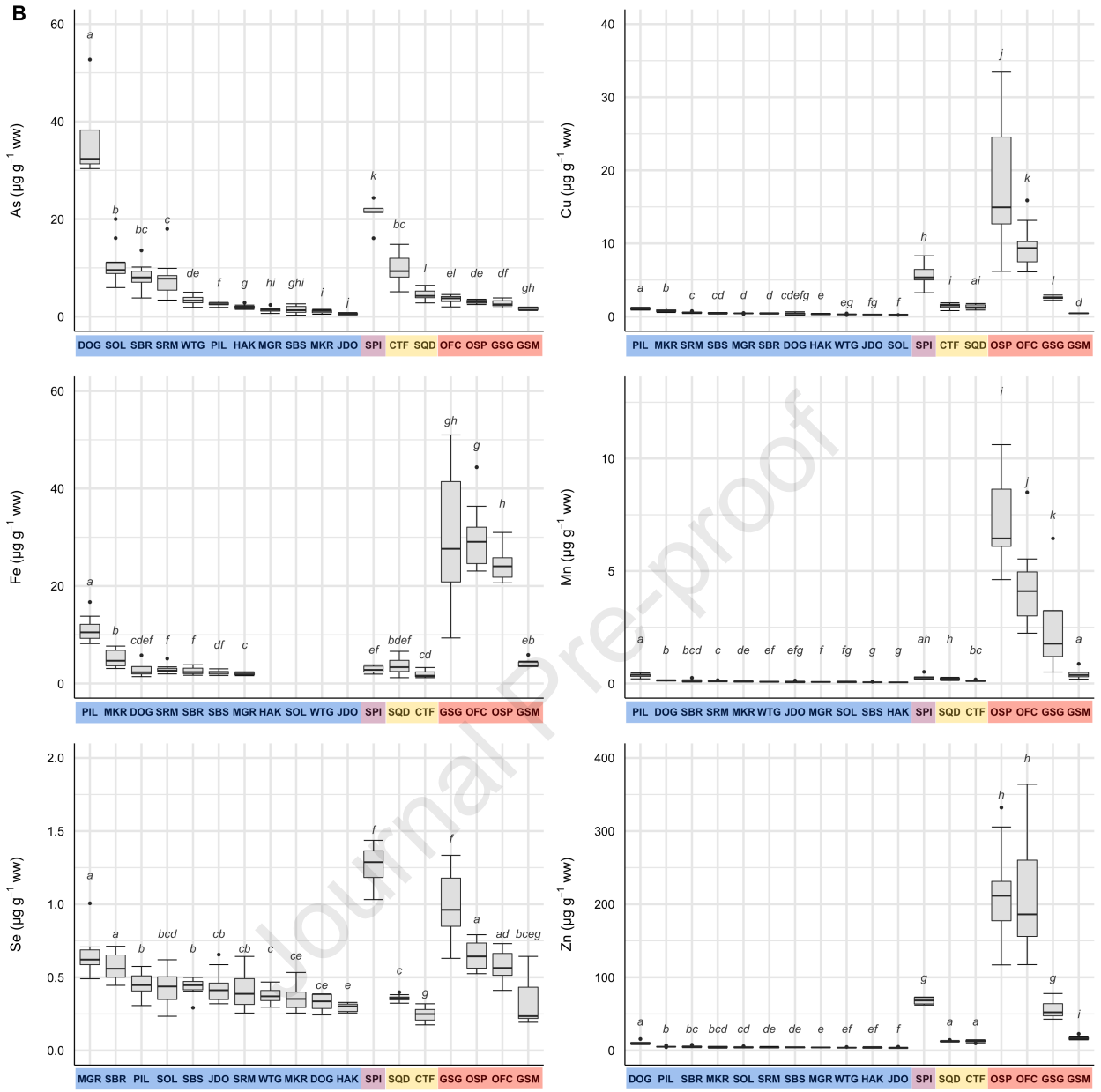
B

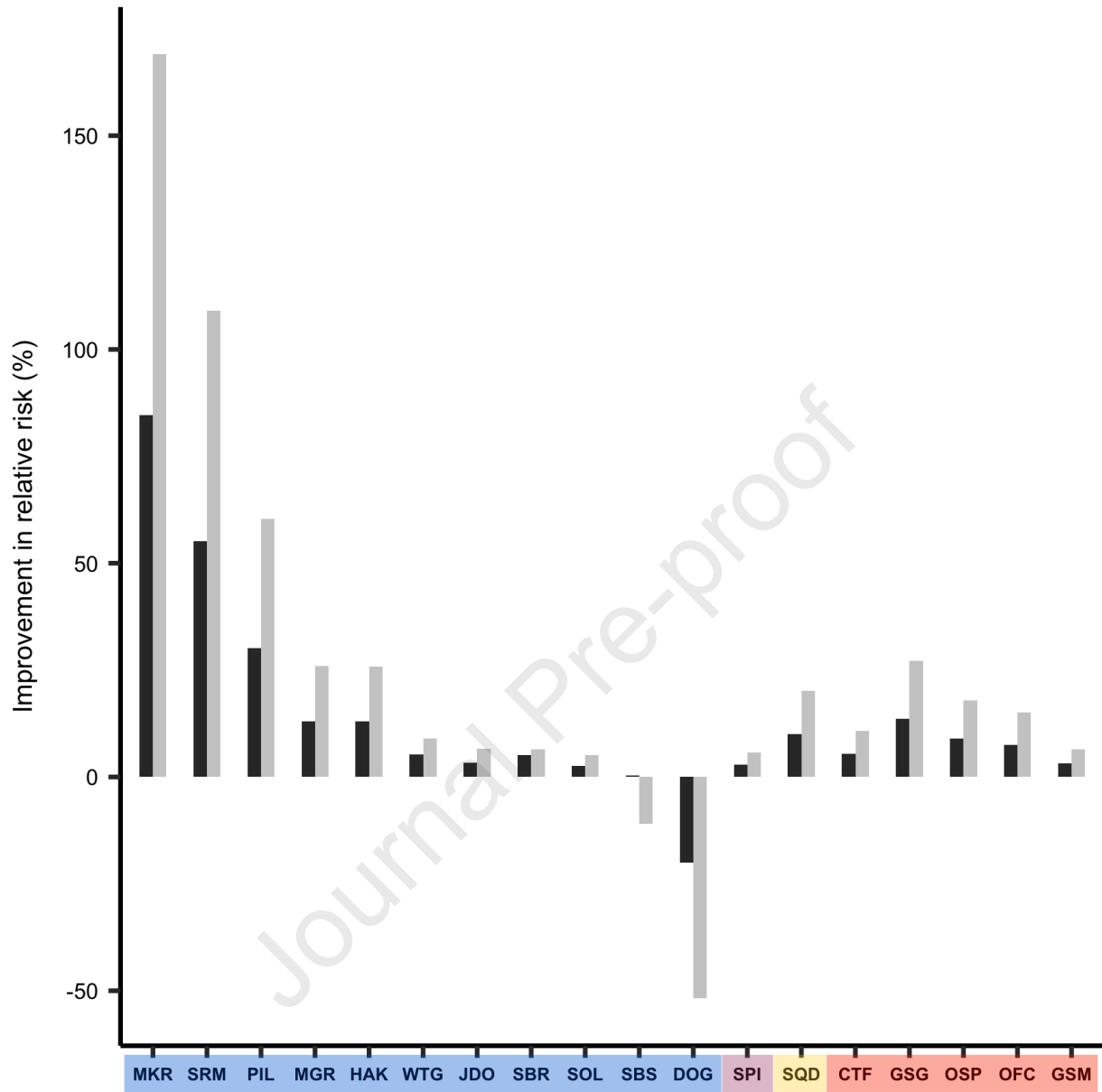












Declaration of interests

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

Journal Pre-proof