The benthic toxic dinoflagellate *Ostreopsis* cf. *ovata* in the NW Mediterranean Sea: Relationship between sea surface temperature and bloom phenology

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Abstract :

Blooms of the toxic benthic dinoflagellate Ostreopsis cf. ovata can induce ecological and human health issues in certain temperate areas. In order to prevent these negative effects, long-term monitoring studies of O. cf. ovata blooms have been conducted in several impacted areas to have a comprehensive understanding of bloom dynamics and efficient tools for risk management.

O. cf. ovata blooms were monitored every summer (from mid-June to the end of August) on five identified sites in Larvotto beach (Monaco, NW Mediterranean Sea), between 2007 and 2019. This time-series represents one of the largest time-series in the world describing blooms of this species.

Bloom phenological features (timing, duration, maximum cell abundance and growth rate), were found to be highly variable throughout the studied period, and were analyzed as a function of different hydroclimatic parameters, including sea surface temperature (SST). The highest net growth rates were related to temperatures ranging between 21°C and 25°C, and did not coincide with maximal temperature records (27.5°C). Such results suggest that, although global warming possibly influences the expansion of O. cf. ovata from tropical to temperate waters, the definite impact of temperature on bloom dynamics might be more complex than a simple facilitation factor for algal growth, at least in NW Mediterranean waters.

Furthermore, monthly SST anomalies calculated over this 13-year survey showed a strong positive correlation between spring SST positive anomalies and the bloom starting date, indicating that blooms occurred earlier in the season when spring SSTs were warmer than usual.

Overall results provide tools to modelers and managers who are facing crucial challenges to predict the distribution and phenology of O. cf. ovata blooms in European coastal waters, moreover in a context of global warming.

Highlights

► Long-term monitoring showed variability in bloom dynamics of *O*. cf. *ovata* in NW Mediterranean Sea.

A strong positive correlation was observed between spring sea surface temperatures and bloom timing.
 Highest growth rates did not coincide with highest sea surface temperatures. ► Sea surface temperature, nutrient loads, wind regimes and rainfall were the main factors affecting *O*. cf. *ovata* bloom phenology.

Keywords : Benthic HABs, Long-term monitoring, Environmental factors, SST anomalies, Temperature niche, Mediterranean Sea

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51	Over the past decades, worldwide reports of Harmful Algal Blooms (HABs) have increased and
52	suggest that these events are more intense and frequent in some areas, and may also be
53	extending to new regions (Hallegraeff, 2010; Wells et al., 2020; Hallegraeff et al., 2021). The
54	causes behind this trend are not clear yet; they may be the result of a combination of changing
55	climate factors (Glibert et al., 2014; Kibler et al., 2015; Moore et al., 2008; Tester et al., 2020;
56	Wells & Karlson, 2018), anthropogenic forcing (Anderson et al., 2002; Glibert et al., 2005;
57	Heisler et al., 2008; Fu et al., 2012; Davidson et al., 2014) and the implementation of worldwide
58	monitoring programs which have improved the capacity of HAB detection (Van Dolah, 2000;
59	Anderson et al., 2012; Hallegraeff et al., 2021).
60	Amongst the organisms causing harmful events, the toxic species belonging to the
61	marine dinoflagellate genus Ostreopsis Schmidt (1901) are found in a large variety of
62	environments, from tropical and sub-tropical environments to temperate areas (Faust et al.,
63	1996; Rhodes, 2011). Ostreopsis spp. was reported in the Mediterranean Sea for the first time
64	in 1972 in Villefranche-sur-Mer, France (Taylor, 1979), and regularly detected thereafter along
65	the Mediterranean coastlines (e.g. Mangialajo et al., 2011; Accoroni and Totti, 2016), with three
66	identified species so far, namely, Ostreopsis cf. ovata (Fukuyo), Ostreopsis cf. siamensis
67	(Schmidt) and Ostreopsis fattorussoi (Accoroni, Romagnoli & Totti). The available data
68	suggest that O. cf. ovata is the most common species in Mediterranean waters. In this area, the
69	first bloom of O. cf. ovata was observed off the Tuscany coast in 1998 (Sansoni et al., 2003).
70	Since then, recurrent blooms have been documented in many other Mediterranean coastal areas

(Blanfuné et al., 2015; Accoroni and Totti, 2016). Moreover, this species was identified to cause 71 noxious episodes reported along the Northern (Brescianini et al., 2006; Barroso García et al., 72 2008; Vila et al., 2008, 2016; Pfannkuchen et al., 2012) and Southern (Illoul et al., 2012) coasts 73 of the Mediterranean Sea. Indeed, O. cf. ovata blooms were linked to human health symptoms 74 including skin irritations after direct contact with the seawater (Tichadou et al., 2010; Tubaro 75 et al., 2011; Pfannkuchen et al., 2012) or respiratory problems from inhalation of aerosols 76 (Durando et al., 2007; Vila et al., 2016). Ostreopsis blooms also have an ecological impact 77 involving mass mortalities of certain invertebrates (Sansoni et al., 2003; Totti et al., 2010) 78 caused either by oxygen limitation due to the excess of microalgal biomass or specific ecotoxic 79 compounds. In Mediterranean Sea, Ostreopsis species produce different toxins that can be 80 transferred in some fish and invertebrates (Aligizaki et al., 2008; Biré et al., 2013, 2015; 81 Brissard et al., 2014) and pose a potential risk of food poisoning in humans too. The Ostreopsis-82 83 produced toxins are palytoxin analogues (Ciminiello et al., 2006, 2012, 2014) and have been related, although not proven yet, to rare and fatal intoxications in the tropics (Tubaro et al., 84 85 2011). Luckily, such problems have not been reported in the Mediterranean yet, but the recurrence of the Ostreopsis blooms have stimulated intense research and monitoring of this 86 taxon in order to prevent human and environmental health risks. 87

88 In the NW Mediterranean Sea, O. cf. ovata blooms are highly seasonal and occur during the summer to autumn season (Vila et al., 2001; Mangialajo et al., 2008, 2011; Totti et al., 89 2010). The existence of such a temporal window suggests that a combination of specific 90 environmental conditions during this period may favor the development of the blooms. Several 91 studies have shown the importance of hydrodynamics on the development of O. cf. ovata 92 blooms (Barone, 2007; Shears and Ross, 2009; Totti et al., 2010). Slight to moderate waves 93 enable cells to settle (Vila et al., 2001), while rough seas and currents would rather favor cell 94 dispersion, which reduces the cell population growth. Relatively low wind intensities (2-4 m s⁻ 95

¹) were registered concurrently to some blooms in the NW Mediterranean (Vila et al., 2016). 96 Concerning nutrient availability, its role on the dynamics of Ostreopsis blooms is still a matter 97 of debate. Indeed, the ability of O. cf. ovata to exploit the full spectrum of mixotrophy has still 98 not been determined, given the large variety of potential sources of inorganic and organic matter 99 which could contribute to Ostreopsis growth (Accoroni et al., 2017; Jauzein et al., 2017; Lee 100 and Park, 2018). For instance, no direct relationship was found between inorganic nutrient 101 concentrations and cell abundances during bloom periods (Vila et al., 2001; Cohu et al., 2011; 102 103 Pistocchi et al., 2011; Accoroni et al., 2012), although variations of nitrate to phosphate (N:P) ratios may be involved in the onset of the bloom (Accoroni et al., 2015a). Regarding salinity, 104 laboratory scale studies showed that O. cf. ovata grew better under high salinity conditions 105 (Pezzolesi et al., 2012), but others suggested that salinity preferences was strain-specific 106 (Tawong et al., 2015). Field observations indicated that O. cf. ovata blooms were associated 107 108 with *in situ* salinity values between 37 and 38 in the NW Mediterranean Sea (Vila et al., 2001; Mangialajo et al., 2008) and with a wider range (31 - 39) in the Northern Adriatic Sea (Accoroni 109 110 et al., 2015a) or in the South-Eastern Mediterranean Sea (Abdennadher et al., 2017). On the other hand, low salinities associated to inflowing freshwater was shown to hinder blooming 111 events (Blanfuné et al., 2015; Carnicer et al., 2015). Sea depth and substratum also appear to 112 be a specific factor when studying the occurrence of blooms, as Ostreopsis spp. cells mainly 113 proliferate in shallow waters (Totti et al., 2010; Cohu and Lemée, 2012) and high abundances 114 were also found hosted on macroalgae blades (Cohu et al., 2013; Accoroni et al., 2015b). 115 Indeed, the presence of macroalgae with ramified structures, combined with a sheltered shallow 116 habitat, contributes to the settlement and growth of O. cf. ovata (Meroni et al., 2018). Among 117 all these environmental factors, it is still unclear if one in particular, or a combination of factors, 118 could have a leading role in O. cf. ovata bloom seasonality, even though its blooms may depend 119

on the development of macroalgal communities which are also modulated by a seasonal regimein temperate latitudes with clear temperature trends

Indeed, temperature is considered as one of the most important environmental factors 122 determining cell biology, growth and reproduction in general terms, and also follows strong 123 seasonal variations that could make it a driver of bloom seasonality. Concerning Ostreopsis 124 spp., several studies suggested that relatively high temperatures, as those occurring in summer, 125 favored its growth and bloom development (Hallegraeff, 2010; Granéli et al., 2011). In the 126 Atlantic coastline of the Iberian Peninsula, David et al. (2012) showed that Ostreopsis spp. 127 blooms could be triggered when sea surface temperatures (SSTs) exceeded a certain threshold 128 (namely, 19.5°C) over a long enough period (three months). In the NW Mediterranean Sea, the 129 occurrence of O. cf. ovata blooms have been reported under a wide range of SST, from 16.8°C 130 to 30°C (Mangialajo et al., 2011; Accoroni and Totti, 2016). Even if major proliferation are 131 132 usually recorded in mid-summer (July) along the coasts of the Ligurian and Catalan Seas coasts, a second bloom often occurs in autumn in these areas (Mangialajo et al., 2008, 2011; Cohu et 133 134 al., 2011; Vila et al., 2016). In the Northern Adriatic Sea, blooms even appear rather later in autumn, between September and October (Monti et al., 2007; Totti et al., 2010; Ninčević 135 Gladan et al., 2019). Overall, these observations suggest a more underlying and complex role 136 of temperature in the occurrence O. cf. ovata blooms, beyond a simple increasing facilitation 137 with increasing temperature. 138

In a context of global warming, the investigation of the environmental niche of *O*. cf. *ovata* is key to help predict bloom events and to plan efficient monitoring programs. A longterm monitoring dataset can provide a good comprehension of the potential link between temperature and *O*. cf. *ovata* bloom events. Most well established HAB monitoring programs consist in sampling the plankton community. However, *Ostreopsis* has a main benthic phase attached to surfaces (macroalgae, rocks, pebbles, sand, and invertebrates) with a self-produced

mucilage. Internal biological rhythms (not well understood yet) and hydrodynamics detach 145 Ostreopsis cells from the substrate and they are found swimming as part of the plankton or in 146 dense aggregates floating at the sea surface. Benthic cells can be considered as the stock or 147 148 reservoir of the bloom, while planktonic cells and floating aggregates can be more related to the health symptoms (Totti et al., 2010). So far, some studies have shown good correlations 149 between the Ostreopsis populations in the plankton and the benthos (Mangialajo et al., 2011; 150 Giussani et al., 2017) but the uncertainties regarding the factors determining their dynamics 151 recommends the sampling of the two fractions. Such combined plankton and benthos 152 monitoring program was launched in Monaco (Ligurian Sea) in summer 2007, in order to assess 153 154 the sanitary risk impact of O. cf. ovata on local recreational waters. This monitoring program is still ongoing and represents the world's largest time-series study ever recorded for O. cf. 155 ovata bloom events. In addition to the data collected on cells abundance in Monaco waters and 156 157 on macroalgae, hydroclimatic parameters were also registered on site and in a neighboring station of the Bay of Villefranche-sur-Mer (France), on a weekly basis and over decades. The 158 159 overall data set enabled to explore the link between biological and ecological information and the local environmental and physical factors in order to describe O. cf. ovata bloom phenology. 160 Data collected over the first two years of this monitoring program already showed a distinct 161 162 pattern concerning the timing of the blooms and specific hydroclimatic conditions. In particular, it was noticed that Ostreopsis blooms occurred earlier after a warmer spring in 2007, and later 163 following a colder summer in 2008 (Cohu et al., 2011). This observation suggested a new link 164 between hydroclimatic factors and O. cf. ovata bloom phenology which has been analyzed in 165 the present study, considering the data recorded since 2007 over thirteen years. 166

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- 168 **2. Materiel and Methods**
- 169 2.1.Sampling site

Monaco is a sovereign city-state located on the French Riviera coast, in the NW Mediterranean 170 Sea. It is a highly touristic area whose population increases considerably during the summer 171 season. The Larvotto beach $(43^{\circ}44.71' \text{ N} - 7^{\circ}26.04' \text{ E})$ is a small sheltered cove surrounded by 172 residential buildings, shops and restaurants. The beach (Fig. 1) is orientated to the South-East 173 and artificial rocky dykes were built as a barrier to protect from waves and swell. Such 174 structures constitute ideal substrates for the development of ramified structured macroalgae, 175 especially Phaeophyceae and Florideophyceae, such as Halopteris scoparia Sauvageau, 176 Dictyota sp. Lamouroux and Ellisolandia sp. Hind & Saunders, which host high densities of 177 benthic microalgae including O. cf. ovata. Thus, neighbors and tourists can be potentially 178 exposed to the health risks associated to O. cf. ovata blooms. So far, sporadic skin and mucosal 179 irritations have been observed (Tichadou et al., 2010). 180

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182 2.2.Sampling process and cell quantification

O. cf. ovata blooms occurring in the NW Mediterranean Sea are always observed in early 183 summer, when the SST increases and wind conditions are relatively calm. In order to quantify 184 the abundance of epiphytic O. cf. ovata cells, a monitoring program started in 2007 and was 185 then conducted every year, generally from mid-June until the end of August. The global 186 estimation of the benthic cell stock in a given area is a complex task as the cell concentration 187 may depend on the presence of certain types of macroalgae and may also vary according to 188 small scale environmental changes. In order to give a reliable estimation of the cell abundance 189 of this area, five different stations were sampled at each sampling date. The sampling campaign 190 191 was carried out once a week during this period, between 9 am and 11 am, in five stations located on rocky substrate areas within the Larvotto Beach (see Fig. 1). The temporal evolution of the 192 bloom was monitored through the estimation of both benthic cell stock and planktonic cell 193

abundance quantification as described by Cohu et al. (2011). Samples of the most representative 194 195 macroalgae (usually Dictyota sp., Ellisolandia elongata and/or Halopteris scoparia) were collected at 0.5-meter depth at each sampling station, by using 250 ml plastic flask filled up to 196 197 the top with the surrounding water. Seawater above the sampled macroalgae, at 0.3-meter depth, was also collected for planktonic cell quantification. Plankton sampling was conducted before 198 collecting the macroalgae in order to avoid releasing Ostreopsis benthic cells initially attached 199 to the macroalgae which could overestimate Ostreopsis cells in the planktonic phase. All 200 201 samples were then fixed with acidic Lugol (1% vol./vol.) and kept dark at 4°C. In the laboratory, macroalgae samples were vigorously shaken in the plastic flasks and the seawater was filtered 202 through a 500 µm meshed filter to remove large macroalgal debris and isolate the epiphytic 203 cells. Two additional rinsing of the macroalgae with 100 ml of 0.2 µm filtered seawater were 204 205 conducted (with additional shaking and percolation steps) to detach as much microalgae cells as possible (Jauzein et al., 2018); all percolated water was merged and the total volume of 206 seawater was noted. Macroalgae were then gently pressed to eliminate excess water and placed 207 208 onto a tin foil for fresh weight measurement. Cell counts were performed on the percolated 209 water of samples following the Utermöhl method (Utermöhl, 1958), by using a light microscope (Axioscope 5, Zeiss, Oberkochen, Germany). A 1 ml Sedgwick Rafter counting chamber was 210 used for the epiphytic Ostreopsis cells and 50 ml sedimentation columns for the planktonic 211 cells. Benthic cell concentrations were expressed as cells per gram of fresh weight of 212 macroalgae (cell g^{-1} FW) and the plankton cells as cell l^{-1} . 213

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2.3.Hydroclimatic parameters

During the sampling period, seawater temperature $(\pm 0.1^{\circ}C)$ was measured *in situ* at each station using a probe. The "Direction de l'Environnement" of Monaco provided hydroclimatic data from their meteorological program, including SST, salinity, pH, wind speed, wind direction and rainfall in Larvotto beach area. Wind speed and direction were used to represent South-North

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and West-East wind intensity vectors. In addition, long-term hydrological data including SST,
nitrate and nitrite measurements from Point B in Villefranche-sur-Mer (43°41.10' N - 7°18.94'
E) were collected by SOMLIT, a French national coastline monitoring network
(http://somlit.fr). Located 12 km away from Monaco, Villefranche Bay can be considered as a
reference area for the western part of the Ligurian Sea, including Monaco. These data were
collected once a week in the bay, all year long, during the whole studied period.

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226 2.4.Calculation and Statistics

227 Cell abundance for each sampling date was calculated as the mean cell abundance considering 228 the five sampling stations. The difference in cell abundance between two sampling dates (i.e. 229 net growth rate = μ) is the result of the combination of several parameters such as cell growth 230 rate, cell dispersion, cell accumulation, predation and cell mortality rates. Net growth rates were 231 calculated between each sampling week by using the following equation:

$$\mu = \frac{LnN2 - LnN1}{t2 - t1}$$

Where μ is the net growth rate (NGR, d⁻¹) between sampling day 1 and sampling day 2, N1 the mean cell abundance at sampling day 1 (cell g⁻¹ FW), N2 the mean cell abundance at sampling day 2 (cell g⁻¹ FW), t1 the date in days at the sampling date 1 and t2 the date in days at the sampling date 2. Note that this formula assumes an exponential growth (as defined in Wood et al., 2005) between the two sampling dates, although the detailed dynamics of the cell abundances is not possible to be tracked.

Data from year 2007 were not included in all the analyses as the monitoring started late in the year (early July) while the bloom was already declining. Starting date, maximum abundances and maximum growth rates could not be estimated for this year. A mean weekly temperature profile was calculated over the thirteen-year monitoring period. Then, for each year, weekly SST anomalies were calculated as differences between temperature values and the mean weekly SST value. Finally, cumulative sums of SST anomalies were performed starting from the first week of each year.

Variations of net growth rate estimated during the phases of bloom development were 246 analyzed as a function of SST using a non-parametric "loess" local polynomial regression 247 model (Cleveland, 1979). The non-parametric Spearman correlation test was used to examine 248 links between the abundance of benthic and planktonic Ostreopsis cells, as well as between 249 SST and O. cf. ovata abundances (in the plankton and the benthos). A Wilcoxon signed-rank 250 (non-parametric) test was applied on SST data collected both in Monaco and Villefranche Bay 251 to ascertain the degree of similarity of the temperatures at both locations. A Principal 252 Component Analysis (PCA) was used to explore the relationships between the different 253 254 phenological features, namely, timing (starting and finalization dates of the bloom), duration, intensity (characterized by the maximum cell abundances) and maximum growth rate of the 255 256 bloom, and environmental parameters (seawater temperature, salinity, winds, rainfall and nutrient concentrations). This PCA was performed using seasonal averages (mean of April, 257 May and June) of environmental factors. These mean estimations were considered as 258 representative of the hydroclimatic environment of the spring season, described as a crucial 259 period before the bloom onset. The data set used for the PCA includes the years between 2008 260 and 2017 since rainfall, wind strength and nutrient concentrations data were not available for 261 the last two years of monitoring. All statistical analyses were performed using R studio (R Core 262 Team, 2019). 263

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3. Results

268 3.1.Interannual variability

Blooms of *O*. cf. *ovata* occurred every year since 2007 at the same monitored site (Fig. 2). All data obtained over the last thirteen years of monitoring were pooled and the mean cell abundances were calculated over the five sampling stations. Results showed a positive correlation between *O*. cf. *ovata* benthic and planktonic abundances of (Spearman test, N = 147, $r_s = 0.805$, p-value < 0.01) (Fig. 3). However, as explained earlier in the Introduction, *O*. cf. *ovata* an epibenthic species, with the benthic fraction constituting the reservoir of the bloom. Therefore, the following analysis will focus exclusively on benthic abundances.

The highest abundance levels of *O*. cf. *ovata* blooms were constantly observed in summer (between June and August). Only two events within the thirteen-year time series showed two successive blooms in the same year: a first maximum peak of benthic cell abundance in summer followed by a second and smaller peak in late summer were observed in 2012 and 2013 (Fig. 2). Highly variable phenological features of the blooms were observed from one year to another (Fig. 2). Nevertheless, three main parameters enabled to distinguish one bloom from another: the timing, the duration and the maximum cell abundance of blooms.

A concentration of 200,000 cells g⁻¹ FW was estimated as the threshold above which the 283 presence of O. cf. ovata was defined as a bloom. Indeed, this level refers to the lowest value of 284 the overall maximum annual abundances measured during the thirteen blooms (212,300 cells 285 g⁻¹ FW measured in 2015). Since samples were collected only once a week, it was difficult to 286 set the exact starting date of the bloom. The average starting day of a bloom was defined in the 287 present study using the abundance threshold (200,000 cells g⁻¹ FW) criterion and was defined 288 on Julian day 191 (\pm 10.6). Delays occurred depending of the year, with blooms appearing as 289 early as 15 days prior to this average starting day (Julian day 177, in 2015), while the latest 290 bloom occurred 25 days later (Julian day 214, 2016). 291

The duration of the event was also variable over the years. When more than 200,000 cells g⁻¹ FW were quantified during two consecutive sampling dates, the event was defined as a two-week bloom. When abundance values above this threshold were registered only once, then the event was defined as a one-week bloom. The blooms lasted in average 18.7 days (\pm 10.5), of which years 2011 and 2017 presented the longest blooming times (35 and 42 days respectively), and years 2007, 2009 and 2015 the shortest blooming times (7 days).

The highest annual maximum abundance values leveled to an average value of 570,878 (\pm 391,871) cells g⁻¹ FW. However, the highest value ever recorded was 1,565,514 cells g⁻¹ FW (174 % more than the average) in 2007 and the lowest value reached 212,300 cells g⁻¹ FW (63 % less than the average) in 2013.

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303 3.2.Relationship between environmental factors and interannual variability of *Ostreopsis*304 cf. *ovata* blooms

305 A multivariate analysis (PCA) was performed on data collected between April and June from the monitoring of environmental parameters, including physico-chemical measurements of 306 water masses (SST, salinity, nutrients) as well as rainfall and wind intensity vectors 307 308 (Supplementary Material 1). The two first axis forming the first factorial space explained 70% of the total inertia of projected variables (Fig. 4), indicating a relatively good representation of 309 310 the relationship between the different environmental parameters. The most important variables for the construction of the first factorial axis of the PCA (46.5 % of the total inertia) were South-311 North winds (0.90), SST (-0.84) and West-Est winds (0.79). The second axis was driven by 312 nitrates (0.80), salinity (-0.62) and nitrites (0.57) and captured 23.6% of the variation. These 313 results indicate that the first factorial axis of the PCA discriminates the years with bloom events 314 according to the SST during the spring season. Indeed, years with high SST during spring seem 315

to be related to low wind conditions. The second factorial axis highlights the nutrient loading 316 of the area and discriminates years according to the nutrient availability during the spring 317 seasons. Indeed, years which experience a high level of rainfall are related to high nitrate 318 concentration and low salinity values. Among the phenological features, the timing of the 319 bloom event (bloom starting day) and the maximum annual growth rate are both correlated to 320 the first axis of the PCA (0.71 and 0.39, respectively), while the intensity of the bloom 321 (maximum cell concentration) is correlated to the second axis (0.41). No clear relation was 322 found between environmental factors and the bloom duration. 323

Throughout the entire time series, the presence of O. cf. ovata benthic cells was 324 observed when SST ranged between 18.2°C and 27.9°C, with an overall temperature average 325 of 24.3 ± 1.8 °C. When processing the entire dataset provided by the thirteen-year survey, 326 blooming periods (i.e. when benthic cell abundance exceeded 200,000 cells g⁻¹ FW) were 327 identified when the mean SST reached 25.0 ± 1.5 °C. When focusing more specifically on each 328 individual year, mean SST values measured during blooming periods ranged between 24.0 \pm 329 330 1.5° C and $25.2 \pm 1.7^{\circ}$ C. At the very beginning of a bloom event, when cell densities first exceeded the threshold of 200,000 cells g^{-1} FW, the average SST was estimated at 24.9 ±1.4°C. 331 Positive net growth rates (NGR) of O. cf. ovata were recorded when the SST varied 332 between 18.2°C and 27.9°C. A negative correlation was found between the temperature and the 333 NGR (Spearman, N = 133, $r_s = -0.27$, p-value < 0.01), but the relation was not linear and highest 334 positive NGRs were not associated to maximal temperature records (Fig. 5). Indeed, the 335 maximum NGR (0.53 d⁻¹) occurred at a temperature of 23.9°C. These variations are confirmed 336 by the local regression loess model, defining the optimal growth temperature at 23.5°C. In these 337 terms, when considering each maximum NGR for each year, the window of optimal growth 338 temperature was estimated between 21.0°C and 24.7°C (Fig. 5). 339

No significant differences were found by comparing summer SST values measured in 340 Larvotto and in Villefranche Bay (Wilcoxon signed-rank test: N = 125, W = 3869, p-value = 341 0.645). Therefore, SST values recorded in Villefranche Bay were used as a proxy for estimating 342 the SST of Monaco area. In contrast to data collected in Monaco, SST measurements in 343 Villefranche Bay were collected weekly and throughout the year, which enabled to describe the 344 annual general hydroclimatic conditions of the area. Between 2007 and 2019, the annual mean 345 SST was estimated at 18.8°C (\pm 4.3) and the respective mean values of the lowest and the 346 highest annual temperatures were estimated at $13.4^{\circ}C$ (±0.4) and $26.3^{\circ}C$ (±0.9), respectively. 347 Annual SST anomalies were cumulated in order to compare the heat accumulated in surface 348 waters between each surveyed year (Fig. 6). Years with high values of SST anomalies were 349 considered as warmer, while years with low values of SST anomalies were considered as cooler. 350 Profiles of cumulated SST anomalies were different between the years but were also found to 351 352 be different at a seasonal and monthly scale, thereby differentiating years with a warmer spring season from years with a cooler spring. Amongst the previously detailed phenological features 353 354 of interest, strong positive correlations were found only between the bloom timing (starting of the event) and the cumulative sum of SST anomalies (Fig. 7A). This was observed for monthly 355 averaged cumulative sum of SST anomalies over the months of April, May and June (prior to 356 357 the bloom event) and during the whole spring season (April-May-June averaged values), but not before then (Fig. 7B). Thus, blooms of O. cf. ovata tend to start earlier in the season when 358 SST anomalies are high. No other significant correlations were observed between SST 359 anomalies and the duration of the bloom nor the maximum abundance of the bloom. 360

361 **4. Discussion**

Understanding the relationship between environmental parameters and the phenology of *O*. cf.
 ovata proliferations in the NW Mediterranean Sea can help predict either the timing, duration
 or intensity of blooms and hence support the management and mitigation of benthic HAB

impacts. Such knowledge cannot be gained by observing one single bloom event (e.g. Vila et 365 al., 2001; Battocchi et al., 2010; Totti et al., 2010; Accoroni et al., 2011; Illoul et al., 2012) but 366 requires pluri-annual time series enabling to study the variability of occurring blooms along 367 several years. Furthermore, long-term surveys as currently presented in this study enabled to 368 characterize the *in situ* environmental niche of the target organism, O. cf. ovata, in the NW 369 Mediterranean Sea. Such a set of data is of great interest for both modelers and managers, 370 focusing on the prediction of HABs in the context of global change and warming of seawater 371 372 surfaces.

In the present study, a multi-parametric approach and the analysis of environmental 373 control parameters for algal growth were both assessed by using the benthic cell number data 374 from O. cf. ovata abundance measurements. Indeed, as already reported by several studies 375 conducted in NW Mediterranean regions (Mangialajo et al., 2008; Cohu et al., 2011; Jauzein et 376 377 al., 2018), a strong correlation was found between concentrations of epiphytic and planktonic cells of O. cf. ovata. Ostreopsis is recognized as a benthic organism adapted to the benthic life 378 379 which is proven by its morphology and ecology (e.g., Hoppenrath et al., 2014). Due to the combination of possible internal biological rhythms and very local sea hydrodynamics, cells 380 can detach from the benthos and circulate as pelagic cell, which explains the observed 381 correlation between benthic and pelagic cell abundances. Epiphytic cells nevertheless 382 represents the main stock, or reservoir, of the O. cf. ovata population able to sustain bloom 383 events. In terms of toxic risks, both benthic and planktonic pools can be involved in direct 384 contact skin irritations, but planktonic cells only would rather be the agents of intoxication 385 events promoted by inhalation (Totti et al., 2010). In this case, Ostreopsis-produced toxic 386 compounds would be spread by aerosolization and wind transport (Ciminiello et al., 2014). 387

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Two main results emerge from the analysis of this long-term dataset and shall be discussed hereafter: (1) Spring SST were driving the timing of the bloom (starting date) and (2) the highest recorded temperatures were not correlated to the highest net growth rates of *O*. cf. *ovata*.

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394 4.1.Interannual variations in bloom phenology

The monitoring of O. cf. ovata blooms which occurred in Monaco during thirteen successive 395 summer seasons, revealed strong yearly variations in terms of phenological features. These 396 397 variations, such as bloom timing, bloom length and bloom intensity, highlight the complexity behind the formation of O. cf. ovata blooms and the possible interactions between 398 environmental parameters which are able to shape bloom events. The PCA analysis carried out 399 400 on this data-set showed that the interannual variability in bloom timing and bloom maximum intensity could be explained by the combination of SSTs, nutrient loadings and wind regimes 401 occurring in the spring season prior to the bloom initiation. Indeed, years which experience high 402 SSTs during the spring season are likely to see O. cf. ovata densities exceed the 200,000 cells 403 g⁻¹ FW threshold early in the summer. High nutrient concentrations, prior and during the bloom 404 405 onset, is also likely to favor intense bloom episodes. Rainfall and freshwater inflows obviously provided a high nutrient load and decreased water salinity, as shown in the PCA. The combined 406 action of SST and nutrient concentrations on O. cf. ovata bloom phenology has also been 407 408 investigated by Accoroni et al. (2015) when studying a time-series of O. cf. ovata blooms in 409 the Northern Adriatic Sea. These authors suggested that the bloom onset could be triggered when the SST reached a certain threshold and simultaneously coinciding with an inflow of P-410 411 riche waters which occurred after a period of relatively high N:P ratio. Similar interconnections were observed by Bucci et al. (2020) regarding the phenology of Alexandrium catenella blooms 412 which occurred in the Bay of Fundy (Gulf of Maine, NW Atlantic). The study showed that the 413

bloom timing and the date exhibiting the maximum cell density (bloom intensity) varied 414 between the years and suggested that both parameters could be linked to a high input of 415 freshwater during the spring and that warmer years were likely to earlier initiate the timing of 416 blooms. In our study, spring season hydrodynamics induced by the strength and the direction 417 of winds did not significantly impact the variability of O. cf. ovata phenological features. This 418 latter observation was described at least at the interannual scale, although it may play a more 419 important role in the bloom development at the seasonal scale. However, these environmental 420 parameters were shown to be related to high SSTs. Indeed, in the NW Mediterranean seas, wind 421 regimes coming from the South are generally warm and thereby increase the water temperature. 422 Our analysis also highlights the importance of the spring season regarding the phenology of 423 recurrent summer bloom events. Even though environmental factors strongly impact the 424 phenology of the bloom when it is occurring, less is known about how hydroclimatic parameters 425 426 are affecting the bloom phenology prior to its development.

Further investigations of factors initiating the bloom revealed a strong positive 427 428 correlation between cumulated SST anomalies and the starting day of the bloom. This relation 429 confirms that O. cf. ovata blooms are likely to occur earlier in the year, when spring SST are warmer than usual. This correlation is even stronger and more significant when approaching 430 431 the summer season. Indeed, from April onwards, temperature anomalies can be cumulated to predict the most favorable period for epiphytic O. cf. ovata abundances to reach the 200,000 432 cells g⁻¹ FW threshold. However, variations in SST during the months of May and June have a 433 strong impact on the bloom timing and may reject the prediction from April. It hence appears 434 that the accuracy for predicting the bloom with SST parameter gets stronger when approaching 435 the bloom event. 436

437 The occurrence of an early bloom can be explained by an early germination of resting
438 *O.* cf. *ovata* cysts and/or by a higher division rate favored by optimal environmental conditions

for algal growth. For instance, Lau et al. (2017) showed that an intense bloom of Alexandrium 439 minutum, which occurred in the northeast of the Peninsular Malaysian waters, was presumably 440 initiated by an active excystment process. Indeed, several environmental parameters have 441 already shown to trigger dinoflagellate cyst germination such as light exposure (Anderson et 442 al., 1987; Moore et al., 2015), nutrient availability (Bravo and Anderson, 1994; Binder and 443 Anderson, 2004) and a mandatory exposed period to dark-cold conditions (Montresor and 444 Lewis, 2006). Temperature also appears to be one of the main factors controlling excystment 445 of resting dinoflagellate cells (Anderson, 1980; Bravo and Anderson, 1994; Ní Rathaille and 446 Raine, 2011; Moore et al., 2015) as dinoflagellate cysts require a certain SST threshold in order 447 to induce excystment. For O. cf ovata blooms in the Mediterranean Sea, no relationship between 448 the bloom starting date and neither SST nor nutrient concentrations was found. Instead, only 449 cumulated temperature anomalies (illustrations of heat accumulation in surface waters) showed 450 451 effects on the bloom timing. Accoroni et al. (2014) indicated that cysts of O. cf. ovata, sampled from the N Adriactic Sea, needed a temperature of 25°C to germinate, which could explain why 452 453 blooms occurring in this area are observed from late summer to beginning of fall, after recording the highest temperatures. Another study conducted by Accoroni et al. (2015a) suggested that 454 the combination of optimal temperature and Nitrate/Phosphate ratio could favor both the 455 initiation of cell excystment and cell proliferation, therefore triggering the bloom onset. In this 456 study monitoring O. cf. ovata blooms in the Conero Riviera (N Adriatic Sea) between 2007 and 457 2012, cells were never observed below 25°C prior to the bloom onset. Conversely, in Monaco, 458 cells of O. cf. ovata have been observed at temperatures as low as 18.2°C, and blooms usually 459 start before the maximum SSTs are reached. These variable results seem to indicate that 460 depending on the geographical location, different SST thresholds might be necessary for 461 triggering the cyst germination process which leads to an O. cf. ovata bloom. 462

O. cf. ovata bloom events also varied in duration and in intensity across the years. 463 During the bloom events, the highest measured abundance of epiphytic cells reached more than 464 1,500,000 cells g⁻¹ FW (in 2007) while the lowest maximum abundance barely exceeded 465 200,000 cells g⁻¹ FW (in 2013). The blooms lasted from less than one week (only one sampling 466 day, in 2009 and 2015) up to 4 weeks (5 consecutive sampling days, in 2011). Length and 467 intensity of blooms are the results of a balance between production of biomass (gross growth 468 rate) versus induction of cell quiescence (encystment) and cell loss. Diverse factors can induce 469 a limitation of microalgal growth or activation of encystment or lysis. These factors can either 470 be abiotic, such as strong hydrodynamics, nutrient depletion or extreme temperatures, or biotic 471 (e.g., grazing, parasitic interactions). Therefore, long lasting and intense bloom episodes could 472 be the result of persisting optimal hydroclimatic conditions over a long enough period which 473 enhances the production of biomass, and this combined to low biotic (grazers, parasites) and 474 475 abiotic (hydrodynamics) pressures which are responsible for cell loss. Temporary cyst formation could be a strategy leading to maintain blooms throughout time. Indeed, O. cf. ovata 476 477 can form temporary cysts from vegetative cells (Accoroni et al., 2014) when biotic or abiotic pressures emerge in order to overcome short periods of stress. This mechanism has already been 478 shown to help maintain blooms of other harmful dinoflagellate such as Alexandrium taylori 479 (Garcés, 2002). Another strategy would be the fueling of the bloom by resting cysts throughout 480 the bloom event, as observed in *Alexandrium* species (Anglès et al., 2012; Lau et al., 2017). By 481 undergoing a slow and wide excystment process, the resting cyst stock would hence 482 continuously provide new vegetative cells to the blooming population, and in fine enable a 483 longer bloom episode. 484

In addition to the numerous environmental factors responsible for *O*. cf. *ovata* bloom phenology, part of the variability observed in bloom dynamics over the years may be the result of the strong small-scale repartition of cells on the benthos. Variations observed in the *O*. cf. *ovata* bloom phenology could be related to a high variability of cell abundances measured over small spatial scales. Indeed, the patchy distribution of epiphytic *O*. cf. *ovata* cells, directly linked to the distribution of their biotic substrate, makes the accurate estimation of the benthic cell stock of a given area challenging. This has already been observed by Cohu et al. (2011), who carried out the first two years of this present monitoring program, and suggested that bloom phenology could be related to the spatial and temporal evolution of the diversity of biotic substrates (Totti et al., 2010), as well as local hydrodynamic conditions.

Finally, although SST is recognized as a determinant factor for the development of 495 benthic dinoflagellate cells (Pistocchi et al., 2011; Parsons et al., 2012), O. cf. ovata abundances 496 497 and growth rates measured during the bloom events are also under the control of other environmental parameters (Carnicer et al., 2015; Accoroni and Totti, 2016). It is generally 498 agreed that O. cf. ovata blooms are favored by low hydrodynamic conditions, since the highest 499 abundance levels are rather recorded in sheltered areas, even though blooms have also been 500 observed in exposed sites (Barone, 2007; Totti et al., 2010; Accoroni and Totti, 2016). By 501 502 comparing bloom events throughout the years, the PCA analysis in this present study showed no straight forward relation between environmental parameters and maximum annual growth 503 rates of O. cf. ovata. 504

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506 4.2.Net growth rate

Throughout the recurrent *O*. cf. *ovata* summer bloom episodes, the highest positive net growth rates did not coincide with the highest recorded SST. This indicates that relatively high temperatures are not necessarily optimal for the growth of *Ostreopsis*, at least in the Monaco area. Instead, the loess regression model indicated that a moderate temperature (23.5°C) was most suitable for *O*. cf. *ovata* growth. Numerous *in situ* studies have already tried to understand the relation between SST and bloom development, often resulting in different conclusions

which seem to depend on the actual studied location (Accoroni and Totti, 2016). For instance, 513 in the NW Mediterranean Sea, the highest abundances of O. cf. ovata were recorded when 514 temperatures were either higher than 26°C (e.g., in the Ligurian Sea (Mangialajo et al., 2008) 515 and in Greece ((Aligizaki and Nikolaidis, 2006)) or below 23°C (e.g., in the Northern Adriatic 516 Sea (Totti et al., 2010; Accoroni et al., 2011)). In most of these studies, the monitored data were 517 reported as cell abundances and did not show any direct correlation between SST and cell 518 concentration (Vila et al., 2001; Totti et al., 2010; Cohu et al., 2011). On the other hand, by 519 estimating the maximum growth rate of O. cf. ovata over successive bloom events in this 520 present study focusing on the NW Mediterranean Sea, a narrower range of optimal growth 521 temperatures was defined between 21.0°C and 24.7°C. This temperature range is in agreement 522 with results obtained from laboratory scale studies, which reported that optimal growth 523 temperatures could be strain specific and part of such a strain specificity referring to a thermal 524 525 niche could depend on the studied area. For instance, Pezzolesi et al. (2012) showed that strains collected in the Tyrrhenian Sea showed maximum growth rates at 20°C whereas highest growth 526 527 rates were observed at 24°C and 30°C when studying strains isolated from the Catalan coast (Spain, Carnicer et al., 2016) and the Adriatic Sea (Granéli et al., 2011), respectively. However, 528 even within the same region of the Mediterranean Sea, temperature niches enabling optimal 529 growth seem to vary depending on the strain. Indeed, a laboratory study carried out by Scalco 530 et al., (2012) on strains collected in the Tyrrhenian and Adriatic seas showed that the optimal 531 temperature for algal growth was strain-specific, varying between 22°C and 26°C depending 532 on the clone. Since O. cf. ovata niche temperatures are strongly variable depending on the 533 location but also depending on the strains across Mediterranean Sea, the impact of temperature 534 on the development of blooms are likely to vary depending on local environmental parameters 535 and clonal diversity. In addition, the highest growth rates of O. cf. ovata do not coincide with 536 the highest observed SSTs, as also reported in most of the in situ and laboratory studies 537

538 conducted in the Mediterranean Sea. This suggests a more complex role of temperature 539 conditions on the phenology of the blooms than a simple increasing bloom facilitation due to 540 increasing temperature. Exceeding too high SST values might even limit the development of 541 the bloom. In the context of ocean warming, this has to be taken into account in the view of 542 implementing long-term previsions of *O*. cf. *ovata* bloom events in the future.

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4.3.Spring SST anomalies and in situ thermal niche: a tool for modelers and managers

Relationships between the phenology of O. cf. ovata blooms and environmental parameters 544 described in this present study are of great interest for predicting and managing HABs. Indeed, 545 546 the correlation between spring SST anomalies and bloom timing represents a useful tool for predicting future O. cf. ovata HABs in the Monaco area. Using SST data recorded during the 547 beginning of the year enabled to predict a temporal window for the beginning of a bloom event 548 549 which can be useful to guide the monitoring effort (Fig. 7A). The *in situ* thermal niche, determined by the relationship between SST and net growth rates during bloom episodes, could 550 also be used by modelers to forecast the future evolution of both the bloom phenology and 551 distribution which can be notably impactful when facing increasing sea temperatures. 552

Modeling the dynamics of coastal HABs is crucial, especially in a context of climate 553 554 change where the occurrence of HABs seems to be increasing in numbers and expanding in space. Many authors are trying to understand and to predict blooms of toxic dinoflagellates 555 (Maguire et al., 2016; Moita et al., 2016; Ruiz-Villarreal et al., 2016) and toxic cyanobacteria 556 (Hamilton et al., 2009; Rigosi et al., 2015) by using models based on ecological observations. 557 Until now, the accuracy in modelling and predicting such complex phenological events seems 558 to be limited by the lack of available data describing both the species itself and the 559 560 environmental parameters, but also by the numerous pending questions regarding the ecology of O. cf. ovata (Asnaghi et al., 2012). Although our study provides data on the actual 561 temperature niche of O. cf. ovata in the NW Mediterranean Sea and the relations between the 562

bloom phenology and spring SST temperatures, more models describing O. cf. ovata blooms 563 still need to be designed, improved and adapted to local environmental variables in order to 564 minimize the risks to be exposed to such events. The determination of the thermal niche of O. 565 cf. ovata in the NW Mediterranean Sea could help forecasting potential long-term variations in 566 the phenology of the blooms but also a possible evolution of the distribution of blooms. On the 567 other hand, the short-term prediction of bloom timing could provide a substantial support to 568 local sanitary authorities in charge of toxic outbreaks, as this could be used to estimate potential 569 risks of exposure weeks before the beginning of the bloom and hence to gain precious time for 570 deploying management strategies. 571

Regarding the impact of climate change, many studies consider global warming as a 572 potential enhancement factor for benthic HABs species to expand their geographical areas, 573 increase the seasonal occurrences and bloom windows (Wells et al., 2015; Gobler et al., 2017; 574 575 Nakada et al., 2018; Gobler, 2020; Tester et al., 2020). In the Mediterranean Sea, climate projections foresees an increase in SSTs, with an average of 0.45°C per decade by 2050 576 577 (Alexander et al., 2018). Geographical expansion towards higher latitudes in the Mediterranean Sea is not possible, since O. cf. ovata is already found in the northern parts of the NW 578 Mediterranean Sea. Moreover, an increase in SST temperature may not necessarily result in 579 580 more intense blooms since this present study has shown that the highest SSTs do not coincide with the highest growth rates. Therefore, the assumption that ocean warming shall favor the 581 development and the expansion of O. cf. ovata species and blooms in the future, seems more 582 complex than previously thought. On the other hand, global warming could favor an earlier 583 bloom, as optimal temperature conditions would be reached earlier in the season. Global 584 warming is not the only consequence of climate change and more specific studies between O. 585 cf. ovata and future ecological conditions in the Mediterranean Sea are needed. 586

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The phenology of O. cf. ovata blooms is governed by a complex assemblage of environmental 589 factors and is still lacking specific knowledge able to describe how these environmental 590 parameters interact and affect the progress of toxic blooms. The present study is based on a 13-591 year time-series of O. cf. ovata blooms in Monaco and reveals that part of the interannual 592 variability of bloom phenological features, such as the bloom timing and the maximum intensity 593 of the bloom, could be explained by a combination of environmental parameters, including SST 594 and nutrient concentrations, especially during the spring season. The optimal growth 595 temperature which favors the O. cf. ovata blooms in Monaco waters has been estimated between 596 597 21°C and 25°C. In addition, this study revealed a clear relation between spring SSTs and the 598 timing of the bloom of O. cf. ovata occurring in the NW Mediterranean region. By processing data from this longest time series dedicated to O. cf. ovata blooms, results suggest that the 599 accumulation of positive SST anomalies induces an early start of the bloom. Considering the 600 increasing trend of SST observed worldwide, due to global climatic changes, such results could 601 be of a great interest for the prediction and management of O. cf. ovata blooms in recreational 602 603 waters. Further data collection in the scope of this time series is needed to improve the present knowledge and to deepen investigations regarding the effect of combined environmental 604 parameters on O. cf. ovata bloom phenology. 605

606 Authors' contribution

K.D., A-S.P., S.M. V.D-S. and R.L. designed and contributed to the conduction of the
monitoring campaign. K.D., A-S.P. and S.M. analyzed the samples. K.D. and S.G. designed
and performed the modelling and statistical analyses. K.D., S.G., E.B., C.J., R.S. and R.L.
analyzed the results and prepared the figures. K.D. and R.L. wrote the first version of the
manuscript. All authors reviewed and accepted the final version of the manuscript.

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623 on HAB.

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629 **References**

630	Abdennadher, M., Zouari, A.B., Sahnoun, W.F., Alverca, E., Penna, A., Hamza, A., 2017. Ostreopsis cf.
631	ovata in the Gulf of Gabès (south-eastern Mediterranean Sea): morphological, molecular and
632	ecological characterization. Harmful Algae 63, 56–67.
633	https://doi.org/10.1016/j.hal.2017.01.009
634	Accoroni, S., Colombo, F., Pichierri, S., Romagnoli, T., Marini, M., Battocchi, C., Penna, A., Totti, C.,
635	2012. Ecology of Ostreopsis cf. ovata Blooms in the Northwestern Adriatic Sea. Cryptogam.

636 Algol. 33, 191–198. https://doi.org/10.7872/crya.v33.iss2.2011.191

- 637 Accoroni, S., Glibert, P.M., Pichierri, S., Romagnoli, T., Marini, M., Totti, C., 2015a. A conceptual 638 model of annual Ostreopsis cf. ovata blooms in the northern Adriatic Sea based on the 639 synergic effects of hydrodynamics, temperature, and the N:P ratio of water column 640 nutrients. Harmful Algae 45, 14–25. https://doi.org/10.1016/j.hal.2015.04.002
- Accoroni, S., Percopo, I., Cerino, F., Romagnoli, T., Pichierri, S., Perrone, C., Totti, C., 2015b. 641 Allelopathic interactions between the HAB dinoflagellate Ostreopsis cf. ovata and 642 643 macroalgae. Harmful Algae 49, 147–155. https://doi.org/10.1016/j.hal.2015.08.007
- 644 Accoroni, S., Romagnoli, T., Colombo, F., Pennesi, C., Di Camillo, C.G., Marini, M., Battocchi, C.,
- 645 Ciminiello, P., Dell'Aversano, C., Dello Iacovo, E., Fattorusso, E., Tartaglione, L., Penna, A., 646 Totti, C., 2011. Ostreopsis cf. ovata bloom in the northern Adriatic Sea during summer 2009: 647 Ecology, molecular characterization and toxin profile. Mar. Pollut. Bull. 62, 2512–2519. 648 https://doi.org/10.1016/j.marpolbul.2011.08.003
- 649 Accoroni, S., Romagnoli, T., Pichierri, S., Totti, C., 2014. New insights on the life cycle stages of the 650 toxic benthic dinoflagellate Ostreopsis cf. ovata. Harmful Algae 34, 7–16. 651 https://doi.org/10.1016/j.hal.2014.02.003
- 652 Accoroni, S., Totti, C., 2016. The toxic benthic dinoflagellates of the genus Ostreopsis in temperate 653 areas: a review. Adv. Oceanogr. Limnol. 7, 1–15. https://doi.org/10.4081/aiol.2016.5591
- 654 Accoroni, S., Totti, C., Razza, E., Congestri, R., Campanelli, A., Marini, M., Ellwood, N.T.W., 2017. 655 Phosphatase activities of a microepiphytic community during a bloom of Ostreopsis cf. ovata in the northern Adriatic Sea. Water Res. 120, 272–279. 656 657
 - https://doi.org/10.1016/j.watres.2017.05.004
- 658 Alexander, M.A., Scott, J.D., Friedland, K.D., Mills, K.E., Nye, J.A., Pershing, A.J., Thomas, A.C., 2018. 659 Projected sea surface temperatures over the 21stcentury: Changes in the mean, variability 660 and extremes for large marineecosystem regions of Northern Oceans. Elem Sci Anth 6, 9. 661 https://doi.org/10.1525/elementa.191
- 662 Aligizaki, K., Katikou, P., Nikolaidis, G., Panou, A., 2008. First episode of shellfish contamination by 663 palytoxin-like compounds from Ostreopsis species (Aegean Sea, Greece). Toxicon 51, 418-664 427. https://doi.org/10.1016/j.toxicon.2007.10.016
- 665 Aligizaki, K., Nikolaidis, G., 2006. The presence of the potentially toxic genera Ostreopsis and Coolia 666 (Dinophyceae) in the North Aegean Sea, Greece. Harmful Algae 5, 717–730. 667 https://doi.org/10.1016/j.hal.2006.02.005
- 668 Anderson, D.M., 1980. Effects of temperature conditioning on development and germination of 669 Gonyaulax tamarensis (dinophyceae) hypnozygotes. J. Phycol. 16, 166–172. 670 https://doi.org/10.1111/j.1529-8817.1980.tb03013.x
- 671 Anderson, D.M., Cembella, A.D., Hallegraeff, G.M., 2012. Progress in Understanding Harmful Algal 672 Blooms: Paradigm Shifts and New Technologies for Research, Monitoring, and Management. 673 Annu. Rev. Mar. Sci. 4, 143–176. https://doi.org/10.1146/annurev-marine-120308-081121
- 674 Anderson, D.M., Glibert, P.M., Burkholder, J.M., 2002. Harmful algal blooms and eutrophication: 675 Nutrient sources, composition, and consequences. Estuaries 25, 704–726. 676 https://doi.org/10.1007/BF02804901
- 677 Anderson, D.M., Taylor, C.D., Armbrust, E.V., 1987. The effects of darkness and anaerobiosis on 678 dinoflagellate cyst germination: Dinoflagellate cyst germination. Limnol. Oceanogr. 32, 340-679 351. https://doi.org/10.4319/lo.1987.32.2.0340
- 680 Anglès, S., Garcés, E., Reñé, A., Sampedro, N., 2012. Life-cycle alternations in Alexandrium minutum 681 natural populations from the NW Mediterranean Sea. Harmful Algae 16, 1–11. 682 https://doi.org/10.1016/j.hal.2011.12.006
- 683 Asnaghi, V., Bertolotto, R., Giussani, V., Mangialajo, L., Hewitt, J., Thrush, S., Moretto, P., Castellano, 684 M., Rossi, A., Povero, P., Cattaneo-Vietti, R., Chiantore, M., 2012. Interannual Variability in 685 Ostreopsis ovata Bloom Dynamic along Genoa Coast (North-Western Mediterranean): A 686 Preliminary Modeling Approach. Cryptogam. Algol. 33, 181–189.
- 687 https://doi.org/10.7872/crya.v33.iss2.2011.181

- Barone, R., 2007. Behavioural trait of *Ostreopsis ovata* (Dinophyceae) in Mediterranean rock pools:
 the spider's strategy. Harmful Algae News 33, 1–3.
- Barroso García, P., Rueda de la Puerta, P., Parrón Carreño, T., Marín Martínez, P., Guillén Enríquez, J.,
 2008. An epidemic outbreak with respiratory symptoms in the province of Almeria (Spain)
 due to toxic microalgae exposure. Gac. Sanit. 22, 578–584.
- Battocchi, C., Totti, C., Vila, M., Masó, M., Capellacci, S., Accoroni, S., Reñé, A., Scardi, M., Penna, A.,
 2010. Monitoring toxic microalgae *Ostreopsis* (dinoflagellate) species in coastal waters of the
 Mediterranean Sea using molecular PCR-based assay combined with light microscopy. Mar.
 Pollut. Bull. 60, 1074–1084. https://doi.org/10.1016/j.marpolbul.2010.01.017
- Binder, B.J., Anderson, D.M., 2004. Physiological and environmental control of germination in
 Scrippsiella trochoidea (dinophyceae) resting cysts1. J. Phycol. 23, 99–107.
 https://doi.org/10.1111/j.0022-3646.1987.00099.x
- Biré, R., Trotereau, S., Lemée, R., Delpont, C., Chabot, B., Aumond, Y., Krys, S., 2013. Occurrence of
 palytoxins in marine organisms from different trophic levels of the French Mediterranean
 coast harvested in 2009. Harmful Algae 28, 10–22. https://doi.org/10.1016/j.hal.2013.04.007
- Biré, R., Trotereau, S., Lemée, R., Oregioni, D., Delpont, C., Krys, S., Guérin, T., 2015. Hunt for
 Palytoxins in a Wide Variety of Marine Organisms Harvested in 2010 on the French
 Mediterranean Coast. Mar. Drugs 13, 5425–5446. https://doi.org/10.3390/md13085425
- Blanfuné, A., Boudouresque, C.F., Grossel, H., Thibaut, T., 2015. Distribution and abundance of
 Ostreopsis spp. and associated species (Dinophyceae) in the northwestern Mediterranean:
 the region and the macroalgal substrate matter. Environ. Sci. Pollut. Res. 22, 12332–12346.
 https://doi.org/10.1007/s11356-015-4525-4
- Bravo, I., Anderson, D.M., 1994. The effects of temperature, growth medium and darkness on
 excystment and growth of the toxic dinoflagellate *Gymnodinium catenatum* from northwest
 Spain. J. Plankton Res. 16, 513–525. https://doi.org/10.1093/plankt/16.5.513
- Brescianini, C., Grillo, C., Melchiorre, N., Bertolotto, R., Ferrari, A., Vivaldi, B., Icardi, G., Gramaccioni,
 L., Funari, E., Scardala, S., 2006. Ostreopsis ovata algal blooms affecting human health in
 Genova, Italy, 2005 and 2006. Euro Surveill. Bull. Eur. Sur Mal. Transm. Eur. Commun. Dis.
 Bull. 11, E060907.3.
- Prissard, C., Herrenknecht, C., Séchet, V., Hervé, F., Pisapia, F., Harcouet, J., Lémée, R., Chomérat, N.,
 Hess, P., Amzil, Z., 2014. Complex Toxin Profile of French Mediterranean Ostreopsis cf. ovata
 Strains, Seafood Accumulation and Ovatoxins Prepurification. Mar. Drugs 12, 2851–2876.
 https://doi.org/10.3390/md12052851
- Bucci, A.F., Thomas, A.C., Cetinić, I., 2020. Interannual Variability in the Thermal Habitat of
 Alexandrium catenella in the Bay of Fundy and the Implications of Climate Change. Front.
 Mar. Sci. 7, 587990. https://doi.org/10.3389/fmars.2020.587990
- Carnicer, O., García-Altares, M., Andree, K.B., Tartaglione, L., Dell'Aversano, C., Ciminiello, P., de la
 Iglesia, P., Diogène, J., Fernández-Tejedor, M., 2016. Ostreopsis cf. ovata from western
 Mediterranean Sea: Physiological responses under different temperature and salinity
 conditions. Harmful Algae 57, 98–108. https://doi.org/10.1016/j.hal.2016.06.002
- Carnicer, O., Guallar, C., Andree, K.B., Diogène, J., Fernández-Tejedor, M., 2015. Ostreopsis cf. ovata
 dynamics in the NW Mediterranean Sea in relation to biotic and abiotic factors. Environ. Res.
 143, 89–99. https://doi.org/10.1016/j.envres.2015.08.023
- Ciminiello, P., Dell'Aversano, C., Dello Iacovo, E., Fattorusso, E., Forino, M., Grauso, L., Tartaglione, L.,
 Guerrini, F., Pezzolesi, L., Pistocchi, R., Vanucci, S., 2012. Isolation and Structure Elucidation
 of Ovatoxin-a, the Major Toxin Produced by *Ostreopsis ovata*. J. Am. Chem. Soc. 134, 1869–
 1875. https://doi.org/10.1021/ja210784u

Ciminiello, P., Dell'Aversano, C., Fattorusso, E., Forino, M., Magno, G.S., Tartaglione, L., Grillo, C., Melchiorre, N., 2006. The Genoa 2005 Outbreak. Determination of Putative Palytoxin in Mediterranean *Ostreopsis ovata* by a New Liquid Chromatography Tandem Mass Spectrometry Method. Anal. Chem. 78, 6153–6159. https://doi.org/10.1021/ac060250j

- Ciminiello, P., Dell'Aversano, C., Iacovo, E.D., Fattorusso, E., Forino, M., Tartaglione, L., Benedettini,
 G., Onorari, M., Serena, F., Battocchi, C., Casabianca, S., Penna, A., 2014. First Finding of *Ostreopsis* cf. *ovata* Toxins in Marine Aerosols. Environ. Sci. Technol. 48, 3532–3540.
 https://doi.org/10.1021/es405617d
- Cleveland, W.S., 1979. Robust Locally Weighted Regression and Smoothing Scatterplots. J. Am. Stat.
 Assoc. 74, 829–836. https://doi.org/10.1080/01621459.1979.10481038
- Cohu, S., Lemée, R., 2012. Vertical distribution of the toxic epibenthic dinoflagellates *Ostreopsis* cf. *ovata*, *Prorocentrum lima* and *Coolia monotis* in the NW Mediterranean Sea. Cah. Biol. Mar.
 55, 373–380.
- Cohu, S., Mangialajo, L., Thibaut, T., Blanfuné, A., Marro, S., Lemée, R., 2013. Proliferation of the toxic dinoflagellate *Ostreopsis* cf. *ovata* in relation to depth, biotic substrate and environmental factors in the North West Mediterranean Sea. Harmful Algae 24, 32–44.
 https://doi.org/10.1016/j.hal.2013.01.002
- Cohu, S., Thibaut, T., Mangialajo, L., Labat, J.-P., Passafiume, O., Blanfuné, A., Simon, N., Cottalorda,
 J.-M., Lemée, R., 2011. Occurrence of the toxic dinoflagellate *Ostreopsis* cf. *ovata* in relation
 with environmental factors in Monaco (NW Mediterranean). Mar. Pollut. Bull. 62, 2681–
 2691. https://doi.org/10.1016/j.marpolbul.2011.09.022
- David, H., Ganzedo, U., Laza-Martínez, A., Orive, E., 2012. Relationships between the Presence of
 Ostreopsis (Dinophyceae) in the Atlantic Coast of the Iberian Peninsula and Sea-Surface
 Temperature. Cryptogam. Algol. 33, 199–207.
- 759 https://doi.org/10.7872/crya.v33.iss2.2011.199
- Davidson, K., Gowen, R.J., Harrison, P.J., Fleming, L.E., Hoagland, P., Moschonas, G., 2014.
 Anthropogenic nutrients and harmful algae in coastal waters. J. Environ. Manage. 146, 206–
 216. https://doi.org/10.1016/j.jenvman.2014.07.002
- Durando, P., Ansaldi, F., Oreste, P., Moscatelli, P., Marensi, L., Grillo, C., Gasparini, R., Icardi, G.,
 Collaborative Group for the Ligurian Syndromic Algal Surveillance, 2007. *Ostreopsis ovata*and human health: epidemiological and clinical features of respiratory syndrome outbreaks
 from a two-year syndromic surveillance, 2005-06, in north-west Italy. Euro Surveill. Bull. Eur.
 Sur Mal. Transm. Eur. Commun. Dis. Bull. 12, E070607.1.
- Faust, M.A., Morton, S.L., Quod, J.P., 1996. Further sem study of marine dinoflagellates: the genus
 Ostreopsis (dinophyceae)1. J. Phycol. 32, 1053–1065. https://doi.org/10.1111/j.0022 3646.1996.01053.x
- Fu, F., Tatters, A., Hutchins, D., 2012. Global change and the future of harmful algal blooms in the
 ocean. Mar. Ecol. Prog. Ser. 470, 207–233. https://doi.org/10.3354/meps10047
- Garces, E., 2002. Role of temporary cysts in the population dynamics of *Alexandrium* taylori
 (Dinophyceae). J. Plankton Res. 24, 681–686. https://doi.org/10.1093/plankt/24.7.681
- Giussani, V., Asnaghi, V., Pedroncini, A., Chiantore, M., 2017. Management of harmful benthic
 dinoflagellates requires targeted sampling methods and alarm thresholds. Harmful Algae 68,
 97–104. https://doi.org/10.1016/j.hal.2017.07.010
- Glibert, P., Seitzinger, S., Heil, C., Burkholder, J., Parrow, M., Codispoti, L., Kelly, V., 2005. The Role of
 Eutrophication in the Global Proliferation of Harmful Algal Blooms. Oceanography 18, 198–
 209. https://doi.org/10.5670/oceanog.2005.54
- Glibert, P.M., Icarus Allen, J., Artioli, Y., Beusen, A., Bouwman, L., Harle, J., Holmes, R., Holt, J., 2014.
 Vulnerability of coastal ecosystems to changes in harmful algal bloom distribution in
 response to climate change: projections based on model analysis. Glob. Change Biol. 20,
 3845–3858. https://doi.org/10.1111/gcb.12662
- Gobler, C.J., 2020. Climate Change and Harmful Algal Blooms: Insights and perspective. Harmful
 Algae 91, 101731. https://doi.org/10.1016/j.hal.2019.101731
- Gobler, C.J., Doherty, O.M., Hattenrath-Lehmann, T.K., Griffith, A.W., Kang, Y., Litaker, R.W., 2017.
 Ocean warming since 1982 has expanded the niche of toxic algal blooms in the North Atlantic
 and North Pacific oceans. Proc. Natl. Acad. Sci. 114, 4975–4980.
- 790 https://doi.org/10.1073/pnas.1619575114

- Granéli, E., Vidyarathna, N.K., Funari, E., Cumaranatunga, P.R.T., Scenati, R., 2011. Can increases in
 temperature stimulate blooms of the toxic benthic dinoflagellate *Ostreopsis ovata*? Harmful
 Algae 10, 165–172. https://doi.org/10.1016/j.hal.2010.09.002
- Hallegraeff, G.M., 2010. Ocean climate change, phytoplankton community responses, and harmful
 algal blooms: a formidable predictive challenge. J. Phycol. 46, 220–235.
 https://doi.org/10.1111/j.1529-8817.2010.00815.x
- Hallegraeff, G.M., Anderson, D.M., Belin, C., Bottein, M.-Y.D., Bresnan, E., Chinain, M., Enevoldsen,
 H., Iwataki, M., Karlson, B., McKenzie, C.H., Sunesen, I., Pitcher, G.C., Provoost, P.,
- 799 Richardson, A., Schweibold, L., Tester, P.A., Trainer, V.L., Yñiguez, A.T., Zingone, A., 2021.
- Perceived global increase in algal blooms is attributable to intensified monitoring and
 emerging bloom impacts. Commun. Earth Environ. 2, 117. https://doi.org/10.1038/s43247021-00178-8
- Hamilton, G., McVinish, R., Mengersen, K., 2009. Bayesian model averaging for harmful algal bloom
 prediction. Ecol. Appl. 19, 1805–1814. https://doi.org/10.1890/08-1843.1
- Heisler, J., Glibert, P.M., Burkholder, J.M., Anderson, D.M., Cochlan, W., Dennison, W.C., Dortch, Q.,
 Gobler, C.J., Heil, C.A., Humphries, E., Lewitus, A., Magnien, R., Marshall, H.G., Sellner, K.,
 Stockwell, D.A., Stoecker, D.K., Suddleson, M., 2008. Eutrophication and harmful algal
 blooms: A scientific consensus. Harmful Algae 8, 3–13.
- 809 https://doi.org/10.1016/j.hal.2008.08.006
- Hoppenrath, M., Murray, S.A., Chomérat, N., Horiguchi, T., 2014. Marine benthic dinoflagellates:
 unveiling their worldwide biodiversity, Kleine Senckenberg-Reihe. E. Schweizerbart'sche
 Verlagsbuchhandlung, Stuttgart.
- Illoul, H., Hernández, F.R., Vila, M., Adjas, N., younes, A.A., Bournissa, M., Koroghli, A., Marouf, N.,
 Rabia, S., Ameur, F.L.K., 2012. The Genus *Ostreopsis* along the Algerian Coastal Waters (SW
 Mediterranean Sea) Associated with a Human Respiratory Intoxication Episode. Cryptogam.
 Algol. 33, 209–216. https://doi.org/10.7872/crya.v33.iss2.2011.209
- Jauzein, C., Açaf, L., Accoroni, S., Asnaghi, V., Fricke, A., Hachani, M.A., abboud-Abi Saab, M.,
 Chiantore, M., Mangialajo, L., Totti, C., Zaghmouri, I., Lemée, R., 2018. Optimization of
 sampling, cell collection and counting for the monitoring of benthic harmful algal blooms:
 Application to *Ostreopsis* spp. blooms in the Mediterranean Sea. Ecol. Indic. 91, 116–127.
 https://doi.org/10.1016/j.ecolind.2018.03.089
- Jauzein, C., Couet, D., Blasco, T., Lemée, R., 2017. Uptake of dissolved inorganic and organic nitrogen
 by the benthic toxic dinoflagellate *Ostreopsis* cf. *ovata*. Harmful Algae 65, 9–18.
 https://doi.org/10.1016/j.hal.2017.04.005
- Kibler, S.R., Tester, P.A., Kunkel, K.E., Moore, S.K., Litaker, R.W., 2015. Effects of ocean warming on
 growth and distribution of dinoflagellates associated with ciguatera fish poisoning in the
 Caribbean. Ecol. Model. 316, 194–210. https://doi.org/10.1016/j.ecolmodel.2015.08.020
- Lau, W.L.S., Law, I.K., Liow, G.R., Hii, K.S., Usup, G., Lim, P.T., Leaw, C.P., 2017. Life-history stages of natural bloom populations and the bloom dynamics of a tropical Asian ribotype of *Alexandrium minutum*. Harmful Algae 70, 52–63. https://doi.org/10.1016/j.hal.2017.10.006
- Lee, B., Park, M.G., 2018. Genetic Analyses of the rbcL and psaA Genes From Single Cells
 Demonstrate a Rhodophyte Origin of the Prey in the Toxic Benthic Dinoflagellate Ostreopsis.
 Front. Mar. Sci. 5, 217. https://doi.org/10.3389/fmars.2018.00217
- Maguire, J., Cusack, C., Ruiz-Villarreal, M., Silke, J., McElligott, D., Davidson, K., 2016. Applied
 simulations and integrated modelling for the understanding of toxic and harmful algal
 blooms (ASIMUTH): Integrated HAB forecast systems for Europe's Atlantic Arc. Harmful Algae
 53, 160–166. https://doi.org/10.1016/j.hal.2015.11.006
- Mangialajo, L., Bertolotto, R., Cattaneo-Vietti, R., Chiantore, M., Grillo, C., Lemee, R., Melchiorre, N.,
 Moretto, P., Povero, P., Ruggieri, N., 2008. The toxic benthic dinoflagellate *Ostreopsis ovata*:
 Quantification of proliferation along the coastline of Genoa, Italy. Mar. Pollut. Bull. 56, 1209–
 1214. https://doi.org/10.1016/j.marpolbul.2008.02.028

- 842 Mangialajo, L., Ganzin, N., Accoroni, S., Asnaghi, V., Blanfuné, A., Cabrini, M., Cattaneo-Vietti, R., 843 Chavanon, F., Chiantore, M., Cohu, S., Costa, E., Fornasaro, D., Grossel, H., Marco-Miralles, F., 844 Masó, M., Reñé, A., Rossi, A.M., Sala, M.M., Thibaut, T., Totti, C., Vila, M., Lemée, R., 2011. 845 Trends in Ostreopsis proliferation along the Northern Mediterranean coasts. Toxicon 57, 846 408–420. https://doi.org/10.1016/j.toxicon.2010.11.019
- 847 Meroni, L., Chiantore, M., Petrillo, M., Asnaghi, V., 2018. Habitat effects on Ostreopsis cf. ovata 848 bloom dynamics. Harmful Algae 80, 64–71. https://doi.org/10.1016/j.hal.2018.09.006
- 849 Moita, M.T., Pazos, Y., Rocha, C., Nolasco, R., Oliveira, P.B., 2016. Toward predicting Dinophysis 850 blooms off NW Iberia: A decade of events. Harmful Algae 53, 17–32. 851 https://doi.org/10.1016/j.hal.2015.12.002
- Monti, M., Minocci, M., Beran, A., Iveša, L., 2007. First record of Ostreopsis cfr. ovata on macroalgae 852 853 in the Northern Adriatic Sea. Mar. Pollut. Bull. 54, 598–601. 854
 - https://doi.org/10.1016/j.marpolbul.2007.01.013
- 855 Montresor, M., Lewis, J., 2006. Phases, stages and shifts in the life cycles of marine phytoplankton. 856 DV Subba-Rao Ed Algal Cult. Analog. Blooms Appl. Sci. Publ. Enfield USA 2006 91–129.
- 857 Moore, S.K., Bill, B.D., Hay, L.R., Emenegger, J., Eldred, K.C., Greengrove, C.L., Masura, J.E., Anderson, 858 D.M., 2015. Factors regulating excystment of Alexandrium in Puget Sound, WA, USA. Harmful 859 Algae 43, 103–110. https://doi.org/10.1016/j.hal.2015.01.005
- 860 Moore, S.K., Trainer, V.L., Mantua, N.J., Parker, M.S., Laws, E.A., Backer, L.C., Fleming, L.E., 2008. Impacts of climate variability and future climate change on harmful algal blooms and human 861 862 health. Environ. Health 7, S4. https://doi.org/10.1186/1476-069X-7-S2-S4
- 863 Nakada, M., Hatayama, Y., Ishikawa, A., Ajisaka, T., Sawayama, S., Imai, I., 2018. Seasonal distribution 864 of Gambierdiscus spp. in Wakasa Bay, the Sea of Japan, and antagonistic relationships with 865 epiphytic pennate diatoms. Harmful Algae 76, 58–65.
- 866 https://doi.org/10.1016/j.hal.2018.05.002
- 867 Ní Rathaille, A., Raine, R., 2011. Seasonality in the excystment of Alexandrium minutum and 868 Alexandrium tamarense in Irish coastal waters. Harmful Algae 10, 629–635. 869 https://doi.org/10.1016/j.hal.2011.04.015
- 870 Ninčević Gladan, Ž., Arapov, J., Casabianca, S., Penna, A., Honsell, G., Brovedani, V., Pelin, M., 871 Tartaglione, L., Sosa, S., Dell'Aversano, C., Tubaro, A., Žuljević, A., Grbec, B., Čavar, M., 872 Bužančić, M., Bakrač, A., Skejić, S., 2019. Massive Occurrence of the Harmful Benthic 873 Dinoflagellate Ostreopsis cf. ovata in the Eastern Adriatic Sea. Toxins 11, 300. 874 https://doi.org/10.3390/toxins11050300
- 875 Parsons, M.L., Aligizaki, K., Bottein, M.-Y.D., Fraga, S., Morton, S.L., Penna, A., Rhodes, L., 2012. 876 Gambierdiscus and Ostreopsis: Reassessment of the state of knowledge of their taxonomy, 877 geography, ecophysiology, and toxicology. Harmful Algae 14, 107–129. 878 https://doi.org/10.1016/j.hal.2011.10.017
- 879 Pezzolesi, L., Guerrini, F., Ciminiello, P., Dell'Aversano, C., Iacovo, E.D., Fattorusso, E., Forino, M., 880 Tartaglione, L., Pistocchi, R., 2012. Influence of temperature and salinity on Ostreopsis cf. 881 ovata growth and evaluation of toxin content through HR LC-MS and biological assays. Water 882 Res. 46, 82–92. https://doi.org/10.1016/j.watres.2011.10.029
- 883 Pfannkuchen, M., Godrijan, J., Marić Pfannkuchen, D., Iveša, L., Kružić, P., Ciminiello, P., 884 Dell'Aversano, C., Dello Iacovo, E., Fattorusso, E., Forino, M., Tartaglione, L., Godrijan, M., 885 2012. Toxin-Producing Ostreopsis cf. ovata are Likely to Bloom Undetected along Coastal 886 Areas. Environ. Sci. Technol. 46, 5574–5582. https://doi.org/10.1021/es300189h
- 887 Pistocchi, R., Pezzolesi, L., Guerrini, F., Vanucci, S., Dell'Aversano, C., Fattorusso, E., 2011. A review on the effects of environmental conditions on growth and toxin production of Ostreopsis ovata. 888 889 Toxicon 57, 421–428. https://doi.org/10.1016/j.toxicon.2010.09.013
- 890 R Core Team, 2019. R: A language and environment for statistical computing. R Found. Stat. Comput. 891 Vienna Austria.
- 892 Rhodes, L., 2011. World-wide occurrence of the toxic dinoflagellate genus Ostreopsis Schmidt.
- 893 Toxicon 57, 400–407. https://doi.org/10.1016/j.toxicon.2010.05.010

894 Rigosi, A., Hanson, P., Hamilton, D.P., Hipsey, M., Rusak, J.A., Bois, J., Sparber, K., Chorus, I., 895 Watkinson, A.J., Qin, B., Kim, B., Brookes, J.D., 2015. Determining the probability of 896 cyanobacterial blooms: the application of Bayesian networks in multiple lake systems. Ecol. 897 Appl. 25, 186-199. https://doi.org/10.1890/13-1677.1 Ruiz-Villarreal, M., García-García, L.M., Cobas, M., Díaz, P.A., Reguera, B., 2016. Modelling the 898 899 hydrodynamic conditions associated with Dinophysis blooms in Galicia (NW Spain). Harmful 900 Algae 53, 40-52. https://doi.org/10.1016/j.hal.2015.12.003 901 Sansoni, G., Borghini, B., Camici, G., Casotti, M., Righini, P., Rustighi, C., 2003. Fioriture algali di 902 Ostreopsis ovata (Gonyaulacales: Dinophyceae): un problema emergente. Biol. Ambiantale 903 17, 17-23. 904 Scalco, E., Brunet, C., Marino, F., Rossi, R., Soprano, V., Zingone, A., Montresor, M., 2012. Growth and 905 toxicity responses of Mediterranean Ostreopsis cf. ovata to seasonal irradiance and 906 temperature conditions. Harmful Algae 17, 25–34. https://doi.org/10.1016/j.hal.2012.02.008 907 Shears, N.T., Ross, P.M., 2009. Blooms of benthic dinoflagellates of the genus Ostreopsis; an 908 increasing and ecologically important phenomenon on temperate reefs in New Zealand and 909 worldwide. Harmful Algae 8, 916-925. https://doi.org/10.1016/j.hal.2009.05.003 910 Tawong, W., Yoshimatsu, T., Yamaguchi, H., Adachi, M., 2015. Effects of temperature, salinity and 911 their interaction on growth of benthic dinoflagellates Ostreopsis spp. from Thailand. Harmful 912 Algae 44, 37–45. https://doi.org/10.1016/j.hal.2015.02.011 913 Taylor, F.J.R., 1979. The description of the benthic dinoflagellate associated with maitotoxin and 914 ciguatoxin, including observations on Hawaiian material. DL Taylor HH Seliger Eds Toxic 915 Dinoflag. Blooms Elsevier Sci. NY 71–77. 916 Tester, P.A., Litaker, R.W., Berdalet, E., 2020. Climate change and harmful benthic microalgae. 917 Harmful Algae 91, 101655. https://doi.org/10.1016/j.hal.2019.101655 918 Tichadou, L., Glaizal, M., Armengaud, A., Grossel, H., Lemée, R., Kantin, R., Lasalle, J.-L., Drouet, G., 919 Rambaud, L., Malfait, P., de Haro, L., 2010. Health impact of unicellular algae of the 920 Ostreopsis genus blooms in the Mediterranean Sea: experience of the French Mediterranean 921 coast surveillance network from 2006 to 2009. Clin. Toxicol. 48, 839-844. 922 https://doi.org/10.3109/15563650.2010.513687 Totti, C., Accoroni, S., Cerino, F., Cucchiari, E., Romagnoli, T., 2010. Ostreopsis ovata bloom along the 923 924 Conero Riviera (northern Adriatic Sea): Relationships with environmental conditions and 925 substrata. Harmful Algae 9, 233–239. https://doi.org/10.1016/j.hal.2009.10.006 926 Tubaro, A., Durando, P., Del Favero, G., Ansaldi, F., Icardi, G., Deeds, J.R., Sosa, S., 2011. Case 927 definitions for human poisonings postulated to palytoxins exposure. Toxicon 57, 478–495. 928 https://doi.org/10.1016/j.toxicon.2011.01.005 929 Van Dolah, F.M., 2000. Marine algal toxins: origins, health effects, and their increased occurrence. 930 Environ. Health Perspect. 108, 133–141. https://doi.org/10.1289/ehp.00108s1133 931 Vila, M., Abós-Herràndiz, R., Isern-Fontanet, J., Àlvarez, J., Berdalet, E., 2016. Establishing the link 932 between Ostreopsis cf.ovata blooms and human health impacts using ecology and 933 epidemiology. Sci. Mar. 80, 107–115. https://doi.org/10.3989/scimar.04395.08A 934 Vila, M., Garcés, E., Masó, M., 2001. Potentially toxic epiphytic dinoflagellate assemblages on 935 macroalgae in the NW Mediterranean. Aquat. Microb. Ecol. 26, 51–60. 936 https://doi.org/10.3354/ame026051 937 Vila, M., Masó, M., Sampedro, N., Illoul, H., Arin, L., Garcés, E., Giacobbe, M.G., Alvarez, J., Camp, J., 938 2008. The genus Ostreopsis in the recreational waters along the Catalan Coast and Balearic 939 Islands (NW Mediterranean Sea): are they the origin of human respiratory difficulties? Proc. 940 12th Int. Conf. Harmful Algae 2008 334–336. 941 Wells, M.L., Karlson, B., 2018. Harmful Algal Blooms in a Changing Ocean, in: Glibert, P.M., Berdalet, 942 E., Burford, M.A., Pitcher, G.C., Zhou, M. (Eds.), Global Ecology and Oceanography of Harmful 943 Algal Blooms. Springer International Publishing, Cham, pp. 77–90. 944 https://doi.org/10.1007/978-3-319-70069-4_5

- Wells, M.L., Karlson, B., Wulff, A., Kudela, R., Trick, C., Asnaghi, V., Berdalet, E., Cochlan, W.,
 Davidson, K., De Rijcke, M., Dutkiewicz, S., Hallegraeff, G., Flynn, K.J., Legrand, C., Paerl, H.,
 Silke, J., Suikkanen, S., Thompson, P., Trainer, V.L., 2020. Future HAB science: Directions and
 challenges in a changing climate. Harmful Algae 91, 101632.
- 949 https://doi.org/10.1016/j.hal.2019.101632
- Wells, M.L., Trainer, V.L., Smayda, T.J., Karlson, B.S.O., Trick, C.G., Kudela, R.M., Ishikawa, A., Bernard,
 S., Wulff, A., Anderson, D.M., Cochlan, W.P., 2015. Harmful algal blooms and climate change:
 Learning from the past and present to forecast the future. Harmful Algae 49, 68–93.
- 953 https://doi.org/10.1016/j.hal.2015.07.009
- Wood, A.M., Everroad, R.C., Wingard, L.M., 2005. Measuring growth rates in microalgal cultures.
 Algal Cult. Tech. 269–288.

956













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