Ecology of *Bathymodiolus puteoserpentis* mussels from the Snake Pit vent field (Mid-Atlantic Ridge)

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Abstract :

Along the northern Mid-Atlantic Ridge (nMAR), in habitats under moderate (<10°C) hydrothermal influence on the Snake Pit vent field (SP), large assemblages dominated by Bathymodiolin mussels remain poorly characterised, contrary to those in warmer habitats dominated by gastropods and alvinocaridid shrimps that were recently described. In this study, we assessed and compared the population structure, biomass, diversity and trophic interactions of two Bathymodiolus puteoserpentis assemblages and their associated fauna at SP. Three sampling units distanced by 30 cm were sampled in 2014 during the BICOSE cruise at the top of the Moose site ("Elan" site), while few meters further down three others, distanced by ~ 1 m were obtained in 2018 during the BICOSE 2 cruise at the edifice's base. We observed a micro-scale heterogeneity between these six sampling units partially explained by temperature variations, proximity to hydrothermal fluids and position on the edifice. Meiofauna dominate or co-dominate most of the sampling units, with higher densities at the base of the edifice. In terms of macrofauna, high abundance of Pseudorimula midatlantica gastropods was observed at the top of the vent edifice, while numerous Ophioctenella acies ophiuroids were found at the base. Contrary to what was expected, the apparent health and abundance of mussels seems to indicate a current climax stage of the community. However, the modification of B. puteoserpentis isotopic signatures, low number of juveniles decreasing over the two years and observations made during several French cruises in the study area raise questions about the fate of the B. puteoserpentis population over time, which remains to be verified in a future sampling campaign.

Highlights

► Assemblage heterogeneity at different scales linked to fluid proximity & location. ► Meiofaunal groups dominated. Macrofaunal disparity between top and base of edifice. ► Low number of mussel juveniles over time questions the future of the population.

Keywords : benthic ecology, associated fauna, habitat, temperature, biodiversity, hydrothermal activity, condition index, food webs, temporal variability

28 1. Introduction

29 Hydrothermal vents sustain islands of productivity and biomass of endemic fauna (Tunnicliffe, 1991; Van 30 Dover, 2002), thanks to the local production of organic matter by chemosynthetic microorganisms. The 31 latter are the major primary producers of hydrothermal ecosystems as they fix inorganic carbon fueling 32 macrofaunal communities dominated by large size taxa such as siboglinid tubeworms, provannid 33 gastropods or bathymodiolin mussels. These large foundation species (Dayton, 1972) live in symbiosis with 34 microorganisms with which they form functional units called holobionts (Zilber-Rosenberg and Rosenberg, 35 2008). These holobionts provide a substratum, a refuge as well as reproduction and feeding grounds to other 36 smaller associated macrofaunal and meiofaunal taxa, therefore contributing to locally enhance species 37 diversity (Goroslavskaya and Galkin, 2011; Govenar and Fisher, 2007; Sarrazin et al., 2020; Van Dover, 38 2002, 2003; Zekely et al., 2006). Vent faunal assemblages are patchily distributed within distinct niches

39 along a physico-chemical gradient from high-temperature fluid emission zones to their peripheries, 40 according to their tolerance to environmental conditions (Tunnicliffe, 1991; Vismann, 2012) and nutritional 41 needs (Govenar et al., 2005; Govenar and Fisher, 2007; Luther et al., 2001; Sarrazin et al., 1997, 2015). 42 Biotic interactions also play a major role in structuring these assemblages. They include positive 43 interactions such as facilitation (Gollner et al., 2015; Micheli et al., 2002; Mullineaux et al., 2000, 2003; 44 Sarrazin et al., 1997) and creation of habitats by engineer species, resulting in resource sharing (Bergquist 45 et al., 2007; Lelièvre et al., 2018; Levesque et al., 2003, 2006) and negative interactions such as predation (Hunt et al., 2004; Micheli et al., 2002), competition (Lenihan et al., 2008) and territoriality (Matabos et 46 47 al., 2015). These different assemblages are strongly influenced by the environmental characteristics at the 48 edifice and vent field scales (Sarrazin et al., 2020). Variations in abiotic conditions may in particular 49 influence the growth conditions of microorganisms, and therefore their production and diversity (Olins, 50 2016: Sievert et al., 1999). Consequently, modifications of the microbial activity can trigger changes at the 51 symbiotic compartment level, by modifying the proportion of endosymbionts and therefore, the availability 52 of energy sources (Lutz and Kennish, 1993), which can impact the fitness of the holobionts. Foundation 53 species may also influence the physico-chemical conditions by modifying fluid flow and composition 54 through their 3D structure and also, the metabolic activity of their symbionts (Govenar and Fisher, 2007). 55 Moreover, a change in the distribution and density of these engineer species may have significant effects 56 on the composition and abundance of their associated fauna (Govenar, 2010).

57 Among these engineer taxa, Bathymodiolus spp. mytilids (Kenk & Wilson, 1985) aggregate in more or less 58 extended patches along the Mid-Atlantic Ridge (MAR). Two symbiont-bearing species of the genus Bathymodiolus are found on the northern MAR: Bathymodiolus azoricus (800-2300m, Cosel & Comtet, 59 60 1999) dominates biological communities at Lucky Strike (LS, 37°09'N) and Menez Gwen (MG, 37°50'N) 61 vent fields, and is present at Rainbow (36°14'N) and Lost City (30°07'N) while Bathymodiolus 62 puteoserpentis (2324-3510m, Cosel, Métiver & Hashimoto, 1994) is present more south at Broken Spur 63 (29°10'N), Snake Pit (SP, 23°22'N), Logatchev (14°45'N), Semyenov (13°31'N) and probably Irinovskoe 64 (13°19'N) vent fields (Franke et al., 2021; Gebruk et al., 2000; Gollner et al., 2021; Maas et al., 1999; Ücker 65 et al., 2021, Van Cosel et al., 1999). Broken Spur is a hybrid zone (O'Mullan et al., 2001) where B. 66 puteoserpentis and hybrids between B. puteoserpentis and B. azoricus co-occur (Ücker et al., 2021). These 67 two species live in symbiosis with two symbiont lineages: sulfide-oxidising (also called thiotrophic) and 68 methanotrophic bacteria (Distel et al., 1995; Duperron et al., 2006, 2016; Fiala-Médioni et al., 2002; Fisher 69 et al., 1993). In hydrothermal ecosystems, sulphur and methane oxidation are the main carbon fixation and 70 energy-acquisition pathways (Conway et al., 1994; Levin and Michener, 2002). Chemoautotrophic sulphur 71 oxidizers use CO_2 as a carbon source while methane oxidizers use methane (Duperron et al., 2006), ensuring 72 holobiont fitness in various environmental conditions. In addition, these mussels have kept the ability to

73 filter-feed and can exploit photosynthetic energy sources by feeding on sinking organic particles (e.g. Page

et al., 1991; Piquet et al., 2022). Nevertheless, the quantity of organic matter available in the deep sea is not

rsufficient to sustain these mussels without the help of their symbionts (Raulfs et al., 2004). *Bathymodiolus*

76 mussels are often colonised by *Branchipolynoe seepensis* polychaetes (Pettibone, 1986), considered either

a commensal (Chevaldonné et al., 1998; Pettibone, 1986; Van Dover et al., 1999) or at least a semi-parasite

78 (Britaev et al., 2003; Britayev et al., 2003,2007; Ward et al., 2004) species.

79 In the present study, we focus on *Bathymodiolus puteoserpentis* mussel assemblages from the Snake Pit 80 (SP) vent field located on the northern MAR. At this vent field, warmer habitats host assemblages 81 dominated by Peltospira smaragdina gastropods (Sarrazin et al., 2022) and Rimicaris exoculata shrimps 82 (Hernández-Ávila et al., 2022; Methou et al., 2022) while B. puteoserpentis inhabits areas with lower 83 hydrothermal influence. These assemblages are distributed according to their proximity to fluid emission 84 and geomorphology of the sulfide structures. The largest known mussel assemblages of SP colonise the 85 Moose Site (23°22'5"N; 44°56'59"W) at a depth of ~3510 meters. The structure and environmental drivers 86 of B. puteoserpentis mussels have not been described in details at SP, contrary to the well-known 87 Bathymodiolus azoricus assemblages found at the northernmost sites Menez Gwen, Lucky Strike and 88 Rainbow (Husson et al., 2017; Sarrazin et al., 2015, 2020). Previous studies have shown that B. 89 puteoserpentis assemblages at SP host 40 taxa, dominated by abundant copepods and nematodes 90 (Goroslavskaya and Galkin, 2011, Zekely et al., 2006), Ophioctenella acies ophiuroids, Pseudorimula 91 midatlantica gastropods, Rimicaris shrimps, and several polychaete species (Goroslavskaya and Galkin, 92 2011; Turnipseed et al., 2004, Zekely et al., 2006). The observed spatial heterogeneity of mussel 93 assemblages was hypothesised to be driven by heterogeneous, fluctuating and abiotic conditions as 94 observed for B. puteoserpentis assemblages at Logatchev (Zielinski et al., 2011) but these remain to be 95 characterised at SP (Turnipseed et al., 2004). Living populations of *B. puteoserpentis* are currently known 96 from a few vent fields (including BS, SP, Logatchev and Semyenov) and probably Irinovskoe (Gollner et 97 al., 2021). Other populations, possibly extinct or declining, were reported at Podeba (17°08'N; (Molodtsova 98 et al., 2017) and Logatchev-4, but these mussels were never observed at the TAG nor Ashadze vent fields. 99 Thus, SP is currently the deepest site of the known bathymetric range of *B. puteoserpentis*. Very low 100 recruitment rate was reported in 2001 in this field, which may suggest mussel populations living in suboptimal conditions (Turnipseed et al., 2004). 101

Overall, and in comparison with its northern and shallower counterpart *B. azoricus*, knowledge on
 population structure of *B. puteoserpentis* as well as on associated living communities remain scarce,
 especially regarding their temporal dynamics. In order to understand how *B. puteoserpentis* assemblages
 are structured in space and time, and compare our results with *B. azoricus* assemblages, we sampled *B.*

puteoserpentis assemblages in 2014 and 2018 at the Moose site at the SP vent field more than 10 years after
those sampled in previous studies (Goroslavskaya and Galkin, 2011; Turnipseed et al., 2003, 2004; Zekely
et al., 2006).

109 Our aim here is to (i) describe B. puteoserpentis assemblages in terms of population structure, mussel 110 biomass and density, (ii) characterise the associated fauna and (iii) examine trophic relationships among 111 the assemblages. In addition, we aim to identify the main environmental drivers structuring B. 112 *puteoserpentis* communities at this site by comparing our results with those of previous studies. We 113 hypothesise that the structure of *B. puteoserpentis* assemblages will differ according to their proximity to 114 hydrothermal emissions, with lower specific abundance of associated fauna and higher mussel biomass in 115 warmer habitats. From our first field observations and a previous study (Turnipseed et al. in 2004), we also 116 hypothesise that *B. puteoserpentis* assemblages from SP may soon start to decline because of the lack of 117 juveniles. Our study is part of a global description and comprehension of hydrothermal fauna assemblages 118 located along the MAR, providing a dataset of their temporal evolution, testing the hypothesis of the 119 presence of declining communities.

120 2. Material and methods

121 **2.1. Study site**

122 The Snake Pit (SP) active vent field (23°22'06"N; 44°56'59"W, 3500 m depth, Figure 1) is located on the 123 MAR at approximately 25 km south of the Kane Fracture Zone, a deep fault that offsets the MAR left-124 laterally over 150 km (Durant et al., 1996; Karson et al., 1987). SP is 600 m long and 200 m wide, composed of recent hydrothermal deposits (< 4000 years, Lalou et al., 1993), and covers around 15 000 m² of the 125 126 seafloor. The active zone is composed of several black smokers separated by slopes and hydrothermal 127 sediment deposits (Brown and Karson, 1988; Karson et al., 1987). The active sulfide sites such as the 128 Beehive ("Les Ruches") or Moose ("Elan") structures (Fouquet et al., 1993) host high-temperature black 129 smoker chimneys, beehives and low-temperature diffuse emission zones as well as inactive zones (Brown 130 and Karson, 1988).



FIGURE 1. Location of the Moose sampling site (« Elan site », 23°22'.06''N; 44°56'.99''W) on the Snake Pit vent field along the Mid-Atlantic Ridge (MAR).

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132 Snake Pit is characterised by a high mineralogical and geochemical diversity (Charlou et al., 2010). The 133 average background seawater temperature is 2.6°C (Hernández-Ávila et al., 2022). The vent field is 134 colonised by Rimicaris exoculata shrimps in the warmest areas with high sulfide concentrations (Methou 135 et al., 2022), Peltospira smaragdina gastropods at intermediate temperatures and sulfide concentrations 136 (Sarrazin et al., 2022) and Bathymodiolus puteoserpentis mussels in lower temperatures and hydrogen 137 sulfide concentrations (Sarrazin et al., 2022). Our study focuses on the Moose site (23°22'5"N; 138 44°56'59"W, 3510 m), which occupies an area of 80 m² on the seafloor with a complex asymmetrical 139 structure composed of active and inactive chimneys as well as beehive-shaped structures, all together 140 resembling the antlers of a moose (Figure 2, a-c).

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142 2.2. Characterisation of environmental conditions and faunal sampling

143 Environmental characterisation and sampling were done during two different cruises: a first cruise on the 2014 144 R/V ROV Victor6000 Pourquoi pas? with the in (BICOSE 1. DOI: 145 https://doi.org/10.17600/14000100) cruise 2018 and а second in (BICOSE 2, DOI: 146 http://dx.doi.org/10.17600/18000004) with the same research vessel but with the manned submersible 147 Nautile. Mussel assemblages were sampled on the Moose site Figure 2 a-c). For each year, three 27 cm-148 long calibration targets CASIMIR were deployed on top of the mussels in order to visually estimate the

149 distance between sampling units (SUs) thanks to visible 7 x 7 mm checkerboards. In 2014, these targets 150 were placed directly on *B. puteoserpentis* assemblage located at the top of the edifice, approximately thirty 151 centimetres from each other (Figure 2 d), defining three sampling areas (SU-1, SU-2, SU-3, Table 1). In 152 2018, these CASIMIR targets were placed on a mussel assemblage at the base of the same edifice, spaced 153 by approximately one metre (Figure 2 e-g), defining three additional sampling areas (SU-4, SU-5, SU-6, 154 Table 1). Moreover, in 2014 only, each CASIMIR was equipped with autonomous MISO probes to measure 155 variations of *in situ* temperatures directly above the mussels (few cms) during 4 days and 14 hours (MISO 156 New, Low-T Logger Manual, precision: $+/-0.22^{\circ}$ C at 25°C, measurement frequency: 30 s). 157 158 The *in situ* chemical analyser CHEMINI (Vuillemin et al., 2009) was used to characterise sulfide and iron 159 concentrations with sampling probes directly above the fauna (few cms) near each CASIMIR target. The 160 submersible temperature sensor, associated with the sampling cannula of the CHEMINI, provided

161 temperature conditions from which we calculated the mean, minimum and maximum. The *in situ* water 162 samplers PEPITO (2014), PIF (2018) and PLUME (2018) designed for the Nautile sampling, collected 163 fluids directly above the fauna for subsequent chemical analyses in the laboratory. pH was measured on 164 fluid samples upon arrival on board. Fe^{2+} values and methane concentrations (obtained from water samples) 165 were only obtained from a single sampling unit, SU-3 for Fe²⁺ and SU-1 for methane and therefore do not 166 appear in Table 1.



FIGURE 2. Sampling of mussel assemblages on the Moose edifice on the Snake Pit vent field. Left (a-b-c) : large views of the Moose edifice taken with the 4K camera of the Nautile showing the top (a), the global structure (b) and the base (c) of the sulphide edifice. (d) Sampling sites during BICOSE 2014 ("CASIMIR C1" corresponds to the sampling unit SU-1, "CASIMIR C2" corresponds to SU-2, "CASIMIR C3" corresponds to SU-3). (e-f-g) Sampling sites during BICOSE 2018 (CASIMIR 1 (e) corresponds to SU-4, CASIMIR 2 (f) corresponds to SU-5, CASIMIR 1 (g) corresponds to SU-6).

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After the environmental characterization, the mechanical arms of the submersibles meticulously removed the mussels and their associated fauna, and placed them in small isotherm sampling boxes. Sampling was completed using the submersible suction samplers to account for the smaller and/or mobile fauna (see sampling protocol in Sarrazin et al., 2015) on all SUs except SU-1 and SU-2 (Table 1). Environmental data are summarised in Tables 1 & 2.

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Mussel assemblage	Year	Dive	PBT	SS	T°C mean (std)	T°C max	T°C min	T°C range	T°C anomaly	рН	H ₂ S (std) (µM)	Fe _{total} (std) (µM)
SU-1	2014	565-02	~	Х	4.3 (0.9)	9.3	2.2	7.1	1.7	8	10.2 (4)	-
SU-2	2014	565-02	~	Х	2.7 (0.2)	4.0	2.2	1.8	0.1	-	0.4 (0.2)	-
SU-3	2014	565-02	~	~	2.6 (0.1)	3.1	2.3	0.8	0	-	0.7 (0.2)	-
SU-4	2018	1922-16	✓	~	3.2 (0.1)	3.6	3.0	0.6	0.6	7.8	n.d.	4.4
SU-5	2018	1922-16	~	~	4.2 (1.3)	6.9	2.6	4.3	1.6	7.5	n.d.	12.3
SU-6	2018	1922-16	✓	~	4.3 (1.3)	6.9	2.6	4.3	1.7	7.6	n.d.	13.2
ТОР	2014	565-02			3.2 (1)	5.5 (3.4)	2.2 (0.1)	3.2 (3.4)	0.6 (1)	8	3.8 (5.6)	-
BASE	2018	1922-16			3.9 (0.6)	5.8 (1.9)	2.7 (0.2)	3.1 (2.1)	1.3 (0.6)	7.6 (0.2)	n.d.	10 (4.8)

174 Table 1. Sampling characteristics and physico-chemical conditions in the two Bathymodiolus puteoserpentis mussel assemblages studied in 2014 175 176 177 178 179 and 2018 on the Moose site from the Snake Pit vent field (Mid-Atlantic Ridge). PBT: small boxes used for faunal sampling. SS: suction samplers used for faunal sampling. The green check mark and the red cross indicate the use or not of the corresponding sampling method. Temperatures (T°C) were obtained by submersible probes (mean + standard deviations). Ranges were calculated by the difference between the maximum and minimum temperatures. Temperature anomalies were obtained by subtracting the established background temperature of 2.6°C at this field (Hernàndez-Avila et al., 2022) from the measured temperatures. Other parameters (pH, H₂S: concentrations of hydrogen sulfide, Fe: concentrations 180 of iron) were obtained by CHEMINI and water sampling tools PEPITO (2014) and PIF/PLUME (2018). The three H₂S concentrations measured in 181 2018 were below the detection threshold n.d.). TOP corresponds to the average of the values measured in 2014 at the top of the edifice (SU-1, SU-182 2, SU-3). BASE corresponds to the average of the values measured in 2018 at the edifice's base (SU-4, SU-5, SU-6). Standard deviations are given 183 in parenthesis.

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Mussel assemblage	T°C mean (std)	T°C max	T°C min	T°C range	
SU-1	6.6 (2.3)	15.8	2.6	13.2	
SU-2	3.8 (0.8)	7.5	2.4	5.0	
SU-3	2.9 (0.2)	3.7	2.5	1.2	

Table 2. Time-series of temperature (T°C) means, standard deviations (std), maximum (max), minimum (min) and range over the *Bathymodiolus puteoserpentis* mussel assemblage located at the top of the Moose edifice (2014) from the Snake Pit vent field (Mid-Atlantic Ridge). These temperature time-series were recorded by MISO probes from 18/01/2014 to 22/01/2014.

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192 2.3. Faunal analyses

On board, mussel individuals were isolated, briefly opened and preserved in 5-litre containers filled with 4% formalin. The rest of the faunal sample (sampling box + suction sampler when performed) was sieved on 300 µm and 20 µm meshes. Large individuals of the 300 µm fraction were picked and 109 individuals (47 in 2014, 62 in 2018), representing key macrofaunal species, were conditioned at -80°C for isotopic analyses and barcoding. The remaining fauna from the 300 µm-fraction was preserved in 96% ethanol and that of the 20 µm-fraction (meiofauna) in 4% formalin for subsequent identification in the lab.

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200 2.3.1 Mussel biomass & volume

201 Mussel biomass and volumes were evaluated for each sampling unit to provide a standardisation mean and 202 allow comparison between samples and other studies. After rinsing with water and draining, each mussel 203 individual was measured (length, width, thickness) using an electronic caliper. The wet weight (WW) was 204 evaluated by weighing whole organisms in aluminum cups and also, the shells and flesh separately. Then, 205 the cups were placed in a 60°C oven for 48 hours in order to eliminate water and obtain the dry weight 206 (DW) that contains both the organic and inorganic materials. Finally, dry tissues were placed in a 550°C 207 oven for seven hours to remove the organic part and obtain the ash weight (AW). The ash-free dry weight 208 (DW - AW = AFDW) corresponds to the quantity of organic matter in the sample. However, two very 209 small individuals were too difficult to open and were not considered in the calculation of biomass. For these

210 individuals, only morphometric measurements were performed. Missing morphometric or biomass values 211 (NA) in the global dataset -due to small size or sometimes broken shells-, were replaced (~8% of the data), 212 thanks to the missMDA R package. An iterative Principal Component Analysis (PCA) was used to replace 213 NAs with coherent and non-random data. To standardise our sampling units, two metrics were used: (1) the 214 estimated area provided by the mussels (m^2) for each SU, calculated by summing the area provided by each 215 individual mussel shell, considering these as being flat ellipsoids (area of two elliptical surfaces 216 $=2(a \times b \times \pi)$ with a = length/2 and b = width/2, (2) the estimated volume occupied by the mussels (dm³) for each SU, calculated by summing the volume occupied by each mussel (ellipsoid volume 217 $=\frac{4}{3}\pi \times a \times b \times c$, with a = length/2; b = width/2; c = thickness/2 (Table 3). Moreover, 218 condition index (CI) of each mussel was estimated with the ratio CI = W/V with W = AFDW(g) =219 220 organic matter and V = shell volume (mL), with the aim of comparing the general physiological 221 condition of individuals between samples and location on the Moose edifice (Table 3). CI is a well-accepted 222 method of determining the physiological condition of bivalves (Fisher et al., 1988; Lenihan, 1999; Lenihan 223 et al., 2008; Smith, 1985).

224 2.3.2 Identification of the associated fauna

225 Once in the lab, all macrofaunal individuals (>300 μ m) were picked individually from each of the sampling 226 units and identified to the lowest taxonomic level possible using a binocular magnifier. To facilitate sorting, 227 all meiofaunal individuals (from the 20 μ m and >300 μ m fractions) were transferred separately to a small 228 volume of 96% ethanol, stained with Phloxine B, which colours the organic matter in pink, and rinsed. For 229 them, taxonomic identification was carried out at higher taxonomic levels (order/class). All fauna (both 230 sieves) belonging to Nematoda, Copepoda and Halacaridae were considered as meiofauna. Juveniles of 231 macrofauna found in the lower 20 μ m fraction were considered as macrofauna.

232 2.3.3 Isotopic analyses

Stable isotope ratios of carbon (${}^{13}C/{}^{12}C$ or $\delta^{13}C$) and nitrogen (${}^{15}N/{}^{14}N$ or $\delta^{15}N$) were used to describe food-233 234 web structure and characterise trophic interactions of mussels and their associated fauna. It is accepted that in natural environments, δ^{13} C can be used to identify the food sources and remains relatively unchanged 235 236 between successive trophic levels, ~1% between each trophic level (DeNiro and Epstein, 1978). The δ^{13} C 237 of local primary producers depend on the isotopic values of their carbon sources and their isotopic 238 fractionation during carbon fixation (Portail et al., 2018). This includes autotrophs using the Calvin-239 Benson-Bassham (CBB) cycle or the reducing tricarboxylic acid (rTCA) cycle as well as methanotrophs. 240 $\delta^{15}N$ is a good indicator of trophic position: successive trophic levels become enriched in $\delta^{15}N$ for 241 approximately 3.4‰ per trophic level (Minagawa and Wada, 1984) or more generally between 2 and 4‰

242 (McCutchan et al., 2003). Both isotopic ratios are expressed according to the δ notation in ‰ (Coplen,
243 2011), in reference to the international references Atmospheric Air for nitrogen and Vienna Pee Dee
244 Belemnite for carbon.

245 Isotopic ratios of carbon and nitrogen were determined on 47 individuals belonging to 10 species in 2014 246 and 62 individuals belonging to 4 species in 2018. The treatment varied according to the taxa. Organisms 247 were dissected to keep their muscles, or, when body size was small, the whole organism was used (Mateo 248 et al., 2008). Some mussels hosted *Branchipolynoe seepensis* polynoids in their mantle cavity: those were 249 isolated and measured. Samples were placed in the oven at 50°C during 48h before being ground in a 250 homogeneous powder using mortar and pestle. Samples containing hard inorganic carbon parts that could 251 not be physically removed (copepods) were conditioned in silver capsules and were decalcified with 252 hydrochloric acid to remove carbonates from their exoskeleton. When possible, isotopic analyses were done 253 on several individuals per taxon in order to take into account intraspecific variability. Stable isotope ratio 254 measurements were performed via continuous flow- elemental analysis - isotope ratio mass spectrometry 255 (CF-EA-IRMS) at University of Liège (Belgium), using a vario MICRO cube C-N-S elemental analyzer 256 (Elementar Analysensysteme GMBH, Hanau, Germany) coupled to an IsoPrime100 isotope ratio mass 257 spectrometer (Isoprime, Cheadle, United Kingdom). IAEA (International Atomic Energy Agency, Vienna, 258 Austria) certified reference materials sucrose (IAEA-C-6; $\delta^{13}C = -10.8 \pm 0.5\%$; mean \pm sd) and ammonium sulfate (IAEA-N-2; $\delta^{15}N = 20.3 \pm 0.2\%$; mean \pm sd) were used as primary analytical standards. Sulfanilic 259 acid (Sigma-Aldrich: $\delta^{13}C = -25.6 \pm 0.4\%$; $\delta^{15}N = -0.13 \pm 0.4\%$; mean \pm sd) was used as a secondary 260 261 analytical standard. Standard deviations on multi-batch replicate measurements of secondary and internal 262 lab standards (amphipod crustacean muscle) analysed interspersed with samples (one replicate of each standard every 15 analyses) were 0.2‰ for both δ^{13} C and δ^{15} N. 263

264 2.3.4 Statistical analyses

All analyses were computed with R (v. 4.2.1, R Core Team, 2022).

Mussel population- The number of mussels analysed for their sizes varied between 18 to 52 and those analysed for their biomass varied between 18 to 51 (Table 3). Mussel biomass and abundance data were pooled by sampling year and compared with the Wilcoxon-Mann-Whitney (WMW) non-parametric test, on three main parameters: shell length, dry weight (DW) and ash-free dry weight (AFDW). Then, a multicomparison of the six sampling units was done with a Kruskal-Wallis test.

Alpha diversity- As they represent the structuring species, *Bathymodiolus puteoserpentis* mussels were not
 taken into account in the statistical analyses of the associated fauna. Standardised to surface mussel area or
 estimated volume, associated fauna abundance datasets were subjected to several statistical tests in order to

evaluate their similarity with the WMW and Kruskal-Wallis (KW) tests, completed when necessary by a Post Hoc Dunn's test. Alpha diversity of each sample was assessed through two indices: the Shannon-Wiener's diversity index (H') which takes into account species richness and the relative abundance of species with a sensitivity to rare species. H' can vary between 0 (absence of diversity) and to a H max that is dataset dependent (high diversity). The Piélou evenness index (J') characterises the homogeneity of a community. It varies between 0 in a community with only one species, and 1 in a community where all species have equal relative abundance.

Beta diversity- Then, in order to compare the two mussel assemblages (2014 at the top of the edifice, 2018 at the base) and especially their associated fauna, the β -diversity was assessed with a PCA with Hellinger transformation on the associated fauna dataset. The comparison analyses of the associated fauna were carried out both by including all the individuals collected (sampling boxes and suction sampling), but also by excluding the individuals recovered by suction sampling for a homogenised comparison of SUs. The relative abundance trends observed being the same, the results presented below include all individuals collected, all methods combined.

288 **3. RESULTS**

289 **3.1.** Physico-chemical conditions

3.1.1. Point temperature measurements- Higher average temperatures were recorded at SU-1, SU-5 and
SU-6 (Table 1). SU-1 also exhibited the highest maximal temperatures, as well as the higher variability
(7.1°C). SU-2, SU-3 and SU-4 exhibited the lowest and more stable temperatures.

3.1.2. Temperature dynamics- In 2014, the MISO probes continuously measured the temperatures on the three sampling units (See Supplementary materials, Figure S1). Mean temperatures were 6.6 ± 2.3 °C on SU-1, 3.8 ± 0.8 °C on SU-2 and 2.9 ± 0.2 °C on SU-3 (Table 2). At the start of the deployment, the temperature values varied between 3-4 °C while at the end of the deployment, they did not exceed 3 °C in all SUs. A peak in temperature reaching 15.8 °C was recorded on SU-1 after 3 days of measurements, dropping rapidly to ~7°C a few minutes later. The other two SUs showed lower temperature variations.

3.1.3. Physico-chemical measurements- The pH values were slightly lower for the SUs sampled in 2018 than for SU-1 (Table 1). pH values at sites SU-2 and SU-3 were missing. Mean H₂S concentrations reached 10.2 μ m in SU-1, which was higher than in all other samples. In 2018, values of Fe_{Total} were similar in SU-5 and SU-6 and three times higher than those recorded in SU-4 (Table 1). Fe²⁺ values (*in situ*) showed an average of 32.2 ± 7.5 μ M on SU-3 (not presented in the table). Similarly, methane concentrations reached 0.33 μ M for SU-1 (0.33 μ M, not presented in the table). Due to these discrepancies in environmental data

acquisition, only temperature variables were used to compare the SUs, the other factors being presented fordiscussion purposes.

307 **3.2. Habitat description-** The habitat of the 2014 samples, located at the top of the Moose edifice, was 308 characterised by the presence of an important aggregate of *Rimicaris* shrimps with few *Pachycara* 309 thermophilum zoarcid fishes, Segonzacia mesatlantica crabs and Phymorhynchus gastropods. At SU-1 and 310 SU-2, the mussels densely covered the edifice surface, forming a thick layer of stacked mussels, thus impairing proper evaluation of the surfaces sampled through imagery. At SU-3 however, the assemblage 311 312 was less thick and the greyish rocky substratum was apparent. The habitat of the 2018 samples, at the base 313 of the edifice, was characterised by dense mussel assemblages, colonising orange and greyish substrata. A 314 few individuals of P. thermophilum and Phymorhynchus gastropods were observed. None to rare 315 filamentous microbial mats were observed near the sampling areas during the two cruises. Contrary to what 316 was expected, mean temperatures were quite similar in the mussel assemblages from the top and base of 317 the edifice despite small scale heterogeneity. However, the closer proximity to fluid emissions of sampling 318 units at the top of the edifice is reflected in the higher hydrogen sulfide concentration measured at SU-1.

319 3.3. Bathymodiolus puteoserpentis: population size structure and biomass analyses- Overall, mussel 320 lengths varied between 5.5 mm and 125.2 mm, with an average length of 84.06 ± 24.01 mm. When data 321 were pooled by sampling year, Wilcoxon-Mann-Whitney (WMW) non-parametric test showed that the 322 median shell lengths did not vary significantly between 2014 (82.81 \pm 27.15 mm) and 2018 (85.41 \pm 20 323 mm, p-value of 0.92). Nevertheless, Kruskal-Wallis test showed that the mean shell lengths varied 324 significantly between SUs (p-value of 9.27e-09), and post hoc Dunn's test indicated that those for SU-1 325 were significantly higher from those of SU-2 and SU-6 (Table 3). Mean shell lengths of SU-2 were 326 significantly smaller than all other SUs, except SU-6 (Table 3). The mean lengths of the four remaining 327 SUs did not differ significantly from each other. The most represented shell sizes were found in the 50-115

328 mm interval. All mussels were larger than 15 mm, except for SU-2 and SU-6, which both included one329 specimen below 15 mm (Figure 3).



FIGURE 3. Size-frequency structure of *Bathymodiolus puteoserpentis* mussels sampled (A) in 2014 and (B) in 2018 during the BICOSE cruises and (C) both years by sampling units on the Moose site at the Snake Pit vent field (Mid-Atlantic Ridge). "N" corresponds to the number of mussels in good condition that could be measured by cruise/by sampling unit. Arrows indicate the average lengths of mussel shells and with standard deviations in parentheses. The size unit in mm was chosen to facilitate comparison with the size structure of *B. puteoserpentis* established by Won *et al.*, 2003a on the Logatchev vent field.

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										Wit	h meiofaunal	taxa	Without meiofaunal taxa		
Sampling units	N mussel (morpho/biomass)	N mussel (total)	Mean WW (g)	Mean DW (g)	Mean AFDW (g)	Mean shell length (mm)	Total area (m²)	Total volume (L)	CI (g/mL)	N taxa	H' index	J' index	N taxa	H' index	J' index
SU-1	43	43	20.1 (±5.7)	4.0 (±1.3)	3,6 (±1.2)	98.4 (±13.8)	0.31	4.00	0.039 (+0.009)	9	1.35	0.62	7	0.93	0.48
SU-2	52*	60	11.5 (±7.5)	2.0 (±1.2)	1,7 (±1.2)	68.8 (±27.5)	0.26	2.80	0.038 (+0.009)	12	1.26	0.51	9	1.49	0.68
SU-3	27	33	17.2 (±9.8)	2.4 (±1.4)	2,3 (±1.3)	84.9 (±29.1)	0.20	2.34	0.033 (+0.006)	12	0.98	0.40	9	1.34	0.61
SU-4	18	23	14.8 (±3.8)	2.2 (±0.7)	2,0 (±0.7)	96.0 (±10.7)	0.16	1.90	0.025 (0.006)	13	1.22	0.48	10	0.58	0.25
SU-5	43	49	14.5 (±6.9)	2.6 (±1.3)	2,4 (±1.2)	88.6 (±17.6)	0.30	3.65	0.032 (+0.008)	14	1.02	0.39	11	0.76	0.32
SU-6	52*	58	13.2 (±8.3)	2.5 (±1.7)	2,3 (±1.6)	79.08 (±22.3)	0.30	3.39	0.039 (+0.012)	12	0.75	0.30	9	1.56	0.71
ТОР	122*	136	15.8 (±8.4)	2.8 (±1.5)	2,5 (±1.5)	82.8 (±27.2)	0.78	9.30	0.037 (+0.009)	16	1.37	0.50	13	1.54	0.60
BASE	113*	130	13.9 (±7.2)	2.5 (±1.4)	2,3 (±1.3)	85.4 (±20.0)	0.75	8.89	0.034 (+0.011)	16	0.93	0.33	13	0.93	0.36

Table 3. Morphometric and biomass characteristics of *Bathymodiolus puteoserpentis* mussels as well as diversity of associated fauna sampled in 2014 and 2018 on the Moose Site from the Snake Pit vent field (Mid-Atlantic Ridge). N morpho/biomass: number of mussels measured and analysed for biomass. N total: total number of mussels sampled, including those used in isotopic analyses. Mean WW: mean wet weight. Mean DW: mean dry weight. Mean AFDW: mean ash-free dry weight = total organic matter. Shell length (mm) = corresponds to the average shell length of mussels. Estimated areas in m² were calculated by summing individual areas of each mussel and corresponds to the colonisation surface offered by the mussels (see M&M section). Estimated volumes in L were calculated by summing individual volumes of each mussel (see M&M section). CI: average condition index of the mussels in the SU. N taxa: number of associated fauna taxa (not including *B. puteoserpentis*). H' corresponds to the Shannon's diversity index, J' corresponds to the Piélou's evenness index, both indexes were calculated with and without meiofauna.*: one small individual not included in biomass measurements. Higher values are highlighted in bold.

Overall, mussel's mean individual wet, dry and ash free dry weights varied between sampling units (Table 3). The quantity of organic matter (AFDW, Table 3) varied from 1.7 ± 1.2 g and 3.6 ± 1.2 g. A Kruskal-Wallis test showed that the AFDW varied significantly between SUs (p-value of 9.84 e-09), and post hoc Dunn's tests indicated that it was significantly higher on SU-1 than all others (Table 3). The condition index was also significantly higher on SU-1 than those of SU-3, SU-4 and SU-5 (Table 3).

349 3.4. Composition and abundance of the associated fauna

A total of 23,207 individuals were sorted: 1885 in the macrofaunal and 21,322 in the meiofaunal (copepods,
 nematodes, Halacaridae, 20 μm and 300 μm) compartments. Overall, meiofaunal taxa dominated the fauna

352 associated with *Bathymodiolus puteoserpentis* mussel assemblages in terms of relative abundance in all SUs

353 except SU-1 where they represented less than 44% of all sorted individuals (Figure 4, Table 4). Nematodes

dominated the meiofauna of all SUs except SU-1 and to a lesser extent SU-5, both dominated by copepods.

355 Macrofaunal taxa were dominated by the limpet *Pseudorimula midatlantica* in sampling units from the top

356 of the edifice in 2014 (from 42.4% to 73.8% of the macrofaunal abundance), while it was almost absent in

357 the sampling units from the base in 2018 (<1.5% of the macrofaunal abundance; Figure 4).



FIGURE 4. Stacked bar graphs showing the composition of associated fauna in *Bathymodiolus puteoserpentis* mussel assemblages sampled on the Moose site at the Snake Pit vent field during the BICOSE cruises (SU-1, SU-2, SU-3 in red for 2014, SU-4, SU-5, SU-6 in blue for 2018). *The left bar graph includes Copepoda, Halacaridae, Nematoda while the right one excludes them for better reading.

- 359 Conversely, the ophiuroid Ophioctenella acies dominated the edifice's base with abundances varying from
- 360 35.8% to 88.5% of the total macrofaunal abundance while only a few individuals were found at the top.
- 361 Densities of associated fauna (macro- and meiofauna) were higher in samples from the base than those from
- the top of the edifice (Table 4).

Class	Family Taxa		SU-1	SU-2	SU-3	SU-4	SU-5	SU-6
Class			Relat. Ab %					
Annolido								
Polychaeta	Ampharetidae		0.00	0.00	0.00	0.08	0.03	0.07
rorychuctu	Amphinomidae	Archinome sp.	0.38	2.39	0.26	0.31	0.09	0.03
	Capitellidae	· · · · · · · · · · · · · · · · · · ·	0.00	7.25	0.26	0.16	0.03	0.00
	Chaetopteridae		0.00	0.06	0.00	0.00	0.00	0.00
	Hesionidae		0.38	1.94	0.00	0.39	0.41	0.24
	Polynoidae		4.89	0.00	0.00	0.00	0.00	0.00
	Polynoidae	Branchinotogluma sp.	0.00	0.00	0.00	0.08	0.18	0.19
	Polynoidae	Branchipolynoe	5.26	1.04	1.05	1.01	1.05	0.52
	Spionidae	seepensis	0.00	0.00	0.00	0.55	0.01	0.01
Arthropoda								
Arachnida	*Halacaridae		0.00	0.06	0.79	0.08	0.03	0.03
*Copepoda			39.47	12.75	8.64	30.05	46.92	33.76
Crustacea	Alvinocarididae	Alvinocaris sp.	0.00	0.00	0.52	0.00	0.00	0.00
		Kimicaris chacei	3.38	0.78	2.30	0.00	0.01	0.00
Febinodormoto								
Stelleroidea	Ophiuridae	Onhioctenella acies	0.00	0.19	1 31	22.72	7.91	0.61
~~~~~		• <i>P</i> ·····						
Mollusca			0.00	0.00	0.00	0.00	0.01	0.00
Gastropoda	Lepetodrindae	Lepetodrilus atlanticus	0.00	0.00	0.00	0.00	0.01	0.00
		P Seudorimula midatlantica	41.55	10.10	9.95	0.25	0.09	0.05
	maannea							
	Neolepetopsidae Paralepetopsis		0.00	0.00	0.00	0.00	0.00	0.01
		ferrugivora						
	Peltospiridae	Peltospira smaragdina	0.00	0.13	0.00	0.00	0.00	0.00
	Raphitomidae	Phymorhynchus sp.	0.00	0.00	0.26	0.16	0.00	0.00
	Skeneidae	Protoura morvatasoni	0.38	0.00	0.26	0.00	0.00	0.00
*Nematoda			4.51	63.24	74.35	44.18	43.24	64.51
Total number of in	dividuals (*considered	as meiofauna)	266	1545	382	1281	7886	11847
Total relative abun	dance meiofauna (%)		43.98	76.05	83.78	74.31	90.19	98.3
Total relative abun	dance macrofauna (%)		56.02	23.94	16.23	25.69	9.82	1.71

Table 4. Relative abundances of associated fauna sampled in mussel assemblages on the Moose Site of the Snake Pit vent field (Mid-Atlantic Ridge) in 2014 and 2018. *Copepoda, Halacaridae and Nematoda are considered as meiofauna (coming from the 20 μm and 300 μm fractions, see M&M). Higher values are highlighted in bold.

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367 Overall, 21 taxa (Class/family/species) were identified across all samples, including 18 for macrofauna (family/species). Total taxonomic richness varied from 9 in SU-1 to 14 in SU-5 (Table 3). SU-1 exhibited 368 369 a co-dominance of *Pseudorimula midatlantica* (41.3%) and copepods (39.5%), as well as other taxa in 370 relatively significant abundance (undetermined polynoids, *B. seepensis*, *Rimicaris chacei*, and nematodes) 371 which contributed to the highest diversity indexes found in this sample despite its lowest number of taxa 372 (Table 3). SU-4 and SU-5 harboured a co-dominance of copepods and nematodes, with highest abundance 373 of O. acies in the macrofaunal compartment. These two SUs had no other well-represented taxa in terms of 374 abundance, as illustrated by their overall lower diversity index values than SU-1. Although it exhibited more 375 taxa, many of which contributed in very low proportions to the total abundance. The other SUs were rather

characterised by a strong dominance of nematodes. SU-2 and SU-3 were similar in terms of abundance. SU6 harboured lower diversity and evenness indexes linked to the very low representation of taxa other than
nematodes and copepods (Table 3). Nevertheless, the diversity is generally low in each of the sampling
units, whether we are interested in species richness or diversity indexes.

380 A total of seven polychaete families were found (in addition to a few undetermined ones of poor condition). 381 They were particularly abundant in samples from the base of the edifice representing from 10% to 63% of 382 the macrofaunal abundance in SU-4 and SU-6 respectively. They include Ampharetidae, Branchinotogluma 383 sp. and Spionidae. *Branchipolynoe seepensis* polychaetes were more abundant in samples from the base, 384 representing from 4% to 31% of the macrofaunal abundance in SU-4 and SU-6 respectively. A high 385 abundance of Capitellidae (30.3%) and to a lesser extent Archinome sp. (10%) polychaetes were found in 386 SU-2. Alvinocaridid shrimps (*Rimicaris chacei*, Alvinocaris markensis) were found almost exclusively in 387 samples from the top of the edifice. Similarly, the small gastropods *Peltospira smaragdina* and *Protolira* 388 thorvaldssoni were only recorded from the top of the edifice although in low numbers. Two other gastropod 389 species (Lepetodrilus atlanticus and Paralepetopsis ferrugivora) were found in low abundance, but at the 390 base. Finally, *Phymorhynchus* sp. was present in two out of six samples.

391 According to the PCA on macrofaunal taxa, the first component explained 37.7% of the distribution of the 392 fauna. SU-1, SU-2 and SU-3 are positively loaded on the first component, while SU-4, SU-5 and SU-6 are 393 negatively loaded on the first component, showing a clear distinction between the samples from the top to 394 those from the base of the edifice, supporting the observed differences in community structure (Figure 5). 395 The PCA showed also similarities between SU-1 and SU-3, probably due to their low faunal abundances 396 (with the exception of several polychaetes and shrimps) and diversities and between SU-5 and SU-6 which 397 exhibited similar macrofaunal dominance (O. acies, Branchipolynoe seepensis). The second component 398 explains 28.6% of faunal distribution.



FIGURE 5. Principal component analysis (PCA) of macrofaunal individuals associated to *Bathymodiolus puteoserpentis* mussel assemblages from the Moose site from the Snake Pit vent field (Mid-Atlantic Ridge). Meiofauna (including nematodes, copepods and halacarids) have been removed from the analysis for better graphical reading. The component 1 explains 37.7% in faunal distribution showing a clear separation between the assemblages sampled in 2014 (right side) and those sampled in 2018 (left side). Axis 2 explains 28.6% in faunal distribution. ALV: *Alvinocaris chacei*; AMP: Ampharetidae; ARC: *Archinome sp.*; B.SEE: *Branchipolymoe seepensis*; BRA: *Branchinotoghuma sp.*; CAP: Capitellidae; CHA: Chaetopteridae; HES: Hesionidae; LEP: Lepetodrilus atlanticus; OPH: Ophioctenella acies; PAR: Paralepetopsis ferrugivora; PHY: Phymorhynchus sp.; PEL: Peltospira smaragdina; POL: Polynoids; PRO: Protolira thorvaldssoni; PSE: Pseudorinula nidatlantica; RIM: Rimicaris chacei; SPI: Spionidae.

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## 400 **3.5.** Isotopic analyses

Bathymodiolus puteoserpentis mussels were distributed in two groups (Figure 6). The first one, which 401 included most of the mussels (90%), showed  $\delta^{13}$ C values ranging from -35.64‰ to -31.09‰ and  $\delta^{15}$ N values 402 403 ranging from -15.42‰ to -4.83‰. The second group included only three mussels from the base of the edifice 404 with much higher values ranging between -16.35% to -12.73% for  $\delta^{13}$ C and +4.53% to +7.10% for  $\delta^{15}$ N. 405 Branchipolynoe seepensis polychaetes exhibited isotopic values very close to those of B. puteoserpentis, ranging between -34.9‰ to -11.1‰ for  $\delta^{13}$ C and from -15.4‰ to +8.9% for  $\delta^{15}$ N (Figure 6). However, a 406 407 few of them differed and appeared to be independent of the mussels. Ophioctenella acies isotopic ratios 408 were much less spread than those of *B. seepensis* and *B. puteoserpentis*, and were grouped between -26.5% and -22.6% for  $\delta^{13}$ C (with an outlier of -12.5%) and between -4.6% and +1.3% for  $\delta^{15}$ N (with an extreme 409 410 value of 6.4‰) (Figure 6). Only three ophiuroids sampled in 2014 were analysed, against 24 for 2018, giving similar results for  $\delta^{13}$ C values but higher  $\delta^{15}$ N values. 411



FIGURE 6. Identification of the main groups of organisms sampled on the Moose site from the Snake Pit vent field during the two BICOSE cruises (BICOSE 2014 represented by empty circles, BICOSE 2018 represented by full circles). Groups were identified according to carbon isotopic values (X axis) and nitrogen isotopic values (Y axis). The dots represent the distribution of all the individuals analysed: *Alvinocaris* sp. (n=2); *Archinome* sp. (n=5), *Branchipolynoe seepensis* (n=4 in 2014, n=12 in 2018); *Bathymodiolus puteoserpentis* (n=14 in 2014, n=17 in 2018); *Branchinotogluma* sp. (n=9); Capitellidae (n=5); *Ophioctenella acies* (n=3 in 2014, n=24 in 2018); *Peltospira smaragdina* (n=2); PHY: Phymorhynchus sp. (n=1); Pseudorimula midatlantica (n=5); Rimicaris chacei (n=6).

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Hydrothermal shrimps showed high  $\delta^{13}$ C values respectively ranging from -23.9‰ to -22.2‰ for 413 414 Alvinocaris sp., and from -23.8 to -17.9% for *Rimicaris chacei*, and  $\delta^{15}$ N values respectively ranging from -1.8 to 2.1‰ and from -0.3 to 5.2‰, respectively. Archinome sp. and Branchinotogluma sp. polychaetes 415 416 showed variable isotopic results: while the former had high values ranging from -19.2% to -17.3% for  $\delta^{13}$ C and from +3.6% to +6.0% for  $\delta^{15}$ N, isotopic ratios of the latter were spread across the entire dataset (from 417 -35.2% to -11.4% for  $\delta^{13}$ C and from -15.6% to +8.5% for  $\delta^{15}$ N) (Figure 6). Capitellidae polychaetes formed 418 419 a distinct intermediate group, near Ophioctenella acies and Alvinocaris sp., with isotopic values ranging 420 from -22.9% to -21.5% ( $\delta^{13}$ C) and from +0.4% to +1.9% ( $\delta^{15}$ N). Finally, gastropods (including 421 Pseudorimula midatlantica and Peltospira smaragdina, only sampled in 2014) showed intermediate values from -29.5% to -23.6% for  $\delta^{13}$ C and from +0.9% to +1.9% for  $\delta^{15}$ N. The only individual of *Phymorhynchus* 422 sp. gastropod sampled in 2014 had low  $\delta^{13}$ C value of -33.5% and low  $\delta^{15}$ N value of -10.8% (Figure 6) close 423 424 to those of symbiotic species despite its carnivorous nature.

# 425 4. Discussion

426 This study provides a detailed description and characterisation of two Bathymodiolus puteoserpentis

427 assemblages colonising the Moose site at the Snake Pit vent field (MAR). Mussel assemblages were suction

428 sampled in triplicate in 2014 at the top of the edifice and in 2018 at its base. The two cruises were carried

429 out with various sampling methods, measuring tools and gear, leading to differences in the physico-chemical

parameters measured. This different sampling effort is due to technical issues and unforeseen events, not allowing an optimal inter-annual comparison. Moreover, the small number of sampling units (n=6) over the two years, makes it difficult to obtain statistically significant results although a clear separation in faunal composition of the two sampling sites (top versus base) was observed. The use of different sampling protocols (sampling in boxes coupled or not with an associated suction sampler) did not seem to influence the results obtained in terms of relative abundance of each taxa, but probably led to an underestimation of faunal abundance and diversity of rarer species on the two sampling units concerned.

## 437 4.1. Habitat characterisation

438 With the exception of SU-1, the temperatures measured above the fauna on the sampling units were 439 relatively uniform, suggesting relatively stable and homogeneous environmental conditions. However, SU-440 1 exhibited higher temperatures and fluctuations, as shown by a recorded maximum temperature of 15.8°C 441 in the 2014 time series. Located on the top of the edifice, SU-1 also exhibited higher hydrogen sulfide 442 values, suggesting a higher influence of hydrothermal fluids. The average temperatures of the other two SUs 443 at the top were similar to those reported for bottom temperature at this field (i.e. 2.6°C, see Hernández-Ávila 444 et al., 2022). At the edifice base, SU-5 and SU-6 had slightly higher temperatures, as well as lower pH and 445 higher total iron concentrations than SU-4. Their temperatures were closer to those measured at SU-1. These 446 patterns are consistent with the small-scale spatial heterogeneity of physico-chemical conditions 447 characterising the vent environment (Chevaldonné et al., 1991; Johnson et al., 1988; Le Bris et al., 2006; 448 Sarrazin et al., 1999). While they can be used to qualitatively compare the habitats of different assemblages, 449 single-point temperature measurements are not fully representative of the temperatures experienced by the 450 organisms over time. In fact, the deployment of the autonomous temperature probes over time in 2014 451 showed higher values than those obtained with the submersible sensors for all temperature variables (mean, 452 standard deviation, maximum, minimum) as also reported in a previous study (Marticorena et al., 2021).

453 In comparison, Bathymodiolus puteoserpentis mussels were found at lower temperatures at the Logatchev 454 vent field. Thus, on the Irina II mound, the average temperatures recorded varied between  $2.7 \pm 0.2$  °C on a 455 sulfide pillar overgrown by mussels and  $3.5 \pm 0.7^{\circ}$ C on the main assemblage without exceeding  $7.4^{\circ}$ C 456 (Zielinski et al., 2011). These results corresponded to those obtained for SU-2, SU-3 and SU-4 but not on 457 the other SUs that had higher mean values (up to 9.3°C on SU-1). B. puteoserpentis sampled at SP (this 458 study) and Logatchev (Zielinski et al., 2011) were found in habitats colder than those of Bathymodiolus 459 azoricus, where the mean temperatures vary between 4.8°C and 10°C (Cuvelier et al., 2011; Desbruyères et 460 al., 2001; Husson et al., 2017).

## 461 **4.2. Population structure of** *Bathymodiolus puteoserpentis*

462 In the assemblages studied here, Bathymodiolus puteoserpentis exhibit large average sizes, with almost all 463 specimens exceeding 5 cm in length. For Bathymodiolus azoricus, Cuvelier et al. (2011) identified different 464 types of assemblages, characterised by specific mussel size ranges, with large ( $\approx 6$  cm), medium (2–6 cm) 465 and small (<2 cm) mussels distributed along a gradient of increasing dilution of the vent fluid (Cuvelier et 466 al., 2009). Since B. puteoserpentis and B. azoricus exhibit similar body size with maximal shell length of 467 12 cm in both species (Van Cosel et al. 1999), the same categories could apply to *B. puteoserpentis*. The 468 mussels observed at Snake Pit would therefore correspond to those found in "Assemblage 1" observed by 469 Cuvelier et al. (2011), characterised by larger mussels.

Our results showed that mean mussel shell lengths did not differ significantly between 2014 and 2018. As 470 471 observed in other MAR mussel assemblages (Cuvelier et al., 2017; Sarradin et al., 1999; Van Audenhaege 472 et al., 2022), the largest mussels were more abundant in the most variable and warmer environment (SU-1) 473 and the smallest in the coldest and least fluctuating environment (SU-2). However, the number of individuals 474 sampled in each SU was too low to statistically support size spatial segregation of mussels, decreasing with 475 increasing distance from hydrothermal input as observed in other studies on B. azoricus (Husson et al., 2017; 476 Van Audenhaege et al., 2022). Moreover, our results report the presence of large mussels in colder habitats 477 (SU-3 and SU-4), questioning the presence of a spatial segregation in relation to environmental conditions 478 for *B. puteoserpentis*.

479 Our video (see Supplementary materials) and *in situ* observations (authors pers. obs.), coupled with our 480 sampling results, denote a general lack of recruitment at the scale of the whole Moose edifice as shown by 481 the predominance of large adult mussels and the near absence of small individuals. The absence of small B. 482 *puteoserpentis* < 5 mm, considered as newly settled juveniles by Van Dover (2002), is consistent with 483 observations by Turnipseed et al. (2004) that reported only 3% of new settlers in July 2001 at the same vent 484 field. This consistent lack of recruitment over the years contrasts with regular observations of new settlers 485 in different *Bathymodiolus* assemblages along the MAR: the presence of small recruits was indeed reported 486 on B. azoricus assemblages at the LS vent field (Cuvelier et al. 2009) as well as on B. puteoserpentis mussel 487 assemblages on the Semeynov vent field (Franke et al., 2021) and recently close to *Bathymodiolus* sp. on 488 Puy des Folles (MAR, ROV SuBastian's 500th Dive at Puy Des Folles Volcano, 2 April 2023, ~10:15 on 489 the video, https://www.youtube.com/live/C152v5tKl20?si=d-GZAV17MmzxjIPb).

490 Several abiotic and biotic factors could limit juvenile settlement at vents, in habitats under strong 491 hydrothermal influence and highly fluctuating conditions. Post-larval stages of mussels have no symbionts 492 when they settle at vents and only acquire them progressively afterwards (Wentrup et al., 2013). As 493 symbionts potentially ensure detoxification allowing vent holobionts to better cope with extreme conditions

494 (Durand et al., 2009, Sun et al. 2022), mussel post-larvae would not be able to handle those prevailing in
495 adult habitat (Husson et al., 2017). However, if larval supply exists, they should be able to settle in peripheral
496 areas with lower hydrothermal influence, which was not observed during our various dives in the area
497 (authors pers. obs.)

498 Biotic factors could also play a role in settlement. Several studies showed a negative correlation between 499 adults and juveniles in dense coastal mussel assemblages (Khaitov, 2013; Lehane & Davenport, 2004, 500 Okamura, 1986), larval settlement being inhibited by adults. Khaitov and Lentsman (2016) showed that 501 *Mytilus* mussel assemblages in the White Sea follow a cyclical demographic structure, comparable to the 502 "endogenous" model proposed by Lukanin et al. (1986) and which is based on density-dependant 503 regulatory processes. In this model, adult mussels have negative impacts on juvenile conspecifics, both 504 directly via the filtration of larvae, and indirectly via the rejection of pellets "polluting" the environment. 505 This leads to endogenous population instability where only the reduction or disappearance of the adult 506 population would allow the installation of new larvae. In our case, we can hypothesise that *B. puteoserpentis* 507 would be currently at a stage where the population is dominated by old large mussels which inhibit 508 recruitment. Competition with conspecifics and/or other species was also observed on the EPR where 509 juveniles of Bathymodiolus thermophilus appear to compete with adult mussels that inhibit their settlement 510 via competition for space and food resources or through intraspecific larviphagy (Comtet and Desbruyères, 511 1998, Lehane & Davenport, 2004). Finally, mussel juveniles could be also removed by grazers such as 512 Lepetodrilid limpets and other associated species as suggested for other settlers in the vent habitat (Micheli 513 et al., 2002, Lenihan et al., 2008).

514 The lack of juveniles would result in an aging of the mussel population, and as mortality occurs, to a decrease 515 in the population density of adults. According to the 'endogenous' model, two scenarios then seem possible: 516 the cycle can restart with the settlement of new recruits or the colony can die out if the larval supply is not 517 sufficient (Khaitov and Lentsman, 2016). During observations at different time-periods carried out during 518 several French cruises (BICOSE in January 2014, BICOSE2 in February 2018, BICOSE3 in November 519 2023, HERMINE in March-April 2017, HERMINE2 in July 2022), mussel recruitment patches have not 520 been observed at Snake Pit, even in areas with low mussel densities, which would suggest a lack of larval 521 supply. This may be due to a lack of reproductive individuals locally, and/or a lack of larval input from 522 distant reproductive populations. B. puteoserpentis from SP are genetically very close to southern 523 populations at the Logatchev, Puy des Folles and Semeynov vent fields that could have contributed to larval 524 supply (Breusing et al., 2016). However, predominant contemporary migration directions remain to be 525 assessed.

## 526 4.3. Community structure of the associated fauna

527 Meiofauna largely dominated the overall abundance of the fauna associated with Bathymodiolus 528 puteoserpentis on the Moose site, with the exception of SU-1, which exhibited a co-dominance between 529 copepods and *Pseudorimula midatlantica* gastropods. Two SUs were clearly different from the others: SU-530 5 and SU-6 harboured very high densities of nematodes and copepods, as illustrated by lower equitability 531 evenness and diversity indexes. Copepods colonised habitats with temperatures around 4°C (SU-5/SU-6) 532 and had lower abundance on the slightly colder SU-3 site. This is similar to what was observed by Husson 533 et al. (2017) at the LS vent field where copepods tended to dominate in warmer habitats. Comparing the 534 three 2014 sampling units, the higher densities of nematodes in lower temperature habitats (SU-2/SU-3) 535 were also in accordance with what was observed at LS (Husson et al., 2017), which could explain their low 536 abundance on the warmer and more fluctuating SU-1. In 2018, copepods were no longer dominant at higher 537 temperatures, replaced by nematodes. We can hypothesise that the latter were attracted by higher 538 concentrations of organic matter at the edifice's base, supported by the presence of brittle stars which may 539 be deposit feeders (see below). Overall, meiofaunal densities were lower at the top of the edifice probably 540 linked to harsher environmental conditions and position on the edifice. However, despite our efforts, it was 541 difficult to gather similar physico-chemical parameters on all SUs to discriminate their environmental 542 conditions and specify the niche of each taxa. Finally, a finer taxonomic resolution of the meiofauna would 543 be necessary to refine compositional differences within and between our sampled assemblages.

544 18 macrofaunal taxa were identified in the 1885 individuals collected in 2014 and 2018. They include 14 545 different families, with three (Polynoidae, Alvinocarididae and Lepetodrilidae) potentially comprising 2-3 546 species/morphotypes. This taxonomic richness is only slightly lower than what was obtained by Turnipseed 547 et al. (2004) who identified 22 macrofaunal taxa (excluding *B. puteoserpentis*) from 16 families in a much 548 larger number of individuals sampled (5244 individuals). Nevertheless, anemones and other species of 549 shrimps, such as *Rimicaris exoculata*, although not retrieved in our samples, were observed close to the 550 mussels (See video in the supplementary materials). Macrofaunal densities, ranging from  $65 \pm 8$  to  $483.2 \pm$ 551 69 individuals per litre, were quite similar, and even higher, to those reported by Turnipseed et al. (2004) 552 with  $146 \pm 120$  individuals per litre. Nevertheless, as the volume calculation method differs between the 553 two studies, this comparison may potentially be biassed.

The dominance of *Ophioctenella acies* brittle stars at the base of the studied edifice suggests the presence of a high accumulation of organic matter. The shape of their teeth indicates that they are adapted to sort and grasp the rich source of organic matter produced by the biological activity of mussels and trapped within the assemblage (Tyler et al., 1995). Stöhr and Segonzac (2005) noticed that brittle stars directly positioned

558 themselves on the syphon of mussels evacuating their pseudo-faeces. O. acies were previously observed in 559 dense patches (about 20 ind. $dm^{-2}$ ) among *B. puteoserpentis* mussels at several hydrothermal fields along the 560 MAR (Tyler et al., 1995). In our study, similar patches have been recorded at the edifice's base (max. of 24 ind.dm⁻² of shell mussels). The activity of mussels holobionts may also impoverish the habitats in sulfides 561 and nitrates while enriching them in ammonia and dissolved organic matter as proposed for B. azoricus at 562 563 two other vent fields of the MAR (LS/MG, Sarradin et al., 2009). These observations support the hypothesis 564 that O. acies appear to prefer habitats less directly subjected to hydrothermal fluids, and potentially 565 accumulating higher concentrations of organic matter than the vertical walls of hydrothermal edifices. Thus, 566 we propose that the less turbulent environment at the edifice base (visual observation) would favour the 567 accumulation of organic matter which may be beneficial to brittle stars. This taxa could be considered as an 568 opportunistic species, developing in more turbid and particle-laden habitats, ingesting both organic-rich 569 particles and microbial-coated sulfide particles as proposed by Tyler et al. (1995).

570 Pseudorimula midatlantica limpets, dominating the top of the edifice, are known to be the most abundant 571 gastropods on SP and Broken Spur vent fields (Goroslavskaya and Galkin, 2011; Turnipseed et al., 2004). 572 Within Bathymodiolus azoricus assemblages at LS, they were found in habitats with higher maximum 573 temperatures than the ophiuroids (10°C versus 6.4°C; Husson et al., 2017). Similar to what we observed in 574 the present study, Turnipseed et al. (2004) listed a spatial segregation between O. acies brittle stars and P. 575 midatlantica limpets on the Moose site. Goroslavskaya and Galkin (2011) also observed a similar pattern. 576 Additional environmental data would be necessary to establish a potential link with species distribution. 577 The quasi-absence of niche overlap between these two macrofaunal dominant species may be linked to 578 competition for resources as they are both detritivorous/bacterivorous. The presence of various 579 microhabitats at vents would allow them to partition their niches according to their respective nutritional 580 needs and physiological tolerances (Lévesque et al. 2006). Beside temperature and sulfide concentrations, 581 other physico-chemical conditions such as substratum and topography could affect vent species distribution 582 (Girard et al., 2020; Sarrazin et al., 1999).

Significant differences in faunal abundances, coupled to the macrofaunal composition disparity, between 583 584 the top and the base of the edifice and between spatially close samples, highlights once again the small-585 scale heterogeneity of faunal distribution at vents. Differences between sampling units close to each other 586 may be partially explained by unconsidered biotic and abiotic factors including the presence of microbial 587 mats, substratum types or the amount of organic matter for example (Sarrazin et al., 2022). Indeed, the 588 fluctuating environmental conditions, proximity of high temperature emissions and relief may have limited 589 the colonisation by certain taxa at the top of the edifice (lower meiofaunal & macrofaunal densities). In fact, 590 the role of currents and topography on the distribution of vent species has been neglected but recent evidence

591 points for a significant one (Girard et al., 2020; Lelièvre et al., 2017). Our results confirm that the abundance, 592 densities and species richness of the associated fauna were slightly higher in the 2018 samples, consistent 593 with our expectations of having more stable and homogeneous conditions at the base of the edifice, and 594 therefore fewer environmental constraints allowing a higher number of taxa to colonise microhabitats as 595 observed elsewhere at vents (Cuvelier et al., 2009, 2011a; Rybakova Goroslavskaya and Galkin, 2015). 596 Finally, the number of taxa is much higher at LS with a total of 79 taxa compared to 44 taxa at SP, including 597 6 taxa of copepods and 7 taxa of nematodes (Goroslavskaya and Galkin, 2011, Turnipseed et al., 2004, 598 Zekely et al., 2006, this study). Whether this represents the reality or is due, in part, to the much greater 599 sampling effort at LS is unknown. Only with a more extensive and comparable number of samples will we 600 be able to accurately compare the two vent fields.

## 601 **4.4. Food web**

602 Vent mussels with their autotrophic symbionts are considered as primary producers. Due to chemosynthetic 603 processes converting inorganic carbon into organic compounds, they can be expected to exhibit carbon depletion and hence, a low  $\delta^{13}$ C value (Ruby et al., 1987). *Bathymodiolus* spp. at MAR vent fields host both 604 605 symbiotic thiotrophic and methanotrophic bacteria located in their gills, which provides them with a high 606 plasticity, the proportion of symbionts changing in accordance with the composition of the chemical 607 compounds in their environment (Riou et al., 2008, Duperon et al. 2016). According to the proportion and 608 activity of each type of symbiont, their contribution to the mussel diet varies, and therefore the final  $\delta^{13}C$  of 609 the host varies. At LS, thiotrophic bacteria fix inorganic carbon from CO₂, leading to  $\delta^{13}$ C values down to -610 30‰, while methanotrophic bacteria fix methane, leading to carbon isotopic ratios typically down to -20‰ 611 (Colaço et al., 2002). In our study, most of the *Bathymodiolus puteoserpentis* mussels were included in the 612 "-30 group" (Colaço et al., 2002) with carbon values very close to -30‰, supporting thiotrophy as the 613 dominant nutritional pathway. These carbon values are in agreement with previous studies carried out on B. 614 puteoserpentis at the same site (Cavanaugh et al., 1992; Colaço et al., 2002) and on B. azoricus sampled on 615 different locations at the LS and MG vent fields, which also displayed average carbon ratios close to -30% 616 (De Busserolles et al., 2009; Portail et al., 2018).

However, isotopic ratios of a few *Bathymodiolus puteoserpentis* individuals collected in 2018 showed more dispersed values, with "marginal" carbon signatures close to -15‰. We hypothesise that this intraspecific variability could indicate a potential fluid resource depletion at the edifice base, inducing the use of a wider range of carbon sources. Portail et al. (2018) assessed the trophic network of *Bathymodiolus azoricus* from northernmost vent fields along the MAR, and identified four potential dominant basal sources: photosynthesis-derived organic matter, autotrophs/thiotrophs using the CBB cycle, autotrophs/thiotrophs

623 using the rTCA cycle, and methanotrophs. Assuming that carbon sources at the Moose site have similar 624 isotopic signature to those from northern MAR vent fields, isotopic signatures of our "marginal" B. 625 puteoserpentis could reflect individuals hosting mainly thiotrophic symbionts using rTCA cycle instead of 626 CBB cycle as in the majority of *Bathymodiolus* mussel. This scenario seems unlikely here as the rTCA cycle 627 is usually favoured in habitats with hypoxic conditions such as those occupied by *Rimicaris exoculata* 628 shrimps (Hügler et al., 2005). Alternatively, they may also indicate the presence of methanotrophic 629 symbionts. However Moose mussels were shown to host low proportions of methanotrophic symbionts 630 (Duperron et al., 2016) reflecting low methane concentrations in their habitat, which is also supported by 631 our data (0.33  $\mu$ M for SU-1). A third possibility is the reliance on filter feeding, as observed in some B. 632 azoricus at LS (De Busserolles et al., 2009; Trask and Van Dover, 1999).

633 As in previous trophic studies (De Busserolles et al., 2009; Portail et al., 2018), the SP trophic network 634 includes holobionts (Bathymodiolus puteoserpentis, Rimicaris sp.) as well as detritivores, bacterivores, 635 scavengers and commensal (Branchipolynoe seepensis) taxa. However, despite a relatively low species 636 richness, carbon isotope ratios are spread over a wide range of values from -35.6‰ (B. puteoserpentis, 2014) 637 to -11.1‰ (B. seepensis, 2018) contrary to what was observed on MG and LS vent fields where the  $\delta^{13}$ C 638 values were between -20 and -30% (Portail et al., 2018). This difference in carbon source distribution 639 between the various vent fields could be explained by several factors such as the physiological diversity among symbiotic and free living microorganisms according to local environmental conditions (Karl, 1995). 640 641 The association between polynoid scale worms and bathymodiolin mussels are frequent both in the Pacific 642 (Chevaldonné et al., 1998; Lindgren et al., 2019; Pettibone, 1984) and Atlantic (Britaev et al., 2003; Britavev 643 et al., 2003, 2007; Chevaldonné et al., 1998; Cuvelier et al., 2011a; Van Dover et al., 1999; Ward et al., 644 2004) oceans. In the present study, the isotopic signatures, with a well-differentiated higher trophic level 645 for Branchipolynoe seepensis polychaetes, confirm that most of them rely on B. puteoserpentis. However, 646 some individuals were occasionally observed outside mussel shells where they normally reside (Sarrazin et 647 al., 2014). The uncommon isotopic ratios of some individuals in the present study could be linked to the 648 three *B. puteoserpentis* specimens having a higher  $\delta^{13}$ C.

## 649 4.5. Climax stage in *B. puteoserpentis* assemblage?

650 Condition indexes, size structures and biomass values exhibited significant differences between samples, 651 with SU-1 standing out in most cases, reflecting small-scale heterogeneity. Overall, even fewer small and 652 intermediate size mussels were sampled in 2018 at the edifice base compared to the 2014 collection at the 653 top. The dominance of *Ophioctenella acies* in this area suggests a high accumulation of organic matter, 654 potentially due to feces and/or decomposition of dead mussels. This may be confirmed by the identification

655 of indicative species in the meiofaunal compartment and/or by the measurement of organic matter between 656 sampling sites. The wide carbon isotopic signatures of mussels in 2018 may be due to a lower availability 657 of reduced chemicals at the edifice base. Hypothetically, this situation could have constrained the mussels 658 to seek for other carbon sources, including material sinking from upper oceanic layers. Thus, we hypothesise that *B. puteoserpentis* mussel assemblages could represent a well-established community that has potentially 659 660 reached a final stage of ecological succession (climax) with few observable recruits at least since 2014. Over 661 time, this may (or not) result in a potential long-term population decline. This hypothesis is supported by 662 the observation of many empty mussel shells at the edifice base during the last French cruises in 2022 and 663 2023.

## 664 **5.** Conclusion

665 This study provides additional knowledge on Bathymodiolus puteoserpentis assemblages at the Snake Pit 666 vent field (MAR). The two sampling cruises allowed us to show that the species richness of the fauna 667 associated with mussel assemblages appears to be relatively low despite a diversified and extended food 668 web, and a high abundance of certain taxa. The local edifice-scale heterogeneity of the fauna was partially 669 explained by variations in environmental conditions such as temperatures, fluid proximity and the location 670 of the assemblage on the edifice. Other non-measured physico-chemical factors such as the availability of 671 food sources, substratum types, topography and currents remain to be evaluated. Regarding the engineer 672 species B. puteoserpentis, very little recruitment was observed in 2014 and 2018 at the edifice and vent field 673 scales. The widening of the food web and uncommon isotopic signatures of some B. puteoserpentis 674 individuals potentially show a search for food diversification. These observations suggest that 675 chemosynthetic pathways may not be carried out at least in some places of the site. Large mussels were still 676 observed 5 years after this study during the BICOSE 3 cruise (2023), which opened the way for an 677 alternative hypothesis: B. puteoserpentis population could have reached and still be in a climax state. This 678 raises questions about the future of this community in the years to come. In fact, various scenarios can be 679 considered, ranging from the potential decline of the community which appears to chronically lack renewal 680 through recruitment, to the complete renewal of the population by a settlement event triggered by the release 681 of larvae at a time when the community feels threatened. A complementary study will have to be carried 682 out to validate or not this hypothesis, while continuing the physico-chemical and taxonomic characterization 683 of these mussel assemblages and their associated fauna.

#### 684 Data availability statement

685 The datasets and video images presented in this study can be found in online repositories. Metadata from

the various cruises are provided here: CAMBON-BONAVITA Marie-Anne (2014) BICOSE cruise, RV

687 Pourquoi pas?, https://doi.org/10.17600/14000100; CAMBON-BONAVITA Marie-Anne (2018) BICOSE

688 2 cruise, RV Pourquoi pas?, https://doi.org/10.17600/18000004.

# 689 Credit author statement

Alicia Veuillot: formal analysis, visualisation, writing – original draft, Florence Pradillon:
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# 715 Supplementary materials



# 716

Supplementary Figure S1. Temperature variations in sampling units SU-1, SU-2 and SU-3 from the top of the Moose
edifice at the Snake Pit vent field (Mid-Atlantic Ridge). Temperatures were taken just above the mussels every 30

records during 4 days and 14 hours with MISO autonomous probes.

Supplementary video. Video of the Moose edifice where we can notice the absence of small mussels on the entire
 sulfide edifice. <u>https://video.ifremer.fr/video?id=12092</u>

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# Highlights

- Assemblage heterogeneity at different scales linked to fluid proximity & location.
- Meiofaunal groups dominated. Macrofaunal disparity between top and base of edifice.
- Low number of mussel juveniles over time questions the future of the population.

# **Declaration of interests**

☑ The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

□ The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

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