

# 1 **Plastic-associated pathogens in marine environments: a meta-analysis.**

2 Jeanne Naudet <sup>a\*</sup>, Emmanuelle Roque d'Orbcastel <sup>b</sup>, Thierry Bouvier <sup>b,c</sup>, Jean-Christophe  
3 Auguet <sup>a</sup>

4 <sup>a</sup> UMR MARBEC, Univ. Montpellier, CNRS, Ifremer, IRD, Montpellier, France

5 <sup>b</sup> UMR MARBEC, Univ. Montpellier, CNRS, Ifremer, IRD, Sète, France

6 <sup>c</sup> Institut Halieutique et des Sciences Marines (IH.SM), Univ. Toliara, Toliara, Madagascar

7 \*Corresponding authors.

8 Email address: jeanne.naudet@gmail.com

## 9 **Submission to *Journal of Hazardous Material***

### 10 **Highlights**

- 11 - The plastisphere displays specific potential human pathogenic bacteria (PHPB)
- 12 - PHPB are more abundant in plastic samples compared to the surrounding water column
- 13 - PHPB are not more abundant nor richer on plastics than on other solid substrates
- 14 - *Vibrionaceae* bacteria are the most abundant PHPB identified in plastic samples
- 15 - Marine substrates are a source of PHPB influencing the plastisphere's composition

### 16 **Abstract**

17 The global accumulation of plastic debris in marine ecosystems is continually increasing.  
18 Understanding the interaction between these debris and the bacterial biofilm on their surface is  
19 essential, particularly with regard to potential human pathogenic bacteria (PHPB). In this meta-  
20 analysis, we re-evaluated 16S rRNA metabarcoding data from 35 plastisphere-related studies,  
21 published up to late 2022. We compared the bacteriomes of plastics and other marine substrates,  
22 identifying PHPB associated with each type of substrate. PHPB were enriched in the  
23 plastisphere compared to the surrounding water but did not appear more abundant nor richer in  
24 than in other marine solid substrates. Nevertheless, we described a distinctive PHPB signature  
25 associated with plastics, including *Vibrionaceae* bacteria (5 % of the whole plastisphere) and  
26 biomarkers such as *Staphylococcus haemolyticus*. The contribution of other marine substrates  
27 to plastic PHPB was quantified and we found that shell and wood substrates were potential  
28 sources of PHPB for plastics. Our results suggest that both plastics and other marine solid  
29 substrates could serve as reservoirs for PHPB. However, plastics convey specific PHPB  
30 communities, and due to their ubiquity and persistence in marine ecosystems, plastic debris  
31 poses a higher risk as fomites compared to other substrates.

32 **Keywords**

33 Plastisphere, Bacterial community, Biofilm, 16S rRNA gene

Preprint not peer reviewed

## 34 **1. Introduction**

35 With on average 500,000 tons of plastics entering marine ecosystems annually (Kaandorp et al.  
36 2023), the introduction of plastic into the oceans is now widely recognized as a major  
37 environmental problem. The prolonged persistence of plastics has led to their increasing  
38 prevalence in marine environments, as highlighted in a recent evaluation by Kaandorp et al.  
39 (2023) which estimates the total amount of buoyant marine plastic litter at over 3 million tons.  
40 Degradation of marine macroplastics under the action of ultraviolet radiation, waves and wind  
41 (Song et al. 2017) leads to the formation of small particles classified into mesoplastics (< 20  
42 mm), microplastics (< 5 mm) and nanoplastics (< 1 mm) (Barnes et al. 2009; Hanke et al. 2013;  
43 Provencher et al. 2017). The high prevalence and the ubiquity of macro- and microplastics in  
44 marine environments raise concerns about the growing likelihood of organisms encountering  
45 and ingesting plastics. Plastics are thus a potential threat to marine wildlife and reared species,  
46 not only by direct effects such as entanglement or mechanical obstruction during ingestion, but  
47 also through indirect effects such as chemical intoxication or biological contamination  
48 (Kershaw 2015; Law 2017; de Sá et al. 2018; Senko et al. 2020).

49 Interactions between plastic surface and microorganisms have been studied for more than a  
50 decade (Harrison et al. 2011) and has generated a rising interest in the scientific community,  
51 with more and more articles addressing this subject (Wright et al. 2021). The biofilm that grows  
52 on the plastic surface, called plastisphere, is rich in various microorganisms such as bacteria,  
53 viruses, archaea, fungi and algae (Zettler et al. 2013; Amaral-Zettler et al. 2020). Surprisingly,  
54 the literature does not always converge to the same conclusions concerning the plastisphere  
55 description. Some authors describe a bacterial community specific to the plastics (Kirstein et  
56 al. 2018; Kesy et al. 2019; Martínez-Campos et al. 2022; Crisafi et al. 2022; Naudet et al. 2023),  
57 while other studies found similar communities between plastic and other substrates  
58 (Oberbeckmann et al. 2016; Pinto et al. 2019; Kelly et al. 2022). A recent meta-analysis  
59 conducted an analysis on Illumina sequencing data from 35 plastisphere-related articles,  
60 including marine (n = 26), freshwater (n = 7), and terrestrial (n = 2) studies. They showed that  
61 the substrate has a lower impact than many other environmental or study design parameters,  
62 such as light regime, environment (marine, freshwater, terrestrial), the type of experiment (in  
63 vitro or in situ, collected or incubated plastics), the primer pair used for amplification, and the  
64 extraction method (Wright et al. 2021). They however pointed out the higher abundance of  
65 plastic-degrading bacteria in plastics than in control samples (Wright et al. 2021).

66 The plastisphere has garnered significant attention, since it serves as a potential host for  
67 pathogens capable of inducing infections in both humans and aquatic animals. Several studies  
68 describe an enrichment in pathogenic bacteria (higher abundance and/or richness) on plastics  
69 compared to environmental samples (Sun et al. 2020; Liang et al. 2023; Naudet et al. 2023),  
70 while others report no discernible differences in pathogen levels across different substrates  
71 (Hou et al. 2021; Kesy et al. 2021a; Delacuvellerie et al. 2022). The *Vibrionaceae* family  
72 emerges as the most frequently cited potential pathogenic bacterial lineage described in the  
73 marine plastisphere (Junaid et al. 2022). Bacteria from this family are ubiquitous in marine  
74 environments (Halder 2012; Ina-Salwany et al. 2019) and play a pivotal role as pioneers of the  
75 plastic colonization, being abundant in biofilms (Kesy et al. 2021a). Ultimately, they can cause  
76 lethal diseases in aquatic species, leading to substantial economic losses in aquaculture (Austin  
77 et al. 2016; Stentiford et al. 2017; Ina-Salwany et al. 2019; Zhang et al. 2020). Additionally,  
78 they can pose a threat to human health (Rivas et al. 2013; Mustapha et al. 2013; Letchumanan  
79 et al. 2019; Schröttner et al. 2020). The evaluation of the colonization of micro- and  
80 macroplastics by bacteria potentially pathogenic for humans (PHPB) is essential to identify this  
81 new environmental danger for humans and minimize the health risks linked to the consumption  
82 of marine products.

83 In this study, we conducted a comprehensive meta-analysis of 16S rRNA gene metabarcoding  
84 data derived from 35 articles exploring the composition of bacterial communities associated to  
85 plastics in marine environments. The primary objective is to provide a comprehensive overview  
86 of the current understanding of the marine plastisphere, specifically delving into the bacterial  
87 pathobiome associated with plastics. Additionally, we aim to compare these communities with  
88 those found on other substrates within marine ecosystems. This comparative analysis seeks to  
89 address the pivotal question of whether plastics pose a higher sanitary risk both for marine  
90 ecosystems and human health compared to other substrates.

## 91 **2. Materials and methods**

### 92 2.1. Search strategy and data selection

93 This systematic research was conducted in November 2022 using computerized bibliographic  
94 databases such as PubMed and Google Scholar. The search included a combination of terms to  
95 retrieve articles on plastic microbiomes in marine environment. Articles of interest were limited  
96 to the following criteria: (1) published in English language, (2) focused on plastics microbiomes  
97 developed in marine ecosystems and natural conditions (*in situ* or *in vitro* with circulating  
98 seawater), (3) used metabarcoding analyses. This search resulted in 54 studies. 16S raw reads  
99 from 36 studies were publicly available and extracted for further analyses, or retrieved on  
100 special request to the authors. The collected metadata is presented in the [suppl. table 1 and 2](#).

101 A total of 1,769 samples were analyzed over the 36 selected studies. The samples had been  
102 collected from seas and oceans all around the world, with a majority of studies originating from  
103 the Mediterranean Sea (27 % of all samples), in the North Sea (16 %) and in the China Sea  
104 (16 %) ([Fig. 1](#)). The number of samples analyzed in each study ranged from four samples  
105 (Crisafi et al. 2022) to 304 samples (Scales et al. 2021; [Fig. 2A](#)). Samples mostly consisted of  
106 plastic substrates (69 % of the sample dataset; [Fig. 2B](#)), compared to environmental substrates  
107 (seawater and sediment), artificial substrates (paint, copper and glass), and natural solid  
108 substrates present in marine ecosystems (shell, stone and wood/cellulose).

109 Most of the studies focused exclusively on macroplastics (>20 mm), on microplastics (<5 mm),  
110 or on mesoplastics (>5 mm and <20 mm), respectively representing 43%, 35% and 8% of all  
111 studies, whilst the remaining 14% sampled plastics on a wide range of size ([Fig. 2C](#)). Most of  
112 the analyzed studies incubated their samples in a marine environment for a known duration  
113 (46 %), while the rest of the studies collected samples from the environment (38 %), or  
114 incubated *in vitro* in an open circuit of seawater (16 %; [Fig. 2D](#)).

115 Several methods were used for the extraction and amplification of the samples according to the  
116 study. Extractions were made using commercial kits or custom extraction protocols, and all 16S  
117 amplifications targeted the V3-V4 regions, using different primers according to the study.

### 118 2.2. Sequence processing per study

119 Raw sequencing data were re-analyzed in R (version 4.2.2), using the dada2 pipeline (Callahan  
120 et al. 2016). Sequences were trimmed and filtered based on read quality profiles. Amplicon  
121 sequence variants (ASVs) were inferred using the dada2 algorithm (Divisive Amplicon

122 Denoising Algorithm) after pooling dereplicated reads from all samples. Forward and reverse  
123 reads were then merged and chimeric sequences were removed. Only the forward reads were  
124 used for studies where the quality of the reverse reads did not allow to process the samples.

### 125 2.3. Combined processing of all studies

126 All studies were merged using the phyloseq package (McMurdie et al. 2013) into a unique  
127 phyloseq object containing the whole ASV count table, taxonomy and sequences. The  
128 taxonomic classification of ASVs was performed with the naive Bayesian RDP classifier  
129 implemented in dada2 and using the SILVA reference database nr\_V132.

130 Reads sums were randomly equalized at 10,000 reads per sample using the phyloseq package  
131 (McMurdie et al. 2013). One study was discarded following this process (Jiang et al. 2018).  
132 After standardization, our final dataset consisted of 356,404 ASVs from 35 articles.

133 Potentially human pathogenic bacteria (PHPB) were identified using a homemade full length  
134 16S rRNA gene database derived from the enhanced infectious disease database (EID2, Wardeh  
135 et al. 2015) containing bacterial species (i.e. cargos) reported to have had interactions with the  
136 human species. This homemade database contains 87,405 full length 16S rRNA from the 878  
137 human bacterial cargos in the EID2 database. All the ASV sequences were blasted against this  
138 database, and only ASVs matching a 16S rRNA sequence with 100 % similarity, 100 %  
139 coverage and more than 200 bp were included in our pathobiome dataset.

### 140 2.4. Statistic workflow

141 The richness (i.e. number of ASVs) and Shannon diversity index H were calculated to assess  
142 the taxonomic diversity of each microbial community. Alpha diversity statistical comparisons  
143 were carried out using a Kruskal-Wallis test followed by a Dunn post-hoc test (p-value  
144 corrected by Bonferroni's method). To assess Beta-diversity of the whole communities, the  
145 ASV count table was agglomerated to merge ASVs that have the same taxonomy at the family  
146 level, overcoming differences in ASVs sequences due to the primer pairs used. Beta-diversity  
147 on the pathobiomes was assessed on ASVs count table merged at the species level. The Bray-  
148 Curtis distance was measured on this new ASV table, with the vegan package in R (Oksanen et  
149 al. 2022), and shown in a principal coordinate analysis (PCoA) plot. Statistical analyses were  
150 performed using permutational analysis of variance (PERMANOVA). Species biomarkers (i.e.  
151 differential abundance of species between sample types) in plastic and seawater samples were  
152 identified by the analysis of bacteriome composition with bias correction (ANCOMBC; p-value

153 corrected by Benjamini–Hochberg method) in the microbiomeMarker package (Cao et al.  
154 2022). The SourceTracker package (Knights et al. 2011) was used to quantify the relative  
155 contribution of other marine substrates to the plastic pathobiome, at the species level.

156 The pathobiome phylogenetic tree was constructed by aligning pathogenic species using  
157 MAFFT (v. 7) with default settings. A maximum likelihood phylogenetic tree was then built  
158 using RAxML (v.1.0.0; Kozlov et al. 2019) and visualized on the online web server iTOL  
159 (v.6.8.1; Ciccarelli et al. 2006).

### 160 3. Results

#### 161 3.1. Composition of the marine plastisphere

162 The total dataset represented 356,404 ASVs. Various factors exerted selection pressure  
163 influencing the composition of bacterial communities. These factors include the source study  
164 (PERMANOVA,  $R^2 = 0.414$ ), geographical location (PERMANOVA,  $R^2 = 0.169$ ), primer pair  
165 used (PERMANOVA,  $R^2 = 0.127$ ), and substrate type (PERMANOVA,  $p < 0.001$ ,  $R^2 = 0.104$ ;  
166 [Suppl. Fig. 1A and 1B](#)). While no significant pairwise differences were observed between  
167 studies (pairwise PERMANOVA,  $p > 0.05$ , [Suppl. table 3A](#)), significant differences were noted  
168 between different substrate types, geographical locations, and primer pairs used (pairwise  
169 PERMANOVA,  $p < 0.05$ , [Suppl. table 3B, 3C, 3D](#)).

170 The plastisphere significantly differed from bacterial communities associated to seawater and  
171 other marine solid substrates as shown in PERMANOVA pairwise comparison ( $p < 0.05$ ,  
172 [Suppl. table 3C, Suppl. Fig. 1A](#)). *Proteobacteria* was the main phylum identified in plastic  
173 samples, followed by *Bacteroidota*, respectively representing 52 % and 18 % of the whole  
174 plastic community. The *Vibrionaceae* family represented 5 % of the bacterial community in  
175 plastic samples, while it represented less than 1 % in stone, seawater, sediment, copper, glass,  
176 and paint samples, but 9 % in the wood/cellulose samples and 29 % of the shell community  
177 ([Suppl. Fig.2](#)).

#### 178 3.2. The plastisphere includes potential human pathogenic bacteria

179 A total of 2,876 ASVs were identified as potential human pathogen bacteria (PHPB) in our total  
180 dataset, originating from 456 different species (and 165 different genera), representing 0.8 %  
181 of the total number of ASVs analyzed and 4 % of the whole dataset reads.

182 The proportion of PHPB reads per sample varied significantly according to the substrate type  
183 ([Fig. 3](#)). Plastics displayed a significantly higher percentage of PHPB reads ( $5.8 \pm 13.3$  %) than  
184 seawater samples ( $1.9 \pm 8.6$  %). With a mean of  $7.3 \pm 3.7$  % of PHPB reads per sample, shell  
185 samples had a significantly higher proportion of PHPB reads than copper, glass, plastic,  
186 seawater and sediment samples. In contrast, sediment samples showed the lowest proportion of  
187 PHPB reads ( $0.09 \pm 0.15$  %; [Fig 3](#)).

188 No significant differences were found between the PHPB richness and the Shannon indices of  
189 plastic and seawater samples ([Fig. 4A and 4B](#)). Plastic community displays a significantly  
190 higher PHBP richness and diversity than sediment, and a lower PHBP richness and diversity



191 than stone, but no significant differences in PHPB richness and diversity with copper, glass,  
192 paint, seawater, shell and wood/cellulose (Fig. 4A and 4B). Surprisingly, while alpha diversity  
193 indices of the total community of sediment samples was the highest of all sample types (Suppl.  
194 Fig. 3), sediment pathobiomes were significantly less rich than plastics, seawater, glass, paint,  
195 shell, stone and wood/cellulose (Dunn post-hoc test,  $p < 0.05$ , Fig. 4A), and the Shannon index  
196 was significantly lower than the one of plastics, seawater, glass, paint, and stone (Dunn post-  
197 hoc test,  $p < 0.05$ , Fig. 4B).

### 198 3.3. Shared and specific pathobiome of the marine plastisphere

199 The phylogenetic tree encompassing all PHPB detected across the various substrates of the  
200 dataset revealed that 98% of PHPB species were identified in plastic samples (Fig. 5). Most of  
201 the PHPB species found in plastics were also detected in the water column. Indeed, 329 species  
202 (72 % of plastic pathobiome species) were shared by plastics and seawater. The other marine  
203 solid substrates did not display any specific PHPB species (Fig. 6A). Some bacterial families  
204 were widespread, as for instance the *Vibrionaceae* and *Pseudomonadaceae* families that were  
205 found in eight out the nine analyzed substrates (Fig. 5). *Vibrionaceae* was the most represented  
206 pathogenic family (33 % of the whole PHPB reads), and was particularly abundant in plastics,  
207 wood/cellulose and shell samples (35 %, 68 % and 95 % of the identified PHPB reads  
208 respectively). The *Vibrionaceae* family also includes the two most abundant PHPB ASVs  
209 identified in plastic samples: *Vibrio alginolyticus* and *Vibrio parahaemolyticus*. The  
210 *Pseudomonaceae*, *Moraxellaceae* and *Bacteroidaceae* were then the most abundant families in  
211 plastic samples (9.1 %, 8.3 %, 5.6 % of plastic PHPB reads).

212 Some families appeared contrariwise quite specific to plastic and seawater samples as for  
213 instance the *Tannerellaceae*, and *Enterococcaceae* families that were detected in plastic,  
214 seawater and glass samples. These families were however detected at low levels of abundance  
215 in plastic samples (Fig. 5). Overall, 101 PHPB species were specific to plastic samples (22 %  
216 of PHPB species identified). The ANCOM-BC approach was used to identify biomarker species  
217 from plastics and seawater (Fig. 6B). Five species were found significantly more abundant in  
218 plastic samples compared to seawater and were identified as plastic biomarkers: *Streptococcus*  
219 *salivarius*, *Psychrobacter nivimaris*, *Cutibacterium avidum*, *Staphylococcus haemolyticus* and  
220 *Vibrio alginolyticus*. Furthermore, pairwise PERMANOVA comparison highlighted significant  
221 differences between PHPB communities from all substrates (copper, glass, paint, seawater,  
222 sediment, shell, stone, and wood/cellulose) and plastic pathobiome (Suppl. table 3E). The

223 geographical location of the samples also significantly affected the pathobiome composition  
224 (PERMANOVA,  $R^2 = 0.114$ ).

#### 225 3.4. Tracking the origin of the plastisphere pathobiome

226 The marine environment in which plastics are immersed can contain many different important  
227 reservoirs and vectors for the transmission of PHPB, but the occurrence and abundance of these  
228 PHPB on plastics, and the source of those microbes have not been extensively described. Using  
229 data from the different substrates analyzed in this meta-analysis, we found that shell and  
230 wood/cellulose were significant sources PHPB for plastics (Fig. 7). Overall, plastics had  
231  $10 \pm 19\%$  shell signal (i.e.; 10% of plastic PHPB reads came from shell pathobiomes) and  
232  $13 \pm 20\%$  wood/cellulose signal. Contribution from natural marine solid substrates could be  
233 high, as for instance in the study of Kelly et al. (2022) where stone signal reached a mean of  
234  $73 \pm 36\%$  in plastic samples ( $n = 15$ ), or in the study of Zhang et al. (2022) where shell signal  
235 reached a mean of  $64 \pm 32\%$  in plastic samples ( $n = 73$ ). However, the source of a large  
236 proportion of PHPB plastic species remained unknown. Surprisingly, the contribution of  
237 sediment and water column microbes to the plastisphere was low ( $5 \pm 7\%$  and  $5 \pm 9\%$   
238 respectively) but was detectable in all the plastic studies (Fig. 7).

239 **4. Discussion**

240 4.1. A better homogeneity between plastisphere study protocols is needed

241 Studies on plastic bacterial communities are raising a growing interest in the scientific  
242 community, with more and more publications each year from all over the world (Wright et al.  
243 2021). Particularly, the interaction between plastic surfaces and potential pathogenic bacteria  
244 has raised a great interest, both for micro and macro-plastics (Fig. 2C; Basili et al. 2020; Kesy  
245 et al. 2021; Delacuvellerie et al. 2022; Lear et al. 2022; Martínez-Campos et al. 2022; Zhang et  
246 al. 2022; Liang et al. 2023). The high number of research developing in this field is  
247 accompanied by a high diversity of study designs to describe plastic bacterial communities.  
248 This diversity participates in increasing knowledge in this field, but also in making more  
249 complex data sharing and comparisons.

250 In this meta-analysis, the 35 analyzed studies presented results obtained following various study  
251 designs, including sampling, DNA extraction and amplification protocols. For instance, the  
252 incubation time of plastic in marine environment is variable according to the study, and  
253 sometimes even not known when plastics were collected directly from the environment (Fig.  
254 2D). Yet, the plastisphere composition evolves with time and the biofilm development kinetics  
255 varies according to the environmental conditions and the substrate, making it complex to  
256 compare various sample communities from different incubation times (Pollet et al. 2018; Xu et  
257 al. 2019; Hou et al. 2021). Moreover, the substrate controls used for comparisons or the number  
258 of samples collected for each substrate varies highly, resulting in a total of 6 samples for copper  
259 to 1,224 for plastics analyzed in our meta-analysis (Fig. 2A and 2B). A higher homogeneity in  
260 the choice of control substrates, while remaining relevant in the study context, could also help  
261 the scientific community to highlight differences in communities found on plastic and other  
262 solid substrates. In addition to that, a better homogeneity between studies protocols during  
263 extraction and amplification, as for instance the primer pair used could improve comparison of  
264 samples across studies. We advise the use of the primer pair 515F-Y/926R described by Parada  
265 et al. (2016) which is largely used in microbiome surveys (e.g. earth microbiome project), is  
266 very accurate and has a good coverage of extremely common marine taxa. (Metcalf et al. 2022).  
267 All these study parameters have been proved to influence the bacterial community identified on  
268 solid surfaces, sometimes more than substrate itself (Wright et al. 2021). Our results were  
269 consistent with this observation. as the study, and more precisely the primer pair used, exerted  
270 a stronger effect on bacterial community composition than the substrate type.

271 Besides study protocols, geographical location and environmental factors, such as water salinity  
272 or temperature, were also described as major drivers of biofilm bacterial composition  
273 (Oberbeckmann et al. 2020; Wright et al. 2021; Lemonnier et al. 2022). Our results made no  
274 exception, as we identified the geographical location as a stronger driver of bacterial  
275 communities than substrate type.

276 Difficulties in contacting study authors and accessing raw data contributed to the limitation of  
277 the dataset used in our meta-analyze. 33 % of the articles that met the inclusion criteria did not  
278 provide raw data or metadata on public databases (n = 18), and 14 studies were discarded from  
279 analysis for this reason. Reviewers and editors should pay more attention to this criterion during  
280 the publication process in order to develop open science.

#### 281 4.2. The plastisphere harbors a bacterial signature including potentially human 282 pathogenic bacteria

283 In our study, we confirmed previous results concerning a specific signature of the plastic  
284 bacterial communities compared to the environment, called “plastisphere”. Potentially human  
285 pathogenic bacteria (PHPB) have been detected in previous plastisphere researches (Zettler et  
286 al. 2013; Y. Zhang et al. 2020; Lear et al. 2022; Liang et al. 2023). In their review, Metcalf et  
287 al. (2022) gathered the results from 31 articles on plastisphere in marine and estuarine  
288 ecosystems, and counted 15 potentially pathogenic bacterial genera identified in these studies.  
289 In our meta-analysis, we gathered and re-analyzed the raw sequencing data from 35 studies,  
290 using a homemade full length 16S rRNA PHPB gene database. We could therefore identify a  
291 wider community of PHPB on plastic samples, composed of 163 PHPB genera, including the  
292 15 listed in Metcalf’s meta-analysis.

293 In our results, plastics display an important percentage of PHPB reads in their associated  
294 biofilm ( $5.8 \pm 13.3$  %) which was significantly higher than in the water column and sediments,  
295 but not significantly higher than in other studied substrates (Fig. 3). Similarly, plastic samples  
296 showed a higher richness in PHPB reads than sediment, but not higher than other substrates  
297 (Fig. 4A). Indeed, shell or stone showed particularly higher abundance (Fig. 3) or higher alpha  
298 diversity indices of PHPB reads than plastics (Fig. 4A and 4B). Variation in PHPB abundance  
299 between marine solid substrates could be explained by differences in the physical nature of the  
300 substrate surface, such as roughness and hydrophobicity (Abdalla et al. 2021; Briand et al.  
301 2022), but also the organic or inorganic nature of a substrate that can provide a nutrient source  
302 for the heterotrophic bacterial communities developing on it (Tobias-Hünefeldt et al. 2021).

303 Finally, the load and composition of PHPB in the surrounding environment can also explain  
304 such variability. Interestingly, sediments, which are documented as potential reservoirs of  
305 pathogens in riverine, estuarine and coastal waters (Perkins 2015; Hassard et al. 2016; B. Zhang  
306 et al. 2020; Frank et al. 2024), harbor less PHPB and with lower diversity than plastic samples.  
307 These results suggest that plastic may in fact be a more effective potential reservoir of PHPB  
308 than sediment. At least from the point of view of abundance, the physical and nutritional  
309 properties of biofilm would therefore provide PHPB a better protection against high salinity,  
310 starvation and top-down controls (Yin et al. 2019), ensuring higher survival of PHPB than in  
311 sediment.

312 Among the ASV biomarkers that were identified as enriched in plastic samples compared to  
313 seawater samples, two species are of particular interest knowing their pathogenicity:  
314 *Staphylococcus haemolyticus* and *Vibrio alginolyticus* (Fig. 6A). *Staphylococcus haemolyticus*  
315 is part of the natural skin microbiota but some strains were identified as responsible of human  
316 infections such as bloodstream infections, ocular infections, epididymo-orchitis, urinary tract  
317 infections (Eltwisy et al. 2022). *Vibrio alginolyticus* belongs to the usual marine flora and is a  
318 common pathogen of marine animals. It causes vibriosis, and becomes more prevalent globally  
319 in recent years (Cao et al. 2022). *Vibrio alginolyticus* was classically considered non-pathogenic  
320 in humans (except in immune-compromised persons, where infections can be severe and  
321 potentially fatal) but is currently qualified as an emerging pathogen (Campanelli et al. 2008).

322 The *Vibrionaceae* family is the most abundant family identified in the whole pathobiome,  
323 representing a third of it. It was moreover detected in all tested substrates except for paint  
324 (Fig. 5). This result is consistent with previous studies (Junaid et al. 2022; Naudet et al. 2023)  
325 and illustrates the ubiquity and the high abundance of these bacteria in marine environments  
326 (Sampaio et al. 2022; Namadi et al. 2023). *Vibrio alginolyticus* and *Vibrio parahaemolyticus*  
327 were the most abundant ASVs detected on plastic samples (Fig. 5). *Vibrios* are well known  
328 pathogens as they can cause human illness (Rivas et al. 2013; Mustapha et al. 2013;  
329 Letchumanan et al. 2019; Schröttner et al. 2020) but also fish diseases, leading to economic  
330 losses in aquaculture (Austin et al. 2016; Stentiford et al. 2017; Ina-Salwany et al. 2019; Zhang  
331 et al. 2020).

332 The high ubiquity of some taxa in marine environments raises the question of the origin and  
333 transfer of PHPB between marine substrates, especially marine plastics. The surrounding  
334 environment, especially seawater and sediment have been suggested to be a bacterial source for  
335 plastic colonization (Jiang et al. 2018; Sun et al. 2020; Bandini et al. 2021). Our results

336 highlighted the potential role of other natural marine substrates in the first colonization of  
337 plastics. Depending on the studies, some substrates indeed made a significant contribution to  
338 plastic pathobiome communities, as for instance shell ( $10 \pm 19\%$ ), wood ( $13 \pm 20\%$ ) and  
339 sediment ( $5 \pm 9\%$ ), as highlighted by the SourceTracker analysis (Fig. 7). Coastal plastic debris  
340 may indeed be in contact with other marine solid debris and share a part of their biofilm  
341 communities. The contribution of seawater to plastic biofilms was low ( $5 \pm 7\%$ ) but was  
342 detected in all studies. The role of seawater may be essential to suspend and convey inherent  
343 bacteria and from other substrates to plastics, and to allow inoculation of new colonizers.  
344 Nonetheless, a large proportion of the plastisphere pathobiome origin remained unexplained  
345 (Fig. 7) suggesting that (i) certain bacteria in the plastisphere come from ancient, terrestrial  
346 plastic contamination before the plastic arrived at sea, (ii) that the plastic PHPB come from the  
347 rare "invisible ASV biosphere" of different marine substrates (water, sediment, etc), (iii) or that  
348 plastic PHPB come from sources not identified in this meta-analysis.

349 4.3. Marine solid substrates including plastic debris could act as fomite for PHPB  
350 communities.

351 The role of plastics as fomites or reservoirs of a specific pathobiome community present on  
352 plastic surfaces is drawing increasing attention in the scientific community (Goldstein et al.  
353 2014; Viršek et al. 2017; Bowley et al. 2021; Junaid et al. 2022). However, the particularity of  
354 plastics as a hub of pathogenic bacteria compared to other solid substrates (wood, glass,  
355 stone...) immersed in aquatic ecosystems is questionable (Oberbeckmann et al. 2016; Kelly et  
356 al. 2022; Metcalf et al. 2022). In this study, non-plastic marine solid substrates, when compared  
357 to plastic samples, exhibited discernible bacterial communities in their biofilms (Suppl. Fig. 1).  
358 Additionally, they harbored equally rich and diverse communities of potential human  
359 pathogenic bacteria (PHPB) compared to plastic samples (Fig. 4A and 4B). The proportion in  
360 PHPB reads measured on these samples were similar to those in plastics, and even higher for  
361 some substrates such as shell or stone (Fig. 3). Moreover, as illustrated in the phylogenetic tree  
362 in figure 5, PHPB ASVs that were very abundant in plastic samples were also detected in other  
363 solid marine substrates to a lesser extent. The few bacterial families that were exclusively  
364 identified in plastic samples were detected with low abundance (Fig. 5), and we don't know  
365 whether we would have found them on other substrates with a sampling as considerable as for  
366 plastics ( $n = 1224$  for plastics,  $n = 545$  for all other substrates). These results highlight that all  
367 these marine solid substrates, including plastic debris, could represent potential fomites of  
368 PHPB in the marine environment. A better homogeneity in the choice of controls used to

369 compare plastisphere communities to other biofilms could help to discriminate the different  
370 roles of these substrates as fomite in marine ecosystems (Metcalf et al. 2022).

371 However, the high mobility of plastic debris in the environment associated to its high  
372 persistence in aquatic ecosystems (Turner et al. 2020) make it a better fomite for bacterial  
373 communities compared to other solid substrates. Plastic debris were already described as  
374 involved in the transfer of harmful microorganisms to marine biota (Lamb et al. 2018). Further  
375 studies on the virulence and the effective transfer of pathogenic bacteria from plastic to  
376 organisms should be led to assess sanitary risk of plastic as vector of pathogens (Beloe et al.  
377 2022). We must also consider that, unlike other natural substrates (i.e. stone, shell), plastic is  
378 the product of industrial production and anthropogenic pollution. This last point makes it a  
379 specific case in which mankind can take action to reduce the risk posed by plastic as a reservoir  
380 of pathogens for humans and marine ecosystems.

381

## 382 5. Conclusion

383 This meta-analysis compiles information from studies on the plastisphere utilizing 16S rRNA  
384 metabarcoding analyses, resulting in an extensive dataset on the marine plastisphere and its  
385 related pathobiome. Our analysis underscores the absence of uniformity in the protocols of  
386 plastisphere investigations and particularly the variability of primer pairs used which  
387 significantly influences the composition of bacterial communities observed and may hinder the  
388 discovery of true plastic bacterial biomarkers. The specific composition of plastic marine  
389 biofilms, and particularly the high abundance of PHPB of the *Vibrionaceae* family was pointed  
390 out by our results. Two ASV biomarkers were identified as enriched in plastic compared to  
391 seawater samples, known for their pathogenicity: *Staphylococcus haemolyticus* and *Vibrio*  
392 *alginolyticus*. Due to their persistence and mobility, plastics could vehicle and act as fomites  
393 for harmful organisms through seas. Further investigations should be made to unveil the  
394 sanitary consequences that these plastics could have in marine environment, both for marine  
395 species and human consumers and to propose solutions for managing this emerging risk.

## 396 List of abbreviations

397 ASV: Amplicon Sequence Variant; PHPB: Potential human pathogenic bacteria

## 398 Acknowledgments

399 We would like to thank the Ecole Normale Supérieure de Lyon, the University of Montpellier,  
400 the French National Centre for Scientific Research (CNRS) and the Ifremer for their financial  
401 support during the PhD work of Jeanne NAUDET. We also have benefited from Datarmor as  
402 the supercomputer of Ifremer's "Computing and data center for the sea", whose computing  
403 power and data storage facilitate work in bioinformatics.

#### 404 **Availability of data and materials**

405 The online version contains supplementary material and is available at xxx

#### 406 **Authors' contributions**

407 JN carried out all practical aspects of the study with guidance and technical support from ER,  
408 JCA and TB. TB coordinates the 3 projects that surrounded JN's thesis. JN wrote the first draft  
409 of the manuscript and all authors contributed substantially to revisions.

410 Full names as follows: Emmanuelle ROQUE D'ORBCASTEL (ER), Jean-Christophe  
411 AUGUET (JCA), Jeanne NAUDET (JN), Thierry BOUVIER (TB).

#### 412 **Competing of interest**

413 The authors declare no competing interests that could have appeared to influence the work  
414 reported in this paper.



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