Sudden peak in tetrodotoxin in French oysters during the summer of 2021: Source investigation using microscopy, metabarcoding and droplet digital PCR

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Abstract :

Tetrodotoxin (TTX) is a potent neurotoxin causing human intoxications from contaminated seafood worldwide and is of emerging concern in Europe. Shellfish have been shown to contain varying TTX concentrations globally, with concentrations typically higher in Pacific oysters Crassostrea gigas in Europe. Despite many decades of research, the source of TTX remains unknown, with bacterial or algal origins having been suggested. The aim of this study was to identify potential source organisms causing TTX contamination in Pacific ovsters in French coastal waters, using three different techniques. Ovsters were deployed in cages from April to September 2021 in an estuary where TTX was previously detected. Microscopic analyses of water samples were used to investigate potential microalgal blooms present prior or during the peak in TTX. Differences in the bacterial communities from ovster digestive glands (DG) and remaining flesh were explored using metabarcoding, and lastly, droplet digital PCR assays were developed to investigate the presence of Cephalothrix sp., one European TTX-bearing species in the DG of toxic C. gigas. Oysters analysed by liquid chromatography-tandem mass spectrometry contained quantifiable levels of TTX over a three-week period (24 June-15 July 2021), with concentrations decreasing in the DG from 424 µg/kg for the first detection to 101 µg/kg (equivalent to 74 to 17 µg/kg of total flesh), and trace levels being detected until August 13, 2021. These concentrations are the first report of the European TTX guidance levels being exceeded in French shellfish. Microscopy revealed that some microalgae bloomed during the TTX peak, (e.g., Chaetoceros spp., reaching 40,000 cells/L). Prokarvotic metabarcoding showed increases in abundance of Rubritaleaceae (genus Persicirhabdus) and Neolyngbya, before and during the TTX peak. Both phyla have previously been described as possible TTX-producers and should be investigated further. Droplet digital PCR analyses were negative for the targeted TTX-bearing genus Cephalothrix.

Graphical abstract



Highlights

► Tetrodotoxin exceeded European guidance value in French oysters (74 µg/kg whole flesh). ► Microscopic analyses suggested co-occurrence of TTX with *Chaetoceros* spp. ► Metabarcoding detected the bacterial family Rubritaleaceae and cyanobacteria *Neolyngbya* spp. before and during the TTX peak.

Keywords : Biotoxin, Bivalves, Cyanobacteria, ddPCR, Metabarcoding, Microalgae, Microbiome

49 **1. Introduction**

50 Tetrodotoxin (TTX) is one of the most potent naturally occurring marine toxins and ingestion of TTX-contaminated seafood results in numerous, sometimes fatal, human intoxications every 51 year (Bagnis et al., 1970; Hwang and Noguchi, 2007; Katikou et al., 2022). TTX has a 52 surprisingly broad distribution across highly diverged marine, freshwater and terrestrial 53 taxonomic groups, including 14 classes of eukaryotes and five bacteria (Bane et al., 2014; Biessy 54 et al., 2019; Chau et al., 2013). Because of its widespread occurrence in unrelated taxa, there is 55 considerable debate about its ecological role and potential source organisms (Williams, 2010). 56 57 TTX has been reported in at least 21 species of edible bivalves and gastropods from ten countries since the 1980's (Biessy et al., 2019). No regulatory limit for TTX in bivalves exists despite a 58 scientific opinion published in 2017 by the European Food Safety Authority (EFSA), 59 recommending a concentration below 44 µg TTX equivalent per kg of shellfish meat, based on 60 a 400 g portion size (Knutsen et al., 2017). In the last decade, many European countries have 61 reported the occurrence of TTXs in bivalve mollusks, with most confirmations from the 62 Netherlands, Spain, Italy, England, Greece and France (Blanco et al., 2019; Bordin et al., 2021; 63 64 Dell'Aversano et al., 2019; Gerssen et al., 2018; Hort et al., 2020; Leão et al., 2018; Réveillon et al., 2021; Turner et al., 2015; Turner et al., 2017; Vlamis et al., 2015). In France, TTX has been 65 66 found in mussels (Mytilus edulis and M. galloprovincialis) and common whelk (Buccinum undatum) from the Mediterranean and Atlantic coasts (Hort et al., 2020). Oysters (Crassostrea 67 68 gigas) and cockles (*Ruditapes philippinarum*) from four sites across the Normandy and Brittany regions also tested positive for TTX with concentrations reaching up to 32 µg/kg and events 69 lasting a maximum of two weeks (Réveillon et al., 2021). Even though temperature is suggested 70 to play a role in TTX occurrence in bivalves across the United Kingdom (Dhanji-Rapkova et al., 71 2023), as would be expected for a bacterial origin, this trend has not been confirmed at the 72 European scale, with lower or similar TTX concentrations in shellfish reported from warmer 73 waters (e.g., Greece and Italy; (Dell'Aversano et al., 2019; Vlamis et al., 2015). 74

To manage the risk associated with TTX in seafood, it is important to understand its origin, how it accumulates and is transported between organisms. The two most common hypotheses regarding the origin of TTX are that 1) it is produced by symbiotic bacteria (endogenous) or that 2) it is exogenously accumulated through the diet (e.g., bacteria, microalgae) (Chau et al., 2011). Exogenous or symbiotic bacteria are commonly suggested as the most likely TTX-producers, with a wide diversity of species and strains implicated. There are reports in the literature of at

least 150 TTX-producing bacterial strains (Katikou, 2019; Magarlamov et al., 2017). Vibrio, 81 Bacillus, Pseudomonas, Actinomyces and Micrococcus are the most commonly reported bacterial 82 genera associated with TTX production, however, only one study has claimed to successfully 83 maintain TTX-producing bacteria in culture (Melnikova et al., 2019). When bacterial cultures 84 reported to produce TTX were analysed using modern highly-specific analytical methods (e.g., 85 liquid-chromatography with tandem mass spectrometry), no evidence of TTX production was 86 found and potentially indicated contamination from the starting material (Biessy et al., 2019; 87 Katikou et al., 2022). There has been evidence of production of TTX-like substance by 88 89 microalgae in one study suggesting that the dinoflagellate Alexandrium tamarense produced TTX (Kodama et al., 1996) and more recently, the presence of high concentrations of the dinoflagellate 90 Prorocentrum minimum (now P. cordatum) was potentially linked to TTX accumulation in Greek 91 shellfish (Vlamis et al., 2015). These links between microalgae and TTX production have not 92 been confirmed to date and despite sparse evidence on microalgal origin, there is a rationale 93 behind the role of microplankton in TTX accumulation in bivalves, in particular with the 94 preferential accumulation of TTX in the digestive systems of marine organisms (Dhanji-Rapkova 95 et al., 2021). This alone forms a basis for studying microalgal and microbial communities in 96 water and the digestive glands of shellfish before, during and after periods of TTX accumulation. 97

98 In 2009, *Paphies australis*, an endemic and abundant clam regularly harvested for food in 99 New Zealand, was reported to accumulate up to 800 µg TTX/kg (McNabb et al., 2014). Recent research investigated the source of TTX in P. australis (Biessy, 2021) and metabarcoding was 100 used to explore the microbiome of P. australis temporally and spatially. The results highlighted 101 a new lead in the search for TTX producers: marine cyanobacteria. Cyanobacteria, especially 102 picocyanobacteria from the Cyanobium, Synechococcus, Pleurocapsa, and Prochlorococcus 103 104 genera were present in the digestive glands and siphons of P. australis containing TTX but were absent from the organs of non-toxic clams (Biessy et al., 2020). Multiple researchers have 105 106 explored the source of TTX at higher trophic levels over the last decades (Noguchi et al., 2006) and it has been hypothesized that different organisms accumulate TTX through the food web 107 (Matsumura, 1998). Some examples include the trumpet shell Charonia sauliae that accumulated 108 109 TTX by ingesting toxic starfish (Noguchi et al., 1982), or the sea slug Pleurobranchaea maculata that likely accumulated TTX from a TTX-containing marine flatworm Stylochoplana sp. found 110 in its stomach (Salvitti et al., 2015), or the non-toxic pufferfish Takifugu obscurus that 111 accumulated TTX in their liver and skin feeding on wild TTX-containing gastropods (Zhang et 112 al., 2020). The accumulation process can be fast, as for the sea slug P. maculata maintained in 113

aquaria that accumulated TTX within one hour of being fed a prepared TTX-containing food 114 source (Khor et al., 2013). Interestingly, Okabe et al. (2019) found DNA from the toxic flatworm 115 Planocera multitentaculata in the intestine of the pufferfish T. niphobles. Further investigation 116 showed that the pufferfish were accumulating TTX from consuming the flatworms' highly toxic 117 eggs. Vlasenko and Magarlamov (2020) recently found high concentrations of TTX in the 118 intestine of the marine nemertean Cephalothrix cf. simula and Malykin et al. (2023) demonstrated 119 that during the spawning season of Cephalothrix species, their eggs can be source of TTX for 120 filter-feeding shellfish. Lastly, high TTX concentrations were detected in the midgut gland of the 121 122 Japanese scallop Chlamys farreri subsp. akazara and the source was likely the larvae of the flatworm, P. multitentaculata (Yasukawa et al., 2023). This is of importance as it highlights that 123 TTX-bearing eggs from TTX-bearing organisms can be filtered and be the source of TTX 124 accumulation in shellfish. In Europe, one TTX-bearing invasive species that has been detected 125 in the last decade is the ribbon worm Cephalothrix simula (Turner et al., 2018). This species is a 126 motile and commonly reported aquatic TTX-bearing species in Europe. Therefore, a detection 127 assay was developed to assess if potentially toxic eggs of C. simula could accumulate in filter-128 129 feeding shellfish, resulting in the accumulation of TTX in the French oysters.

To date, TTX concentrations in French shellfish were found to peak for a few weeks in 130 early summer (Réveillon et al., 2021) but data remain sparse for determining environmental 131 132 dynamics due to low frequency of the national surveillance of emerging toxins (Amzil et al., 2023; Hort et al., 2020). However, the reasons for TTX variability over time and its source 133 organisms in shellfish remain unknown. The overarching goal of the present study was to 134 investigate the source of TTX in a French shellfish production site. Non-toxic adult Pacific 135 oysters, Crassostrea gigas, purchased from a farm, were placed in cages in the Loperhet estuary 136 where a previous peak of TTX had been observed. The approach of the study was to link TTX-137 concentrations observed in the oysters by liquid chromatography coupled to tandem mass 138 spectrometry (LC-MS/MS) to potential source organisms by: 1) characterising the eukaryotic 139 microalgae in the surrounding waters and the bacterial communities in different organs of 140 C. gigas, sourced weekly or fortnightly from the targeted estuary, using microalgal cell counts 141 142 and metabarcoding techniques respectively; and to 2) develop a targeted assay to detect potential sources of TTX in C. gigas. In this case, a droplet digital PCR assay (Hindson et al., 2011) was 143 developed to detect genetic material from Cephalothrix sp. 144

146 **2. Material and Methods**

147 **2.1.** Field experimental set-up and sampling

Triploid Pacific ovsters Crassostrea gigas (n = 91) grown from spat and of similar size were 148 purchased from an oyster farm (Vrennig, Aber Wrach, France), and were placed in oyster bags 149 at Loperhet, France (48.34388889, -4.300555556; Figure 1) on April 1, 2021. Oysters were 150 harvested 13 times, fortnightly from 11 May 2021, weekly in June and July 2021 (except on the 151 18th of June when the tides did not allow for sampling) and fortnightly in August and September 152 2021. At each sampling point, C. gigas (n = 6) were harvested for metabarcoding analysis (n = 6)153 3) and LC-MS/MS analysis of TTX analogues (3 individuals were pooled to have n = 1 LC-154 MS/MS analysis per sample). Immediately upon arrival in the laboratory, three individuals were 155 156 aseptically dissected for genetic and toxin analysis into digestive glands (DG, including the intestine) and remaining flesh (RF), which were placed into individual five mL sterile screw cap 157 tubes that were filled with RNAlaterTM (ThermoFisher, MA, USA) and kept cold (4 °C) until 158 159 further analysis.

Salinity and water temperature were recorded as part of the REPHY programme (2022) atthe "Pointe du Château" site (Figure 1).



Figure 1 | Map of sampling sites. Loperhet = estuary where oysters were placed for the study;
 Pointe du Château = site where water temperature and salinity were recorded; Lanvéoc large =
 site where water samples were collected for microscopy analysis.

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2.2. Toxin extractions and LC-MS/MS analysis

Extraction of TTXs from shellfish was performed following the method described in Réveillon et al. (2021). For each oyster sample (n = 3), DG and RF were extracted separately. Briefly, acetic acid (500 μ L of 1% Hipersolv CHROMANORM for LC-MS, AventorTM, UK) was added to *C. gigas* (200 mg) before grinding with glass beads (Ø 150–250 μ m, VWR) in a mixer mill (10 min, 30 Hz; Retsch MM400 mixer Mill, ThermoFisher). After centrifugation (15,000 × g, 10 min), 250 μ L of the supernatant was subsequently collected and filtered with a 3 kDa cut-off (Pall Corporation, NY, USA) before analysis.

LC-MS/MS analyses were performed using an Ultra-Fast Liquid Chromatography system 175 (Nexera, Shimadzu, Kyoto, Japan) coupled to a triple-quadrupole mass spectrometer (5500 176 QTRAP, Sciex, MA, USA). The parameters of both LC and MS systems were the same as 177 178 described in Réveillon et al. (2021), as were the limits of detection (LOD) and quantification (LOQ) (5 and 12.5 µg/kg of total flesh). Matrix effect and extraction recovery were defined to 179 estimate the total yield (matrix effect x extraction recovery) that was used to correct TTX 180 181 concentration in DG (yield 74%, corrective factor of 1.35) and RF (yield 129%, corrective factor 0.775). 182

Concentrations of TTX obtained in DG and RF were converted into μ g/kg of total flesh equivalent by using the percentage of DG and RF mass over the total flesh in addition to the concentration of TTX obtained in each tissue (for example: DG of the oyster sample of June 24 represented 14.2% of total mass and had a concentration of 424 μ g/kg while RF represented 85.8% and had a concentration of 16 μ g/kg, corresponding finally to (424.3 x 0.142) + (16 x (0.858) = 60 + 14 = 74 μ g/kg of total flesh equivalent).

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2.3. Microscopy analysis of microalgae in plankton

Water samples were collected at the site 'Lanvéoc large' (48.309405, -4.44852, Figure 1)
using a 5-L Niskin bottle at the subsurface (0–1 m) for light microscopy counts of microalgae.
Samples were fixed with acidic Lugol's Iodine solution (0.1% final concentration) (Andersen and
Throndsen, 2003) and stored at 4 °C until examination. Total assemblages were counted using
an inverted microscope (Zeiss, Axio Observer) following REPHY guidelines (REPHY, 2022).
Samples were gently homogenized before settling in 10 mL sub-sample for > 12 h in counting

chambers (Utermohl, 1958) (Hydro-Bios, Alenholz, Germany). The lower limit of quantification
was 100 cells/L. Data is publicly available (REPHY, 2022).

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199 **2.4.** Molecular analysis

200 2.4.1. DNA extractions, PCR and metabarcoding

Dissected organs were removed from RNAlater[™] and were individually placed in the tube 201 containing bashing beads of the DNA extraction kit (DNeasy Powersoil Pro kit, Qiagen, MD, 202 USA), using sterile technique. The DNA was then extracted following the manufacturer's 203 instructions using an automated homogenizer (1600 MiniG Automated Tissue Homogenizer and 204 Cell Lyser, SPEX SamplePrep, NJ, USA) and a robotic workstation for DNA extraction 205 206 (QIAcube, Qiagen). Negative extraction controls were performed every 23 samples. PCR amplification was undertaken targeting the V3 and V4 regions of the 16S ribosomal RNA 207 208 (rRNA) gene using the primer sets as described in Klindworth et al. (2013): Bact341F- 5'- CCT ACG GGN GGC WGC AG-3' and Bact785R- 5'-GAC TAC HVG GGT ATC TAA TCC-3'. 209 Primer sets were modified to include IlluminaTM overhang adaptors following the dual-indexing 210 method from Kozich et al. (2013). 211

PCR reactions were undertaken in triplicate with 450 nM of each primer, 25 µL of 2X 212 MyFiTM Mix (Bioline, UK), ca. 5 ng of DNA, and sterile water for a total reaction volume of 50 213 µL. The cycling conditions were: 95 °C for 5 min, followed by 30 cycles of 94 °C for 30 s, 54 214 °C for 30 s, 72 °C for 45 s, and a final extension of 72 °C for 7 min. Triplicates of PCR products 215 were pooled and visualized on 1.5% agarose gel with Red Safe[™] DNA Loading Dye (Herogen 216 217 Biotech, USA) and UV illumination. PCR negatives (RNA/DNA-free water Life Technologies) were run to assess for contamination during the PCR steps. The PCR products were purified, 218 219 cleaned of primer dimers and normalized using SequalPrep Normalisation plate (ThermoFisher, MA, USA), and submitted to Auckland Genomics (University of Auckland, New Zealand) for 220 221 library preparation. Sequencing adapters and sample-specific indices were added to each amplicon via a second round of PCR using the NexteraTM Index kit (Illumina Inc., USA). 222 223 Amplicons were pooled into a single library and paired-end sequences $(2 \times 250 \text{ bp})$ generated on a MiSeq® instrument. The sequencing libraries were prepared following the Illumina 16S 224 225 Metagenomics Library Prep manual with the exception that after the indexing PCR, five µL of each sample (including three water samples acting as sequencing blank) were pooled and a single 226 227 clean-up was undertaken on the pool instead of samples being individually cleaned. Quality

control was undertaken using a bioanalyzer before the library was diluted to 4 nM and denatured. 228 A 15% PhiX spike was used, and the final loading concentration was 7 pM. Sequence data were 229 automatically demultiplexed using MiSeq® Reporter (version 2, Illumina Inc.), and forward and 230 reverse reads assigned to samples. Raw sequence reads were deposited in the National Centre for 231 Biotechnology Information (NCBI) short read archive under the accession number 232 233 PRJNA1029659 (reviewer link: https://dataview.ncbi.nlm.nih.gov/object/PRJNA1029659?reviewer=eb944as8eah290qvflpdcijc 234

235 <u>fn</u>).

236 2.4.2. Bioinformatics

Raw reads were processed, subsequent to primers being removed with cutadapt (Martin, 237 2011), using the DADA2 package (Callahan et al., 2016) within R. Reads were truncated to 228 238 and 230 bp and filtered with a maxEE (maximum number of "expected errors") of 2 and 4 for 239 forward and reverse reads respectively (reads not reaching this threshold were discarded). 240 DADA2 constructs a parametric error matrix (based on the first 10^8 bps in the dataset), the 241 samples are dereplicated and sequence variants for the forward and reverse reads are inferred 242 based on the derived error profiles from the samples. Singletons observed in the inference step 243 were discarded. Subsequently, paired-end reads were merged with a maximum mismatch of 1 bp 244 and a required minimum overlap of 10 bp. Forward and reverse reads, which did not merge were 245 not included in further analysis. Chimeras were removed using the function 246 247 removeBimeraDenovo. The resulting chimera-checked, merged Amplicon Sequence Variants 248 (ASV) were classified taxonomically using the rdp classifier (Wang, Garrity, Tiedje, & Cole, 2007) with a bootstrap of 70 against the Cyanoseq (Lefler et al., 2023) version of the SILVA 249 250 v138 reference database (Pruesse et al., 2007). This database updated the taxonomic classification of the cyanobacteria, compared with the default SILVA database. The results were 251 252 parsed into a table using the *phyloseq* package (McMurdie and Holmes, 2013), and reads assigned 253 as eukaryotes, chloroplasts and mitochondria were removed. Negative controls were assessed, 254 and the maximum read number of each contaminating ASVs in a control sample was subtracted from the samples. To facilitate comparison, samples were subsampled to an even depth of 8214 255 256 sequences.

Multivariate patterns were assessed with PERMANOVA undertaken using adonis2 in vegan (Oksanen et al., 2019) with one factor and three levels (Pre-TTX, Peak TTX and post-TTX) using Bray-Curtis distance matrices. Composition plots were plotted using ggplot and ggnested (Wickham, 2016) at the genus level showing only those taxa that were present at a relative abundance of 2% (bacteria) or 0.5% (cyanobacteria) in at least one sample. ASVs that were present in higher abundance in the TTX peak period were identified by subtracting the maximum number of reads for the respective ASVs present in samples outside the TTX period away from those samples during the peak TTX period. ASVs with read counts greater than 0 were determined to be more abundant in the TTX period and plotted in a similar fashion as the previous composition plots.

267 2.4.3. Droplet digital PCR assay development

Droplet digital PCR was conducted using an automated droplet generator (QX200 Droplet 268 Digital PCR System, Bio-Rad) to determine the presence or absence of potential TTX sources, 269 Cephalothrix sp., in the digestive glands of C. gigas samples. DNA extracts were the same as 270 used for metabarcoding analyses. Primers and gBlocks to detect Cephalothrix sp. from the 271 digestive glands of C. gigas were designed for this study (Table 1). Each ddPCR reaction 272 included 1 µL of 450 nM of each primer, 10 µL Bio-Rad ddPCR Evagreen supermix, 2 µL DNA, 273 and sterile water for a total reaction volume of 22 µL. The Bio-Rad QX200 droplet generator 274 partitioned each reaction mixture into nanodroplets by combining 20 µL of the reaction mixture 275 with 20 µL of Bio-Rad droplet oil. After processing, this resulted in a total nanodroplet volume 276 of 40 µL, which was transferred to a PCR plate for amplification using the following cycling 277 protocol: hold at 95 °C for 5 min, 45 cycles of 95 °C for 30 s, 60 °C for 60 s, one final cycle of 278 signal stabilisation of 4 °C for 5 min and 90 °C for 5 min for the Cephalothrix sp. assay. The 279 plate was then analysed on the QX200 instrument. For each ddPCR plate run, at least one 280 negative control (RNA/DNA-free water Life Technologies), and one positive control (using the 281 282 gBlocks designed for the assay, Table 1) were included.

Table 1 | Sequences of primers and gBlock designed and used in this study.

Target	Sequence	NCBI accession number	Targeted gene region
Cephalothrix sp,			
Fw_Cephalothrix	GTAGTTGACCTCCTTCAGGTATTG	NC_01282 1.1	Mitochondrial cytochrome c oxidase subunit III (Cox3)
R_Cephalothrix	TCCTCCTAAACGATCTCCTTC		
gblock_C_simula	CCCTAGGATGGAGTTAGGTTGTAGTTGAC CTCCTTCAGGTATTGAGCCTTTAAATCCTT TTTCGGTACCTTTATTGAATACTGCTGTTC TTTTGGCTTCTGGAATTACTGTAACCTGAT CTCATCATTCTATAATAGAAGGAGATCGT TTAGGAGGAATTCAA GGTTTAATTTTAAC		

287 **3. Results**

288 **3.1.** Tetrodotoxin analysis and environmental parameters

Tetrodotoxin was detected in C. gigas samples for seven weeks (three weeks above the LOQ) 289 over the sampling period (24 June to 13 August 2021), with a rapid accumulation between 11 (< 290 limit of detection) and 24th June (peak at 424 and 16 µg/kg in DG and RF, respectively) which 291 then decreased for the following six weeks (Figure 2A). On June 24, 81% of the total TTX budget 292 293 was in the digestive gland. The TTX in the total flesh equivalent was 74 µg/kg on June 24. TTX was the only congener in all positive samples. Temperature and salinity were relatively stable 294 295 over the TTX peak (< 1 °C and 2 salinity units for six weeks prior and three weeks after TTX peak, while the temperature increased by ~2 °C after July 15, Figure 2B). A Distance-based 296 ReDundancy Analysis (dbRDA) showed no significant relationship between TTX concentrations 297 and environmental parameters (data not shown). 298



Figure 2 | A) Tetrodotoxin concentrations in the oysters *Crassostrea gigas* collected from the Loperhet site over a five-month period, determined using liquid chromatography-mass spectrometry (LC-MS/MS; n = 3 individuals blended together for n = 1 LC-MS/MS analysis). B). Temperatures (solid line = average daily temperatures, dotted line = maximum daily temperatures) and salinity at the Pointe du Château site (data retrieved from REPHY, (2022)).

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306 3.2. Microscopy analysis

The phytoplankton counts from the Lanvéoc station (17 km from the cages deployed, Figure 307 1) showed that dinoflagellates and diatoms were abundant during the study period (Figure 3). 308 Abundance of some species increased (>1,000 cells per litre threshold was used for this study) 309 just prior to the peak in TTX concentration (displayed in grey, Figure 2). This includes the 310 dinoflagellate genera Diplopsalis, Gyrodinium, Katodinium, Protoperidinium and Tripos, and 311 the order of Peridianiales, and diatoms of the genera Chaetoceros and Pseudo-nitzschia. 312 313 Interestingly, diatoms from the genus Chaetoceros bloomed (reaching 40,000 cells per litre) the same week as C. gigas tested positive for TTX for the first time. 314



Figure 3 | Microscopic observations and counts (cells per litre) of A) dinoflagellates and B) diatoms
species in the Lanvéoc large station near the study site. The grey area represents the period of quantifiable
TTX concentrations. Data retrieved from REPHY (2022).

3.3. Metabarcoding results

323 **3.3.1.** Bacterial analysis

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A total of 2,165 ASVs were identified in the 16S rRNA gene sequencing with samples rarefied to an even depth of 8,214 reads. Multivariate analysis (PERMANOVA) showed a nonsignificant difference in community composition amongst time periods (pre-TTX peak, during TTX peak and post TTX-peak) for bacteria in both the digestive gland (Bray–Curtis: P = 0.088, F = 1.508) and the remaining flesh (Bray–Curtis: P = 0.275, F = 1.182; Figure 4).





The bacterial composition showed that the genus *Mycoplasma* (family: Mycoplasmataceae) 335 was the dominant component in the microbiomes of C. gigas over the whole period, with the 336 exception of 24 June 2021, ranging from 17% to 93% in the digestive gland and between 25% 337 and 91% in the rest. The family Vibrionaceae was also present in the microbiome throughout the 338 experiment in both the digestive gland and the rest. Changes were noted in the genera of 339 340 Vibrionaceae with Aliivibrio being less prevalent after the TTX peak (grey area, Figure 5) and with Vibrio having a higher relative abundance during the peak (especially in the rest) and 341 afterwards. There was a large increase in the genus Persicirhabdus (family: Rubritaleaceae) at 342 the start of the peak in TTX concentrations in the digestive gland (grey area, Figure 5A) reaching 343 a relative abundance of 54% on June 24, and with a peak in relative abundance (41%) in the rest 344 prior to the TTX peak (Figure 5B). Other genera that increased with the TTX peak included 345 Sulfuvorum at the start of the peak and the genera Geminobacterium, Haloferula, Myxosarcina, 346 Neolyngbya, Pseudahrensia, Roseibacilus and Pseudoaltermonas mid-TTX peak in the digestive 347

348 glands (30 June 2021, Figure 5A).





Figure 5 | Stacked barplots showing the relative abundance of bacterial 16S ribosomal RNA sequences in the digestive gland (A) and the rest (B), at the genus level (>2% proportion in at least one sample), that are present in *Crassostrea gigas* during the study period. The grey area corresponds to the period when *C. gigas* tested positive for TTX.

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3.3.2. Cyanobacterial composition

The cyanobacterial community in C. gigas showed that in the digestive glands, 356 Parasynechococcus was the dominant genus although it had a low relative abundance (2%) at 357 the beginning of the TTX peak and reached a maximum relative abundance of 94% post TTX 358 peak (Figure 6A). Parasynechococcus was not as abundant in the rest of C. gigas, with the 359 exception of 9 September 2021, where it had a relative abundance of 96%. In the rest samples, 360 361 Zarconia was observed to be the dominant cyanobacterial genus prior to the TTX peak (mean relative abundance: 76%; Figure 6B) although this genus was not prevalent in the digestive gland 362 363 during the experiment. Species of the Neolyngbya genus appeared just before TTX increased, especially in the digestive gland. 364





Figure 6 Stacked barplots showing the relative abundance of cyanobacterial sequences in the digestive gland (A) and the rest (B), at the genus level (>2% proportion in at least one sample), that are present in *Crassostrea gigas* during the study period. The grey area corresponds to the period when *C. gigas* tested positive for TTX.

371 **3.3.3.** TTX-bearing bacterial composition

We investigated the ASVs that were more prevalent when *C. gigas* tested positive for TTX, by subtracting the core species present before and after the TTX peak period from those present during the peak of TTX (Figure 7). The most abundant bacterial genus in TTX-bearing *C. gigas*' digestive glands at the start of the peak was *Persicirhabdus* (4,294 reads) and disappeared completely after June 30. (Figure 7A). The other genera present were less abundant and varied between the sampling days although *Vibrio* increase during the TTX peak period, especially in the samples of remaining flesh.





Figure 7 | Stacked barplots showing the relative abundance of bacterial 16S ribosomal RNA
sequences for the digestive gland (A) and the rest (B), at the genus level (>1% of sample read number),
that were only present in the tetrodotoxin-bearing *Crassostrea gigas* during the TTX peak.

383 3.4. Droplet digital PCR

The assay developed in this study targeting the genera of interest, *Cephalothrix*, was successful and amplified the designed gBlock. No *Cephalothrix* sp. DNA was detected in the digestive glands of *C. gigas*. Because of the lack of nemertean tissue samples acquired for this assay development, the specificity of the assay was not tested and should be considered as a limitation of this assay development.

389

4. Discussion

391 In this study, we aimed at identifying potential sources of TTX contamination in French Pacific oysters C. gigas that were placed in an estuary previously known for TTX presence over 392 a short period of time in summer 2019 (Loperhet, Brittany, France, unpublished data). Seasonal 393 differences in bacterial communities present in the digestive glands and other organs of C. gigas 394 were explored, as well as phytoplankton differences in the surrounding waters, aimed at 395 identifying any relationships with TTX-presence. To achieve this, the bacterial communities 396 were characterized using metabarcoding in C. gigas placed in cages and harvested 13 times, 397 weekly or fortnightly for five months, starting at the end of spring and finishing at the start of 398 399 autumn. We also developed one detection assay for one of the TTX-bearing and invasive genera 400 present in Europe, the ribbon worm Cephalothrix sp., to understand if it could be the source of TTX accumulation in European shellfish. 401

TTX was detected in C. gigas for a period of seven weeks (three weeks only > LOQ), with a 402 rapid accumulation, from $<5 \mu g/kg$ (i.e., LOD) on the 11th of June to 74 $\mu g/kg$ on the 24th of June 403 $(424 \,\mu g/kg \text{ in the digestive gland and only } 16 \,\mu g/kg \text{ in the remaining flesh})$ and a rapid depuration 404 (50 μ g/kg six days later and <5 μ g/kg four weeks later in the digestive glands). These results 405 classify C. gigas as a rapid detoxifier for TTX (Bricelj and Shumway, 1998; Guéguen et al., 406 2008; Lassus et al., 2000), and showed no migration between the organs, with the majority (i.e. 407 >90% when considering LOD/LOQ) of TTX being in the digestive glands, similar to distribution 408 between organs already reported for C. gigas (Dhanji-Rapkova et al., 2021). Oysters showed 409 410 highest TTX levels in the digestive gland indicating that TTX in oysters likely comes from an exogenous source, feeding in particular, as has been previously established in other shellfish 411 studies (Biessy, 2021; Biessy et al., 2021; Dhanji-Rapkova et al., 2023). 412

Analysis of the eukaryotic phytoplankton in the surrounding waters showed that the 413 concentration of some microalgal species increased just prior to the peak in TTX concentrations. 414 In particular, diatoms from the genus Chaetoceros bloomed the same week that TTX 415 concentrations peaked, reaching 40,000 cells per litre. The diatom genus Chaetoceros is one of 416 the most diverse and widespread groups among marine phytoplankton. Its species are distributed 417 worldwide, and they often dominate in coastal ecosystems (Evensen, 1975; Guillard and Kilham, 418 419 1977). The genus includes over 200 different species, some of which are important bloom formers in both oceanic and coastal habitats (Assmy et al., 2008). Many Chaetoceros species are 420 421 used as feed algae for seafood aquaculture (Gladue and Maxey, 1994) and there has not been any report of them producing harmful toxins to date. Chaetoceros species have only been reported to 422 cause mechanical damage in fish due to the barbed setae becoming embedded in the lamellar 423 epithelium of fish gills with resultant production of excessive amounts of mucus, resulting in 424 asphyxia (Rolton et al., 2022). It is not likely that blooms of Chaetoceros spp. have ever been 425 426 tested for TTX production, but this is worth investigating in the future, in this study's location 427 and around the same time of year when peaks in TTX concentrations have been observed. Culturing and testing for TTX should also be extended to the dinoflagellate genera *Diplopsalis*, 428 429 Gyrodinium, Katodinium, Protoperidinium and Tripos, and the diatoms from the genera 430 Pleurosigma and Pseudo-nitzschia that were also present prior to the peak in TTX concentrations. 431

A parallel approach of this study was to investigate the bacterial and cyanobacterial 432 microbiome of the oysters using metabarcoding. The bacterial communities of both DG and RF 433 were dominated by species from the genus Mycoplasma, a genus that is also consistently 434 associated with bivalves, often in high abundances (Aceves et al., 2018; Biessy et al., 2020; 435 436 Lokmer and Wegner, 2015; Pierce and Ward, 2018). The family Vibrionaceae (Proteobacteria) was also abundant in all organs, which corroborates with previous studies of TTX-bearing 437 bivalves (Biessy et al., 2020; Milan et al., 2018) and non-TTX-bearing bivalves (Aceves et al., 438 2018; Cho, 2019; Dubé et al., 2019). Vibrionaceae were present throughout the study and did not 439 have a correlation with the peak in TTX concentration. Of note is the increase in the abundance 440 441 of bacteria from the Rubritaleaceae family, in particular the genera Persicirhabdus before and during the TTX peak. Bacteria from the Rubritaleaceae are often associated with shellfish 442 microbiomes (Biessy et al., 2020; Cho, 2019), showing different proportions likely due to 443 differences in geographical origins (del Rio-Lavín et al., 2023). In a previous study, we found 444 445 that Verrucomicrobia (four ASVs from the Rubritaleaceae family) were significantly and

positively correlated to TTX concentrations in TTX-bearing clams (Biessy et al., 2020). To the
extent of our knowledge, species from the Rubritaleaceae family, in particular the genus *Persicirhabdus*, have not been tested for TTX production and are a great candidate for future
work to identify the producer of TTX.

450 Previously, cyanobacteria have been hypothesized as potential TTX producers. Strains of the genera Pleurocapsa, Cyanobium and Prochlorococcus were detected in TTX-bearing clams in 451 New Zealand, and the genus Synechococcus was found to be significantly correlated to higher 452 TTX concentrations in these clams (Biessy et al., 2020). Another study showed that 453 cyanobacteria from the Calotrichaceae lineage were found centred in the microbial co-occurrence 454 455 network of toxic pufferfish along Vibrio species (Li et al., 2020). It was hypothesised that Vibrio and cyanobacteria may form a symbiotic relationship in the gut environment of toxic puffer fish 456 457 (Li et al., 2020). In this study, species of Neolyngbya were detected in the oyster's digestive glands just before and at the start of the peak in TTX concentrations before completely 458 459 disappearing when the TTX concentrations started to decrease. Neolyngbya spp. belong to the family Oscillatoriaceae and are described as marine filamentous cyanobacteria that are well 460 461 known as producers of a rich diversity of secondary metabolites and toxins (Caires et al., 2018), and should be tested for TTX production in the future. Cyanobacterial ecotoxicology is 462 463 particularly well documented in freshwater habitats and are known to produce a wide range of 464 toxins (Jakubowska and Szelag-Wasielewska, 2015; Rastogi et al., 2015), including paralytic shellfish toxins (PSTs). PSTs are potent neurotoxins produced by cyanobacteria and marine 465 dinoflagellates and are known to exert the same toxic effect as TTX through an interaction with 466 voltage gated sodium channels (Narahashi, 1988). Both toxins have been detected simultaneously 467 in several aquatic species but toxicological studies on cyanobacteria have not been performed to 468 the same extent in marine environments (Bane et al., 2014; Frazão et al., 2010). Marine 469 cyanobacteria are thus an interesting lead for future studies aiming to identify TTX-producing 470 471 bacterial species. It is worth noting that the eukaryotic communities (using metabarcoding targeting the V4 18S rRNA gene region) were also investigated in this study to identify potential 472 473 eukaryotic TTX-producers. However, despite using C. gigas blocking primers previously 474 developed by Clerissi et al. (2020), over 80% of sequence reads belonged to the class Bivalvia and thus the data was not robust enough to include in this study. It is also important to note that 475 476 the metabarcoding analyses from this study was limited by the frequency of sampling. This study 477 was designed to guarantee to capture the yearly peak of TTX but if this study was to be repeated, 478 oysters should be harvested daily or at least weekly around the time of the TTX peak, and during

a shorter period (Mid-June until the end of July). Another limit to metabarcoding is the wellknown challenge of taxonomic reference database and limited taxonomic resolution (Keck et al.,
2023) and these were hopefully mitigated by the use of up-to-date databases such as *Cyanoseq*(Lefler et al., 2023), in particular for cyanobacteria taxonomy, and bioinformatic pipelines.

Lastly, we developed a droplet digital PCR assay aimed at investigating the presence or 483 absence of Cephalothrix, a genus of interest. This genus is known to be TTX-bearing and has 484 been detected in Europe in the last decades (Turner et al., 2018). It is possible that the eggs of 485 this species, after spawning events, could accumulate in filter-feeding shellfish. This is 486 particularly relevant because spawning of some organisms can be very regular (e.g., within a 487 488 week for some fish (Cushing, 1969) and within a month for some marine worms (Bybee et al., 2007)), which is the case here with TTX concentrations in French shellfish found to peak for a 489 490 few weeks in early summer (Réveillon et al., 2021). Due to the lack of positive tissue samples available for assay development, gBlocks were designed to use as positive controls. Cephalothrix 491 492 sp. was not detected in the digestive glands of C. gigas and was thus not likely to have been the source of TTX contamination in the oysters. It is worth mentioning that due to the lack of positive 493 494 and negative Nemertean tissues available, this assay is not fully developed (i.e., sensitivity, crossreactivity with other species) and this result should be interpreted with care. This assay, once 495 496 fully developed and validated, could be of broad interest globally to detect this invasive and toxic 497 species.

498 **5.** Conclusion

499 The aim of this study was to investigate the potential source of TTX in French Pacific oysters Crassostrea gigas that were placed in an estuary previously reported to encounter TTX 500 501 accumulation in shellfish. The microscopic observations of algae present during the TTX peak 502 showed that a sudden bloom in Chaetoceros species co-occurred with the sudden appearance of TTX in oysters. Metabarcoding analysis suggested that prokaryotic microbiome components 503 correlate with previous studies indicating that 1) bacteria from the Rubritaleaceae (in particular 504 from the genus Persicirhabdus) were positively correlated to TTX concentrations in TTX-505 bearing shellfish and that 2) marine cyanobacteria, Neolyngbya in particular for this study, should 506 be investigated as potential TTX producers. Culturing, isolating and testing these bacterial and 507 508 cyanobacterial species and genera for toxin production would be a logical next step in the search 509 to identify the source of TTX in marine organisms.

511 Data Availability Statement

512 The datasets presented in this study are available in online repositories. The names of the 513 repository/repositories and accession number(s) can be found in the article.

514 Author Contributions

LB, KS, KNM, DR and PH conceived and designed the experiments. LB, LT, KNM, LL, AT and VS performed the experiments and TTX analysis. LB and JP undertook the metabarcoding analysis. All authors helped with the data interpretation and experiments, and the editing and writing of the manuscript.

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523 **Conflict of Interest**

524 The authors declare that the research was conducted in the absence of any commercial or 525 financial relationships that could be construed as a potential conflict of interest.

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