
Total mineralization of 2-ethylhexyl nitrate by bacterial cocultures

Nicolau Elodie ^{1,2}, Della Giustina Gabriela ¹, Jouanneau Yves ², Marchal Rémy ^{1,*}

¹ Department of Biotechnology, Institut Français du Pétrole (IFP), 1-4 Avenue de Bois Préau, 92852, Rueil-Malmaison, France

² CEA, IRTSV, LCBM, Grenoble, France

* Corresponding author : Rémy Marchal, email address : remy.marchal@ifp.fr

Abstract :

2-Ethylhexyl nitrate (2-EHN) is a widely-used chemical which is commonly added to diesel oil to boost its cetane index. The 2-EHN molecule is recalcitrant to biodegradation but still utilized as sole carbon source by *Mycobacterium austroafricanum* IFP 2173. The incomplete degradation of 2-EHN by this strain results in the accumulation of an intermediary metabolite i.e. 4-ethylidihydrofuran-2(3H)-one (4-EDF). The study aimed at isolating 4-EDF degraders in order to achieve total mineralization of 2-EHN in cocultures with *M. austroafricanum* IFP 2173. Bacterial isolates were obtained from diesel-contaminated soil by enrichment in serial cultures supplemented with 4-EDF, the degradation of which was monitored by CO₂ measurements. Two strains were isolated and identified as *Bacillus cereus* and *Burkholderia* sp., respectively. Complete mineralization of 2-EHN was achieved by associating *M. austroafricanum* IFP 2173 with either bacterial isolate in cocultures. In the context of environmental acceptability, efficient degradation of a potentially persistent pollutant by a bacterial consortium is demonstrated.

Keywords : 2-EHN, Biodegradability, Cocultures, Cetane index

47
48 **INTRODUCTION**
49

50 2-Ethyhexyl nitrate (2-EHN) is added at 0.05 % to 0.4 % to diesel oil to improve ignition of
51
52 fuel oil in diesel engines (Guibet 1999 ; Bornemann *et al.* 2002). The annual production of 2-
53
54 EHN associated with the worldwide consumption of diesel oil is about 100 000 tons.
55

56
57 Considering the widespread market of 2-EHN and the-risks of accidental release, the question
58
59 of its biodegradability is relevant from an environmental standpoint. The degree of persistence
60

1
2
3 of contaminants in natural environment is influenced by various factors including their
4
5 chemical structure, the presence of microorganisms able to degrade them, and environmental
6
7 conditions suitable for microbial biodegradation activities (Philp *et al.* 2005).
8
9

10 The biodegradability of chemicals is usually assessed using tests conducted aerobically in
11
12 liquid cultures (OECD 1993). A substance is considered readily biodegradable if the level of
13
14 degradation reaches 70 % in terms of chemical oxygen demand or 60 % in terms of biological
15
16 oxygen demand over a period not exceeding 28 days. A substance may be not readily but
17
18 nevertheless inherently biodegradable, if complete degradation is obtained in a special-
19
20 purpose "inherent test" of high degrading power (Pagga 1997).
21
22

23
24 2-EHN was found not readily biodegradable according to the American Chemistry Council
25
26 Petroleum Additives Panel (Health, Environmental, and Regulatory Task Group, 2006).
27
28 However, 2-EHN was recently shown to be degraded using particular microbial populations
29
30 originating from a treatment plant supplied with wastes from petroleum refineries (Solano-
31
32 Serena *et al.* 2009). Moreover, *Myco. austroafricanum* IFP 2173 was found to partially
33
34 degrade 2-EHN with concomitant accumulation of 4-ethylidihydrofuran-2(3H)-one (4-EDF)
35
36 (Nicolau *et al.* 2008). This six-carbon lactone was not further metabolized by strain IFP 2173.
37
38 The aim of this study was to isolate microorganisms capable of utilizing 4-EDF as sole source
39
40 of carbon, then associate such isolates with *Myco. austroafricanum* IFP 2173 as a means to
41
42 achieve total mineralization of 2-EHN.
43
44
45
46
47
48
49

50 MATERIALS AND METHODS

51 Strain isolation and liquid cultures

52
53 The soil sample used for strain isolation originated from the upper layer of a diesel-polluted
54
55 site. Its moisture content was 27% according to the norm ISO11465; pH was 6.0 according to
56
57 the norm ISO10390; diesel oil content was 10 g kg⁻¹ of soil (dry weight).
58
59
60

1
2
3 Cultures on 4-EDF were performed in 15-ml test tubes closed with Teflon-coated stoppers
4 and sealed with aluminium caps. Unless otherwise indicated, 1 μ l of 4-EDF was used as
5 carbon source in 2 ml of mineral salts medium (Bouchez *et al.* 1995). Tubes were incubated at
6 30°C with shaking (150 rpm). Biodegradation was monitored at time intervals by analysis of
7 the CO₂ evolved in headspace. Measurement of endogenous respiration was performed
8 similarly in control tubes with no added 4-EDF. Experiments were carried out in duplicates
9 and abiotic controls supplemented with mercuric chloride (0.2 g l⁻¹) were run under similar
10 conditions.
11

12 Cultures on 2-EHN were performed in 120-ml serum bottles as previously described (Nicolau
13 *et al.* 2008). At the end of the incubation period, 10 ml of MTBE (methyl *tert*-butyl ether) was
14 introduced into the flasks. After overnight extraction at 4°C, residual 2-EHN was analysed.
15 CO₂ production by the cultures was quantified after acidification to pH 2 with H₃PO₄. The net
16 amount of CO₂ produced was calculated as the difference between the amount of CO₂ found
17 in test flasks and that found in the control flasks without 2-EHN. The mineralization yield was
18 calculated as the carbon ratio between the net CO₂ produced and the theoretical amount of
19 CO₂ generated from complete oxidation of the consumed 2-EHN.
20

21 CO₂, 2-EHN and 4-EDF were quantified by gas chromatography (GC) as previously described
22 (Nicolau *et al.* 2008).
23

24 **Nucleic acid extraction and phylogenetic analyses**

25 Genomic DNA of isolates was extracted from the bacterial pellet of a 5-ml Tween 80-grown
26 culture (Solano-Serena *et al.* 2004).
27

28 16S ribosomal DNA was PCR-amplified using primer F8 (5'-
29 AGAGTTTGATYMTGGCTCAG-3'), 1492R (5'-CGGTTACCTTGTTACGACCT-3')
30 (Grabowski *et al.* 2005). Amplicons were cloned in pCR2.1 TOPO vector (Invitrogen), then
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

1
2
3 sequenced on both strands. Strains were identified using Blast on the NCBI server
4
5 (http://www.ncbi.nlm.nih.gov/BLAST/).
6

7
8 16S DNA sequences were aligned with the CLUSTALX program. Trees were constructed
9
10 with neighbour-joining algorithms, gaps being removed from the analysis.
11

12 **Chemicals**

13
14 2-EHN (CAS Number 27247-96-7) was obtained from Sigma Aldrich. Mineral salts were
15
16 from VWR (France). 4-EDF was prepared by chemical synthesis from furane-2(5H)-one
17
18 according to the procedure described by Alexakis *et al.* (1997).
19
20
21
22
23

24 **RESULTS**

25 **Strain selection on 4-ethyl-dihydrofuran-2-(3H)-one**

26
27 In order to isolate microorganisms endowed with ability to degrade 4-EDF , enrichments were
28
29 performed in liquid cultures using a sample of diesel-contaminated soil as starting inoculum.
30
31 Ten mg of soil were initially incubated in liquid medium with 0.5 g l⁻¹ 4-EDF as substrate.
32
33 Bacterial growth was monitored by CO₂ measurement in the headspace of the cultures (Figure
34
35 1). CO₂ evolution became detectable after six days of incubation and stopped after 15 days.
36
37 Two subcultures were successively performed using a 21 day-old culture as inoculum (10 %
38
39 v/v). Figure 1 shows that lag phase before start of growth was shortened in subcultures
40
41 compared to the initial culture.
42
43
44
45
46
47

48 Bacterial isolation from subcultures was performed by plating diluted samples on agar plates.
49
50 Two morphological types of isolates were obtained on agar plates, designated S1 and S2. In
51
52 both cases, colonies were white in colour and showed round and regular shape. They were
53
54 identified by their 16S ribosomal DNA sequence as members of *Bacillus cereus* (strain S1)
55
56 and *Burkholderia* sp. (strain S2), respectively.
57
58
59
60

1
2
3 **2-EHN biodegradation in cocultures of bacterial isolates and *Myco. austroafricanum* IFP**
4
5 **2173**
6

7
8 *Bacillus cereus* S1 and *Burkholderia* sp. S2 were incubated either separately or in
9
10 combination with *Myco. austroafricanum* IFP 2173 in mineral salts medium containing 2-
11
12 EHN. CO₂ evolution in the culture headspace was monitored over 17 days (Figure 2). As
13
14 expected, a moderate CO₂ production was observed with *Myco. austroafricanum* IFP 2173
15
16 alone, corresponding to the partial mineralization of the substrate. A much higher level of
17
18 CO₂ evolution was observed in cocultures where strain IFP 2173 was associated with either
19
20 *Bacillus cereus* S1 or *Burkholderia* sp. S2. Over the same period of time, pure cultures of
21
22 *Bacillus cereus* S1 and *Burkholderia* sp. S2 were unable to produce significant amounts of
23
24 *Bacillus cereus* S1 and *Burkholderia* sp. S2 were unable to produce significant amounts of
25
26 CO₂ at the expense of 2-EHN suggesting that they did not utilize this compound as carbon
27
28 source.
29

30
31 Quantitative determination of 2-EHN in cultures by the end of the incubation allowed to
32
33 calculate mineralization yields (Table 1).
34

35
36
37
38 In abiotic controls, high recovery rates (> 85%) were obtained, indicating that substrate losses
39
40 were low over the experimental time period. *Myco. austroafricanum* IFP 2173 utilized 2-EHN
41
42 extensively but substrate conversion reached only 23%, due to accumulation of 4-EDF (data
43
44 not shown). High biodegradation rates and mineralization levels were obtained in cocultures
45
46 of strain IFP 2173 with *Bacillus cereus* S1 or *Burkholderia* sp. S2, and no accumulation of 4-
47
48 EDF was observed. In pure cultures of *Bacillus cereus* S1 and *Burkholderia* sp. S2, 2-EHN
49
50 was not significantly degraded and a low level of CO₂ probably arising from respiration on
51
52 endogenous substrate was detected.
53
54
55
56

57
58
59 **DICUSSION**
60

1
2
3 We isolated two strains able to degrade 4-EDF, a by-product that is released by *Myco.*
4
5 *austrofricanum* IFP 2173 when grown on 2-EHN. The isolation procedure was easy and
6
7 yielded two phylogenetically unrelated bacterial strains suggesting that the microbial ability to
8
9 utilize lactones like 4-EDF is widespread in soil. Since many lactones have negative impact
10
11 on human health, the biodegradability of 4-EDF by *Bacillus cereus* S1 and *Burkholderia* sp.
12
13 S2 may reduce health hazard associated with the use of 2-EHN and improve the
14
15 environmental acceptability of this chemical.
16
17

18
19 The concerted attack on 2-EHN in cocultures exemplifies the extended degradation
20
21 capabilities of bacterial associations compared to individual microorganisms. This finding has
22
23 important consequences in terms of environmental safety. In case of accidental release of 2-
24
25 EHN in the environment, it may be expected that the chemical be degraded through natural
26
27 attenuation involving commensalism processes, provided that adequate indigenous bacterial
28
29 species co-exist on polluted sites.
30
31
32
33
34
35

36 REFERENCES

- 37
38 Alexakis, A., Vastra, J. and Mangeney, P. (1997) Acceleration of the conjugate addition of
39
40 diethyl zinc to enones by either Cu(OTf)₂ or trivalent phosphorus ligands.
41
42 *Tetrahedron Letters*, **38**, 7745-7748.
43
44
45 American (the) Chemistry Council Petroleum Additives Panel; Health, E., and Regulatory
46
47 Task Group. (2006). High production volume (HPV) Challenge program; final
48
49 submission for Nitric acid, 2-Ethylhexyl ester
50
51
52 Bornemann, H., Scheidt, F. and Sander, W. (2002) Thermal decomposition of 2-ethylhexyl
53
54 nitrate (2-EHN). *J Chem Kinet*, **34**, 34-38.
55
56
57
58
59
60

- 1
2
3 Bouchez, M., Blanchet, D. and Vandecasteele, J.-P. (1995) Degradation of polycyclic aromatic
4 hydrocarbons by pure strains and by defined strains associations: inhibition
5
6 phenomena and cometabolism. *Appl Microbiol Biotechnol*, **43**, 156-164.
7
8
9
- 10 Guibet, J. C. and Faure, E. (1999) *Fuels and Engines: technology, energy, environment*,
11 Technip editions, Paris.
12
13
14
- 15 Nicolau, E., Kerhoas, L., Lettere, M., Jouanneau, Y. and Marchal, R. (2008) Biodegradation
16 of 2-ethylhexyl nitrate by *Mycobacterium austroafricanum* IFP 2173. *Appl Environ*
17 *Microbiol*, **74**, 6187-93.
18
19
20
21
- 22 OECD (1993) *Guidelines for the testing of chemicals*, OECD; Part 3. Paris .
23
24
25
- 26 Pagga, U. (1997) Testing biodegradability with standardized methods. *Chemosphere*, **35**,
27 2953-2972.
28
29
30
- 31 Philp, J. C., Bamforth, S. M., Singleton, I. and Atlas, R. M. (2005) Environmental pollution
32 and restauration: a role for bioremediation. In *Bioremediation: applied microbial*
33 *solutions for real-world environmental cleanup*(Eds, Atlas, R. M. and Philp, J. C.)
34 ASM Press, Washington, DC, pp. 1-48.
35
36
37
38
- 39 Solano-Serena, F., Marchal, R., Heiss, S. and Vandecasteele, J. P. (2004) Degradation of
40 isooctane by *Mycobacterium austroafricanum* IFP 2173: growth and catabolic
41 pathway. *J Appl Microbiol*, **97**, 629-39.
42
43
44
45
- 46 Solano-Serena, F., Marchal, R., Lebeault, J. M. and Vandecasteele, J.-P. Distribution in the
47 environment of degradative capacities for gasoline attenuation. (2000)
48
49
50
51
52
- 53 Solano-Serena, F., Nicolau, E., Favreau, G., Jouanneau, Y. and Marchal, R. (2009)
54
55 Biodegradability of 2-ethylhexyl nitrate (2-EHN), a cetane improver of diesel oil.
56
57
58
59
60

1
2
3 Grabowski, A., Nercessian, O., Fayolle, F., Blanchet, D. and Jeanthon, C. (2005) Microbial
4
5 diversity in production waters of a low-temperature biodegraded reservoir. *FEMS*
6
7
8 *Microbiol Ecol*, 54, 427-43.
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

For Peer Review

1
2
3 **TABLE**
4
5
6
7

8 Table 1 : Mineralization rates of 2-EHN in separate cultures of *Myco. austroafricanum* IFP
9
10 2173, *B. cereus* S1 and *Burkholderia* sp. S2, and in combinations.
11
12

Strains	Substrate	Substrate	Mineralization yield (%)*
	recovery in	degradation in test	
	abiotic flasks	flasks	
	(%)	(%)	
<i>B. cereus</i> S1	97	9,1	2,1
<i>Burkholderia</i> sp. S2	90	2,1	7,2
<i>Myco. austroafricanum</i> IFP 2173	95	> 99.5	23.1
S1 + IFP 2173	85	99,4	100
S2 + IFP 2173	89	99,1	95,4

13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35 * Mineralization yields were calculated with respect to the substrate consumed
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54
55
56
57
58
59
60

FIGURES

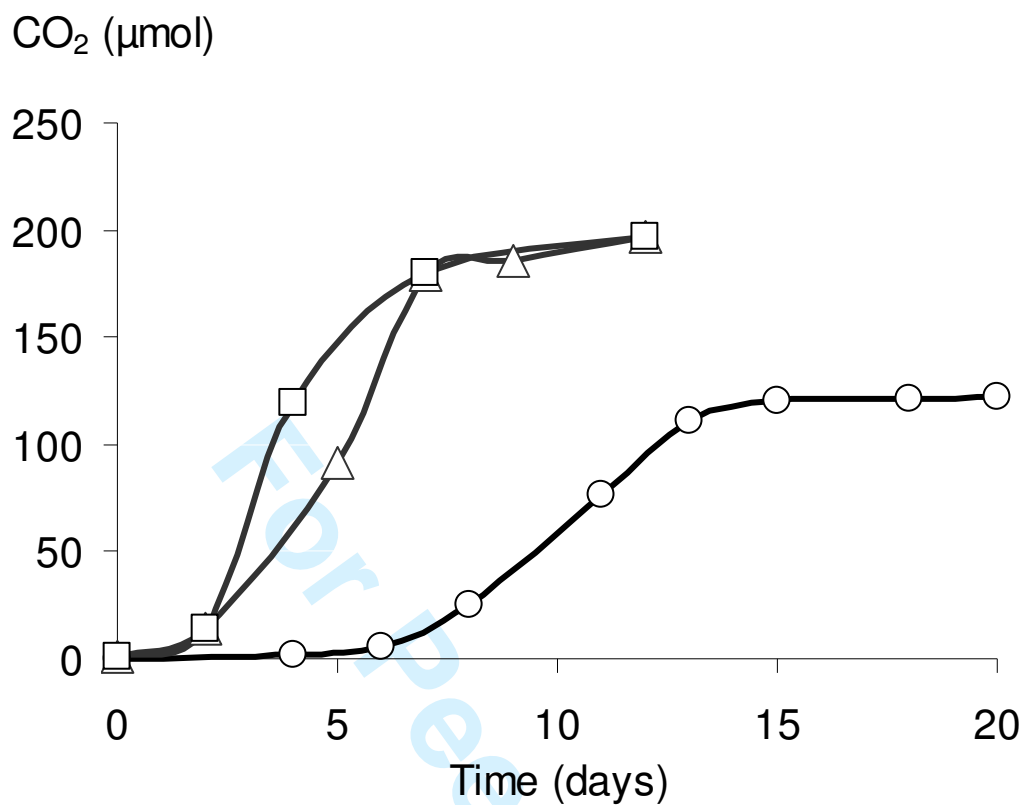


Figure 1: Mineralization curves of successive enrichment cultures grown on 4-EDF.

○ initial culture, △ first subculture and □ second subculture. Strains S1 and S2 were isolated from the first and second subcultures, respectively.

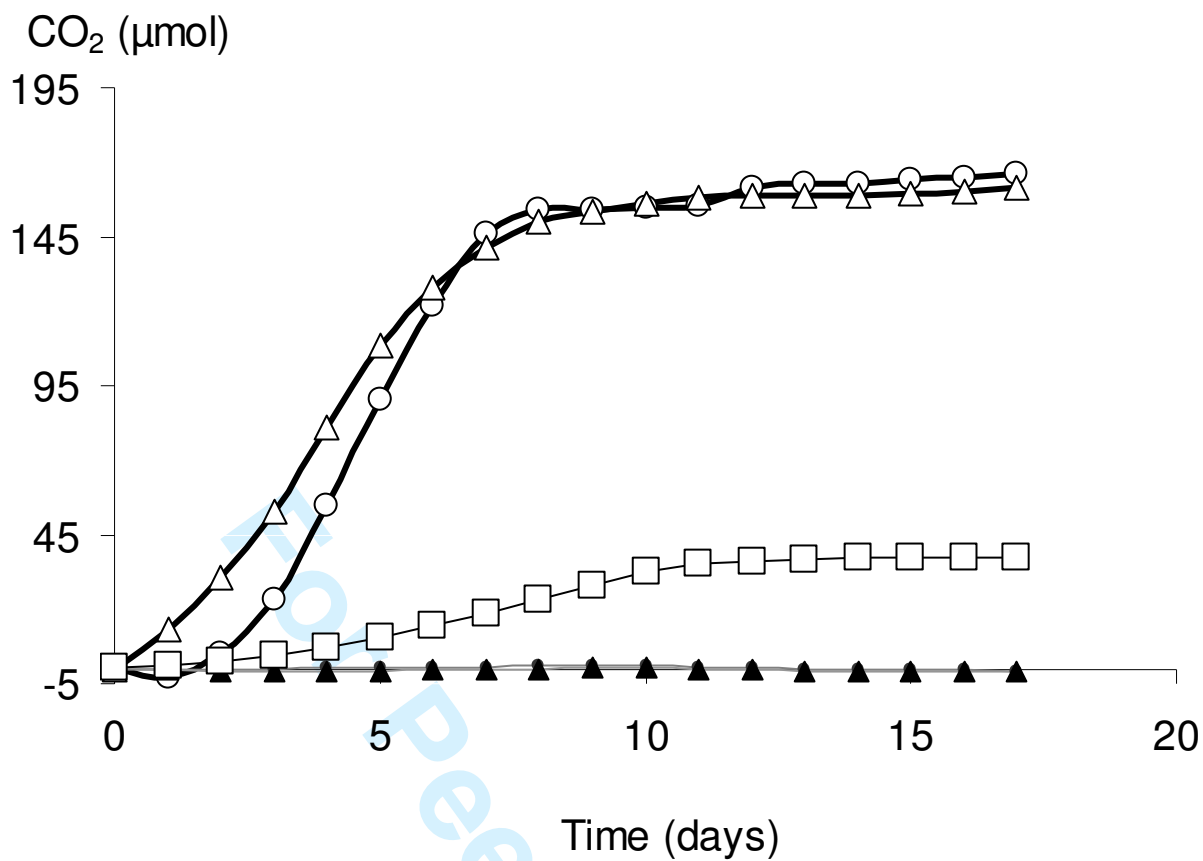


Figure 2 : Mineralization of 2-EHN in pure cultures and combinations with strain IFP2173.

○ Strain S1 with *Myco. austroafricanum* IFP 2173; △ strain S2 with *Myco. austroafricanum* IFP 2173; □ *Myco. austroafricanum* IFP 2173 alone; controls of strains S1 ● and S2 ▲, alone.