

1 **A Year-long Field Study of Buried Plastics Reveals**

2 **Underestimation of Plastic Pollution on Hawaiian Beaches**

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21 **Abstract**

22 Global models estimate that two-thirds of floating ocean plastic has accumulated in coastal
23 areas since the 1950s, with Hawai'i's windward shores particularly vulnerable due to their
24 proximity to the North Pacific Garbage Patch. Our quarterly surveys revealed that 91% of
25 recovered plastic particles were buried below the surface (deeper than 2 cm), with most particles
26 being small fragments (93%) ranging from 5.4 to 7.9 mm. This study offers new insights into
27 subsurface plastic pollution, exposing a previously hidden vertical distribution. We observed
28 significant variations in plastic abundance across depths, beaches, and sampling periods, along
29 with a positive correlation between particle size and sand grain size. Additionally, through
30 reconciliation science, we critically reflect on the cultural impacts of our research, emphasizing
31 the importance of aligning plastic pollution studies with local community values and
32 environmental stewardship.

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41 **1. Introduction**

42 Plastic pollution has become one of the most pressing environmental challenges globally
43 (Calil, J *et al.*, 2021). Since the onset of synthetic polymer production in the 1950s, global
44 models of plastic pollution predict that approximately two-thirds of the plastic mass released
45 into the ocean has either stranded or settled around the world’s shorelines (Kaandorp *et al.*,
46 2023; Lebreton *et al.*, 2019; Onink *et al.*, 2021). Furthermore, plastic debris transported by
47 ocean currents and wind can travel vast distances across the world’s ocean far away from its
48 source, eventually converging in one of the five subtropical gyres, where it persists as “legacy
49 plastics.” The accumulation of floating plastic debris in these gyres is commonly referred to as
50 “Garbage Patches” (Eriksen *et al.*, 2014; Lebreton *et al.*, 2018; Maximenko *et al.*, 2012; van
51 Sebille *et al.*, 2020). Shorelines near these Garbage Patches, particularly those of islands that
52 are exposed to the large current systems that form the gyres, are particularly susceptible to
53 receive “legacy plastics” pollution (Barnes *et al.*, 2018; Pham *et al.*, 2023). While large-scale
54 models are valuable in identifying high accumulation areas, they often lack the resolution
55 needed to fully describe the temporal and spatial variations in plastic accumulation as well as
56 the state of plastics (e.g. size, shape, type, degradation level) in these regions (Critchell *et al.*,
57 2015; Critchell & Lambrechts, 2016; Hardesty *et al.*, 2017; Melvin *et al.*, 2021). This limitation
58 hinders our ability to develop effective mitigation strategies tailored to the specific challenges
59 these areas face.

60 Hawai‘i is heavily impacted by plastic pollution due to its proximity to the North Pacific
61 Subtropical Gyre, where the high concentration of floating plastic debris is known as the North
62 Pacific Garbage Patch, NPGP, (Berg *et al.*, 2024; Carson *et al.*, 2013; Kaandorp *et al.*, 2023;
63 Lebreton *et al.*, 2018). Beaches on the windward (east) side of the islands, which are exposed

64 to the trade winds and are aligned with the NPGP, receive a significant amount of weathered
65 marine debris that has traveled across the Pacific Ocean (Brignac *et al.*, 2019; Berg *et al.*, 2024;
66 Carson *et al.*, 2013). This has been confirmed by studies showing that a significant number of
67 identifiable items on Hawaiian beaches originate from non-local sources, as evidenced by
68 foreign labels, and the severely degraded, biofouled surfaces suggesting long-term exposure to
69 the marine environment (Brignac *et al.*, 2019; Carson *et al.*, 2013; Connors *et al.*, 2024).
70 Although numerous studies and field surveys have investigated the state and abundance of
71 plastic debris on Hawaiian beaches, these studies have focused on the beach surface layers, and
72 to our knowledge, there are no reports investigating the distribution of plastic particles deeper
73 than 10 cm in the sand columns of Hawaiian beaches (Agustin *et al.*, 2015; Berg *et al.*, 2024;
74 Blickley *et al.*, 2016; Brignac *et al.*, 2019; Carson *et al.*, 2011, 2013; Cooper & Corcoran, 2010;
75 McDermid & McMullen, 2004; Ribic *et al.*, 2012; Young & Elliott, 2016).

76 The handful of studies that have investigated buried plastic—in beaches in Brazil, in the
77 Indian Ocean, and on the Azores—have highlighted significant quantities in the deeper
78 sediment layers (Fauziah *et al.*, 2015; Kusui & Noda, 2003; Lavers *et al.*, 2019; Pham *et al.*,
79 2023; Taniguchi *et al.*, 2016). For instance, Pham *et al.*, (2023) revealed that buried plastic in
80 sand column cores (10.1-100 cm) from sandy beaches across the Azorean archipelagos
81 accounted for 84% of the total abundance of plastic particles found in the sampled sand-core,
82 with concentrations varying across the horizontal stretch of the beach, with the highest
83 concentration at the backshore of the beach (comprising the berm and foredune). Another study
84 by, Lavers *et al.*, (2019) on the Cocos Islands (Australia) revealed that between 10-70 times
85 more debris items per m² are buried compared to items visible on the surface of beaches.
86 Moreover, plastic particles were discovered down to 2 m depth on a sandy beach in Brazil
87 (Taniguchi *et al.*, 2016; Turra *et al.*, 2014). These findings highlight that limiting surveys to

88 plastic particles on the beach surface could underestimate the total abundance of plastic
89 pollution in littoral environments.

90 Considering that the quantity and type of plastic found on beach surfaces are often used to
91 get a proxy for the extent of ocean plastic pollution levels, with oceanic insular areas being
92 particularly valuable for monitoring global pollution trends (Barnes *et al.*, 2018; Pham *et al.*,
93 2023; Serra-Gonçalves *et al.*, 2019), including the subsurface concentration would provide a
94 more comprehensive assessment of the state of ocean plastic pollution. Furthermore, repeated
95 measures of beach plastic, with appropriate frequencies, could reflect changes in the abundance
96 of debris at sea (Ryan *et al.*, 2009). As noted by Pham *et al.* (2023), tracking buried plastic
97 particles over time may provide a more accurate representation of total plastic abundance on
98 beaches and serve as a more stable indicator of ocean plastic pollution, as these buried plastics
99 are less susceptible to disturbance from e.g. wind and beach clean-ups. Thus, a better
100 understanding of the quantity of plastic particles in the deeper layer of the sand column could
101 be useful in assessing ocean plastic pollution and the effectiveness of upstream mitigation
102 efforts, especially in zones which are predicted to receiving significant amount of the “legacy
103 plastics.”

104 In this study, we investigate the distribution and concentration of plastic debris within the
105 sand columns of beaches on the windward side of O‘ahu. Through quarterly surveys of plastic
106 debris (> 0.5 mm) across different depths (up to 1 m) in 60 x 60 cm quadrats, conducted from
107 November 2022 to February 2024, we aim to assess changes in observed plastic abundance in
108 the sand column. By examining plastic concentrations at various depths and beaches, we also
109 provide first insights into the total abundance and distribution of plastic on beaches of Hawai‘i.

110 This data can refine our understanding of beach plastics on the shores of Hawai‘i and contribute
111 to more comprehensive monitoring of broader trends of plastic pollution in the Pacific Ocean.

112 This study also includes an initially unplanned but transformative collaboration with a
113 Kanaka ‘Ōiwi (Native Hawaiian) leader, Kimeona Kāne, Chairperson of the Waimānalo
114 Neighborhood Board and cultural practitioner, which began in the later stages of the research.
115 This collaboration led to critical realizations about the potential harm our research methods
116 might have caused, such as disinterring ancestral burial remains, and prompted a reflection on
117 how our scientific approach failed to address important cultural considerations. We draw from
118 and seek to extend Liboiron *et al.*'s (2021) concept of reconciliation science, which calls for
119 integrating such reflections into the scientific process. As Liboiron *et al.* argue, these reflections
120 should be named in scientific works, “*Rather than dividing these reflections into a separate*
121 *‘opinion’ piece or social science paper...*” In line with this approach, we provide a critical
122 reflection on our research methods and findings through the lens of this collaboration, aiming
123 to prevent further harm by highlighting our missteps. These insights are discussed in the Results
124 and Discussion sections.

125 **1. Methods**

126 *1.1. Study area*

127 The field sampling for this study was conducted on O‘ahu which is one of the eight main
128 islands of the Hawaiian archipelago in the North Pacific Ocean and is the third largest and most
129 populated island in Hawai‘i (Fletcher *et al.*, 2012). O‘ahu has approximately 107 km of sandy
130 beach that is separated into four regions: north, east, south, and west (Fletcher *et al.*, 2012). The
131 three beach sites used for this study; Kahuku (21°42'09.0"N 157°57'36.0"W), Kokololio
132 (21°37'41.2"N 157°55'15.6"W), and Waimānalo (21°20'06.0"N 157°41'45.6"W), are all located

133 on the east side of O‘ahu (as shown in Fig. 1), also known as the windward side due to the
134 predominant easterly trade winds that consistently hit O‘ahu’s eastern coast and expose the
135 shoreline to short-period trade wind waves year-round. In addition to trade wind waves the
136 windward coast is also affected by large refracted North Pacific swells during winter (Fletcher
137 *et al.*, 2012). Shallow fringing reefs protect much of East O‘ahu’s shoreline from the full energy
138 of large waves. However, beaches behind these protective reefs are typically low-lying and
139 narrow, prone to inundation during large waves and storms. Furthermore, Hawai‘i is in a micro-
140 tidal zone with a tidal range of about 1 m but large winter swells can cause variations in beach
141 width by up to two-thirds (Fletcher *et al.*, 2012).

142 The combination of the east side of O‘ahu facing and being near the North Pacific
143 Subtropical Convergence Zone, where debris accumulates to form the NPGP, and the influence
144 of northeast and easterly trade winds, brings significant amounts of plastic debris to the shores
145 on the windward side of O‘ahu (Cooper & Corcoran, 2010; Kubota, 1994). Studies comparing
146 the leeward and windward sides of O‘ahu also suggest that the windward side is more prone to
147 accumulate plastic originating from the open ocean (Brignac *et al.*, 2019). The beaches were
148 selected due to their locations on the windward of O‘ahu, and their varying beach characteristics
149 and plastic accumulation patterns, with Kahuku known for accumulating the largest amounts
150 of plastics at the surface (Brignac *et al.*, 2019; Young & Elliott, 2016). A summary of the key
151 characteristics of Kahuku, Kokololio, and Waimānalo, including their positions, dimensions,
152 and sand size as classified by the Wentworth scale (Wentworth, 1922), and notable features, is
153 provided in Table 1.

154 **Table 1:** Key characteristics of the three studied beaches; Kahuku, Kokololio, and Waimānalo,
155 including their geographical position, approximate beach length, subaerial width, sand grain

156 size (classified by the Wentworth scale; Wentworth, 1922), and any notable environmental or
 157 physical features relevant to the study.

Beach	Position	Length (m)	Width (m)	Sand size (Wenworth)	Notable Features
Kahuku	North East	~600	5-15	Coarse	Located in James Campbell National Wildlife Refuge (low human presence). Exposed to strong winter North Pacific swells. Dynamic shoreline changes.
Kokololio	East	~650	10-30	Coarse to Medium	Situated in a bay. Affected by trade wind waves year-round and large refracted North Pacific swells during winter.
Waimānalo	South East	~6,500	20-40	Medium to very fine	Largest of the three beaches, located in a bay. Exposed to easterly trade winds and winter swells.

158 *1.2. Beach plastic sampling*

159 All sites considered in this study were sampled every three months over 15 months, starting
 160 in November 2022, with subsequent samplings in February 2023, May 2023, August 2023,
 161 November 2023, and February 2024. Sampling was conducted using triplicate quadrats (60 x
 162 60 cm) placed 10 m apart (measuring from the middle of the quadrats) and parallel to the
 163 shoreline along the drift line (the berm). The drift line was defined as the line of debris
 164 accumulation above the high tide line (as shown in Fig. S1 in the SI). All three quadrats were
 165 sampled on the same day, resulting in three separate sampling days per beach (with 1-7 days
 166 between the sampling days, most of which were conducted during weekends to allow more
 167 volunteers to join). The exact locations of the quadrats were marked by their distance from the
 168 vegetation line and documented with photographs. The quadrats were positioned 5-10 m from

169 the locations of the previous quadrats from the preceding sampling campaign three months
170 prior, using the previous location information. The distance varied to avoid disturbing local
171 wildlife (e.g., monk seals) and beachgoers. The same quadrat was never resampled during the
172 field surveys. All sets of quadrats were sampled within the same sections of the beaches, within
173 an approximately 20-50 m range, to maintain consistency and avoid significant variations
174 across different beach areas.

175 The quadrats used were four-sided hollow boxes with dimensions of 60 x 60 cm and a width
176 of 20 cm and were made on O‘ahu by Pac Pro Hawai‘i (<https://www.ppg-hi.com/>) using 16-
177 gauge galvanized steel with welded corners. Inside the metal frame, markings at 2 cm and 10
178 cm were added as guides for collecting different sand layers (see Fig. S6 for the metal frame in
179 the SI).

180 The sand of the surface layer was collected within the metal box frame down to the 2 cm
181 marking and added to a 5-gallon bucket (22.7 L), which had a marking of 7.2 L (corresponding
182 to the volume of sand from 60 x 60 x 2 cm), serving as a secondary guideline for the quantity
183 of sand needed to be collected. The collected sand in the bucket was weighed, and the
184 temperature and moisture content of the sand were measured using a liquid-in-glass
185 thermometer and a high-frequency moisture meter DML300L by MeterTo. The mass of the dry
186 sand was then calculated by subtracting the mass attributed to moisture from the total mass of
187 the sand of each layer. The sand was then transferred to the top bin of the Buoyancy Separation
188 Device (BSD) developed on O‘ahu by Seed.World (<https://www.seed.world/>). The BSD is a
189 wheelbarrow equipped with a top bin (see Fig. S8 in the SI) and two battery-powered hoses that
190 circulate ocean water from the main body of the wheelbarrow (under the top bin) to the top bin.
191 The water from the hoses is used to stir the sand and to separate the buoyant debris from the

192 sand. By positioning the BSD at a tilted angle, the floating debris moves towards the lower end
193 of the wheelbarrow and is collected in a 250- μ m mesh bag, while the heavier cleaned sand
194 remains in the top bin. Once the sand had been stirred until no more floating debris was
195 observed, the clean sand was returned to the beach. This method facilitated the effective in-
196 field separation of sand and floating debris, allowing only the removal of floating debris. After
197 removing the first 2 cm of sand, the quadrat content of the next 10 cm of sand was exhumed to
198 a depth of 12 cm, representing the 2-12 cm layer. This sand was collected into two 5-gallon
199 buckets with markings at 18 L each to guide the collection of the next 10 cm layer of sand,
200 totaling 36 L (volume of 60 x 60 x 10 cm). The floating debris was collected as described above.
201 The process was repeated, collecting buried floating debris in subsequent 10 cm increments
202 down to a depth of 1 m. Buoyant debris from each depth layer was collected into separate mesh
203 bags, resulting in 11 mesh bags per quadrat. Efforts were made to avoid collecting too much
204 natural debris. On the occasional encounters with small marine animals (e.g., crabs) within the
205 sand column were returned to the beach if found in the quadrats.

206 Overall, 594 samples were collected (11 depth layers, 6 sampling time sets, 3 beaches, and
207 3 quadrats per beach). After each quadrat was sampled and the plastic removed, the 'clean' sand
208 was replaced to restore the area. A detailed field survey method, with step-by-step pictures, are
209 provided in the SI.

210 1.3. *Plastic sampling processing*

211 The mesh bags with the floating debris collected per stratified layer were air-dried for at
212 least a week after collection. The content for each dried mesh bag was weighed to the nearest
213 0.001 g (Mettler-Toledo, AE 240-S Dual Range Balance) and then spread out on a mat. The
214 plastic debris was manually separated from the natural debris and the total amount of plastic

215 particles collected per sample was weighed to the nearest 0.001 g. The plastic debris was then
216 laid out on a blue background sheet alongside a reference coin of 37 mm diameter and were
217 photographed with a resolution of at least 10 pixels in diameter. The image was processed using
218 the Segmentation Model developed by The Ocean Cleanup, where the model workflow is
219 described by Royer *et al.*, 2024. This model classifies each plastic particle into four classes
220 (hard fragment, pellet, line, and foam), into 12 colors (black, white, blue, green, red, orange,
221 salmon, yellow, lightblue, lightgreen, indigo, turquoise, and lightgray), and measures the size
222 of the particles (minimum and maximum length).

223 The plastic counts (n) provided by the Segmentation Model processing and the mass (in kg)
224 of the measured dry weight (DW) of the sand were used to calculate the concentration ($n \cdot \text{kg}^{-1}$
225 DW) of plastics in the sand column sampled at the three beach sites. The total counts of plastic
226 particles per stratified layer and the concentration ($n \cdot \text{dm}^{-3}$) of plastic particles in terms of the
227 volume of sand are provided in S9 and S10, respectively, in the SI.

228 1.4. Polymer identification

229 Attenuated Total Reflectance/Fourier Transform Infrared spectroscopy (ATR/FTIR) was
230 used to determine the bulk polymer identity of sampled plastic particles collected during field
231 surveys. If the sample size exceeded 60 particles, a random selection of 60 particles was
232 subjected to polymer ID analysis. The spectra were recorded between 4000 cm^{-1} and 550 cm^{-1}
233 with a resolution of 4 cm^{-1} and 16 scans on a Nicolet iS5 FTIR spectrometer equipped with a
234 KBr beam splitter and a diamond laminate iD5 ATR module. The main peaks in the spectra
235 were identified, using the OMNIC™ Spectra Software, to determine the functional groups
236 present and establish the polymer identity, following the method described by (Jung *et al.*,
237 2018). For samples requiring further verification or those that could not be identified manually,

238 spectral libraries within the OMNICTM Spectra Software were consulted, provided the search
239 score was ≥ 0.90 (on a scale from 0 to 1). Samples that remained unidentified after these steps
240 were labeled as UnI (UnIdentified).

241 1.5. Data analysis

242 To determine whether plastic counts (the dependent variable) differed among the fixed
243 effects - beach sites, months, and depth - while accounting for quadrats as a random effect, a
244 generalized linear mixed-effects model (GLMM) was employed following similar approach to
245 that employed by Pham *et al.*, 2023. The GLMM accommodates both fixed effects, which were
246 'Beach', 'Month', and 'Depth', and random effects, specifically 'Quadrat' in our case. It should
247 be noted that the number of Months used in our model were six to represent the six field surveys.
248 The random effect for quadrats allows for variability between sampling units to be accounted
249 for, acknowledging that each quadrat might have inherent differences affecting plastic counts.
250 The model was fitted using a negative binomial distribution to address overdispersion in the
251 count data. The 'lme4' package (using R software) was used for fitting the linear and GLMM
252 models. To assess the significance of the fixed effects, an ANOVA was conducted using Wald
253 Chi-square (χ^2) tests using the 'car' package (using R software). This analysis tested the
254 influence of each fixed effect on the dependent variable, 'Counts'. Specifically, the categories
255 tested were 'Beach', 'Month', and 'Depth'. The ANOVA provided a statistical assessment of
256 how plastic counts varied across these categories. All data processing and graphical
257 representations were performed using R software version 2024.04.2.

258 1.6. Sand grain size

259 To measure the sand grain size at each beach, we used the Citizen Science tool
260 'SandSnap' (<https://sandsnap-erdccchl.hub.arcgis.com/>), a web-based application designed to

261 collect sand grain information from a wide range of locations worldwide. Users take a photo of
262 the sand with a U.S. coin in the frame and upload the image for automated sand grain analysis.
263 The application calculates nine gradation metrics (D_{10} , D_{16} , D_{25} , D_{50} , D_{65} , D_{75} , D_{84} , D_{90} , and
264 mean), and the results are added to the SandSnap database, which can be viewed on the data
265 viewer at <https://sandsnap-erdch1.hub.arcgis.com/> (McFall *et al.*, 2024). For our study, we
266 used a U.S. quarter for the SandSnap processing and used the value D_{50} which is the commonly
267 used data point to represent sand grain size. The D_{50} value, or the median grain diameter,
268 represents the size at which 50% of the grains in a sediment sample are smaller and 50% are
269 larger. SandSnap has reported a 22% error in D_{50} values (McFall *et al.*, 2024). Pictures were
270 taken for each depth layer along the sand column in 1-2 quadrats, and this was conducted during
271 the last two field campaigns (November 2023 and February 2024). Our data entries are
272 summarized in the table S1 in the SI. SandSnap provided an easy and convenient method to
273 collect many data points without the need to extract sand from beaches.

274 1.7. *Permits*

275 The beach site at Kahuku was located just outside the James Campbell National Wildlife
276 Refuge, requiring a permit for access. This permit was secured through the James Campbell
277 National Wildlife Refuge prior to the commencement of fieldwork and remained valid for the
278 duration of the study. No additional permits were necessary for plastic removal activities on the
279 beach, as confirmed by the Department of Land and Natural Resources (DLNR) and the
280 Division of Aquatic Resources (DAR), provided that only plastic was removed.

281 1.8. *Analysis of Methods for Reconciliation Science*

282 In our effort to align with reconciliation science as conceptualized by Liboiron *et al.*
283 (2021), we critically analyzed our methods through collaboration with cultural practitioner

284 Kimeona Kāne. Reconciliation science, as described by Liboiron *et al.*, (and how we interpret
285 it), emerges from the recognition that scientific research often perpetuates colonial frameworks
286 and extractive practices that often overlook or undermine local cultural practices and culture-
287 ecological interconnections. Liboiron *et al.* advocates for embedding reflections of research
288 relations within the scientific study and reporting, rather than relegating them to separate
289 afterthought pieces. This encouragement, together with our collaboration with Kimeona Kāne
290 provided guidance for how we reflected on our work in Hawai‘i, especially after realizing that
291 our initial methods failed to adequately consider the cultural and ecological implications for the
292 local Hawaiian community.

293 **2. Results & Discussion**

294 *2.1. Plastic concentration in the sand column*

295 From the 594 samples collected at Kahuku, Kokololio, and Waimānalo during the six field
296 campaigns, a total of 77,033 plastic particles were recovered across all three beaches. Each
297 plastic sample was photographed, and the images were processed through the Segmentation
298 Model (Royer *et al.*, 2024) which provided detailed information on the plastic particle count
299 for each sample, along with additional data such as particle lengths, colors, and classifications.
300 An example of the output image from the Segmentation Model is shown in Fig. 3.

301 Plastic particles were found at depths down to 1 m across all three beaches, and the sand
302 column sampling revealed that the majority (91%) of the total abundance of sampled plastic
303 particles were located below the surface, ranging from 2 to 102 cm deep. Specifically, at
304 Kahuku, 75% of the plastic particles recovered were buried; at Kokololio, 91%; and at
305 Waimānalo, 92% of the total plastic particles were found below the surface layer. This higher
306 abundance of plastics in the deeper layers aligns with findings from studies investigating the

307 distribution of plastics below the sand surface. This higher abundance of plastic particles in the
308 deeper layers is consistent with the findings of the handful of studies that have investigated
309 buried plastics on sandy beaches (Pham *et al.*, 2023; Lavers *et al.*, 2019; Taniguchi *et al.*, 2016;
310 Turra *et al.*, 2014). These findings suggest that solely examining surface layers may
311 significantly underestimate the total abundance of plastic pollution on beaches. For example,
312 in our study, the surface layer contains only 8-25% of the total plastic particles, depending on
313 the beach.

314 Furthermore, using the plastic particle counts for each sample, we investigated the
315 concentration of plastic particles per dry weight of sand along the sand column as shown in Fig.
316 4. While the highest concentration of plastic particles is generally found at the surface, the sand
317 column analysis at Kokololio and Waimānalo revealed notable exceptions. Several deeper
318 layers, specifically at depths of 52-62 cm, 62-72 cm, and 72-82 cm, showed significantly higher
319 concentrations of plastic particles, which are particularly apparent during the February 2024
320 sampling at Waimānalo and the February 2023 sampling at Kokololio, with some layered
321 samples revealing over 2000 particles at these depths. To date, it remains unclear how plastics
322 become buried at such depths (2-102 cm). Whereas the observed plastic abundance within the
323 1-meter sand column during May 2023 and August 2023 was relatively low across all three
324 beaches. To further investigate the observed variations in plastic abundance, we conducted
325 statistical analyses to identify significant factors contributing to these changes, as detailed
326 below.

327 To investigate whether factors such as depth within the sand column significantly affect
328 plastic particle counts, we fitted a GLMM with a negative binomial distribution to account for
329 overdispersion. The fixed effects in the model included 'Beach', 'Month', and 'Depth', while

330 'Quadrat' was treated as a random effect. Results from the GLMM summary revealed that the
 331 variance for quadrat (random effect) was low (variance = 0.03, Std. Dev. = 0.17), indicating
 332 minimal variability between quadrats in plastic counts and that the variation in plastic counts is
 333 likely driven by the fixed effects rather than quadrat-level differences. ANOVA was conducted
 334 to test the influence of each fixed effect; 'Beach', 'Month', and 'Depth', on the dependent
 335 variable, 'Counts'. The results showed that all three factors - Beach, Month, and Depth -
 336 significantly contributed to the variation in plastic particle counts, as summarized in Table 2.

337 **Table 2.** Analysis of deviance with χ^2 of fixed categorical variables: depth (n = 11), beach (n =
 338 3), and Month (n = 6) in GLMM negative binomial of plastic counts.

Variable	χ^2	Df	<i>p</i>
Beach	267.751	2	< 0.001
Month	120.985	5	< 0.001
Depth	56.155	10	< 0.001

339 In line with ANOVA results, and as observed in Fig. 4 the beach has a significant impact
 340 on the observed plastic concentration in the sand column. Significantly fewer plastic particles
 341 were consistently recovered at Kahuku compared to Kokololio and Waimānalo. For instance,
 342 Kahuku had a total of 4,269 particles recovered across all field surveys. In contrast, Kokololio
 343 had nearly ten times as many particles, with a total abundance of 41,840 plastic particles, and
 344 Waimānalo had about seven times more than Kahuku, with a total of 30,924 particles.
 345 Interestingly, Kahuku and the neighboring beach Kahuku are commonly known to be the most
 346 polluted beaches on O'ahu (Brignac *et al.*, 2019; Young & Elliott, 2016), more so than
 347 Waimānalo and Kokololio. These findings highlight that a beach with high surface pollution
 348 may not necessarily correspond to high pollution levels in the sand column.

349 Additionally, we observed significant fluctuations in the number of plastic particles at all
350 three beaches when sampling during different months (Fig. 4). To better visualize these
351 temporal trends, we plotted the total abundance of buried plastics (2-102 cm) and surface
352 plastics (0-2 cm) against the survey months (Fig. 5). As shown in Fig. 5, a decrease in the total
353 abundance of plastic particles (in logarithmic scale) was observed in the 1 meter sand column
354 at all three beaches from February 2023 to May 2023, followed by an increase from August
355 2023 to November 2023, and another rise into February 2024, (with the exception for Kahuku
356 in February 2024).

357 The quarterly surveys revealed variations in the observed abundance of plastic particles
358 within the sand column down to 1 meter from the surface. Higher abundances were typically
359 observed during winter months and lower abundances during summer months across all three
360 beaches. The variation in observed plastic concentrations across different sampling months may
361 be attributed to processes such as fresh sand deposition, erosion, or the plastic being washed
362 away. Our ongoing research effort focuses on sediment erosion, accretion, and deposition at
363 these sites to better understand whether the burial of plastics in the sand column are driven by
364 beach erosion, deposition dynamics, and other environmental factors. This work will help
365 clarify the processes affecting the observed plastic concentrations in the deeper beach sediment
366 layers.

367 2.2. *Plastic Sizes*

368 The size distribution of the recovered plastic particles was measured using the Segmentation
369 Model. This approach allows for a more continuous distribution of particle sizes compared to
370 traditional size ranges (Royer *et al.*, 2024). By rounding up the maximum length computed by
371 the Segmentation Model to 0.1 mm, we investigated the distribution of particle lengths by

372 calculating the mean max length (mean length = m , in mm) of the particles (counts per sample
373 = n) at each depth and beach, as illustrated in the box plots in Fig. 6. To avoid skewedness of
374 results the lines were removed and the results with lines (representing 3% of all particles
375 sampled) are also shown in Fig. S14 in the SI. The average particle lengths per layer in the sand
376 column were measured to be in the range of 5.9-7.5 mm at Kahuku, 6.6-7.6 mm at Kokololio,
377 and 5.5-6.6 mm at Waimānalo. The sizes of the plastic particles did not exhibit a clear trend
378 with increasing depth in the sand column.

379 During the field surveys, we observed that the sand grain size also varied among the
380 three beaches, with Waimānalo having the finest sand. To investigate any possible correlation
381 between the sand grain size and the mean length of the plastic particles, we measured the grain
382 size using SandSnap (McFall *et al.*, 2023) during the last two field survey campaigns
383 (November 2023 and February 2024). The output files retrieved from SandSnap are available
384 in Table S1 in the SI. Using SandSnap D_{50} values, we determined the sand composition at each
385 beach according to the Wentworth scale: Kahuku has coarse sand (larger D_{50} values), Kokololio
386 has medium to coarse sand (medium D_{50} values), and Waimānalo has medium to fine sand
387 (smaller D_{50} values) (Wentworth, 1922). For each depth in the sand column, we measured the
388 D_{50} value of the sand grain size and correlated it with the mean maximum length of the plastic
389 particles at that depth through a linear regression model. Fig. 7 illustrates a potential positive
390 correlation ($m = 1.79$, $R^2 = 0.18$, with a p-value of 5.36×10^{-5} from Wald test) in that as sand
391 grain size increases, the mean particle length tends to increase as well. As highlighted in the
392 colored circles, Waimānalo (blue), with its finer sand, generally contained smaller plastic
393 particles. Kokololio (green), with coarser sand than Waimānalo but finer than Kahuku (red),
394 contained plastic particles that were slightly larger than those at Waimānalo but smaller than
395 those at Kahuku. The plastic particle sizes at Kahuku exhibited greater variability compared to

396 Kokololio and Waimānalo, however in general slightly larger plastic particles were found in
397 the sand columns at Kahuku.

398 It could be assumed that finer sands, such as Waimānalo, tend to retain smaller plastic
399 particles, while coarser sands, like those at Kahuku and Kokololio, are less effective at retaining
400 these smaller particles. This relationship could suggest that beach sediment characteristics
401 influence how plastics are retained and distributed within the sand column. This is in line with
402 Rodrigues *et al.* (2024) which showed that grain size was an important factor explaining
403 microplastic concentration on sandy beaches across the Azores archipelagos. This differential
404 retention capacity could explain why Kahuku, with a larger sand grain size, shows subsurface
405 plastic particle abundance. Nonetheless, the linear regression analysis revealed that sand grain
406 size accounts only for 18% of the variance in mean particle length, indicating that other factors
407 also contribute to variations in particle length.

408 Furthermore, the sand grain sizes at different beaches are also results of different beach
409 characteristics; for example, coarser sand grain sizes, such as at Kahuku, are typically a result
410 of high-energy beaches with steep slopes, while on sheltered beaches with low-energy
411 conditions, finer and smaller sand grain sizes are more likely to be present (Jaubet *et al.*, 2021;
412 McFall, 2019). Beach dynamics, the steepness of the beach, the exposure to waves, and the
413 wind are factors that are also likely to influence the plastic concentration on the surface and in
414 the sand column of the beach. Kahuku can be classified as a high-energy beach, in which the
415 beach face changes dramatically over short periods, with dynamic and large erosion and
416 accretion events which could also explain the low plastic particle concentration in the sand
417 column.

418 2.3. *Polymer ID, Plastic Class & Color*

419 From the FTIR/ATR analyses, we observed that polyethylene (PE) and polypropylene (PP)
420 were the dominant polymers recovered within the sand columns, accounting for 89% of the
421 total plastic particles analyzed. On average, 65% of the particles recovered were PE and 24%
422 were PP, as shown in Fig. 8(A). The other polymers identified were polystyrene (PS), polymers
423 composed of PP and PE mixtures (PP/PE), ethylene-vinyl acetate (EVA), and nylon, which all
424 together made up 10% of the total abundance of particles subjected to FTIR/ATR analyses. It
425 should be noted that only buoyant polymers were recovered using the BSD in the field to
426 separate the plastic from the sand. However, as referenced in Brignac *et al.*, more than 90% of
427 particles recovered on the windward side of O'ahu are less dense plastic materials and
428 predominantly buoyant plastics.

429 The Segmentation Model classified the majority (94%) of the total abundance of the
430 sampled plastic particles as hard fragments, as illustrated in Fig. 8(B). The next dominant class
431 was line, which accounted for 3% of the total abundance of recovered particles. The
432 Segmentation Model also depicted that the most prevalent colors were light blue (28%), light
433 grey (27%), and turquoise (19%), as observed in Fig. 8(C). Most of the particles collected were
434 off-white, which the Segmentation Model may have recognized as light blue and light grey.

435 The polymer ID, class, and color composition distribution per beach and along the sand
436 column is in the Fig. S15 in the SI, and it shows that the dominant plastic classes, colors, and
437 polymer types remained consistent with depth in the sand column and across different beaches,
438 showing no significant variation.

439 2.4. Reflection of Research Approach in Hawaiian Context

440 2.4.1. Research Methods and Data Collection

441 In conducting this research, we prioritized quantifiable metrics – such as plastic particle
442 abundance and distribution – and did not include cultural considerations. Such metrics can be
443 advantageous for refining models of plastic pollution in high-accumulation zones. At the same
444 time, the pursuit of this data risks causing unintended harm, particularly in Hawai‘i, where
445 cultural practices, worldviews, and the environment are deeply interconnected. One example
446 is the burial of iwi (ancestral bones) on the beach, which are considered sacred in Native
447 Hawaiian traditions. Disturbance of iwi is a violation of cultural and spiritual practices (Hall,
448 2010). Such violations are akin to desecrating a grave, representing a profound disrespect for
449 both ancestors and the living communities who care for them. It was through ongoing
450 conversations with cultural practitioner Kimeona Kāne that we began to recognize the potential
451 harm our research could inflict due to oversights in our methodology, highlighting the need to
452 engage with community and cultural practitioners at the outset of the research.

453 Furthermore, our focus on physical metrics and disconnect from the culture led us to
454 overlook how plastics might affect not only the environment but also the cultural practices tied
455 to the stewardship of the land and ocean. For instance, we did not account for the potential
456 impacts of plastics on culturally significant species like sand turtles (mole crabs) and other
457 intertidal species living in the sand column. These species, once abundant and used as fishing
458 bait, also play a role in maintaining ecological balance. In Hawaiian worldviews, including the
459 principle of mālama ‘āina (to care for and live in harmony with the land), there is an emphasis
460 on the interconnectedness of all life forms and the responsibility to maintain the balance
461 between people, the environment, and its living beings. For example, ensuring that sand turtles
462 remain undisturbed by pollution and continue to thrive is essential for this balance. Our study,
463 by not foregrounding the principle of mālama ‘āina, missed an opportunity to incorporate
464 culturally grounded understandings of ecological balance into our scientific framing. However,

465 as Ngata reminds us, that merely addressing the limitations of Western scientific paradigms
466 cannot be resolved with “more research” or “better research”—for instance, by conducting
467 additional studies on the plastic pollution impact on sand turtles. Indigenous communities have
468 long been hindered by colonial barriers that prevent them from implementing their own
469 knowledge. The solution may be as straightforward as stepping aside, breaking down colonial
470 barriers and letting Indigenous scientists and communities, who already possess vital
471 knowledge and solutions, to lead the way (Ngata & Liboirion, 2021). Ngata’s argument
472 emphasizes the necessity of letting Indigenous scientists to ask the questions that matter most
473 to their communities—questions such as who is harmed, who is displaced, and whose futures
474 are altered by the presence of plastics.

475 None of the initial researchers were from Hawai‘i or Oceania, and collaboration with a
476 cultural practitioner and Kanaka ‘Ōiwi leader began only toward the end of the study, who
477 played a transformative role in illuminating the broader ecological and cultural impacts of
478 plastic pollution, for example by highlighting the significance of beaches as cultural space and
479 on the significance of mālama ‘āina. As a result, the research questions, methods, and
480 interpretations were shaped without local cultural perspectives. As scholars such as O’Neil
481 (2016) have highlighted, the creation of data is always shaped by subjective choices: what to
482 measure, how to measure it, and which variables are considered important. Co-producing
483 research with the local communities where it is conducted not only has the potential to reduce
484 the risk of unintended harm but also enables the creation of knowledge that is more culturally
485 relevant and context specific. In this way, integrating community perspectives from the outset
486 of shaping the research inquiries leads to research that is both more representative and more
487 respectful of the place and people it aims to study and ultimately serve.

488 While we recognize that our initial approach did not engage communities from the
489 beginning, we view this as an opportunity for growth, improvement, and (un)learning. The
490 implications of not incorporating local perspectives carry significant socio-ecological costs for
491 Indigenous communities and, ultimately, for all of us (Stein *et al.* 2024). As also advocated by
492 Arshad-Ayaz *et al.* (2020), by naming our missteps and failures, we hope to interrupt ingrained
493 habits in our research practices and learn from them, rather than continue reproducing them. In
494 our commitment to advancing more ethical and accountable research in our field, we
495 acknowledge that our research should have been developed in collaboration with the
496 community in which it was conducted—Hawai‘i. We recognize that if plastic pollution research
497 aims to support the communities most affected, it should be co-produced or designed in
498 collaboration with those communities. Moving forward, we will strive to adopt more
499 accountable, reciprocal, and culturally aware research practices. We draw inspiration from
500 work carried out by scholars such as Alegado *et al.* (2021) and Winter *et al.* (2020), who have
501 contributed to rethinking research partnerships through the development of the Kūlana Noi‘i
502 framework, which encourages researchers to engage in deeper, reciprocal relationships with
503 communities. We also engage with the perspectives and insights of scholars like Liboiron and
504 Ngata, who advocate for anti-colonial research practices in plastic pollution research (Ngata &
505 Liboiron, 2021; Peryman, *et al.* 2024, Liboiron, 2021a; Liboiron, 2021b).

506 2.4.2. *Permitting for Field Work*

507 For the removal of plastics from the beaches at Waimānalo and Kokololio, no permits were
508 required. However, for Kahuku at James Campbell National Wildlife Refuge, a permit was
509 obtained due to its protected status. While our initial focus was on meeting regulatory
510 requirements, we now recognize the need for a more proactive approach that ensures our

511 research respects the cultural and ecological relationships of the sites studied. The Kapa‘akai
512 analysis, a legal framework used to assess the impact of proposed actions on Native Hawaiian
513 cultural resources, could serve as a model for integrating cultural considerations into
514 environmental research (University of Nations, 2020; McGuire & Mawyer 2023). Similarly,
515 tools like the Mai Ka Po Mai management plans can help researchers align their projects with
516 Indigenous cultural practices, leading to more respectful, place-based research (OHA, 2021;
517 Quioco, *et al.* 2023). This would represent a step toward more responsible, culturally informed
518 research practices in Hawai‘i. However, beyond adopting frameworks and as advocated by the
519 Kūlana Noi‘i it is important to engage with the community of the study area early and build
520 mutually respectful relationships on common goals with local communities from the beginning
521 of shaping the research inquiry (Alegado *et al.* 2021).

522 2.4.3. *Research as a source of waste and pollution*

523 While efforts were made to minimize waste and pollution generation, we acknowledge that
524 research activities can still produce waste. Equipment and materials used during fieldwork, such
525 as rubber mallets, sampling mesh bags, and plastic containers, can inadvertently contribute to
526 plastic waste at study sites through wear and tear, breakage, or shedding. For instance, the
527 rubber mallets occasionally broke pieces when hitting the metal frame into the sand with and
528 while visible fragments were collected, some pieces may have been unintentionally left behind.

529 3. Conclusion

530 This study provides the first insights into plastic concentration and distribution along the
531 sand columns of O‘ahu’s windward beaches, revealing that 91% of the total plastic abundance
532 is located below the surface layer, extending down to 1 m. These findings underscore the
533 necessity of considering subsurface plastic pollution when evaluating the extent of plastic

534 contamination in coastal zones. The majority of plastic particles identified throughout the sand
535 columns were small hard fragments, predominantly composed of polypropylene and
536 polyethylene. This substantial yet often invisible plastic load, characterized primarily by small
537 particles, indicates a more severe plastic pollution problem on the beaches of Hawai‘i. It further
538 emphasizes the challenges posed by these micro-sized pollutants in sandy environments and
539 highlights the urgent need for upstream remediation efforts to prevent plastic from reaching the
540 shores of Hawai‘i. We also observed higher plastic abundance in the 1-meter sand column
541 during winter months and lower concentrations during summer months, which requires further
542 investigation to better understand this variation. Furthermore, the potential positive correlation
543 between plastic particle length and sand grain size could further describe plastic particle
544 dependencies on beach characteristics. Ongoing research is focused on exploring the correlation
545 between plastic burial in the sand column and beach dynamics, with the goal of refining models
546 for plastic pollution transportation and accumulation. By deepening our understanding of
547 plastic quantities and fluxes in subsurface sand layers, this research aims to improve
548 assessments of ocean plastic pollution and the effectiveness of upstream mitigation strategies.

549 Our study also critically evaluates our methods and the cultural impacts of our work. By
550 prioritizing quantifiable metrics over cultural and ecological considerations, we risk harming
551 significant sites and cultural practices. This reflection underscores the importance of integrating
552 local perspectives and values into research methodologies from the outset of shaping research
553 inquiries. Such an approach not only reduces the potential for unintended harm but also leads
554 to a more holistic understanding of plastic pollution's impacts. It further allows for the creation
555 of sustainable solutions tailored to the unique needs of the ocean ecosystems and coastal
556 communities our research seeks to serve.

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751 **CRedit authorship contribution statement**

752 **Astrid E. Delorme:** Conceptualization, Formal analysis, Methodology, Investigation,
753 Resources, Writing – original draft, Visualization, Project administration, Funding acquisition.

754 **Olivier B. Poirion:** Segmentation Model Development, Writing – review & editing. **Laurent**
755 **Lebreton:** Methodology, Supervision, Writing – review & editing. **Pierre-Yves Le Gac:**
756 Writing – review & editing, Supervision, Project administration. **Kimeona Kāne:**
757 Conceptualization, Writing – review & editing. **Sarah-Jeanne Royer:** Methodology,
758 Investigation, Writing – review & editing, Supervision, Project administration, Funding
759 acquisition.

760 **Declaration of competing interest**

761 The authors declare that the research was conducted in the absence of any commercial or
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774 Early Career Ocean Professionals (ECOPs), and volunteers for sample collection assistance.

775 **Data Availability**

776 Data related to this manuscript will be made available on the Zenodo Community ‘Horizon-
777 Europe-STORAGE’ (<https://zenodo.org/communities/horizon-europe-storage/>) once
778 published.

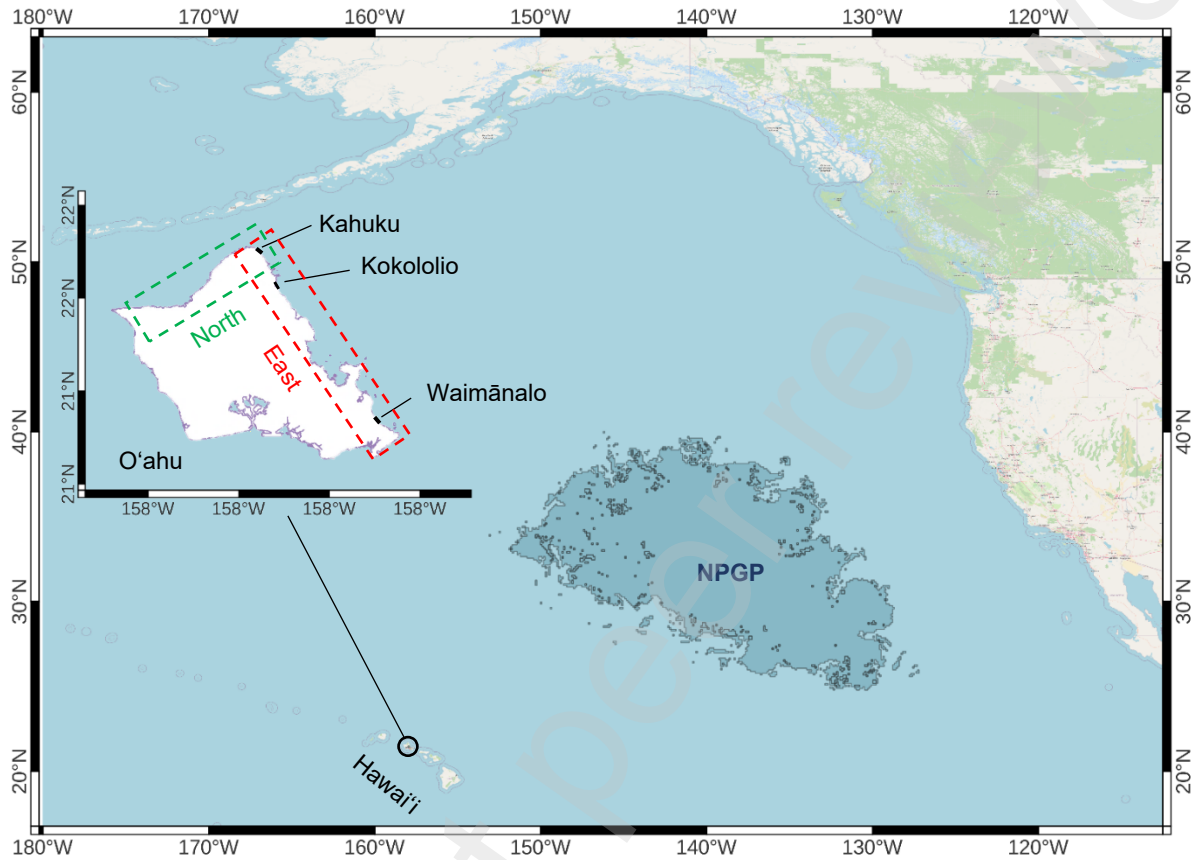


Fig. 1. Predicted location of the North Pacific Garbage Patch (NPGP) for June 2023, based on the numerical model presented in Lebreton *et al.* (2018), showing its position relative to the Hawaiian Islands. The main map highlights the NPGP's proximity to Hawai'i, with a focus on the central Pacific region. The inset map of O'ahu indicates the locations of Kahuku, Kokololio, and Waimānalo beach sites, and the East and North sides of the islands are highlighted in the red and green dashed boxes, respectively. Map created using QGIS version 3.34.9 (www.qgis.org).

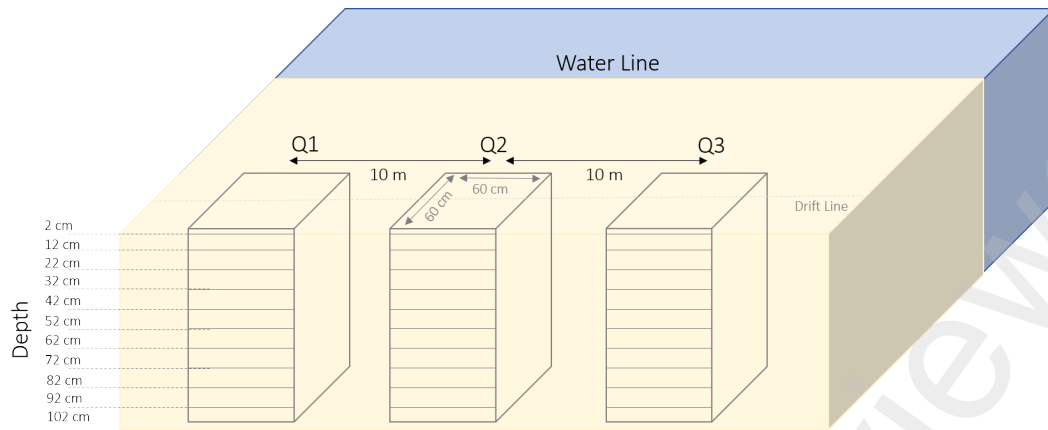


Fig. 2. Illustration of stratified sand column sampling for plastic debris at the drift line of the beach sites. Buoyant plastic debris was collected inside triplicate quadrats, Q1, Q2, and Q3, of the dimensions 60 cm × 60 cm, from stratified 11 layers with the first one being the surface layer of 2 cm depth, then 10 layers of 10 cm depth down to 102 cm.

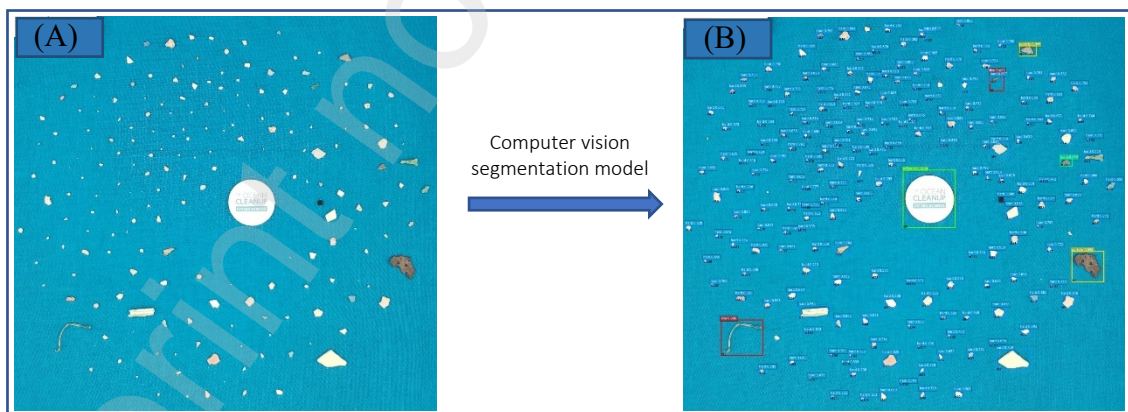


Fig. 3: (A) Image of plastic sample collected from quadrat 2 at Kokokolio in November 2022 from the sand column layer at a depth of 92-102 cm. The image is processed using the Segmentation Model developed by The Ocean Cleanup (Royer *et al.*, 2024), with (B) the output image of the model processing.



Fig. 4. The average concentration of plastic particles, expressed as counts (n) per dry weight (DW) of sand, along the sand column. The sand column is divided into 11 layers: the first 2 cm corresponds to the surface layer, followed by 10 cm strata extending down to 102 cm. The vertical distribution of plastic particle concentration is shown for the three beaches; Kahuku (red), Kokololio (green), and Waimānalo (blue), and across the six sampling months. The

average concentration and standard deviation error are calculated from the three quadrats sampled per sampling day.

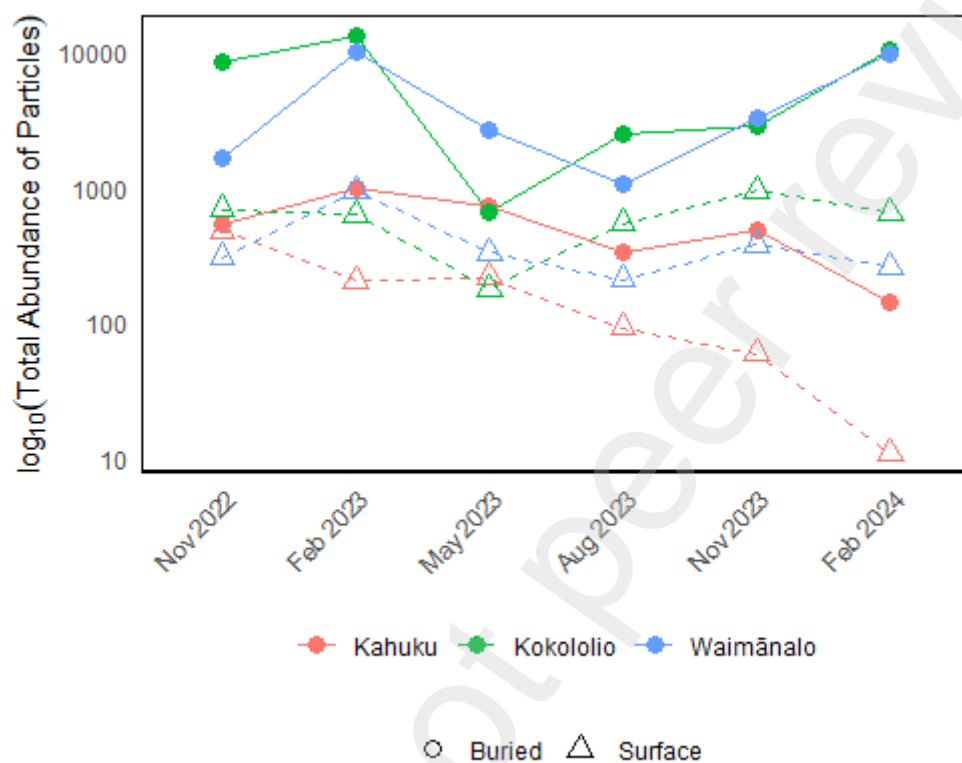


Fig. 5. Variation in the observed total abundance of plastic particles in the surface layer (Δ), at 0-2 cm, and buried in the sand column (\circ), at 2-102 cm, at all three beaches: Kahuku (red), Kokololio (green), and Waimānalo (blue).

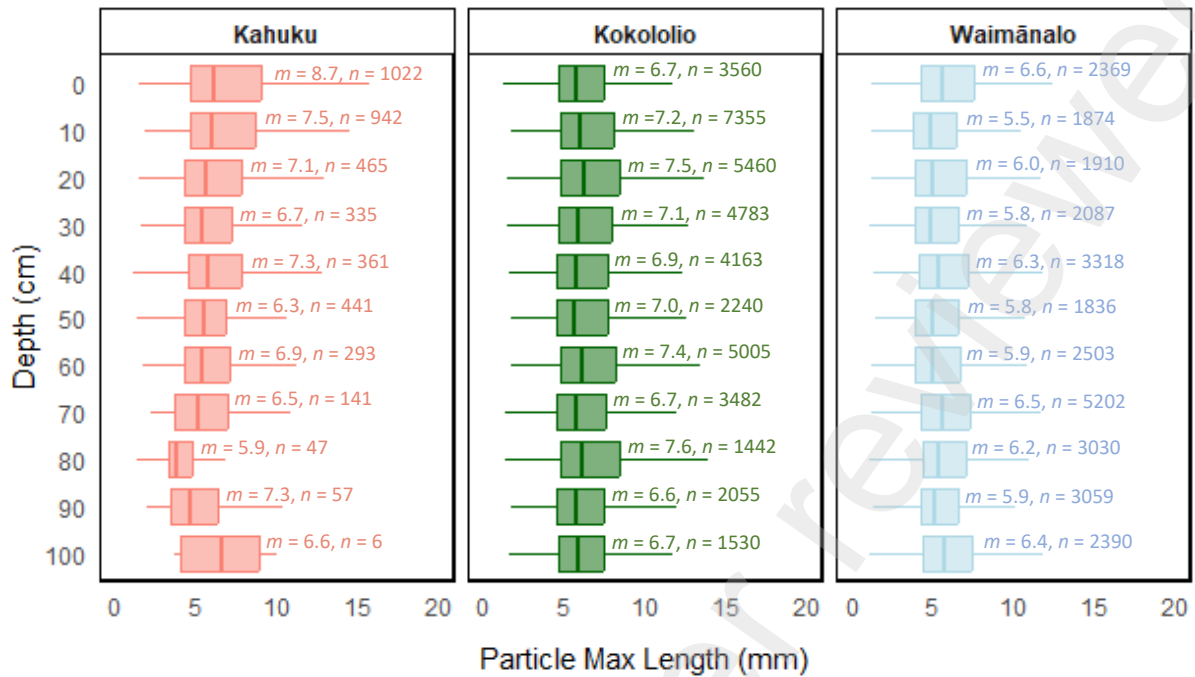


Fig. 6. Boxplots represent the distribution of maximum particle lengths in mm showing the mean (m) max. length (in mm) and counts of plastic particles per layer in the sand column (n), excluding the lines, collected at the three beach sites; Kahuku (red), Kokololio (green), and Waimānalo (blue) and across the six sampling months. Outliers have been removed for clarity.

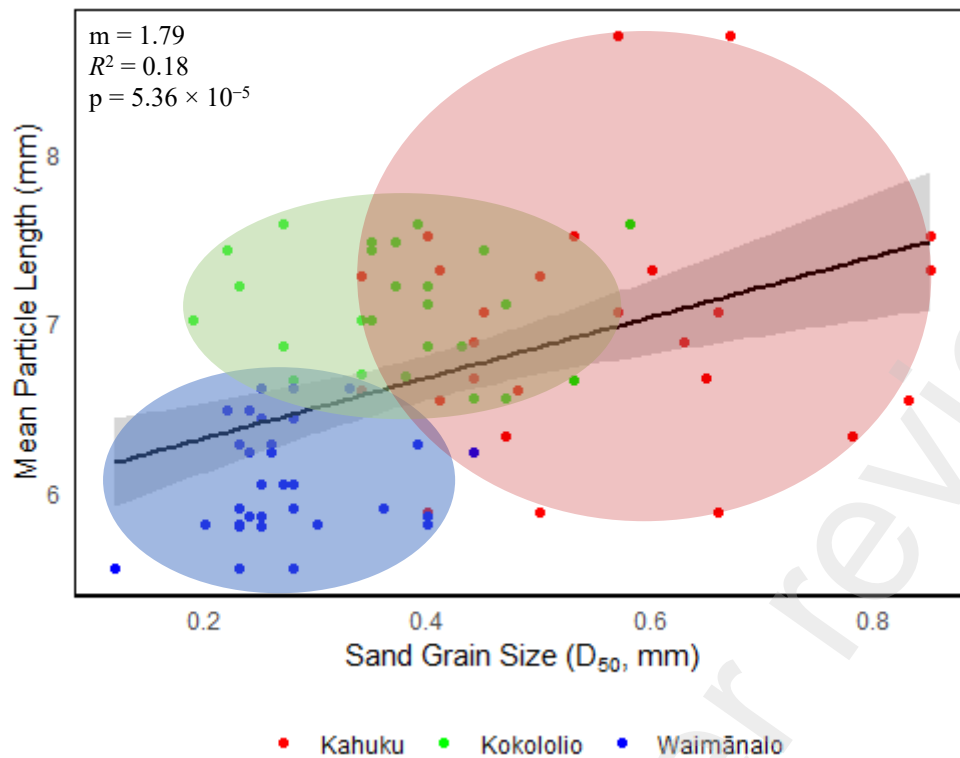


Fig. 7. Potential correlation between sand grain size and the mean particle lengths of sampled plastic particles at each depth strata in the sand column, with sand grain size measured using SandSnap at each layer in the sand column. The solid black line represents the regression line prediction, and the confidence interval is in the shaded area. The plastic particles with the sand grain size measured at the different beaches are highlighted in shaded circles by color: Kahuku (red), Kokololio (green), and Waimānalo (blue).

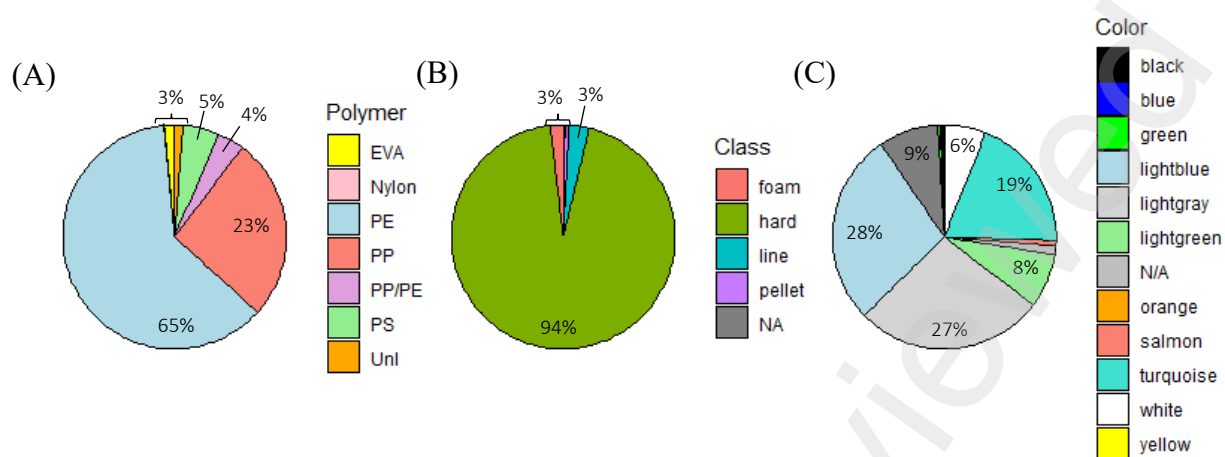


Fig. 8. Pie charts representing (A) the relative abundance of plastic classes, (B) the distribution of plastic colors, and (C) the composition of polymer types across Kahuku, Kokololio, and Waimānalo. The class and color compositions reflect the relative abundance of the total abundance of plastic particles sampled during field campaigns, as analyzed using the Segmentation Model. Polymer compositions were determined from ATR/FTIR analysis of the randomly selected aliquots of plastic particles from each sample collected.