
Mercury Levels in Tissues (Cartilage, Skin, and Muscle) of the Greenland Shark (*Somniosus microcephalus*): Potential Contamination Sources and Implications for Health and Conservation

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Abstract :

The jaws of the Greenland shark have high levels of mercury. Hg of cartilage in comparison with skin and muscle from the same specimen makes it possible to apprehend the distribution of the pollutant in the body. The level of the pollutant between jaw, skin and fresh meat (muscle) shows a strong correlation. The muscle is the most contaminated element in comparison with the skin and cartilage. The species presents the highest levels among different groups of sharks and the results are in accordance with previous studies. Marine ecosystems in the Arctic are globally contaminated by mercury (atmosphere, water, sediments, food web). The methylmercury reduces blood calcium levels, directly affecting the metabolism of cartilage cells. Even if cases of malformations could not be observed in the Greenland shark, numerous cases for other shark species have been documented in connection with heavy metals (e.g., Hg), and in particular for aplacental viviparous with potential morphological anomalies on embryos. The situation of the Greenland shark is worrying due to the conservation status, the fishing catches, the tardive sexual maturity and reproduction, the climate change and the level of mercury on its organism. The results incite to recommend ecological, environmental and fisheries management measures.

Keywords : Greenland shark, Mercury, Potential abnormalities, Conservation, Arctic

28 **1. Introduction**

29

30 The situation of the Greenland shark is worrying. The species holds a conservation classification of Near
31 Threatened (NT) in Europe (Burgess et al. 2015). However, the reported catches by the Icelandic and
32 Greenlandic fleets surged to more than 43 tonnes in 2010 and 120 tonnes in 2020 (Tonnes Live Weight; FAO
33 areas 14, 14b, 14b2, 5a, 5a1, 5a2) (ICES, 2022).

34 The Greenland shark, inhabiting the cold waters of the Arctic and North Atlantic Oceans, including the Celtic
35 and North Seas, faces bycatch mainly in Arctic regions, notably by halibut or shrimp fisheries (David et al.,

36 2013). Greenland Shark bycatch ranged from 34.4% for longline, 32.8% for trawl and 18.3% for gillnet
37 respectively in NAFO subarea 0 (Bryk et al., 2018).

38 The Greenland shark displays aplacental viviparity (Nielsen et al., 2020). Genetic investigations have identified
39 genetically mixed individuals in the Canadian Arctic, sub-Arctic, and temperate Eastern Atlantic areas,
40 suggesting introgressive hybridization with the Pacific sleeper shark (*Somniosus pacificus*) Bigelow and
41 Schroeder, 1944 (Walter et al., 2017). Insights into Greenland shark sexual maturity showed male and female
42 body lengths-at-maturity (Total Length) of 2.84 ± 0.06 m and 4.19 ± 0.04 m, respectively (Nielsen et al. 2020). The
43 species with a large long-live are particularly vulnerable to exposure to mercury biomagnified (Rumbold et al
44 2014). Several scientific studies confirm the climate change in the Arctic and the polluted environment on
45 mercury included food webs with potential effects of shark health (Braune et al., 2015; MacMeans, 2015;
46 Rodríguez-Romero et al., 2018; Schaefer et al., 2020).

47 Natural phenomena such as volcanoes, rock erosion, but also forest fires, whether from natural, accidental or
48 human causes, without forgetting industrial activities, especially the exploitation of coal and gold, emit mercury
49 (Hg) into the environment (Veiga et al., 2006; Dabrowski et al., 2008; Witt et al., 2009). Mercury has the
50 particularity of lasting up to two years in the atmosphere before the abiotic oxidation into Hg^{2+} (Ariya et al.,
51 2015). Hence, the Arctic Ocean is susceptible to mercury contamination, which may stem from sources that are
52 notably remote (Ariya et al., 2004).

53

54 Industrial activities associated with resource exploitation in the Arctic, including gold mining, have continued to
55 expand in recent years (Mered, 2020). Gold processing involves the use of mercury and heavy metal pollution of
56 water (Fashola et al., 2016). Kirk et al. 2008 demonstrated the methylation of Hg through analyses of Arctic
57 waters. Concentrations of methylated Hg and Dimethylmercury (Me₂Hg) were often low in surface waters
58 (23.8 ± 9.9 and 4.7 ± 4.4 pg L⁻¹, respectively), but they increased with depth (maximum: 178 and 170 pg L⁻¹,
59 respectively; mean: 70.3 ± 37.3 and 56.8 ± 37.8 pg L⁻¹, respectively). Water serves as a significant source of
60 gaseous Hg (gaseous elemental Hg, GEM) in the atmosphere, especially during the ice-free period (Braune et al.,
61 2015). Methylmercury (MeHg) by photodegradation phenomenon and Me₂Hg on the surface can contribute to
62 producing GEM deep in the water column (Mason et al. 1998; Chen et al 2003). Subsequently, it undergoes
63 transformation or is reduced to elementary Hg and released back into the atmosphere. Another portion settles as
64 sediments, while yet another part transforms into monomethylmercury (MeHg) and dimethylmercury (Me₂Hg)
65 (Braune et al., 2015). The deposition of land-based mercury on marine sediments causes increased microbial Hg

66 methylation and the transfer of MeHg to benthic and pelagic food webs (Ferreira Araujo et al., 2022). The
67 contribution of mercury stems, in part, from the potential release of mercury from permafrost due to the erosion
68 of rivers and coastlines resulting from Arctic warming (Schaefer et al., 2020).

69

70 The marine environment is not exempt from this concern. Different mercury forms are present, including
71 elemental (Hg⁰), inorganic (Hg²⁺), and organic (CH₃Hg⁺) forms (Clarkson, 1997). Among these,
72 methylmercury (CH₃Hg⁺), the hazardous variant of mercury, gradually accumulates within organisms of marine
73 species including sharks and undergoes an amplification process (known as bioaccumulation) as it moves up the
74 trophic web, from lower to higher levels (Harding et al., 2018, Biton-Porsmoguer et al., 2018; 2022). This
75 susceptibility is linked to their extended lifespans and high trophic level (Gelsleichter and Walker 2010), which
76 can lead to potential morphological abnormalities (Casarini et al., 1997; Rosa et al., 2004; Moore, 2015;
77 Cabanillas-Torpoco et al., 2023).

78 The Greenland shark exhibits elevated mercury levels, among the highest among marine species analysed in the
79 Canadian Arctic (Chételat and Braune, 2012). The presence of mercury has been demonstrated previously in
80 different organs of the Greenland shark (only in muscle and liver) (McMeans et al., 2007, 2010, 2015; Chételat
81 and Braune, 2012; Corsolini et al., 2014). The biological characteristics of the Greenland shark mentioned
82 above, including longevity, make it a preferential candidate for study the level of contaminating elements in their
83 organism. Further investigations of mercury production have a particular importance in marine arctic
84 environment particularly in zones, which enhance uptake in the food web to evaluate source-receptor impacts
85 (AMAP, 2021). The objective of this study are 1) to measure and confirm mercury concentration within the
86 cartilaginous tissue of the species, specifically targeting the upper and lower jaws; 2) to analyse and compare
87 mercury levels present in skin and muscle samples extracted from the same individual specimens; 3) to examine
88 potential sources of mercury contamination and investigate its potential effects on the morphology and
89 reproductive patterns of sharks; 4) to provide recommendations for conservation strategies aimed at reducing
90 bycatch and minimizing the presence of mercury in the delicate Arctic polar ecosystem. In accomplishing these
91 objectives, the study aims to contribute valuable insights into mercury accumulation dynamics, potential
92 consequences for the species, and imperative conservation actions required in the Arctic environment.

93

94 **2. Materials and methods**

95

96 *Sampling*

97 The sampling phase was conducted aboard the bottom trawler "Steinunn SF-10." The vessel operated from
98 Hornafjörður, Iceland, during the year 2010. On August 8, 2010, a female Greenland shark was captured in the
99 Greenland Sea, near Denmark Strait (68°016 N - 22°764 W) within FAO fishing area 5.2 (Fig.1). Measuring 434
100 cm in total length, the specimen was procured and retained on the ship. Samples of both upper and lower jaws
101 muscle and skin were meticulously extracted using knives and scalpels (Fig.1). The tools underwent thorough
102 cleaning with soapy water, disinfection using 90% ethanol, and subsequent rinsing with distilled water between
103 each sample collection. After extraction, all samples were placed within hermetically sealed plastic bags and
104 accurately referenced. Subsequently, the muscle sample was immersed in 70% ethanol, while the jaw
105 (cartilaginous tissue) and skin samples were dried and were securely stored in airtight plastic bags.

106

107 *Mercury analysis*

108 The total mercury (THg) concentration of the tissue samples was determined using cold vapor atomic absorption
109 spectroscopy (CVAAS) with thermal decomposition and gold amalgamation. The instrument used was the Hydra
110 IIC direct mercury analyzer by Teledyne Leeman Labs (Hudson, NH). Calibration curves were constructed using
111 the Certified Reference Materials (CRM) BCR 414 (phytoplankton), DORM-4 (fish protein), EM-CE278K
112 (mussels) EMR-CE464 (Tuna fish) ranging from 0.276 mg.kg⁻¹ to 5.240 mg. kg⁻¹ dw of THg. Instrument
113 linearity was established across two separate ranges; a low range (0 ng to 30 ng) and high range (75 ng to 200
114 ng). The high calibration range has a correlation coefficient (r^2) value of 0.9992 and the low calibration range
115 has an r^2 value of 0.9996. The limit of quantitation (LOQ) was set as 10 times the standard deviation of the y-
116 axis intercept of the method calibration curve.

117 A portion of each tissue sample was placed in a nickel boat and weighed on an analytical balance (± 0.0001 g).
118 The amounts of upper and lower jaws, skin and muscle were 0.0127 g dry mass, 0.0175 g dry mass, 0.0213 g dry
119 mass and 0.0142 \pm 0.004 g dry mass, respectively. Given the unconventional method of preserving muscle in
120 alcohol for THg analysis and the sufficient sample quantity available for analysis, a total of three separate
121 analyses were conducted. The provided result is the average of the three measurements. Samples were first
122 introduced into the decomposition furnace where they were dried (300 °C, 30 s) and combusted (800 °C, 150 s),
123 then into the catalyst furnace (600 °C, 60 s), and finally through the drying tube and gold amalgamation trap
124 (600 °C, 30 s) before entering the spectrometer. Samples were not homogenized to avoid unnecessary
125 contamination from additional handling and due to the limited available tissue in some cases. Samples were not
126 digested before analysis. For quality assurance, a CRM (EMR-CE464 Tuna Fish) measurement was performed at
127 the end of the six analyses.

128

129 3. Results and discussion

130

131 Elevated THg levels were analysed in all tissues of the shark's body, ranging from 1.332 mg kg⁻¹ dw (upper
132 jaw) to 3.225 mg kg⁻¹ dw (muscle) (Fig.2; Table 1). A linear concentration increase trend is confirmed in THg
133 between the lower and upper jaws (cartilage), skin and fresh meat (muscle) ($R^2=0.9996$) (Fig.2).

134 The distribution of mercury levels across different shark tissues reveals that muscle accumulates higher levels
135 compared to the skin and cartilage. This disparity may be attributed to variations in the rate of cell turnover,
136 especially notable in cartilage (Marconi et al. 2020) and skin (Meyer and Seegers, 2012), but also with the fact
137 that mercury in fish included sharks is mostly in the methylated form and has a protein affinity (Chouvelon et al.,

138 2018). The mercury levels analysed in this study are consistent with those observed from other specimens of
139 *Somniosus microcephalus* (Table 1). However, it appears that the varying mercury levels obtained are not
140 correlated with the length of the individuals ($P>0.05$, $n=19$, MacMeans et al., 2010). A multiple test on muscle
141 tissue indicated that concentrations of mercury were higher on Greenland shark, 5.930 mg kg^{-1} , $n = 19$
142 (Macnean, 2017), 3.225 mg kg^{-1} , $n = 1$ (this study) in comparison with specimens of Carcharhiniformes, $1.520 \pm$
143 1.900 mg kg^{-1} , $n = 1739$; Lamniformes, $2.580 \pm 4.790 \text{ mg kg}^{-1}$, $n = 508$; others species of Squaliformes, $1.610 \pm$
144 1.040 mg kg^{-1} , $n = 415$; and Myliobatiformes, $0.383 \pm 0.350 \text{ mg kg}^{-1}$, $n = 195$ (Tiktak et al., 2020). Several
145 factors may explain higher THg values in Greenland shark (i) their feeding at a high trophic position (TP) and
146 (ii) foraging on slightly more contaminated offshore resources in the Arctic (MacMeans et al., 2010). The lack of
147 a correlation between mercury levels and the age (size) of the sharks can be attributed to (i) mercury
148 accumulation in larger, presumably older Greenland sharks not being solely driven by bioaccumulation and (ii)
149 larger specimens not necessarily feeding at a higher TP than smaller sharks (MacMeans et al., 2010).

150

151 The limitation of the method used

152 The measurement of mercury in a sample preserved in ethanol, such as the muscle sample in this case, is not
153 standard. Usually, measurements are carried out on fresh (MacMeans et al., 2010) or freeze-dried samples
154 (Biton-Porsmoguer et al., 2018; 2022). Here, the measurements were conducted 13 years after the collection, and
155 the sample preservation method could not be controlled. Alcohol preservation can be used for maintaining tissue
156 samples for various analyses, including mercury measurements. However, certain factors need to be considered.
157 Ethanol itself may contain trace amounts of mercury. Hence, it is crucial to use high-purity ethanol for
158 preservation and ensure that the ethanol used does not introduce additional mercury into the samples.
159 Here, ethanol (70% USP-grade Isopropanol (Isopropyl Alcohol) and 30% USP-grade purified water) was used. It
160 is also reasonable to assume that mercury concentrations could change during preservation, likely due to the loss
161 of mass caused by protein and water loss in samples preserved in ethanol. Hence Gibbs et al. (1974) observed a
162 slight increase in mercury concentrations in fish after a short period of preservation in ethanol. However, when
163 preserved in ethanol, fish samples typically experience an initial loss of mass due to deshydration, followed by
164 a stabilisation (Shields and Carlson, 1996). In this case, the muscle sample was preserved in ethanol for 13 years.
165 It is highly likely that any initial dehydration induced by the alcohol has reached a state of stabilization. Utilizing
166 a conversion equation holds merit in the endeavour to retrocalculate the mass of ethanol-preserved samples to
167 their initial state under fresh conditions. Nevertheless, it is important to note that these conversion equations tend

168 to possess limited reliability, typically capable of offering only coarse approximations of the live condition
169 (Billy, 1982). Consequently, we have chosen to refrain from their application. We acknowledge that the
170 concentrations measured in the muscle of the ethanol-preserved shark are prone to being higher than the actual
171 concentrations determined based on fresh weight. Shields and Carlson's study (1996) indicates that this
172 overestimation could potentially reach up to 20%. This uncertainty adds to the measurement uncertainties.

173 The selection of analytical methods for measuring mercury in alcohol-preserved samples could vary from those
174 employed for fresh or frozen samples. Cold vapor atomic absorption spectrometry (CVAAS) has been previously
175 employed for mercury analysis in alcohol-preserved samples (Levengood et al., 2013), thereby confirming the
176 applicability of the method used in this study. The authors analysed the composition of the alcohol in the jars used
177 for specimen storage. Mercury was detected at a very low concentration in the ethanol, confirming that there are
178 minimal mercury transfers from the samples to the alcohol.

179

180 *Sources of mercury in arctic region for sharks*

181 Several studies have highlighted that prey serves as a primary source of mercury in marine organisms,
182 particularly for top predators like sharks, owing to the phenomenon of bioaccumulation (McMeans et al., 2015;
183 Biton-Porsmoguer et al., 2018; Li et al., 2023). The main prey of the Greenland shark are fish (mainly Teleosts
184 and Elasmobranch as rays), mammals (seals, whales and bears), squids and crustaceans (shrimp and crabs)
185 (McMeans et al., 2010; Nielsen et al. 2013; 2019). The species more contaminated by mercury (muscle) in
186 Arctic were elasmobranch and mammals (as rays and seals, 1.2-2 THg $\mu\text{g kg}^{-1}$ dw) and teleosts (as ling, blue
187 ling, tusk, 0.12-0.94 THg $\mu\text{g kg}^{-1}$ dw) (McMeans et al., 2010; 2015).

188 The results of this study demonstrate 1) the presence of mercury in the cartilage and 2) a strong correlation with
189 the level of THg in cartilage with the skin and the muscle. Numerous studies have focused on the implications of
190 mercury for human health (Ha et al., 2017). It is noteworthy that the flesh of the Greenland shark is consumed in
191 Iceland, known as "Hákarl," and is subjected to a distinctive fermentation process followed by drying for a
192 duration of four to five months (Travel Food Atlas, 2020). The maximum allowable mercury level for shark meat
193 intended for human consumption is set at 1.0 mg/kg of fresh meat¹. Considering the mercury levels analysed in
194 this study, the shark flesh appears unsuitable for human consumption and could potentially pose a health risk.
195 (Guzzi et al., 2021). Moreover, mercury can also have effect on sharks.

¹ Commission Regulation (EU) 2022/617 of 12 April 2022 amending Regulation (EC) No 1881/2006.

196

197 *Potential effects of mercury on shark morphology and reproduction*

198 The Elasmobranchs (sharks, skates and chimaeras) are characterized by a cartilaginous skeleton (Seidel et al.
199 2021). The vertebral centra of elasmobranchs are composed of calcified cartilage, mainly comprised of the
200 calcium phosphate mineral hydroxyapatite [Ca₁₀(PO₄)₆(OH)₂], which is embedded within an organic matrix of
201 proteins, including proteoglycan and collagen (Urist, 1961; Porter et al., 2006). Mercury leads to metabolic
202 disorders and disrupts the proper assembly of the cytoskeleton in marine species. The toxicity of mercury also
203 causes morphological alterations on the development various fish organs, including the gills and olfactory
204 organs. Malformations can involve the craniofacial and skeletal systems with stunted growth, curved spine, and
205 eye deformities (Weis and Weis, 1977; Weis et al., 1981; Jagoe et al., 1996; Ribeiro et al., 1996; Oliveira
206 Ribeiro et al., 2000, Devlin, 2006; Adams et al., 2010).

207

208 The effects of mercury on bone metabolism have also been demonstrated for teleosts and in particular for the
209 activity of osteoclasts (cell responsible for the dissolution and absorption of bone) and osteoblasts (cell which
210 secretes the substance of bone) (Suzuki et al., 2004). The methylmercury reduces blood calcium levels
211 (calcemia), directly affecting the metabolism of bone cells at the scale (Suzuki et al., 2004). Osteoclasts on bone
212 and those on cartilage also have similar basic molecular characteristics (Larroure et al., 2021). Yachiguchi et
213 al., (2014) reported a decrease in TRAP (Tartrate-Resistant Acid Phosphatase) and ALP (alkaline phosphatase)
214 expression in marine fish exposed to Methylmercury or inorganic mercury. The ALP and TRAP are known as
215 marker enzymes for osteoblasts and osteoclasts, respectively (Bonuci and Nanci 2001). Mercury would therefore
216 inhibit the activity of both osteoclasts and osteoblasts (Rodríguez and Mandalunis 2018).

217 The observation of sharks during boarding and the study of the literature did not make it possible to identify
218 malformations in Greenland shark. However, many observations of morphological abnormalities in sharks have
219 been documented for many years. Thirteen studies have been published since 1963 reporting malformations on
220 the blue shark (*Prionace glauca*) (Linnaeus, 1758) in different marine areas (North and South Atlantic Ocean,
221 Pacific Ocean, California Gulf, Mediterranean Sea) (Cabanillas-Torpoco et al., 2023). Hoenig and Walsh (1983)
222 noted four cases of vertebral lesions in three species of sharks, Sandbar shark *Carcharhinus plumbeus* (Nardo,
223 1827, Limon shark *Negaprion brevirostris* (Poey, 1868) and Sand tiger shark *Carcharias Taurus* Rafinesque,
224 1810). The spines had fused centra, ribs, and neural arches, additional deposition and erosion of calcified
225 material in the centra, and in one case compression of the centra (Hoenig and Walsh 1983). Missing fins in an

226 adult tawny nurse shark, *Nebrius concolor* (Lesson, 1831) were reported by Taniuchi and Yanagisawa (1987),
227 abnormal bicephalism in blue shark and tope shark (*Galeorhinus galeus*) (Linnaeus, 1758) by Ramirez-Amaro et
228 al. (2019). The complete absence of pelvic fins in a milk shark (*Rhizoprionodon acutus*) (Rüppel, 1837) was
229 reported in the Persian–Arabian Gulf (Moore 2015). The causes of such deformities were unknown.

230 The environmental degradation or pollution (effects of contaminants) can explain the malformation
231 presence in sharks (Casarini, 1997; Rosa et al., 2004; Mancini et al., 2006). In this context, the presence of high
232 levels of mercury (Hg) have been confirmed by several studies (Barrera-García et al., 2012; Rodríguez-Romero
233 et al., 2018) and genetic alterations abnormalities in a blue shark embryo related with initial stages of
234 development (Mancini et al., 2006; Rodríguez-Romero et al., 2018). Moreover, Van Hees and Ebert (2017) have
235 demonstrated the maternal mercury transfer on the leopard shark *Triakis semifasciata* Girard, 1855 (an
236 aplacental viviparous species as Greenland shark). Embryos were found with potentially harmful mercury
237 concentrations in their muscle tissues.

238

239

240 **4. Conclusion**

241

242 Marine ecosystems in the Arctic are globally contaminated by mercury (atmosphere, water, sediments,
243 food web). Hg analyses performed on the specimen showed a contamination in different parts of his organism
244 (cartilage, skin and muscle) with potential risks for the individual and human health. Recent studies
245 demonstrated the maternal mercury transfer for aplacental viviparous and connected mercury with potential
246 morphological anomalies on sharks. Arctic temperatures will increase more than the global average
247 (Johannessen et al., 2004). Estimated aerobic scope for polar species will indicate their adaptive capacity and
248 predictions of food web shifts (Steiner et al., 2019). Studies on mercury levels on elasmobranch should to be
249 encouraged to analyse the state of embryos and better understand the causes of malformations. These results
250 could be very interesting to list the species most affected and for their survival rate and the conservation of the
251 species.

252 However, considering the substantial amount of data that corroborates the elevated mercury levels found across
253 various tissues of the Greenland shark (including cartilage), as well as the observed morphological anomalies in
254 numerous other species (raising suspicions of environmental pollution), coupled with the declining population in
255 Europe (Burgess et al., 2015), shouldn't precautionary measures be implemented as a principle of safeguard?

256 In order to protect the environment, the precautionary approach shall be widely applied by States according to
257 their capabilities. Where there are threats of serious or irreversible damage, lack of full scientific certainty,
258 explained in this case in part by the difficulty of studying this Green land shark at different biological stages but
259 also by the geographical area and sampling conditions, shall not be used as a reason for postponing cost-effective
260 measures to prevent environmental degradation (Principle 15 of the 1992 Rio Declaration States) (United Nation,
261 1992).

262 One of the objectives of this work was to provide recommendations for conservation strategies aimed at reducing
263 bycatch, detailed by gear previously. The proposal measures could be, 1) Islandic and Greenland fleets must take
264 the necessary measures to develop the turtle excluder devices and Nordmore grates and release specimen still
265 alive; 2) Protect the central Arctic Ocean beyond national jurisdiction with a precautionary fisheries moratorium;
266 3) Contribute effectively to the fight against global warming and reduce any industrial production that is a source
267 of mercury (and any other pollutant) in the Arctic Ocean and 4) Initiate research programs to better understand
268 the areas of movement and distribution of the species, which will inevitably be reduced by climate change and in
269 order to limit any fishing activity there.

270

271

272 **Declaration of competing interest**

273

274 The authors have no competing interests to declare that are relevant to the content of this article.

275

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277

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286 **References**

287

288 Adams, D.H., Sonne, C., Basu, N., Dietz, R., Nam, D.H., Leifsson, P.S., Asger, L. Jensen, A.L., 2010. Mercury
289 contamination in spotted seatrout, *Cynoscion nebulosus*: An assessment of liver, kidney, blood, and nervous
290 system health. *Sci Total Environ*, 408 (23): 5808-5816. <https://doi.org/10.1016/j.scitotenv.2010.08.019>.

291

292 Arctic Monitoring and Assessment Programme (AMAP). 2021. 2021 Mercury Assessment. Summary for police-
293 makers. Arctic Monitoring and Assessment Programme. Arctic Council. AMAP. 16p.

294

295 Ariya, P.A., Dastoor, A.P., Amyot, M., Schroeder, W.H., Barrie, L., Anlauf, K., Raofie, F., Ryzhkov, A.,
296 Davignon, D., Lalonde, J., Steffen, A., 2004. The Arctic: a sink for mercury, *Tellus B: Chem Phys*
297 *Meteorol*, 56:5, 397-403. [https://doi: 10.3402/tellusb.v56i5.16458](https://doi:10.3402/tellusb.v56i5.16458).

298

299 Ariya, P.A., Amyot, M., Dastoor, A., Deeds, D., Feinberg, A., Kos G., Poulain, A., Ryjkov, A., Semeniuk, K.,
300 Subir, M., Toyota, K., 2015. Mercury physicochemical and biogeochemical transformation in the atmosphere
301 and at atmospheric interfaces: a review and future directions. *Chem Rev*, 115(10): 3760-802. [https://doi:](https://doi:10.1021/cr500667e)
302 [10.1021/cr500667e](https://doi:10.1021/cr500667e).

303

304 Barrera-García, A., O'Hara, T., Galván-Magaña, F., Méndez-Rodríguez, L.C., Castellini, J.M., Zenteno-Savín,
305 T., 2012. Oxidative stress indicators and trace elements in the blue shark (*Prionace glauca*) off the east coast of
306 the Mexican Pacific Ocean. *Comparative Biochemistry and Physiology Part C: Toxicol Pharmacol*, 156(2): 59-
307 66. [https://doi: 10.1016/j.cbpc.2012.04.003](https://doi:10.1016/j.cbpc.2012.04.003)

308

309 Billy, A.J. 1982. The effects of formalin and isopropyl alcohol on length and weight measurements
310 of *Sarotherodon mossambicus* Trewavas. *J. Fish Biol*, 21: 107-112. [https://doi.org/10.1111/j.1095-](https://doi.org/10.1111/j.1095-8649.1982.tb02828.x)
311 [8649.1982.tb02828.x](https://doi.org/10.1111/j.1095-8649.1982.tb02828.x)

312 Biton-Porsmoguer, S., Banaru, D., Boudouresque, C.F., Dekeyser, I., Bouchoucha, M., Marco-Miralles, F.,

313 Harmelin-Vivien, M., 2018. Mercury contamination of the blue shark (*Prionace glauca*) and the shortfin mako

314 (*Isurus oxyrinchus*) in the North-eastern Atlantic Ocean: Implication for fishery management. *Mar. Pollut. Bull.*

315 127: 131-138. [https://doi: 10.1016/j.marpolbul.2017.12.006](https://doi:10.1016/j.marpolbul.2017.12.006).

- 316 Biton-Porsmoguer, S., Banaru, D., Harmelin-Vivien, M., Béarez, P., Bouchoucha, M., Marco-Miralles, F.,
317 Marquès, M., Lloret, J., 2022. A study of trophic structure, physiological condition and mercury
318 bioaccumulation in swordfish (*Xiphias gladius*): evidence of unfavourable conditions for the swordfish
319 population in the Western Mediterranean. *Mar. Pollut. Bull.*, 176 (113411): 1-10.
320 <https://doi.org/10.1016/j.marpolbul.2022.113441>.
321
- 322 Bonucci, E., Nanci, A., 2001. Alkaline phosphatase and tartrate-resistant acid phosphatase in osteoblasts of
323 normal and pathologic bone. *Ital J Anat Embryol*, 106 (2 Suppl 1): 129-33.
324
- 325 Braune, B., Chételat, J., Amyot, M., Brown, T., Clayden, M., Evans, M., Fisk, A., Gaden, A., Girard, C., Hare,
326 A., Kirk, J., Lehnher, I., Letcher, R., Loseto, L., Macdonald, R., Mann, E., McMeans, B., Muir, D., O'Driscoll,
327 N., Poulain, A., Reimer, K., Stern, G., 2015. Mercury in the marine environment of the Canadian Arctic: Review
328 of recent findings, *Sci Total Environ*, (509–510): 67-90. <https://doi.org/10.1016/j.scitotenv.2014.05.133>.
329
- 330 Bryk, J.L., Hedges, K.J., Treble, M.A., 2018. Summary of Greenland Shark (*Somniosus microcephalus*) catch in
331 Greenland Halibut (*Reinhardtius hippoglossoides*) fisheries and scientific surveys conducted in NAFO Subarea
332 0. Scientific Council Meeting. Serial n° N6831. NAFO SCR Doc. 18/01. 17p.
333
- 334 Burgess, G.H., Sherrill-Mix, S.A., Kyne, P.M., 2015. *Somniosus microcephalus* (Europe assessment). *The IUCN*
335 *Red List of Threatened Species* 2015: e.T60213A48917268. Accessed on 25 February 2023.
336
- 337 Cabanillas-Torpoco, M., Abbatepaulo, F., Rodrigues, L., Marquez, R., Oddone, M.C., Cardoso, L.G., 2023.
338 Teratological records in blue shark *Prionace glauca* embryos from the South-western Atlantic Ocean. *J Mar*
339 *Biolog Assoc U.K.*, 103, e12, 1–5. <https://doi.org/10.1017/S0025315422000996>
340
- 341 Casarini, L.M., Gomes, U.L. and Tomas, A.R.G., 1997. Would Santos harbour dredged material dumping be a
342 reason of teratogeny on *Raja agassizi*? In Congresso Latino-Americano sobre Ciências do Mar Colacmar.
343 Caderno de resumos 7, 152–153. Santos: SBEEL.
344

- 345 Chételat, J., Braune, B.M., 2012. Canadian Arctic contaminants assessment report III – 2012: mercury in
346 Canada's North / Northern Contaminants Program. Canada. Aboriginal Affairs and Northern Development
347 Canada. "Editors: John Chételat, Birgit Braune. Xxiii, 276p.: col. Ill., maps. R74-2/1-2013E-PDF.
348
- 349 Chen, J., Pehkonen, S.O., Lin, C.J., 2003. Degradation of monomethylmercury chloride by hydroxyl radicals in
350 simulated natural waters. *Water Res* 37: 2496–504.
351
- 352 Chouvelon, T., Cresson P., Bouchoucha, M., Brach-Papa, C., Bustamante, P., Crochet, S., Marco-Miralles, F.,
353 Thomas, B., Knoery J., 2018. Oligotrophy as a major driver of mercury bioaccumulation in medium-to high-
354 trophic level consumers: A marine ecosystem-comparative study, *Environ Pollut*, 233: 844-854.
355 <https://doi.org/10.1016/j.envpol.2017.11.015>.
356
- 357 Clarkson, T.W., 1997. The toxicology of mercury. *Crit Rev Clin Lab Sci*. 34(4): 369-403. [https://doi:](https://doi.org/10.3109/10408369708998098)
358 [10.3109/10408369708998098](https://doi.org/10.3109/10408369708998098).
359
- 360 Corsolini, S., Ancora, S., Bianchi, N., Mariotti, G., Leonzio, C., Christiansen, J.S., 2014. Organotropism of
361 persistent organic pollutants and heavy metals in the Greenland shark *Somniosus microcephalus* in NE
362 Greenland. *Mar. Pollut. Bull.* 87 (1–2): 381-387. <https://doi.org/10.1016/j.marpolbul.2014.07.021>.
363
- 364 Cresson, P., Fabri, M.C., Bouchoucha, M., Brach Papa, C., Chavanon, F., Jadaud, A., Knoery, J., Miralles, F.,
365 Cossa, D., 2014. Mercury in organisms from the Northwestern Mediterranean Slope: importance of food
366 sources. *Sci. Total. Environ*, 497-498: 229–238.
367
- 368 Dabrowski, J.M., Ashton, P.J., Murray, K., Leaner, J.J., Mason, R.P., 2008. Anthropogenic mercury emissions in
369 South Africa: Coal combustion in power plants. *Atmos Environ*, 42 (27): 6620-6626.
370 <https://doi.org/10.1016/j.atmosenv.2008.04.032>
371
- 372 Davis, B., VanderZwaag, D.L., Cosandey-Godin, A., Hussey, N.E., Kessel, S.T. and Worm, B., 2013. The
373 Conservation of the Greenland Shark (*Somniosus microcephalus*): Setting Scientific, Law, and Policy
374 Coordinates for Avoiding a Species at Risk. *J Int Wildl Law Policy*, 16:4, 300-330.

- 375 <https://doi.org/10.1080/13880292.2013.805073>
- 376
- 377 Devlin, E.W., 2006. Acute toxicity, uptake and histopathology of aqueous methyl mercury to fathead minnow
- 378 embryos. *Ecotoxicology*, 15(1): 97-110. <https://doi: 10.1007/s10646-005-0051-3>
- 379
- 380 Ehnert, S.L., 2017. Mercury Accumulation and Effects in the Brain of Atlantic Sharpnose Sharks
- 381 (*Rhizoprionodon Terranova*). UNF Graduate Theses and Dissertations. 736.
- 382 <https://digitalcommons.unf.edu/etd/736>
- 383
- 384 Fashola, M.O., Ngole-Jeme, V.M., Babalola O.O., 2016. Heavy Metal Pollution from Gold Mines:
- 385 Environmental Effects and Bacterial Strategies for Resistance. *Int J Environ Res Public Health*. 2016 Oct
- 386 26;13(11):1047. doi: 10.3390/ijerph13111047.
- 387
- 388 Ferreira Araujo, B.F., Osterwalder, S., Szponar, N., Lee, D., Petrova, M.V., Pernov, J.B., Ahmed, S.,
- 389 Heimbürger-Boavida, L.E., Laffont, L., Tesserenc, R., Tananaev, N., Nordstrom, C., Magand, O., Stupple, G.,
- 390 Skov, H., Steffen, A., Bergquist, B., Aspmo Pfaffhuber, K., Thomas, J.L., Scheper, S., Petäjä, T., Dommergue,
- 391 A., Sonke, J.E., 2022. Mercury isotope evidence for Arctic summertime re-emission of mercury from the
- 392 cryosphere. *Nat Commun* 13, 4956. <https://doi.org/10.1038/s41467-022-32440-8>
- 393
- 394 Gibbs, R.H. Jr, Jarosewich, E., Windom, H.L. 1974. Heavy metal concentrations in museum fish specimens:
- 395 effects of preservatives and time. *Science*, 184(4135):475-7. doi: 10.1126/science.184.4135.475.
- 396
- 397 Guzzi, G., Ronchi, A., Pigatto, P., 2021. Toxic effects of mercury in humans and mammals,
- 398 *Chemosphere*, 263:127990, <https://doi.org/10.1016/j.chemosphere.2020.127990>.
- 399
- 400 Ha, E., i Basu, N., Bose-O'Reilly, S., Dórea, J.G., McSorley, E., Sakamoto, M., Chan, H.M., 2017. Current
- 401 progress on understanding the impact of mercury on human health, *Environ Res*, 152: 419-433.
- 402 <https://doi.org/10.1016/j.envres.2016.06.042>
- 403
- 404 Harding, G., Dalziel, J., Vass, P., 2018. Bioaccumulation of methylmercury within the marine food web of the
- 405 outer Bay of Fundy, Gulf of Maine. *PLoS ONE* 13(7): e0197220. <https://doi.org/10.1371/journal.pone.0197220>

406

407 Hoenig, J.M. and Walsh A.H., 1986. Skeletal lesions and deformities in large sharks. *J Wildl Dis*, 19(1): 27-33.

408

409 ICES., 2022. Eurostat/ICES data compilation of catch statistics. Official Nominal Catches 2006-2020.

410 Copenhagen. Format: Archived dataset in .xlsx and .csv formats. [https://www.ices.dk/data/dataset-](https://www.ices.dk/data/dataset-collections/Pages/Fish-catch-and-stock-assessment.aspx)

411 [collections/Pages/Fish-catch-and-stock-assessment.aspx](https://www.ices.dk/data/dataset-collections/Pages/Fish-catch-and-stock-assessment.aspx). Accessed on 15 September 2023

412

413 Jagoe, C.H., Faivre, A., Newman, M.C., 1996. Morphological and morphometric changes in the gills of

414 mosquitofish (*Gambusia holbrooki*) after exposure to mercury (II), *Aquat Toxicol*, 34 (2):163-183.

415 [https://doi.org/10.1016/0166-445X\(95\)00033-Z](https://doi.org/10.1016/0166-445X(95)00033-Z)

416

417 Johannessen, O.M., Bengtsson, L., Miles, M.W., Kuzmina, S.I., Semenov, V.A., Alekseev, G.V., Nagurnyi,

418 A.P., Zakharov, V.F., Bobilev, L.P., Pettersson, L.H., Hasselmann, K. and Cattle, H.P., 2004. Arctic climate

419 change: observed and modelled temperature and sea-ice variability. *Tellus A: Dyn Meteorol Oceanogr*, 56(4),

420 p.328–341.DOI: <https://doi.org/10.3402/tellusa.v56i4.14418>

421

422 Kirk, J.L., St. Louis, V.L., Hintelmann, H., Lehnher, I., Else, B., Poissant, L., 2008. Methylated mercury species

423 in marine waters of the Canadian high and sub-Arctic. *Environ Sci Technol*, 42:8367–73.

424

425 Larroure, Q.C., Cribbs, A.P., Rao, S.R., Philpott, M., Snelling, S.J., Knwoles, H.J., 2021. Loss of mutual

426 protection between human osteoclasts and chondrocytes in damaged joints initiates osteoclast-mediated cartilage

427 degradation by MMPs. *Sci Rep* 11, 22708. <https://doi.org/10.1038/s41598-021-02246-7>

428

429 Levensgood, J.M., Soucek, D.J., Taylor, C.A., Gay, D.A., 2013. Mercury in small Illinois fishes: historical

430 perspectives and current issues. *Environ Monit Assess*, 185: 6485–6494. [https://doi.org/10.1007/s10661-012-](https://doi.org/10.1007/s10661-012-3040-z)

431 3040-z

432

433 Li, Z., Pethybridge, H.R., Wu, F., Li, Y., 2023. Mercury bioaccumulation in thresher sharks from the eastern

434 tropical Pacific: Influences of body size, maturation stage, and feeding habitat, *Sci Total Environ*, 872, 162248.

435 <https://doi.org/10.1016/j.scitotenv.2023.162248>.

436

- 437 Mancini, P.L., Casas, A.L. and Amorim, A.F., 2006. Morphological abnormalities in a blue shark *Prionace*
438 *glauca* (Chondrichthyes: Carcharhinidae) foetus from southern Brazil. *J Fish Biol* 69, 1881–1884.
439
- 440 Marconi, A., Hancock-Ronemus, A., Gillis, A.J., 2020. Adult chondrogenesis and spontaneous cartilage repair in
441 the skate, *Leucoraja erinacea*. *eLife* 9:e53414. <https://doi.org/10.7554/eLife.53414>
442
- 443 Mason, R.P., Fitzgerald, W.F., Morel F.M.M., 1994. The aquatic biogeochemistry of elemental mercury. *Geochi.*
444 *Cosmochim. Acta* 58: 3191-3198.
445
- 446 Mason, R.P., Rolffhus, K.R., Fitzgerald, W.F., 1998. Mercury in the North Atlantic. *Mar Chem*, 61: 37–53.
447
- 448 McMeans, B.C., Borgå, K., Bechtol, W.R., Higginbotham, D., Fisk, A.T., 2007. Essential and non-essential
449 element concentrations in two sleeper shark species collected in arctic waters. *Environ Pollut*, 148 (1): 281-290.
450 <https://doi.org/10.1016/j.envpol.2006.10.039>.
451
- 452 McMeans, B.C., Svavarsson, J., Dennard, S., Fisk, A., 2010. Diet and resource use among Greenland sharks
453 (*Somniosus microcephalus*) and teleosts sampled in Icelandic waters, using $d^{13}C$, $d^{15}N$, and mercury. *Can J Fish*
454 *Aquat Sci*, 67: 1428–1438.
455
- 456 McMeans, B.C., Arts, M.T., Fisk, A.T., 2015. Impacts of food web structure and feeding behavior on mercury
457 exposure in Greenland Sharks (*Somniosus microcephalus*). *Sci Total Environ*, 509–510: 216-225.
458 <https://doi.org/10.1016/j.scitotenv.2014.01.128>.
459
- 460 Mered, M., 2020. Les mondes polaires. Presse universitaire de France, Paris. 521pp.
461
- 462 Merly, L., Lange, L., Meÿer, M., Hewitt, M.A., Koen, P., Fischer, C., Muller, J., Schilack, V., Wentzel, M.,
463 Hammerschlag, N., 2019. Blood plasma levels of heavy metals and trace elements in white sharks (*Carcharodon*
464 *carcharias*) and potential health consequences, *Mar Pollut Bull*, 142: 85-92.
465 <https://doi.org/10.1016/j.marpolbul.2019.03.018>
466

- 467 Meyer, W. and Seegers, U., 2012. Basics of skin structure and function in Elasmobranch: a review. J Fish Biol,
468 80 (5): 1940-1967. <https://doi.org/10.1111/j.1095-8649.2011.03207.x>
469
- 470 Moore, A.B.M., 2015. Morphological abnormalities in elasmobranchs. J Fish Biol, 87: 465–471.
471
- 472 Nielsen, J., Hedeholm, R.B., Simon, M., Steffenson, J.F., 2014. Distribution and feeding ecology of the
473 Greenland shark (*Somniosus microcephalus*) in Greenland waters. Polar Biol 37: 37-46.
474 <https://doi.org/10.1007/s00300-013-1408-3>
475
- 476 Nielsen, J., Hedeholm, R.B., Lynghammar, A., McClusky, L.M., Berland, B., Steffensen, J.F., Christiansen, J.S.,
477 2020. Assessing the reproductive biology of the Greenland shark (*Somniosus microcephalus*). PLoS ONE
478 15(10): e0238986. <https://doi.org/10.1371/journal.pone.0238986>
479
- 480 Oliveira Ribeiro, C.A., Guimaraes, J.R., Pfeiffer, W.C., 1996. Accumulation and distribution of inorganic
481 mercury in a tropical fish (*Trichomycterus zonatus*). Ecotoxicol. Environ. Saf. 34: 190–195.
482
- 483 Oliveira Ribeiro, C.A., Pelletier, E., Pfeiffer, W.C., Rouleau, C., 2000. Comparative Uptake, Bioaccumulation,
484 and Gill Damages of Inorganic Mercury in Tropical and Nordic Freshwater Fish, Environ Res, 83 (3): 286-292.
485 <https://doi.org/10.1006/enrs.2000.4056>
486
- 487 Porter, M.E., Beltrán, J.L., Koob, T.J., Summers, A.P., 2006. Material properties and biochemical composition
488 of mineralized vertebral cartilage in seven elasmobranch species (Chondrichthyes). J Exp Biol. 209(Pt 15):2920-
489 8. <https://doi:10.1242/jeb.02325>
490
- 491 Ramírez-Amaro, S., Fernández-Peralta, L., Serna, F., Puerto, M.A., 2019. Abnormalities in two shark species,
492 the blue shark, *Prionace glauca*, and the school shark, *Galeorhinus galeus* (Elasmobranchii: Carcharhiniformes),
493 from the Canary Islands, eastern tropical Atlantic. Acta Ichthyol Piscat 49 (3): 295–303.
494
- 495 Rodríguez, J., Mandalunis, P.M., 2018. A Review of Metal Exposure and Its Effects on Bone Health. J
496 Toxicol.:4854152. <https://doi:10.1155/2018/4854152>.
497

- 498 Rodríguez-Romero, J., Simeón-de la Cruz, A., Ochoa-Díaz, M.R. and Monsalvo-Spencer, P., 2018. New report
499 of malformations in blue shark embryos (*Prionace glauca*) from the western coast of Baja California Sur,
500 Mexico. *J Mar Biolog Assoc U.K.*, 99(2): 1-6. [https://doi: 10.1017/s0025315418000127](https://doi.org/10.1017/s0025315418000127)
501
- 502 Rosa, R.S., Mariano, E.F., Sampaio, C.L.S., 2004. Má-formação em *Rhinobatus percellens* Jord and Evern,
503 1896; Rhinobatidae na Baía de Todos os Santos, BA. In Reunião da Sociedade Brasileira para Estudo em
504 Elasmobrânquios SBEEL N. Caderno de resumos, pp. 165–166. Recife: SBEEL.
505
- 506 Rumbold, D., Wasno, R., Hammerschlag, N., Volety, A., 2014. Mercury accumulation in sharks from the coastal
507 waters of southwest Florida. *Arch Environ Contam Toxicol*, 67 (3): 402-412. [https://doi.org/10.1007/s00244-](https://doi.org/10.1007/s00244-014-0050-6)
508 [014-0050-6](https://doi.org/10.1007/s00244-014-0050-6)
509
- 510 Seidel, R., Jayasankar, A.K., Dean, M.N., 2021. The multiscale architecture of tessellated cartilage and
511 its relation to function. *J Fish Biol*, 98:942–955. <https://doi.org/10.1111/jfb.14444>.
512
- 513 Schaefer, K., Elshorbany, Y., Jafarov, E., Schuster, P.F., Striegl R.G., Wickand K.P., Sunderland, E.M., 2020.
514 Potential impacts of mercury released from thawing permafrost. *Nat Commun*, 11, 4650.
515 <https://doi.org/10.1038/s41467-020-18398-5>
516
- 517 Shields, P.A., Carlson, S.R. 1996. Effects of formalin and alcohol preservation on lengths and weights of
518 juvenile sockeye salmon. *Alaska Fishery Research Institute*, 3: 81–93.
- 519 Ste-Marie, E., Watanabe, Y.Y., Semmens, J.M., Marcoux, M., Hussey, N.E., 2022. Life in the slow lane: field
520 metabolic rate and prey consumption rate of the Greenland shark (*Somniosus microcephalus*) modelled using
521 archival biologgers. *J Exp Biol*, 225, jeb242994. <https://doi.org/10.1242/jeb.242994>
522
- 523 Steiner, N.S., Cheung, W.W.L., Cisneros-Montemayor, A.M., Drost, H., Hayashida, H., Hoover, C., Lam, J.,
524 Sou, T., Sumaila, U.R., Suprenand, P., Tai, T.C., VanderZwaag, D.L., 2019. Impacts of the Changing Ocean-Sea
525 Ice System on the Key Forage Fish Arctic Cod (*Boreogadus Saida*) and Subsistence Fisheries in the Western
526 Canadian Arctic—Evaluating Linked Climate, Ecosystem and Economic (CEE) Models. *Front Mar Sci* 6: 179.
527 [https://doi: 10.3389/fmars.2019.00179](https://doi.org/10.3389/fmars.2019.00179)
528

- 529 Suzuki, N., Yamamoto, N., Watanabe, A., Kambegawa, A. and Hattori, A., 2004. Both mercury and cadmium
530 directly influence calcium homeostasis resulting from the suppression of scale bone cells: The scale is a good
531 model for the evaluation of heavy metals in bone metabolism. *J Bone Miner Metab*, 22 (5): 439-446.
532
- 533 Taniuchi, T. and Yanagisawa, F., 1987. Albinism and lack of second dorsal fin in an adult tawny nurse shark,
534 *Nebrius concolor*, from Japan. *Jpn J Ichthyol*, 34: 393–395.
535
- 536 Tiktak, G.P., Butcher, D., Lawrence, P.J., Norrey, J., Bradley, L., Shaw, K., Preziosi, R., Megson, D., 2020. Are
537 concentrations of pollutants in sharks, rays and skates (Elasmobranchii) a cause for concern? A systematic
538 review, *Mar Pollut Bull*, 160, 111701. <https://doi.org/10.1016/j.marpolbul.2020.111701>.
539
- 540 Travel Food Atlas., 2020. "Hákarl: Iceland's Rancid Fermented Shark
541 Delicacy" <https://travelfoodatlas.com/hakarl-iceland-smelly-fermented-shark-delicacy>. Accessed on 15
542 September 2023
543
- 544 Trischitta, F., Faggio, C., Torre, A., 2012. Living with high concentrations of urea: They can!. *Open J Anim*
545 *Sci*, 2: 32-40. <https://doi.or/10.4236/ojas.2012.21005>
546
- 547 United Nations., 1992. Report of the United Nations Conference on Environment and Development.
548 *A/CONF.151/26 (Vol. I)*. 5pp.
549 [https://www.un.org/en/development/desa/population/migration/generalassembly/docs/globalcompact/A_CONF.](https://www.un.org/en/development/desa/population/migration/generalassembly/docs/globalcompact/A_CONF.151_26_Vol.I_Declaration.pdf)
550 151_26_Vol.I_Declaration.pdf. Accessed on 15 September 2023
551
- 552 Urist, M.R., 1961. Calcium and phosphorus in the blood and skeleton of the Elasmobranchii. *Endocrinol* 69,
553 778-801. <https://doi.org/10.1210/endo-69-4-778>
554
- 555 Van Hees, K.E., Ebert, D.A., 2017. An evaluation of mercury offloading in two Central California
556 elasmobranchs. *Sci Total Environ* 590–591, 154-162. <https://doi.org/10.1016/j.scitotenv.2017.02.191>
557

- 558 Veiga, M.M., Maxson, P.A., Hylander, L.D., 2006. Origin and consumption of mercury in small-scale gold
559 mining. *J Clean Prod* 14 (3–4), 436-447. <https://doi.org/10.1016/j.jclepro.2004.08.010>
560
- 561 Walter, P.R., Roy, D, Hussey, N.E., Stelbrinck, B., Kovacs, K.M., Lydersen, C., MacMeans, N.C., Svarvasson,
562 J., Kessel, S.T., Biton Porsmoguer, S., Wildes, S., Tribuzio, C.A., Campana, S.E., Petersen, S.D., Dean Grubbs,
563 R., Heath, D.D., Hedges, K.J., Fisk, A.T., 2017. Origins of the Greenland shark (*Somniosus microcephalus*):
564 impacts of the ice-olation and Introgression. *Ecol Evol* 7(19), 8113-8125. <https://doi.org/10.002/ece3.3325>
565
- 566 Weis, P., Weis, J.S., 1977. Methylmercury teratogenesis in the killifish, *Fundulus heteroclitus*. *Teratology*, 16,
567 317-325. <https://doi.org/10.1002/tera.1420160311>
568
- 569 Weis, J.S., Weis, P., Herbe, M., Vaidya, S., 1981. Methylmercury tolerance of killifish (*Fundulus heteroclitus*)
570 embryos from a polluted vs non-polluted environment. *Mar Biol* 65, 283-287.
571 <https://doi.org/10.1007/BF00397123>
572
- 573 Witt, E.L., Kolka, R.K., Nater, E.A., Wickmann, T.R., 2009. Influence of the Forest Canopy on Total and
574 Methyl Mercury Deposition in the Boreal Forest. *Water Air Soil Pollut* 199, 3-11.
575 <https://doi.org/10.1007/s11270-008-9854-1>
576
- 577 Yachiguchi, K., Sekiguchi, T., Nakano, M., Hattori, A., Yamamoto, M., Kitamura, K., Maeda, M., Tabuchi, Y.,
578 Kondo, T., Kamauchi, H., Nakabayashi, H., Srivastav, A.K., Hayakawa, K., Sakamoto, T., Suzuki, N., 2014.
579 Effects of inorganic mercury and methylmercury on osteoclasts and osteoblasts in the scales of the marine teleost
580 as a model system of bone. *Zoolog Sci.* 31(5), 330-7. <https://doi.org/10.2108/zs130265>

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583 FIGURES CAPTIONS

584

585 Fig. 1. Position of the Greenland shark fished and sampled off Northwest Iceland marked by red cross. A. Skin.
586 B. Upper jaw. C. Fresh meat. D. Lower jaw. Female. Total Length 434 cm. Greenland Sea. August 2010.

587

588 Fig.2. Hg level (mg kg⁻¹ dw) in selected tissues of the Greenland shark: cartilage (upper and lower jaw), skin and
589 muscle (fresh meat).

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595 TABLES

596

597 Table 1. Range of total Hg levels (mg kg^{-1} ww) in the Greenland shark (*Somniosus microcephalus*) in the Arctic

598 and adjacent seas; - No data. dw: dry wet; ww: wet weight; dry weight (dw) concentration = 5 ww concentration

599 (Cresson et al., 2014).

600

Regions	Greenland shark							References
	TL (cm)	upper jaw	lower jaw	muscle	skin	liver	ww/dw	
Greenland Sea	434	1.332	1.985	3.225	2.625	-	dw	<i>This study</i>
Cumberland Sound	236-325*	-	-	-	-	4.92±0.58	dw	McMeans et al. 2007
Islandic waters	415.6±25.2	-	-	5.93±0.59	-	5.32±2.73	dw	McMeans et al. 2010
Pangnirtung, Baffin Bay	293±1.1	-	-	1.715±0.457	-	-	ww	Chételat and Braune 2012
NE Greenland fjords	-	-	-	4.10-6.91	-	-	dw	Corsolini et al. 2014
Labrador Sea	273.3±31.5	-	-	3.54±1.02	-	-	dw	McMeans et al. 2015

601 TL = Total Length of individuals (cm); *fork length (cm); Hg = Total Hg concentration (mg kg^{-1})

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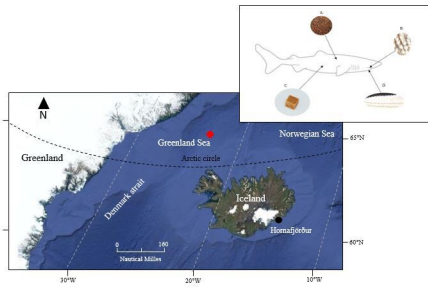
605

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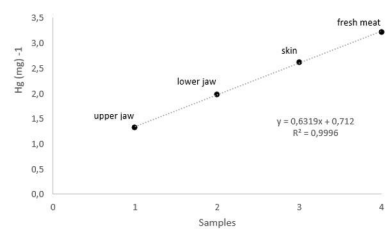
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Declaration of competing interest

The authors have no competing interests to declare that are relevant to the content of this article.

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