Multiscale analysis of living benthic foraminiferal heterogeneity: Ecological advances from an intertidal mudflat (Loire estuary, France)

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Abstract :

An unprecedented sampling effort on the Loire estuary allowed a multi scale approach to identify parameters controlling density variations of benthic foraminifera. Indeed, the distances between the samples analysed for this study vary from 1 cm to hundreds of kilometres. To catch this range of distance variations, a model called Scale Variance Analysis was build describing the participation of each scale to the total observed variance. The SVA model requires, for each scale, the stability of relative variance. A comparison with the Moran's Index and experimental variogram is proposed showing coherent conclusions with the SVA analysis. The analysis shows that in order to maximize information on foraminiferal density variation, sampling campaigns should be designed with stations distant from few meters to 1 km, with a particular focus on the hectometre scale. A range of scale too rarely investigated in the community of benthic foraminifera ecology. Next, based on two intertidal mudflat stations separated of few hundred meters, the present study shows that for Ammonia tepida, the scale dependant preponderant parameters is the Chl a concentration in the top first centimetre. Contrastingly, the indicators of food quality such as the lability index and the oxygen penetration depth do not seem to affect A. tepida densities. This high quantity, low quality diet is interpreted as an opportunistic behaviour that is indirectly confirmed by a kinetic approach. This approach compares the deep infaunal microhabitat density with the shallow infaunal microhabitat density. The identical ratio indicates quick saturation of the available resources.

Highlights

▶ A new indicator to quantify spatial heterogeneity among different scale, the Scale Variance Analysis (SVA), is proposed. ▶ It revels higher heterogeneity of Ammonia tepida between 1 and 1000 m, in agreement with Moran's Index and experimental variogram. ▶ At the hectometer scale, direct correlation of *A. tepida* density with ChI a is observed. ▶ Equilibrium between deep and shallow infaunal population is rapidly reached for *A. tepida* probably due to rapid saturation of environmental capabilities.

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2 Introduction

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43 Benthic foraminifera are ubiquitous in sediment, giving them the potential to be bio-44 indicators of ecosystem functioning in all marine environments including transitional areas 45 such as intertidal mudflats (e.g. Debenay et al., 2001; Schönfeld et al., 2012). Moreover, their 46 distribution evolves monthly to seasonally (Alve and Murray, 2001; Kitazato et al., 2000) which fits with the frequency of most survey sampling and smooths the influence of 47 48 environmental parameters that changes more rapidly (e. g. semi-diurnal tides). However, a 49 comprehensive framework of their spatial heterogeneity is still lacking, while it could be an 50 important step towards a standardization of the sampling strategy required for the 51 optimization and generalization of their use as bio-indicators (Schönfeld et al., 2012). 52 Additionally, assuming that a consequence has similar heterogeneity that its cause, multiscale 53 analysis of heterogeneity would improve our understanding of the preponderant parameters controlling species distribution (e. g. Talley 2007). 54

55 Since the 50's, ecological studies take advantage of some mining engineer geostatistical 56 methods to express the spatial distribution variability using synthetic indices (Legendre and 57 Fortin, 1989). For example, the Moran's Index (Moran, 1950) has been applied on benthic 58 foraminifera by Hohenegger et al. (1993) and Thibault de Chanvalon et al. (2015) to identify 59 scales of patchiness. However, in the case of our study from the Loire estuary, to compare 60 samples acquired at different spatial resolutions, and with a different sampling size, scale 61 variance analysis is better-suited (Moellering and Tobler, 1972). In the Loire estuary, 62 Mojtahid et al. (2016) documented the spatial distribution patterns of living foraminifera at a 63 kilometric to decametric scale using a Van Veen grab sampler while Thibault de Chanvalon et al. (2015) described foraminiferal distribution based on 1 cm³ samples from the intertidal 64

mudflat "Les Brillantes". New data from 6 sites on the "Les Brillantes" mudflat distant from
few metres to hundreds of metre are here gathered with the Mojtahid et al. (2016) and the
Thibault de Chanvalon et al. (2015) datasets.

68 In the present paper, we will illustrate the role of such geostatistical tools to characterise the 69 preponderant factor controlling some species distribution following a three steps demarche 70 consisting in i) exploring, ii) identifying and iii) validating the causality relationship. In the 71 case of the Ammonia tepida density distribution in the Loire estuary, we will i) determine the 72 most significant scale to assess spatial foraminiferal density variation based on multiscale 73 analyses, in order to ii) identify the preponderant mechanism controlling foraminifera density assuming it has a similar heterogeneity. This second step will be achieved looking at direct 74 75 correlation with foraminifera density. To validate this deterministic approach we will iii) test the steady state between foraminifera density and the controlling factor based on vertical 76 77 distribution analysis. Hence, with depth, different microhabitat characterised by different 78 feeding time, exchange their population due to mobility or bioturbation. We will demonstrate 79 that equilibrium between the density of these microhabitats indicate steady state regime at all 80 depths.

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3 Materials and methods

83 **3.1 Study area**

84 "Les Brillantes" mudflat is located in the inner part of the Loire River estuary (Figure 1), the outlet of a 117,045 km² - drainage basin composed of both sedimentary and granitic rocks. 85 The mean discharge of the Loire River is 900 m³s⁻¹, varying from 120 m³s⁻¹ in summer to over 86 5000 m³s⁻¹ during winter flood, and leading to a high seasonal variability of water salinity at 87 "Les Brillantes". The Loire estuary is macrotidal and hyper-synchronous (Le Floch, 1961) 88 89 with a tidal range from 2 to 7 m, producing large intertidal areas ("Les Brillantes" is 1350 ha) 90 and important sediment resuspension. Therefore, sediment grain size characteristics at "Les 91 Brillantes" is quite homogeneous with silty-clay unimodal deposits with a median size of 92 between 10 and 20 µm (Coynel et al., 2016). The 12 hours tidal cycling produces large daily 93 changes of salinity and structuration of water column, especially when it chimes with low 94 flow (that mostly varies seasonally, see Thibault de Chanvalon et al., 2016) and high tidal 95 intensity (whose main cycle lasts 2 weeks). Depending on the period of the year, the moment 96 of the lunar cycle and the time of the tidal cycle, salinity at the sampling point can vary from 97 35 to 0 and was about 20 in May 2013 and 14 in September 2012.

Two stations on the unvegetated slikke were chosen to study the spatial variability at the metre scale in "Les Brillantes" mudflat (Figure 1): Site 1 is located 20 m offshore from a 1 mhigh-eroded cliff while Site 2 is 500 m offshore. The main difference between the two stations lies in the longer emersion time for site 1, the closest to the shore. According to Benyoucef (2014), Site 1 is characterized by a denser microbiofilm (*i.e.* microphytobenthos composed of diatoms).

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105 **3.2** Sampling strategies

106 In this study, the estimation of metric heterogeneity for micro and meiofaunal (i.e. 107 foraminifera) composition is based on three replicate interface cores (triplicates cores) from 108 the same site, distant from each other by few meters. For all the other measured parameters 109 (oxygen profiles, macrofauna), a dedicated core was sampled from each site. A similar 110 vertical sampling resolution was used for all analyses *i. e.* cores with inner diameter of 8.2 111 cm, were sliced every 2 mm until 2 cm and every 5 mm until 5 cm with the exception of 112 macrofauna of which the abundance was determined for the full core depth. To minimise the 113 temporal variability, the foraminifera samples were acquired in September 2012 and May 114 2013, simultaneously to those from Mojtahid et al. (2016, sampled in September 2012) and 115 Thibault de Chanvalon et al., 2016.

116

3.3 Biological compartment

The core triplicate dedicated to foraminifera was sliced few hours after recovery and 117 incubated overnight in Cell-TrackerTM Green (Invitrogen Detection Technologies) / 118 dimethylsulfoxide (DMSO) mixture (final concentration of 1 µmol L⁻¹) then preserved in 10% 119 120 formaldehide / 3.8% borate mixture. This method was chosen for its accuracy at 121 discriminating living from dead foraminifera since it reacts with enzymes to produce a 122 fluorescent compound (see details in Bernhard et al., 2006). Only the larger (>150 µm) 123 fraction including adult specimens was conserved for identification (see detailed procedure in 124 Langlet et al., 2014). Then, only foraminifera fluorescing continuously and brightly under an 125 epifluorescent binocular (Olympus SZX12 with a fluorescent light source Olympus URFL-T) 126 were picked out, counted and determined. Note that, the species that we refer to as Ammonia 127 tepida in the following text corresponds to the phylotype T6 according to the recent 128 classification of Ammonia sp. (Richirt et al., 2019). This is a common Ammonia phylotype in 129 the European intertidal mudflats (Bird et al., 2020).

For aminiferal oxygen uptake (FOU) is calculated with the equation (1), with $R_i(T_{13})$ being the respiration rate of the species i from laboratory measurement at 13 °C in pmol O₂ ind⁻¹ d⁻¹ from Geslin et al., 2011, the exponential being the Arrhenius temperature correction (with T_A a constant in °C determined by Bradshaw, 1961) and d°_i the measured areal density of living for aminifera in the oxic layer in ind m⁻²:

$$FOU = \sum_{i} R_{i}(T_{13}) \exp(\frac{T_{A}}{T_{13}} - \frac{T_{A}}{T_{obs}}) d_{i}^{O}$$
(1)

135

We calculated the average living depth ALD_x, initially proposed by Jorissen et al. (1995) to
describe quantitatively the microhabitats distribution, following equation (2):

$$ALD_x = \frac{\sum_i n_i D_i}{\sum_i n_i} \tag{2}$$

With n_i the number of specimens in interval i, D_i the midpoint of sample interval and x the
lower boundary of the deepest sample.

Another core triplicate was dedicated to microphytobenthos (MPB) and frozen *in situ* by liquid nitrogen. Pigments extraction used a cold mixture (4°C) of 90% methanol/0.2M ammonium acetate and 10% ethyl acetate (90/10 vol/vol) and measurement performed by HPLC (see Méléder et al., 2005 for details). To assess organic matter quality, we used the lability index, LI = Chl *a*/(Chl *a* + Pheo *a*), with Pheo *a* corresponding to the total amount of phaeophorbides a and pheophytins a, respectively due to grazing and microbial activity (*e.g.* Bianchi and Findlay, 1991; Cartaxana et al., 2003).

147 Finally, each core of the triplicate dedicated to the macro-invertebrates was homogenized
148 over its full depth (35cm) sieved at 1 mm and preserved in 4% formaldehyde before species
149 identification and counting in the >1 mm fraction.

150 **3.4 Oxygen fluxes**

Dissolved oxygen vertical profiles were measured in a separate core in the dark, within few hours after sampling using a Clark-type microelectrode with a 50 μ m thick tip (OX50, Unisense, Denmark) connected to a multimeter (Unisense) in a temperature controlled bath. Twelve and 10 oxygen profiles were measured in September 2012 and 24 and 4 profiles in May 2013 at stations S1 and S2 respectively. Diffusive O₂ uptake (DOU) was estimated with the PROFILE software by fitting the measured oxygen concentration with concentration from diffusion-reaction models (see details in Berg et al., 1998).

158 **3.5 Statistical analysis**

159 3.5.1 Moran's Index

Patchiness effect was explored using spatial correlograms built using the Moran's Index (I), computed with R (package "spdep" following Bivand and Wong, 2018 and Fortin and Dale, 2005, Equation (3)). This index calculates the similarity of pair values for one neighbourhood compared to the global mean of the dataset, a neighbourhood being defined by a weighted (wd) function of the distance (l_{ij}) between the pair values (x_i, x_j). Here, we used a weighted function sensitive only to the scale of the distance, *i. e.*:

$$I(d) = \frac{\sum_{i}^{n} \sum_{j \neq i}^{n} w_{d}(l_{ij})(x_{i} - \bar{x})(x_{j} - \bar{x})}{\sum_{i}^{n} (x_{i} - \bar{x})^{2}} \times \frac{n_{d}}{\sum_{i,j}^{n} w_{d}(l_{ij})}$$
(3)

$$w_{d}(l_{ij}) = \begin{cases} 1, & 10^{d} < l_{ij} < 10^{d+1} \\ 0, & \text{otherwise} \end{cases}$$
(4)

ith

and d, the scale of interest, n, the number of samples and n_d the number of samples forming at least one pair. Significance of values is estimated based on Monte-Carlo analyses provided in the "spdep" package (function moran.mc, done with 9999 simulations). This function compares the I value obtained from the original dataset with a distribution produced by many simulated I values. First, these simulated I values were obtained by random distribution of all density values. Second, to take into account that in some case, very few sample formed pairs, the simulated I values were obtained by exchanging randomly 10% of the samples forming pairs with random samples from the dataset.

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3.5.2 Scale Variance Analysis

Scale variance analysis (SVA) decomposes the total variance of a dataset to identify the 175 contribution of each scale to the variance (Moellering and Tobler, 1972; Wu et al., 2000). The 176 177 SVA compares each sample to a local mean which complements the Moran's Index, in which 178 samples are compared to the global mean. This approach requires a priori explicit definition 179 of scales of interest and a priori delimitation of all regions, necessarily nested over the 180 different scales. By convention, for a dataset hierarchized over k scales of interest, the scale 1 181 is the size of the initial samples, that are gathered in local regions belonging to the scale 2. 182 Then, the mean of each local regions is treated as sample of the scale 2 and are gathered again 183 in intermediate regions belonging to the scale 3. The process is repeated until the scale k, 184 covering the extent over which the sampling has been done. On each scale, the concept of 185 "scale variances" is introduced which corresponds to the variance of samples of scale h over a 186 region of scale h+1.

187 The following details are inspired from Moellering and Tobler (1972) but using different 188 writing. For samples of the scale h, gathered in regions belonging to the scale h+1, the scale 189 variance, $V_i^{h \rightarrow h+1}$, is defined according to the equation (5).

$$V_i^{h \to h+1} = \frac{1}{n^h} \sum_{j=1}^{n^h} (x_i^{h+1} - x_{i,j}^h)^2$$
(5)

With x_i^{h+1} , the mean value of all samples nested in the group i; $x_{i,j}^h$ the different sample (whose size is belonging to the scale h) value constituting the group i and n^h the number of 192 sample $x_{i,j}^h$ constituent the group i. For simplicity, we here assume that n^h does not depend of 193 i, *i.e.* all groups of size belonging to the scale h+1, are constituted by the same number of 194 sample from the scale h. Then, for being representative of the importance of the variance of a 195 certain scale (h) over the whole dataset one need to look at the mean of the scale variances of 196 the scale h, according to equation (6 and 7)

$$\overline{V^{h \to h+1}} = \frac{1}{N^{h+1}} \sum_{i=1}^{N^{h+1}} V_i^{h \to h+1}$$
(6)

W

 $N^{h+1} \sum_{i=1}^{k} n^{p}$ $N^{h+1} = \prod_{p=h+1}^{k} n^{p}$ (7)

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Equation (7) describes that N^{h+1} is equal to the number of group whose size belonging to the scale h+1. Thus, one can demonstrate (see Appendice 1) that for a dataset hierarchized over k scale of interest, the variance (VAR) can be decomposed into the sum of the mean of the scale variances (equation (8)):

 $VAR = V^{1 \to k} = \sum_{h=1}^{k-1} \overline{V^{h \to h+1}}$ (8)

γ	n	1
	υ	1

202 However, SVA requires a complete dataset with all values of scale 1 totally enumerated (Moellering and Tobler, 1972). In our case, the data range over 8 orders of magnitude from 203 204 samples at the cm scale to a sampling area of hundreds of kilometres. Assuming 5 samples per group, an exhaustive sampling would require $5^7 = 78.125$ analyses. To overcome the 205 206 analytical limitation to produce so many analyses, a supplementary assumption is required: 207 the scale stability of relative variance. For each scale, this assumption assumes that all groups i are characterized by the same relative scale variance (RV^{h->h+1}). The RV^{h->h+1} is defined as 208 the square of the relative standard deviation *i. e.* for any i: 209

$$RV^{h \to h+1} = \frac{V_i^{h \to h+1}}{(x_i^{h+1})^2}$$
(9)

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This is a heavy assumption but it allows the calculation of any mean scale variance as soon as both the relative scale variance and the sum of the square of the mean of all higher scale are known (see Appendice 2) according to equation (10):

For any
$$h \le k-2$$
 $\overline{V^{h \to h+1}} = \frac{RV^{h \to h+1}}{N^k} \prod_{i=h+1}^{k-1} (1 + RV^{i \to i+1}) \sum_{j=1}^{N^k} (x_j^k)^2$ (1)

This relation indicates that the means of the scale variances for a certain scale can be calculated as soon as the relative scale variance is known. Lastly, this relation indicates that direct comparison of the relative scale variance from different scales of the same dataset is not meaningful and that comparison of the mean scale variance has to be preferred.

217 3.5.3 Experimental variogram

To complement Moran's Index, that compares samples to the global mean, and SVA, that compares samples to a local mean, an experimental variogram was built. In this case, each pair of samples is compared to the square of their difference without referring to any external mean. Then, the gamma value (γ) is computed as the half of the mean of the values that belong to a certain distance, according to equation (11).

$$\gamma(d) = \sum_{i}^{n} \sum_{j \neq i}^{n} w_{d}(l_{ij}) (x_{i} - x_{j})^{2} \times \frac{1}{2 \sum_{i,j}^{n} w_{d}(l_{ij})}$$
(11)

W

$$w_d(l_{ij}) = \begin{cases} 1, & 10^d < l_{ij} < 10^{d+1} \\ 0, & \text{otherwise} \end{cases}$$

ith

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224 3.5.4 Application to the Loire estuary dataset

225 Moran's index, SVA and experimental variogramare calculated based on the average 226 density of *A. tepida* in the first centimetre depth measured in this study, in Thibault de 227 Chanvalon et al. (2015) and in the study of Mojtahid et al. (2016). This combined dataset is 228 represented in Figure 1 and is not regularly distributed in the Loire area. For example, for the 229 Moran's index calculation, each class of distance covering scales from centimetre to 230 hectometre (hundreds of meters) are represented by less than 72 sample pairs. The scale 231 variance analysis (SVA) was calculated based on regions delimited arbitrary by the black 232 lines on the Figure 1. The column "available sampling / exhaustive sampling" in Table 1 233 summarized the number of regions per scale and compared it with an exhaustive sampling as 234 theoretically requested. The scale levels 3, 4 and 5 does not contain enough samples to be 235 gathered into at least one group. Thus, the mean of the scale variance for scale 3, 4 and 5 236 could not be calculated directly and was estimated by the difference between the global 237 calculated variance and all the means of the scale variance. Moreover, all information for 238 close (< 100 m) samples belong to or are close to stations 1 and 2. Nevertheless, such 239 limitations are very frequent in foraminiferal dataset and the spatial recovery obtained with 240 this combined dataset is rare in the literature motivating the pursuit of the spatial heterogeneity analysis. 241

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3.6 Model of microhabitat equilibrium

243 The relation between deep infaunal and shallow infaunal foraminiferal faunas is modelled 244 using a dynamic 2 boxes-model (whose equations are detailed in Figure 2A) based on typical 245 assumptions drawn from ecological studies (Levin, 1976). The shallow infaunal box is 246 characterised by a population (pop_{sh}), a first order mortality rate (k_{d_sh}) and a reproduction rate 247 described with the Verhulst equation, that is, a first order rate (k_p) decreasing to zero as the 248 population saturates (pop_{sat}) available resources. The deep infaunal box is characterised by a 249 population (pop_{de}) and a first order mortality rate ($k_{d de}$). The transfer between the 2 boxes 250 follows a first order rate (kech), roughly estimating biomixing. The ratio of deep over shallow 251 infaunal population, $\alpha = pop_{de} / pop_{sh}$, predicted by this model after an important increase of 252 environment capability (by 500 fold in this example) is shown in Figure 2B. First the shallow 253 population increase, hence α decrease. After a short delay, deep infaunal population increase 254 too, leading to an equilibrium between exponentially growing shallow and deep infaunal 255 population and α reaching a plateau (α_{exp} , on Figure 2B). The higher k_{ech} is, the faster the first 256 plateau is reached. Finally, once the population reaches the limits of the environment 257 capabilities another equilibrium is observed between the two populations that is characterised 258 by a second plateau (α_{sat} on Figure 2B). The higher k_p is, the faster the second plateau is 259 reached, other parameters having much less influence on the rate of α changes. The α value at 260 the plateau is defined as:

$$\alpha_{sat} = \frac{k_{ech}}{k_{ech} + k_{d_de}}$$
(1)
2)

$$\alpha_{exp} = \begin{cases} 1 - \frac{b}{2} & \text{if } b \ll 2\\ \frac{1}{b} & \text{if } b \gg 2 \end{cases}, \text{ with } b = \frac{k_p + k_{d_de} - k_{d_sh}}{k_{ech}}$$

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4 Results

264 **4.1 Environmental parameters**

Table 2 summarizes most of the environmental parameters extracted from public survey databases (banque HYDRO, SYVEL and SHOM networks, see glossary) and previous publications (Benyoucef, 2014; Thibault de Chanvalon et al., 2016).

The two campaigns were characterized by contrasted river discharge (150 m³ s⁻¹ in 268 September versus 1200 m³ s⁻¹ in May), organic carbon content in the top of sediment (2.1 % 269 versus 2.8 %), salinity (22 versus 8, respectively) and temperature (17 °C versus 13 °C). 270 271 Contrastingly, nutrients showed mainly spatial variation with a higher concentration of dissolved phosphorus at Station 1, closer to the shore, with 15.4 and 14.2 µmol L⁻¹ for 272 September 2012 and May 2013 respectively, compared to 2.7 and 4.8 µmol L⁻¹ for Station 2 273 274 (Thibault de Chanvalon et al., 2016). Finally, the oxygen penetration showed both important spatial and temporal variability with a lower value in May 2013. 275

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4.2 Biological parameters

277 Table 3 indicates that the three studied biological compartments differ significantly 278 between the two stations and less clearly between the two campaigns. Station 1 shows higher 279 abundances of microphytobenthos (the average of the two campaign for chlorophyll A is 340 mg m⁻² in Station 1 versus 180 mg m⁻² in Station 2), macrofauna (770 ind m⁻² in Station 1 280 versus 290 ind m^{-2} in Station 2 on average) and living foraminifera (78 ind / 10 cm³ versus 24 281 282 ind / 10 cm³) than Station 2. While Ammonia tepida (>70%) dominates foraminiferal 283 communities in both stations, macrofaunal assemblages switch from a dominance of the 284 polychaetes *Hediste diversicolor* (>75%) at Station 1 to a dominance of both the bivalve 285 Scrobicularia plana and the polychaetes Heteromastus filiformis at Station 2. The main seasonal variation visible in both stations is for macrofauna with an increase of H. filiformis 286

associated to a decrease of *S. plana* in May 2013. However, in Station 1, but only there, foraminifer density decreases significantly in May 2013 with, for example, a near disappearance of *H. germanica* going from 36 ind / 10 cm³ down to 2 ind / 10 cm³ in the first top centimetre. The lability index (LI), that is higher than 0.9 in both stations and seasons, indicates important *in situ* autotrophic activity of microphytobenthos.

4.3

4.3 Vertical distribution

The densities of benthic foraminifera and Chl a concentration presented in Figure 3 show in several cases an exponential decrease of densities with depth. The high density of *H*. *germanica* in the top 2 millimetres in September 2012 at Station 1 (67 ind /10 cm³) appears concomitantly to particularly high densities of *A. tepida* (829 ind /10 cm³) and Chl a (201 mg m⁻²). However, in detail, the exponential decrease associated to this high density appears more progressive than for *A. tepida* and Chl a, with a minimum reached at 1.4 mm depth for *H. germanica* versus 0.6 mm depth for the others.

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4.4 Foraminifera aerobic respiration rates

301 The respiration rates (RR) estimated for the foraminiferal population for each season and 302 station as well as the relative foraminiferal contribution to DOU are shown in Table 4. The estimated respiration rates are 3357 \pm 117 pmol O₂.ind⁻¹.d⁻¹ and 2154 \pm 75 pmol O₂.ind⁻¹.d⁻¹ 303 for A. tepida and 685 ± 134 pmol O₂.ind⁻¹.d⁻¹ and 439 ± 86 pmol O₂.ind⁻¹.d⁻¹ for H. germanica 304 305 at respectively 17°C (September) and 13°C (May). The maximal relative contribution of the 306 foraminiferal fauna to DOU was 2.3%, at station S1 in September 2012 mostly carried by the 307 dense population of A. tepida. For all other samplings, foraminiferal respiration rates (sum of 308 A. tepida and H. germanica respiration rates) are much lower and varies from 0.085 to 0.099 mmol $O_2 m^{-2} d^{-1}$. 309

4.5 Statistical analysis

311 Figure 4A shows that the Moran's index is significantly higher than zero for all scales 312 below 0.1 km, which indicates that foraminiferal densities are grouped into patches of 313 hundreds of metre's size. The negative value for Moran's index between 1 and 10 km (Figure 314 4A) indicates that most of the difference between environments occurs between 1 and 10 km 315 for A. tepida densities. SVA results presented in Figure 4B (black dots) show that most of the 316 variance (average scale variance > 10%) comes from the scales between 1 m and 1 km but the 317 lack of data (see Table 1) prevents us from a better accuracy. Interestingly, scale variance 318 analysis quantifies that scales between 1 cm and 1 m counting for 3.6 times less to the overall 319 variance than the scales between 1 m and 1 km. Stability of SVA, even based on our sparse 320 dataset, is illustrated by the white dots of Figure 4B. It shows that the SVA processed without 321 the particularly dense Station 1 does not modify significantly the results. The Figure 4C 322 shows the square differences of paired samples in grey dots. The variogram calculated for 323 each distance range is particularly high between 100 m and 1 km.

324

5 Discussion

325 **5.1** Critical scale of heterogeneity identified by multiscale analysis:

326 In "Les Brillantes" mudflat, our data show an overall high surface foraminiferal density (up 327 to 829 ind. / 10 cm³) and very low diversity (only two different species identified) (Figure 3). High surface density (over 100 ind. / 10 cm³) of foraminiferal fauna is commonly reported in 328 329 intertidal mudflat surfaces from estuaries (Debenay et al., 2006; Thibault de Chanvalon et al., 330 2015) or inlets (Alve and Murray, 1994; Cesbron et al., 2016; Goldstein et al., 1995) while a 331 very low diversity is more typical of macrotidal estuaries. Indeed, only the species that are the 332 most tolerant to large daily salinity variations can grow in macrotidal estuaries (Murray, 2006), especially on non-vegetated mud. These later are for instance known for their absence 333 334 of agglutinated species (Berkeley et al., 2008).

335 Debenay and Guillou (2002) demonstrated that the estuarine compatible species colonize 336 successive areas along the salinity gradient. However, they did not identify the preponderant 337 forcing among all the parameters covarying with salinity. For example, in the Loire estuary, 338 the regular dredging of the navigation channel has also been invoked to explain the extreme 339 poverty of the diversity of the foraminifera fauna with only three living species reported over 340 the whole salinity gradient (Mojtahid et al., 2016). In addition to the constrains specific to 341 estuarine environments, more common forcing such as grain size, food availability or food 342 quality would also modify foraminiferal growth opportunities and produce, in fine, an 343 irregular surface density distribution such as that illustrated in Figure 1. In an attempt to catch 344 such a variability, deterministic models (e.g. TROX model from Jorissen et al., 1995) build on 345 predefined forcing, estimate compliance between a species and an environment while they 346 hardly quantify the density variations. This issue is particularly critical in dynamic 347 environments where kinetic effects, such as new colony settling (Alve, 1999; Weinmann and 348 Goldstein, 2017) may induce changes in hydrodynamic dispersal or hysteresis associated to 349 transient environmental changes, prevail over saturation of the environment capabilities. A 350 complementary approach used in ecological survey, based on geostatistical models (e.g. 351 Talley, 2007) proposes at first, to synthetize spatial patterns in order to infer causality as a 352 second step. From the three different geostatistical models chosen for this study (Moran's 353 Index, SVA and experimental variogram, Figure 4) one common picture appears: most of the 354 density variation comes from the scales between the metre scale, that gathers all paired 355 samples distant from 1 to 10 m and the hectometre scale, that gathers all paired samples distant 356 from 100 to 1000 m. The Moran's Index (Figure 4A) underlines particularly the hectometre 357 scale where the Index decreases and crosses the zero line, changing from a distribution with 358 almost similar densities (Moran Index above 0) to a distribution with contrasted or random 359 densities (Moran Index value below or equal to zero, Figure 4A). The variogram plot (Figure 360 4C) confirms the preponderant role of the hectometre scale with the highest gamma value 361 calculated. Sadly, the lack of paired samples distant from 1 m to 100 m could hide unexpected 362 changes and prevents us from being more precise about the most significant scale. Identical 363 limitation is visible for the SVA that equally distributes the missing variance into the scales 364 with missing paired samples (from metre to hectometre scale). However, SVA model predicts 365 preponderant role of at least one of these scales by the difference between the overall variance 366 to the variance attributed to the other scales. At all events, in order to maximize information 367 on foraminiferal density spatial distribution, we recommend designing future sampling 368 campaigns with stations distant from 1 metre to 1 kilometre, with a particular focus on the hectometre scale. 369

The geostatistical models are apparently in contradiction with the importance habitats succession along salinity gradient (Debenay and Guillou, 2002) since the SVA attributes only 0.9 % of the total variance to the scale of salinity changes (over 100 km), the minimum 373 attributed to any scale. Qualitative analysis solves this discrepancy. For example, a qualitative 374 SVA coded with values equal to 0 when no Ammonia tepida is observed, to 1 when A. tepida 375 is a minor species (<10%) and equal to 2 when A. tepida is a major species (>10%) leads to 376 drastically increase the importance of the estuarine scale (black diamonds on the Figure 4B). 377 In the qualitative analysis, the scale over 10 km produces 41% of the total variance and 378 therefore in strong agreement with the importance of the salinity gradient. Taken together, 379 these results highlight the efficiency of deterministic models for qualitative predictions, 380 understood as the order of magnitude of foraminiferal population densities and their lacks to 381 quantify predictions. Geostatistical models represent promising tools to cross this gap 382 especially when performed in combination with deterministic models. For example, scale 383 analyses can by hyphened with studies of environmental processes in order to associate one 384 preponderant process to each scale of important variation. This exercise is proposed in the 385 following discussion.

5.2 Limiting factors at the Les Brillantes mudflat scale

387 The focus on Les Brillantes mudflat allows investigation of processes explaining density 388 variations over few hundreds of meters, a critical scale identified from the geostatistic models. 389 On one hand, our results show that the two stations present very few qualitative differences, 390 *i.e.* changes over order of magnitudes - the most significant being *H. germanica* in September 391 2012 (Table 3) with high density in Station 1 probably produced by optimal conditions for 392 development of propagules and/or reproduction during September weak riverine influence. 393 This sensitivity tends to position H. germanica downstream from A. tepida in the estuarine 394 succession as observed by Debenay et al., 2006 and Mojtahid et al., 2016 while Alve and 395 Murray, 1994, Debenay et al., 2000 and Debenay and Guillou (2002) observations state for 396 the opposite. On the second hand, quantitative differences between stations are observed on 397 every variable with 1.5 to 5 fold more abundance of microphytobenthos, meiofauna and 398 macrofauna at Station 1 (Table 3) and up to 5 fold faster respiration (Table 4). The longer 399 emersion time, hence the longer light time exposure, and the higher nutrient input, probably 400 streaming for the grazing land of the shore via a small channel (Table 2) might favour primary 401 production compared to Station 2 and consequently may support higher density of fauna.

402 The exponential relation observed between Chl a and A. tepida (Figure 5A) indicates a 403 possible deterministic relation between primary production and A. tepida at the hectometre 404 scale and makes Chl a a good limiting factor for deterministic models at this scale. However, 405 such a relation owes a lot to the opportunistic character of A. tepida, understood as the ability 406 for a species to saturate rapidly the capabilities of an environment. Prolonging this 407 interpretation, we can estimate that all parameters varying differently than Chl a have 408 negligible effect on A. tepida density. Surprisingly, the co-varying parameters LI and the 409 OPD, reputed to trace organic matter lability, evolved differently underlying the specific diet 410 regime of A. tepida. Indeed, this species is known in the literature for being carnivorous (Dupuy et al., 2010), predating on metazoan classes (Chronopoulou et al., 2019) and thus may
ignore variation of primary production quality. Oppositely, the literature indicates that *H*. *germanica* feeds mostly on diatoms notably to steal their chloroplast (Pillet et al., 2011;
Cesbron et al., 2017; Jauffrais et al., 2018; LeKieffre et al., 2018). It seems that this so-called
"kleptoplasty" specialisation turns into a disadvantage when facing opportunistic species in
low quality high quantity food environments.

418 **5.3** Vertical distribution of foraminifera: Microhabitat vs Bioturbation

419 5.3.1 Biomixing and chemotaxis forcing

420 The fine vertical sampling resolution (Figure 3) allows a precise description of the typical 421 exponential vertical decrease of shallow infaunal microhabitat (Buzas et al., 1993). The very 422 shallow density maximum indicates a favourable environment, supposedly a reproduction 423 layer and/or propagule spawning event, due to high oxygen concentration and/or fresh organic 424 matter (Berkeley et al., 2007; de Stigter et al., 1999; Geslin et al., 2004). The progressive 425 decrease with depth is usually associated with the biomixing produced by macrofauna 426 bioturbation (e.g. Alve and Bernhard, 1995; Saffert and Thomas, 1998; Thibault de 427 Chanvalon et al., 2015), a predation-related strategy (De Stigter et al., 1998; Loubere, 1989) 428 or the occurrence of oxygen oases around animal burrows (Goldstein et al., 1995; Steineck 429 and Bergstein, 1979). However, the very steep decrease of A. tepida (minimum reached at 0.8 430 cm depth) and the systematic slight re-increase at depth (except Station 2 in May 2013, Figure 431 3) producing a shallow minimum density, corresponds to a specific pattern, likely produced 432 by the combination of biomixing and chemotaxis (BC model, Thibault de Chanvalon et al., 433 2015). In this BC model, when buried close enough to the surface, the foraminifera detect the 434 oxygenated layer and move back to the surface while, when buried deeper than their 435 pseudopod length, the foraminifera are trapped at depth in a dormancy stasis, as observed by 436 LeKieffre et al. (2017). The shallow minimum density corresponds to the chemotaxis range of 437 the foraminifera. Oases model has been discarded in these stations because of the absence of 438 correlation at the centimetre scale between deep living foraminifera and burrow traces 439 (Thibault de Chanvalon et al., 2015).

In September 2012, when surface densities were high enough, *H. germanica* densities presented a similar pattern than *A. tepida* but with a less steep decrease since the shallow minimum is reached at 1.4 cm depth for *H. germanica* versus 0.6 mm for *A. tepida* (Figure 3). On the line with the BC model, we interpret this observation as a wider chemotaxis range for *H. germanica*, maybe related to its pseudopod length, which feels necessary to move back to the surface once buried deeper than *A. tepida*. However, this difference could also come from *H. germanica* kleptoplasty, as proposed by Cesbron et al. (2017) in order to interpret similar observations. In this case, *H. germanica* would be less sensitive to oxygen depletion and tolerates being buried deeper before moving back to the surface.

449 5.3.2 **Deep and shallow infaunal comparison**

450 Based on the BC model, specimens' behaviour depends on their position compared to the 451 shallow minimum, with upper specimens being more active and especially able to reproduce 452 and growth while lower specimens are probably in dormancy stasis. When taking into account 453 density results from the two sampling stations in September and in May, a linear positive 454 correlation between these two populations appears (Figure 5B). It is a significant result as the line crosses the origin with a high R^2 value (0.95) for A. tepida. The relation for H. germanica 455 456 is less convincing since the lower observed population induces higher uncertainties. However, 457 for A. tepida, this relevance is highlighted by other biological parameters measured in this 458 study such as Chl a (Figure 5B) that does not follow any linear relation.

459 The model of microhabitat equilibrium (§3.6 and Figure 2) details explicitly, despite 460 evident oversimplifications, how the shallow and deep infaunal population interaction can be 461 describe by their ratio, so called α , which depends of the intrinsic species dynamics 462 (propagule spawning / reproduction and mortality rates), biomixing rate and the delay since 463 the last change of available resources. After a certain point, function of biomixing and 464 reproduction rate, the model shows that the time does not influence the α value anymore, slow 465 growing species being in exponential growth while faster growing species, such as 466 opportunists, having already saturated the environment capabilities. The constancy of the α 467 value found for A. *tepida* ($\alpha = 1.1$, Figure 5B) despite a 5 fold change of density indicates that 468 during each campaign, depth repartition of foraminifera population has reached an 469 equilibrium since the last change of available resources, more likely a saturation equilibrium. 470 Hence biomixing is fast relatively to foraminifera resource changes. This is not the case for 471 Chl a probably because its main resource (available light) changes too fast compared to 472 biomixing events while foraminifera populations average short term variations of food 473 availability. Moreover, equation (11) indicates that an α about 1 indicates high biomixing rate 474 compared to mortality in anoxia, a first step to estimate biomixing rate using foraminifera 475 vertical distribution. Taken together, analysis of vertical distribution confirms the steady state 476 reached between surface resources and A. tepida density at surface and at depth and the 477 importance of Chl a concentration at Les Brillantes. While the bioturbation intensity was 478 expected to be a supplementary depth cause due to mortality increase at depth, this effect was 479 not found to be significant since the highest density of foraminifera (including *H. germanica*) 480 matches with the highest density of macrofauna (figure 5A).

481

6 Conclusion

482 Because of the several extreme conditions characterizing intertidal mudflat habitats, 483 amongst which we can cite the risk of burial in anoxic sediments and the large daily salinity 484 variation, the two species observed in "Les Brillantes" mudflat developed contrasting skills. 485 H. germanica suffers from freshwater conditions during river flood periods but seems to get 486 longer-range chemotaxis to face anoxia while A. tepida appears to be much less sensitive to 487 freshwater inputs and favours dormancy as a strategy to overcome burial into anoxic depths. 488 These differences could come from their different feeding strategies, H. germanica having a 489 more specific diet while A. tepida, feeding from different sources, is emancipated from 490 primary producer dependency and shows an opportunistic behaviour.

491 Geostatistic models confirm the effectiveness of average salinity to describe qualitatively 492 the habitats distribution. However, they indicate that the density of foraminifera in these 493 habitats are controlled by other parameters, such as Chl a, that varies over distance from 1 m 494 and 1 km. These distances are often underrepresented in publications looking at foraminiferal 495 heterogeneity and require supplementary investigations to state about their importance. Thus, 496 we recommend to future models to fit geostatistical and deterministic approaches, for 497 example, by associating a particular preponderant mechanism to each scale characterized by 498 high heterogeneity.

499

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505

507

5088 Glossary

509 SYVEL (Surveillance system of the Loire estuary) network maintains 6 high frequency 510 stations between Nantes and Paimboeuf for physicochemical parameters of subsurface waters 511 (temperature, salinity, dissolved oxygen concentration, and turbidity). Founded by the region 512 Pays de la Loire.

513 SHOM is a french national military service for marine and coastal geographic information.

514 Banque HYDRO is a national public gathering hub of river flows, alimented mainly by 515 numerous national services.

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9 Bibliography

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Alve, E., 1999. Colonization of new habitats by benthic foraminifera: a review. Earth-Sci.
Rev. 46, 167–185. https://doi.org/10.1016/S0012-8252(99)00016-1

Alve, E., Bernhard, J.M., 1995. Vertical migratory response of benthic foraminifera to
controlled oxygen concentrations in an experimental mesocosm. Mar Ecol Prog Ser 116, 137–
151. http://dx.doi.org/10.3354/meps116137

Alve, E., Murray, J.W., 2001. Temporal Variability in Vertical Distributions of Live
(stained) Intertidal Foraminifera, Southern England. J. Foraminifer. Res. 31, 12–24.
https://doi.org/10.2113/0310012

Alve, E., Murray, J. w, 1994. Ecology and taphonomy of benthic foraminifera in a
temperate mesotidal inlet. J. Foraminifer. Res. 24, 18–27.
https://doi.org/10.2113/gsjfr.24.1.18

529 Benyoucef, I., 2014. Télédétection visible proche-infrarouge de la distribution spatio530 temporelle du microphytobenthos estuarien (Ph.D. thesis). Université de Nantes.

Berg, P., Risgaard-Petersen, N., Rysgaard, S., 1998. Interpretation of measured
concentration profiles in sediment pore water. Limnol. Oceanogr. 43, 1500–1510.
https://doi.org/10.4319/lo.1998.43.7.1500

Berkeley, A., Perry, C.T., Smithers, S.G., Horton, B.P., 2008. The spatial and vertical
distribution of living (stained) benthic foraminifera from a tropical, intertidal environment,
north Queensland, Australia. Mar. Micropaleontol. 69, 240–261.
https://doi.org/10.1016/j.marmicro.2008.08.002

Berkeley, A., Perry, C.T., Smithers, S.G., Horton, B.P., Taylor, K.G., 2007. A review of the
ecological and taphonomic controls on foraminiferal assemblage development in intertidal
environments. Earth-Sci. Rev. 83, 205–230. https://doi.org/10.1016/j.earscirev.2007.04.003

541 Bianchi, T.S., Findlay, S., 1991. Decomposition of Hudson Estuary Macrophytes:
542 Photosynthetic Pigment Transformations and Decay Constants. Estuaries 14, 65.
543 https://doi.org/10.2307/1351983

Bird, C., Schweizer, M., Roberts, A., Austin, W.E.N., Knudsen, K.L., Evans, K.M.,
Filipsson, H.L., Sayer, M.D.J., Geslin, E., Darling, K.F., 2020. The genetic diversity,
morphology, biogeography, and taxonomic designations of Ammonia (Foraminifera) in the
Northeast Atlantic. Mar. Micropaleontol. 155, 101726.
https://doi.org/10.1016/j.marmicro.2019.02.001

549 Bivand, R.S., Wong, D.W.S., 2018. Comparing implementations of global and local 550 indicators of spatial association. TEST 27, 716–748. https://doi.org/10.1007/s11749-018-551 0599-x Bradshaw, J.S., 1961. Laboratory experiments on the ecology of foraminifera. Cushman
Found Foram Res Contr 12, 87–106.

Buzas, M.A., Culver, S.J., Jorissen, F.J., 1993. A statistical evaluation of the microhabitats
of living (stained) infaunal benthic foraminifera. Mar. Micropaleontol. 20, 311–320.
https://doi.org/10.1016/0377-8398(93)90040-5

- Cartaxana, P., Jesus, B., Brotas, V., 2003. Pheophorbide and pheophytin a-like pigments as
 useful markers for intertidal microphytobenthos grazing by Hydrobia ulvae. Estuar. Coast.
 Shelf Sci. 58, 293–297. https://doi.org/10.1016/S0272-7714(03)00081-7
- 560 Cesbron, F., Geslin, E., Kieffre, C.L., Jauffrais, T., Nardelli, M.P., Langlet, D., Mabilleau, G., Jorissen, F.J., Jézéquel, D., Metzger, E., 2017. Sequestered Chloroplasts in the Benthic 561 562 Foraminifer Haynesina Germanica: Cellular Organization, Oxygen Fluxes and Potential 563 Ecological Implications. J. Foraminifer. Res. 47, 268-278. 564 https://doi.org/10.2113/gsjfr.47.3.268
- 565 Chronopoulou, P.-M., Salonen, I., Bird, C., Reichart, G.-J., Koho, K.A., 2019.
 566 Metabarcoding Insights Into the Trophic Behavior and Identity of Intertidal Benthic
 567 Foraminifera. Front. Microbiol. 10. https://doi.org/10.3389/fmicb.2019.01169

568 Coynel, A., Gorse, L., Curti, C., Schafer, J., Grosbois, C., Morelli, G., Ducassou, E., Blanc, 569 G., Maillet, G.M., Mojtahid, M., 2016. Spatial distribution of trace elements in the surface 570 sediments of a major European estuary (Loire Estuary, France): Source identification and 571 evaluation of anthropogenic contribution. J. Sea 118, 77–91. Res. 572 https://doi.org/10.1016/j.seares.2016.08.005

573 De Stigter, H., Jorissen, F., Van der Zwaan, G., 1998. Bathymetric distribution and 574 microhabitat partitioning of live (Rose Bengal stained) benthic foraminifera along a shelf to 575 bathyal transect in the southern Adriatic Sea. J. Foraminifer. Res. 28, 40–65.

de Stigter, H.C., van der Zwaan, G.J., Langone, L., 1999. Differential rates of benthic
foraminiferal test production in surface and subsurface sediment habitats in the southern
Adriatic Sea. Palaeogeogr. Palaeoclimatol. Palaeoecol. 149, 67–88.
https://doi.org/10.1016/S0031-0182(98)00193-X

Debenay, J.-P., Guillou, J.-J., 2002. Ecological transitions indicated by foraminiferal
assemblages in paralic environments. Estuaries 25, 1107–1120.
https://doi.org/10.1007/BF02692208

583 Debenay, J.-P., Guillou, J.-J., Redois, F., Geslin, E., 2000. Distribution Trends of 584 Foraminiferal Assemblages in Paralic Environments, in: Martin, R.E. (Ed.), Environmental 585 Micropaleontology. Springer US, pp. 39–67.

586 Debenay, J.-P., Guillou, J.-J., Tsakiridis, E., de Casamajor, M.-N., 2001. Bioindicateurs 587 d'impact dans les ports et les estuaires: les foraminifères. Rev. Fr. Génie Civ. 5, 1105–1122.

Dupuy, C., Rossignol, L., Geslin, E., Pascal, P.-Y., 2010. Predation of Mudflat MeioMacrofaunal Metazoans by a Calcareous Foraminifer, Ammonia tepida (cushman, 1926). J.
Foraminifer. Res. 40, 305–312. https://doi.org/10.2113/gsjfr.40.4.305

Fortin, M.-J., Dale, M.R.T., 2005. Spatial analysis a guide for ecologists. Cambridge
University Press, Cambridge, N.Y.

Geslin, E., Heinz, P., Jorissen, F., Hemleben, Ch., 2004. Migratory responses of deep-sea
benthic foraminifera to variable oxygen conditions: laboratory investigations. Mar.
Micropaleontol. 53, 227–243. https://doi.org/10.1016/j.marmicro.2004.05.010

596 Goldstein, S.T., Watkins, G.T., Kuhn, R.M., 1995. Microhabitats of salt marsh foraminifera: St. Catherines Island, Georgia, USA. Mar. Micropaleontol., Selected papers 597 598 from the Fifth International Symposium Foraminifera 17-29. of 26, https://doi.org/10.1016/0377-8398(95)00006-2 599

Hohenegger, J., Piller, W.E., Baal, C., 1993. Horizontal and vertical spatial
microdistribution of foraminifers in the shallow subtidal Gulf of Trieste, northern Adriatic
Sea. J. Foraminifer. Res. 23, 79–101. https://doi.org/10.2113/gsjfr.23.2.79

Jauffrais, T., LeKieffre, C., Koho, K.A., Tsuchiya, M., Schweizer, M., Bernhard, J.M.,
Meibom, A., Geslin, E., 2018. Ultrastructure and distribution of kleptoplasts in benthic
foraminifera from shallow-water (photic) habitats. Mar. Micropaleontol., Benthic
Foraminiferal Ultrastructure Studies 138, 46–62.
https://doi.org/10.1016/j.marmicro.2017.10.003

Jorissen, F.J., de Stigter, H.C., Widmark, J.G.V., 1995. A conceptual model explaining
benthic foraminiferal microhabitats. Mar. Micropaleontol. 26, 3–15.
https://doi.org/10.1016/0377-8398(95)00047-X

Kitazato, H., Shirayama, Y., Nakatsuka, T., Fujiwara, S., Shimanaga, M., Kato, Y., Okada,
Y., Kanda, J., Yamaoka, A., Masuzawa, T., Suzuki, K., 2000. Seasonal phytodetritus
deposition and responses of bathyal benthic foraminiferal populations in Sagami Bay, Japan:
preliminary results from ^aProject Sagami 1996±1999^o. Mar. Micropaleontol. 15.

Langlet, D., Baal, C., Geslin, E., Metzger, E., Zuschin, M., Riedel, B., Risgaard-Petersen,
N., Stachowitsch, M., Jorissen, F.J., 2014. Foraminiferal species responses to in situ,
experimentally induced anoxia in the Adriatic Sea. Biogeosciences 11, 1775–1797.
https://doi.org/10.5194/bg-11-1775-2014

Legendre, P., Fortin, M.-J., 1989. Spatial pattern and ecological analysis. Vegetation 80,
107–138.

621 LeKieffre, C., Jauffrais, T., Geslin, E., Jesus, B., Bernhard, J.M., Giovani, M.-E., Meibom,

622 A., 2018. Inorganic carbon and nitrogen assimilation in cellular compartments of a benthic

623 kleptoplastic foraminifer. Sci. Rep. 8, 1–12. https://doi.org/10.1038/s41598-018-28455-1

LeKieffre, C., Spangenberg, J.E., Mabilleau, G., Escrig, S., Meibom, A., Geslin, E., 2017.
Surviving anoxia in marine sediments: The metabolic response of ubiquitous benthic
foraminifera (Ammonia tepida). PloS One 12, e0177604.

Levin, S.A., 1976. Population Dynamic Models in Heterogeneous Environments. Annu.
Rev. Ecol. Syst. 287–310.

Loubere, P., 1989. Bioturbation and sedimentation rate control of benthic microfossil taxon abundances in surface sediments: A theoretical approach to the analysis of species microhabitats. Mar. Micropaleontol. 14, 317–325. https://doi.org/10.1016/0377-8398(89)90016-9

Méléder, V., Barillé, L., Rincé, Y., Morançais, M., Rosa, P., Gaudin, P., 2005. Spatiotemporal changes in microphytobenthos structure analysed by pigment composition in a
macrotidal flat (Bourgneuf Bay, France). Mar. Ecol. Prog. Ser. 297, 83–99.
https://doi.org/10.3354/meps297083

Moellering, H., Tobler, W., 1972. Geographical Variances. Geogr. Anal. 4, 34–50.
https://doi.org/10.1111/j.1538-4632.1972.tb00455.x

Mojtahid, M., Geslin, E., Coynel, A., Gorse, L., Vella, C., Davranche, A., Zozzolo, L.,
Blanchet, L., Bénéteau, E., Maillet, G., 2016. Spatial distribution of living (Rose Bengal

- stained) benthic foraminifera in the Loire estuary (western France). J. Sea Res.
 https://doi.org/10.1016/j.seares.2016.02.003
- 643 Murray, J.W., 2006. Ecology and applications of benthic foraminifera. Cambridge
 644 University Press, Cambridge.
- Pillet, L., de Vargas, C., Pawlowski, J., 2011. Molecular Identification of Sequestered
 Diatom Chloroplasts and Kleptoplastidy in Foraminifera. Protist 162, 394–404.
 https://doi.org/10.1016/j.protis.2010.10.001
- 648 Richirt, J., Schweizer, M., Bouchet, V.M.P., Mouret, A., Quinchard, S., Jorissen, F.J., 2019.
- 649 Morphological Distinction of Three Ammonia Phylotypes Occurring Along European Coasts.
- 650 J. Foraminifer. Res. 49, 76–93. https://doi.org/10.2113/gsjfr.49.1.76
- Saffert, H., Thomas, E., 1998. Living foraminifera and total populations in salt marsh peat
 cores: Kelsey Marsh (Clinton, CT) and the Great Marshes (Barnstable, MA). Mar.
 Micropaleontol. 33, 175–202. https://doi.org/10.1016/S0377-8398(97)00035-2
- 654 Schönfeld, J., Alve, E., Geslin, E., Jorissen, F., Korsun, S., Spezzaferri, S., 2012. The
- FOBIMO (FOraminiferal BIo-MOnitoring) initiative—Towards a standardised protocol for
 soft-bottom benthic foraminiferal monitoring studies. Mar. Micropaleontol. 94–95, 1–13.
- 657 https://doi.org/10.1016/j.marmicro.2012.06.001
- 658 Steineck, P.L., Bergstein, J., 1979. Foraminifera from Hommocks salt-marsh, Larchmont
- 659 Harbor, New York. J. Foraminifer. Res. 9, 147–158.
- Talley, T.S., 2007. Which spatial heterogeneity framework? Consequences for conclusions
 about patchy population distributions. Ecology 88, 1476–1489.

Thibault de Chanvalon, A., Metzger, E., Mouret, A., Cesbron, F., Knoery, J., Rozuel, E.,
Launeau, P., Nardelli, M.P., Jorissen, F.J., Geslin, E., 2015. Two-dimensional distribution of
living benthic foraminifera in anoxic sediment layers of an estuarine mudflat (Loire estuary,
France). Biogeosciences 12, 6219–6234. https://doi.org/10.5194/bg-12-6219-2015

- 666 Thibault de Chanvalon, A., Mouret, A., Knoery, J., Geslin, E., Péron, O., Metzger, E.,
- 667 2016. Manganese, iron and phosphorus cycling in an estuarine mudflat, Loire, France. J. Sea
- 668 Res. 118, 92–102. https://doi.org/10.1016/j.seares.2016.10.004
- Weinmann, A.E., Goldstein, S.T., 2017. Landward-directed Dispersal of Benthic
 Foraminiferal Propagules At Two Shallow-water Sites in the Doboy Sound Area (Georgia,
- 671 U.S.A.). J. Foraminifer. Res. 47, 325–336. https://doi.org/10.2113/gsjfr.47.4.325
- Wu, J., Jelinski, D.E., Luck, M., Tueller, P.T., 2000. Multiscale Analysis of Landscape
 Heterogeneity: Scale Variance and Pattern Metrics. Geogr. Inf. Sci. 6, 6–19.
 https://doi.org/10.1080/10824000009480529

10 Tables

10.1 Table 1: Scale Variance Analysis

Scale level (h)	Characteristic scale distances (10 ^{h-3} m)	Family size* (n ^h)	Available sampling / exhaustive sampling	Relative scale variance $(RV^{h \rightarrow h+1})$	$\frac{1}{N^h} \sum_{i=1}^{N^h} (x_i^h)^2$	Mean of the scale variance $(\overline{V^{h \to h+1}})$	$\frac{\overline{V^{h \to h+1}}}{VAR} \ge 100$	$\frac{\overline{V^{h \to h+1}}}{VAR} \ge 100^{\dagger}$
8	100 km		1/1		5.904			
7	10 km	7	7/7	2.318		14.38	0.9	0.6
6	1 km	7	7/49	3.513		74.19	4.6	2.5
5	100 m	7	0/343				26.6	28.2
4	10 m	7	0/2 401				26.6	28.2
3	1 m	7	2/16 807				26.6	28.2
2	1 dm	5	106/84 035	0.083		114.85	7.1	4.6
1	1 cm	7	7/588 245	0.083		125.54	7.7	7.7

678 *Number of samples of order h required to constitute a sample of order h+1, † calculated after exclusion of S1 from the dataset

680 **10.2 Table 2 : Geochemical parameters**

		Sept 2012	May 2013
	Average flow (m ³ s ⁻¹)	150	1200
	Salinity ²	22 ± 5	8 ± 7
	Water Temperature (°C) ²	17 ± 0.5	13 ± 1
	Tidal coefficient ³	50 ± 10	80 ± 20
	Grain size nomination ⁴ (0-5cm)	silty clay	silty clay
C 1	Oxygen Penetration Depth (mm) (± SD)	2.3 (±0.4, n=12)	1.9 (±0.2, n=24)
51	Dissolved phosphorus in 0-1 cm ⁵ (μ mol L ⁻¹)	15.4	14.2
	Total Organic Carbone in 0-1 cm ⁵ (%)	2.3	2.8
S2	Grain size nomination ⁴ (0-5cm)	silty clay	silty clay
	Oxygen Penetration Depth (mm) (± SD)	4.7 (±0.7, n=10)*	1.4 (±0.2, n=4)
	Dissolved phosphorus in 0-1 cm ⁵ (μ mol L ⁻¹)	2.7	4.8
	Total Organic Carbone in 0-1 cm ⁵ (%)	2.1	2.8

- 681 *Most of the profiles show bioturbation ¹measured at Mont-Jean sur Loire (banque
- 682 HYDRO) ²GIP Loire ³SHOM, ⁴from Benyoucef (2014) ⁵from Thibault de Chanvalon et al.
- 683 (2016)

684

10.3 Table 3: Biological parameters

			Sept 2012			May 2013			
			Mean	SD	n	Mean	SD	n	
	Microphytobenthos	Chl a from 0 to 1 cm (mg.m ⁻²)	384	121	3	295	185	3	
S1	$(mg m^{-2})$	Lability Index from 0 to 1 cm	0.969	0.007	3	0.973	0.007	3	
	Manafaunal danaity	Hediste diversicolor	635		1	631		1	
	(ind m^{-2})	Heteromastus filiformis	16		1	25		1	
	(ma m)	Scrobicularia plana	159		1	70		1	
	D	A. tepida from 0 to 1 cm	245	3	2	123	21	2	
	Foraminiferal	A. tepida from 1 to 5 cm	53	7	2	31	8	2	
	(ind $/ 10 \text{ cm}^{-3}$)	<i>H. germanica</i> from 0 to 1 cm	36	2	2	2	1	2	
	(ind. / 10 cm)	<i>H. germanica</i> from 1 to 5 cm	6		1	4	1	2	
	Foraminifera	A. tepida	1.54	0.17	2	1.74	0.03	2	
	$ALD_5(cm)$	H. germanica	1.66		1	3.45	0.06	2	
	Microphytobenthos	Chl a from 0 to 1 cm (mg.m ⁻²)	166	25	3	198	22	3	
	$(mg m^{-2})$	Lability Index from 0 to 1 cm	0.945	0.008	3	0.978	0.011	3	
	Manafaunal danaitu	Hediste diversicolor	25		1	51		1	
	(ind m^{-2})	Heteromastus filiformis	83		1	159		1	
	(ma m)	Scrobicularia plana	162		1	108		1	
S 2		A. tepida from 0 to 1 cm	46	12	3	60	12	2	
	Foraminiferal	A. tepida from 1 to 5 cm	11	1	3	10		1	
	(ind / 10 cm^{-3})	<i>H. germanica</i> from 0 to 1 cm	5	1	3	5	1	2	
		H. germanica from 1 to 5 cm	4	1	3	5		1	
	Foraminifera	A. tepida	1.80	0.27	3	1.50		1	
	ALD ₅ (cm)	H. germanica	2.76	0.21	3	2.67		1	

10.4 Table 4: Respiration rate calculation

Sampling date	Station	Species	Total nur foraminit oxic zone (ind. 50c	mber of fera in e m ⁻²)	f the	$\begin{array}{c} \text{DOU} \\ (\text{mmolO}_2 \text{ m}^{-2} \text{ d}^{-1}) \end{array}$		⁻¹)	RR by foraminiferal population (mmolO ₂ m ⁻² d ⁻¹)			Foraminiferal contribution to DOU %	
			Mean	SD	n	mean	SD	n	mean	SD	n	mean	SD
September	S 1	Ammonia tepida	1053	77	2	24	10.9	12	0.557	0.152	2	2.3	0.6
2012		H. germanica	108	28	2	24	10.9	12	0.009	0.005	2	0	0
	S2	Ammonia tepida	144	38	3	9.1	4.5	10	0.097	0.025	3	1.1	0.3
		H. germanica	12	4	3	9.1	4.5	10	0.002	0	3	0	0
May	S 1	Ammonia tepida	217	1	2	71	31	7	0.093	0	2	0.1	0
2013		H. germanica	2	1	2	71	31	7	0	0	2	0	0
	S2	Ammonia tepida	198	36	3	56	15	3	0.085	0.015	3	0.2	0
		H. germanica	2	1	3	56	15	3	0	0	3	0	0

11 Figures

11.1 Figure 1: Bathymetry of the Loire estuary with surface density of *A. tepida* from Mojtahid et al. (2016) sampled in September
 2012, Thibault de Chanvalon et al. (2015) sampled in May 2013 and this study (both May 2013 and September 2012). Black lines
 indicate regions used in the scale variance analysis for the scale 7 (A1 to A7) and the scale 6 (B1 to B7). Bottom right insert focus
 on the "Les Brillantes" mudflat (Map produced on R using leaflet package, bathymetry from the SHOM).



694 11.2 Figure 2: A. Deterministic model to explain these relations based on biomixing 695 and chemotaxis forcing.

 v_{prod} , v_{death} , $v_{sh->de}$ and $v_{de->sh}$ correspond to the rate of reproduction, the rate of death, the rate of exchange from shallow to deep microhabitat and the rate of exchange from deep to shallow microhabitat respectively. k_p , k_{d_sh} , k_{d_de} and k_{ech} are the associated parameters while pop_{sh} and pop_{de} are the population of shallow and deep microhabitat. B. Example of a representative result from the model following a 500 fold increase of environment capability , *i. e.* 500 fold pop_{sat} increase.

702

11.3 Figure 3: Vertical distribution in the sediment column of living benthic705 **foraminifera and Chlorophyll a.**

709 11.4 Figure 4: Geostatistical model processes with the dataset from Figure 1 (black 710 dots).

711 Error bars in Figure A corresponds to twice the standard deviation of the distribution 712 obtained when the samples are randomly distributed (number of simulation is 104; star 713 attribution is based on p-value; <=0.01***; <=0.05**; <=0.1*). InFigure B open circle 714 represents the scale variance analysis (SVA) after exclusion of S1 from the dataset (open 715 circle) and using qualitatively transformed dataset (black diamonds, see text for details). In 716 the experimental variogram (Figure C), grey dots correspond to the square of the difference of 717 each possible pairs, plotted against their distance. The black dots correspond to the mean and 718 the error bars to a third of the standard deviation for each scale of distance between samples.

11.5 Figure 5: A. Evolution of biological parameters in Les Brillantes according to
Chl a (data from Table 2). B. Relation between deep infaunal and shallow
infaunal population of *A. tepida* and *H. germanica* in Les Brillantes.

726 **12 Appendices**

12.1 Appendix 1: Demonstration that the variance of a dataset hierarchized over k scale is equal to the sum of the mean scale variance of each scale.

Based on the definition of the variance and on the formalism previously described, samples' value are the x_i^1 (they belong to the scale 1) and the global mean is x^k , we have :

$$VAR = V^{1 \to k} = \frac{1}{N^1} \sum_{i=1}^{N^1} (x_i^1 - x^k)^2$$
1

$$V^{1 \to k} = \frac{1}{N^1} \left(\sum_{i=1}^{N^1} (x_i^1)^2 - N^1 (x^k)^2 \right)$$

The samples can be gathered in n^{k-1} group whose extension belong to the scale k-1. There is N¹ / n^{k-1} samples per group and the x_j^{k-1} are the means of each groups and the x_i^1 are now written as $x_{i,j}^1$, with the subscript j indicates to which groups the sample belong.

$$V^{1 \to k} = \frac{1}{N^{1}} \left(\sum_{i=1}^{N^{1}} (x_{i,j}^{1})^{2} - \sum_{j=1}^{n^{k-1}} \frac{N^{1}}{n^{k-1}} (x_{j}^{k-1})^{2} + \sum_{j=1}^{n^{k-1}} \frac{N^{1}}{n^{k-1}} (x_{j}^{k-1})^{2} - N^{1} (x^{k})^{2} \right)$$

$$V^{1 \to k} = \frac{1}{N^{1}} \sum_{j=1}^{n^{k-1}} \left(\sum_{i=1}^{n^{k-1}} (x_{i,j}^{1})^{2} - \frac{N^{1}}{n^{k-1}} (x_{j}^{k-1})^{2} \right) + \frac{1}{n^{k-1}} \left(\sum_{j=1}^{n^{k-1}} (x_{j}^{k-1})^{2} - n^{k-1} (x^{k})^{2} \right)$$

$$V^{1 \to k} = \frac{1}{N^{1}} \sum_{j=1}^{n^{k-1}} \frac{N^{1}}{n^{k-1}} V_{j}^{1 \to k-1} + V^{k-1 \to k}$$

$$V^{1 \to k} = \frac{1}{n^{k-1}} \sum_{j=1}^{n^{k-1}} V_{j}^{1 \to k-1} + V^{k-1 \to k}$$
A31

The property expressed by A3 is true also for any j and p, $V_i^{1 \rightarrow k-p}$, especially

$$V_j^{1 \to k-1} = \frac{1}{n^{k-2}} \sum_{l=1}^{n^{k-2}} V_{j,l}^{1 \to k-2} + V_j^{k-2 \to k-1}$$
A41

735 Once you inject A4 into A3 you get,

$$V^{1 \to k} = \frac{1}{n^{k-1}} \frac{1}{n^{k-2}} \sum_{j=1}^{n^{k-1}} \sum_{l=1}^{n^{k-2}} V^{1 \to k-2}_{j,l} + \frac{1}{n^{k-1}} \sum_{j=1}^{n^{k-1}} V^{k-2 \to k-1}_{j} + V^{k-1 \to k}$$
A54

Then using the property expressed by A3 for p = 2 you get

$$V^{1 \to k} = \frac{1}{n^{k-1}} \frac{1}{n^{k-2}} \frac{1}{n^{k-2}} \sum_{j=1}^{n^{k-1}} \sum_{l=1}^{n^{k-2}} \sum_{m=1}^{n^{k-3}} V_{j,l,m}^{1 \to k-3} + \frac{1}{n^{k-1}} \frac{1}{n^{k-2}} \sum_{j=1}^{n^{k-1}} \sum_{l=1}^{n^{k-2}} V_{j,l}^{k-3 \to k-2} + \frac{1}{n^{k-1}} \sum_{j=1}^{n^{k-1}} V_{j}^{k-2 \to k-1} + V^{k-1 \to k}$$
Determine this for all a setting to 2

737 Repeating this for all p until p=k-2

$$V^{1 \to k} = \frac{1}{N^2} \sum_{j=1}^{n^{k-1}} V_j^{1 \to 2} + \dots + \frac{1}{N^{k-2}} \sum_{j=1}^{N^{k-2}} V_j^{k-3 \to k-2} + \frac{1}{N^{k-1}} \sum_{j=1}^{N^{k-1}} V_j^{k-2 \to k-1} + V^{k-1 \to k}$$

$$VAR = V^{1 \to k} = \sum_{h=1}^{k-1} \overline{V^{h \to h+1}}$$
A71

738

739 **12.2** Appendix 2: Demonstration of the expression of the mean of the scale variance

740 **as a function of the relative scale variance**

741 Based on the relative scale variance definition, for any i :

$$RV^{h \to h+1} = \frac{1}{(x^{h+1})^2} \left(\frac{1}{n^h} \sum_{i=1}^{n^h} (x_i^h)^2 - (x^{h+1})^2\right)$$
B1

742 Once reorganized, we get:

$$\sum_{i=1}^{n^{h}} (x_{i}^{h})^{2} = (1 + RV^{h \to h+1})n^{h}(x^{h+1})^{2}$$

743 By sum over all the j group from the scale h+1,

$$\sum_{j=1}^{N^{h+1}} \sum_{i=1}^{n^{h}} (x_{i}^{h})^{2} = \sum_{j=1}^{N^{h+1}} (1 + RV^{h \to h+1})n^{h} (x^{h+1})^{2}$$
$$\sum_{i=1}^{N^{h}} (x_{i}^{h})^{2} = (1 + RV^{h \to h+1})n^{h} \sum_{j=1}^{N^{h+1}} (x_{j}^{h+1})^{2}$$
B2

- The equation B2 express the sum of the square at the scale h, as a function of the sum of the
- square at the scale h+1. Injecting p times the equation B2 into itself, we get:

$$\sum_{i=1}^{N^{h}} (x_{i}^{h})^{2} = \prod_{i=h}^{h+p-1} (1 + RV^{i \to i+1}) n^{i} \sum_{j=1}^{N^{h+p}} (x_{j}^{h+p})^{2}$$
B3

746 On another hand, taking the definition of the mean scale variance, we have:

$$\overline{V^{h \to h+1}} = \frac{1}{N^{h+1}} \sum_{i=1}^{N^{h+1}} (x_i^{h+1})^2 R V^{h \to h+1}$$
B4

747 Injecting B3 into B4, with p=k-1-h,

$$\overline{V^{h \to h+1}} = \frac{RV^{h \to h+1}}{N^{h+1}} \prod_{i=h+1}^{k-1} (1 + RV^{i \to i+1}) n^i \sum_{j=1}^{N^k} (x_j^k)^2$$
$$\overline{V^{h \to h+1}} = \frac{RV^{h \to h+1}}{N^k} \prod_{i=h+1}^{k-1} (1 + RV^{i \to i+1}) \sum_{j=1}^{N^k} (x_j^k)^2$$
B5

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