



Discovery of the breeding colonies of a critically endangered and elusive seabird, the Mascarene Petrel (*Pseudobulweria aterrima*)

Descubrimiento de colonias de anidación de un ave marina elusiva y críticamente amenazada, *Pseudobulweria aterrima*

[Claire-Cécile Juhasz](#)¹, [Jérôme Dubos](#)¹, [Patrick Pinet](#)², [Yahaïa Soulaïmana Mattoir](#)^{1,3}, [Patxi Souharce](#)⁴, [Christophe Caumes](#)⁵, [Martin Riethmuller](#)³, [Fabien Jan](#)³ and [Matthieu Le Corre](#)¹

ABSTRACT. Seabirds are the most endangered group of birds and among them, the gadfly petrels (genera *Pseudobulweria* and *Pterodroma*) are the most threatened and least known. The Mascarene Petrel (*Pseudobulweria aterrima*) is endemic to Réunion Island and is one of the rarest birds in the world. This species was considered extinct in the mid-20th century but was rediscovered in 1970. The population is thought to be in decline because of predation by invasive predators, habitat destruction, and light-induced mortality. The first goal of this paper is to detail the methods that we used to discover the breeding colonies of this species and to determine the threats at these sites. The second goal is to present characteristics of the colonies we found, the threats occurring at these colonies, and the first conservation actions implemented at these sites. We first conducted an island-scale acoustic survey using autonomous recording units (ARUs) to locate the breeding colonies. We then used infrared thermal binoculars to precisely locate the places where birds displayed and landed. Because all discovered breeding sites were on vertical cliffs, we abseiled these cliffs to access the nests. Once burrows were discovered, we deployed infrared camera traps to determine the presence of alien predators or competitors (rats, cats, tenrecs) and to study the behaviors of the birds at the colony. The large-scale acoustic survey revealed the presence of 17 vocally active sites, 16 of which were investigated with infrared thermal binoculars. We observed petrel landings at five of these sites. Two of them were accessible and we abseiled to find the nests. We found 14 occupied burrows at one of these sites and eight at the other. Camera traps revealed the presence of rats and tenrecs at both sites, and cats were detected close to the colonies. The two colonies are on tall vertical cliffs covered with native vegetation, at elevations of 650 m a. s. l. and 1250 m a. s. l., respectively. These findings allowed us to implement conservation actions, such as invasive mammal control, and to start long-term monitoring and applied research for conservation. We are confident that the methods we developed could be used with great success at other places where finding colonies of a cryptic, rare, and nocturnal seabird is particularly challenging.

RESUMEN. Las aves marinas son el grupo de aves más amenazado y entre ellas los géneros *Pseudobulweria* y *Pterodroma* son los más amenazados y poco conocidos. *Pseudobulweria aterrima* es endémica de la isla de Reunión y es una de las aves más raras del mundo. Esta especie fue considerada extinta a mediados del siglo 20 pero fue redescubierto en 1970. Se supone que la población está en disminución debido a la depredación por especies invasoras, destrucción del hábitat y la mortalidad inducida por la luz. El primer objetivo de este estudio es detallar los métodos que utilizamos para descubrir las colonias de reproducción de esta especie y determinar las amenazas en estos sitios. El segundo objetivo es presentar las características de las colonias encontradas, las amenazas ocurriendo en estas colonias y las primeras acciones de conservación implementadas en estos sitios. Primero realizamos en toda la isla un monitoreo acústico utilizando Unidades de Grabación Automática (ARU) para ubicar las colonias. Luego, utilizamos binoculares térmicos infrarrojos para ubicar con precisión los lugares donde las aves realizaron despliegues y aterrizaje. Dado que todos los sitios de reproducción descubiertos estuvieron en acantilados verticales, accedimos a los nidos por medio de rapel. Una vez que las madrigueras eran descubiertas, desplegamos cámaras trampa infrarrojas para determinar la presencia de depredadores no nativos o competidores (ratas, gatos, tenrecs) y para estudiar los comportamientos de las aves en la colonia. El monitoreo acústico de gran escala reveló la presencia de 17 sitios vocalmente activos. Entre estos, 16 fueron investigados con binoculares térmicos infrarrojos. Observamos aterrizajes de petreles en 5 de estos sitios. Dos de estos fueron accesibles y los nidos encontrados utilizando técnicas de rapel. Encontramos 14 madrigueras ocupadas en uno de estos sitios y 8 en el otro. Las cámaras trampa revelaron la presencia de ratas y tenrecs en los dos sitios y los gatos fueron detectados cerca de las colonias. Las dos colonias se encuentran en grandes acantilados verticales cubiertos por vegetación nativa a una elevación de 650 msnm y 125 msnm respectivamente. Estos resultados nos permitieron implementar acciones de conservación como el control de mamíferos invasores y comenzar un monitoreo a largo plazo e investigación aplicada para la conservación. Estamos seguros que los métodos que desarrollamos pueden ser utilizados con una alta probabilidad de éxito en otros lugares donde el encuentro de colonias de aves marinas crípticas, raras y nocturnas sea particularmente retador.

Key Words: *autonomous recording unit; conservation; endemic seabird; infrared thermal binoculars; Pseudobulweria aterrima; tropical island*

INTRODUCTION

Seabirds are the most endangered group of birds (Croxall et al. 2012, Dias et al. 2019). More than 30% of the 359 seabird species are considered globally endangered (Dias et al. 2019). A global study investigating population trends of seabirds worldwide has shown a decline of 69.7% in monitored colonies between 1950 and 2010 (Paleczny et al. 2015). This decline is due to numerous man-related threats occurring at sea and on land. The threats affecting most seabird species are predation by invasive alien species at breeding colonies, fishery bycatches, overfishing, hunting/trapping, and climate change/severe weather (Dias et al. 2019). Among seabirds, pelagic species such as albatrosses and petrels are the most endangered (Paleczny et al. 2015, Dias et al. 2019).

The gadfly petrels (genera *Pseudobulweria* and *Pterodroma*) constitute one of the most threatened and least known groups of seabirds. Gadfly petrels are threatened by invasive alien species at breeding colonies and light pollution (Rodríguez et al. 2017, Dias et al. 2019). Eradication or control of invasive alien species is one of the major priorities in seabird conservation (Dias et al. 2019), but represents a huge challenge, particularly when several alien species are present simultaneously. For instance, the cumulative effects of feral cat (*Felis catus*) predation on adults and rat (*Rattus* spp.) predation on eggs and chicks result in rapid population declines (Dias et al. 2019, Rodríguez et al. 2019). The mass mortality of fledglings caused by artificial lights is a major issue in urbanized islands and archipelagos such as the Canaries, Hawaii, and Réunion Island (Rodríguez et al. 2017). Reduction of light pollution combined with rescue programs of grounded birds are currently the only solutions to significantly reduce this mortality (Troy et al. 2013, Gineste et al. 2017, Rodríguez et al. 2017, 2019).

Among gadfly petrels, the genus *Pseudobulweria* includes the most threatened seabird species in the world (International Union for Conservation of Nature [IUCN] 2021). It includes five species, of which one is extinct (*P. rupinarum*), three are on the verge of extinction (the critically endangered *P. becki*, *P. macgillivrayi*, and *P. aterrima*), and one is near threatened (*P. rostrata*; Warham 1996, Croxall et al. 2012, IUCN 2021). These species are distributed in the tropical Indian and Pacific oceans (Gangloff et al. 2012). The three critically endangered species are endemic to one island or a group of islands (Gangloff et al. 2012). Because of their rarity and very secretive way of life (nocturnal behaviors, burrow nesting on remote and inaccessible cliffs), the biology of these birds is extremely poorly known. Consequently, the breeding sites of all of these species except the Tahiti Petrel (*P. rostrata*) are still unknown.

The Mascarene Petrel (*P. aterrima*) is one of the rarest, least known, and most endangered birds of the world (Jouanin 1987, Warham 1990, Le Corre et al. 2003, IUCN 2021). This critically endangered petrel is endemic to Réunion Island and is the only representative of the *Pseudobulweria* genus in the Indian Ocean. Although eight birds were collected in 1880 and 1891 (Jouanin 1970), the species was believed to be extinct by the mid-20th century (Jouanin 1987). It was rediscovered in 1970 with a dead grounded bird. Two more grounded Mascarene Petrels were found (dead or dying) in 1973 and 1995 (Jouanin 1987, Attie et al. 1997). In 1996, Le Corre et al. (2002, 2003) discovered the

extent of the impact of urban light on the mortality of fledgling petrels and shearwaters at Réunion Island. This very important mortality is due to the fact that most petrel and shearwater colonies at Réunion Island are located inland and most cities are coastal, so birds fly over illuminated areas each time they commute from their colonies to the sea (Le Corre et al. 2002, 2003, Chevillon et al. 2022).

Since that time, a rescue operation based on large-scale public awareness campaigns has been made annually by a local non-governmental organization (the Société d'Études Ornithologiques de La Réunion). These campaigns have resulted in the finding of 58 Mascarene Petrels between 1996 and 2021, of which 49 were released successfully (Chevillon et al. 2022). A first estimate of the population size on the basis of sightings at sea between 1978 and 1995 indicated a range of 45 to 400 pairs, with an average estimate of 250 breeding pairs (Attie et al. 1997). The population is believed to have declined because of predation by rats and cats at breeding colonies, habitat destruction (Virion et al. 2020), and light-induced mortality (Le Corre et al. 2003, Chevillon et al. 2022). However, no colony was known until 2016 (this study).

The first objective of this article is to present the different methods used to locate the breeding colonies of the Mascarene Petrel, because we are confident that these methods could be used with success at other places where finding a cryptic, rare, and nocturnal seabird is particularly challenging. Our second objective is to present, for the first time for this species, the main characteristics of its colonies, the threats that are occurring there, and the first conservation measures that we implemented.

METHODS

Automated acoustic survey and record analysis

From July 2015 to April 2017, 21 autonomous recording units (ARUs; six SM4, 11 SM3, and four SM2+ from Wildlife Acoustics, Inc.) were deployed at 68 locations. Recording sites were chosen on the basis of previous acoustic surveys conducted by Riethmuller et al. (2012) and the location of grounded birds attracted to artificial lights (Le Corre et al. 2003, Virion et al. 2020, Chevillon et al. 2022). ARUs were tied to trees within 300 meters of the targeted cliffs. During the first season of searching (July 2015 to April 2016), the ARUs were programmed to record from 8 PM to 5 AM (i.e., from 1h00–2h00 after sunset to 0h30–2h00 before sunrise). During the second season (July 2016 to April 2017), the ARUs were programmed to record from 2 AM to 4 AM (i.e., from 3h30–5h00 before sunrise to 1h30–3h00 before sunrise) on the basis of knowledge acquired during the previous season.

In order to describe temporal changes in vocal activity at the scale of a year and at the scale of a night, we placed an ARU at a site (named Malabar; MB) where Mascarene Petrels were acoustically detected in 2014 (M. Le Corre, P. Pinet, M. Riethmuller, and J. Dubos, *unpublished data*; see also Results). This ARU was deployed from July 2015 to January 2016 and from July 2016 to April 2017, with the same settings as previously described. First, we calculated the mean number of calls per hour from 1 November 2015 to 12 December 2015. These dates were selected because previous observations suggested that they correspond to a period of intense vocal activity (Riethmuller et al. 2012). This allowed us to determine the time of night when birds are most active

vocally. Second, we described the annual pattern of changes of vocal activity during this part of the night using all records obtained on this site.

ARUs were set at 16 kHz level (8 kHz for each microphone) with high pass filter at 220 Hz, because the fundamental frequency of Mascarene Petrel calls is approximately 3 kHz (Riethmuller, *personal communication*). ARUs used for acoustic surveys were deployed on average for 54 days (minimum: 3; maximum: 213) at a given location, except the ARU at the MB site. Batteries and SD cards were changed at least every two months. The direction of each microphone was recorded and used to determine the direction of the vocalizations. Recordings were retrieved in WAV format. We developed a recognizer with the Kaleidoscope Pro software (Wildlife Acoustics, version 4.5.4) to automatically detect the calls. Each record with positive automated detections was subsequently manually analyzed to determine the direction of the calls and to count the calls.

Nocturnal observations with infrared thermal binoculars

We used infrared thermal binoculars (“Matis” type, Safran) with a video recorder (Espionner Angel Eye Mini DVR 2.5”) to detect petrels at night. Binoculars were used in December 2015 (three weeks), from February to May 2016 (15 weeks), and from October 2016 to March 2017 (24 weeks). Nocturnal observations were made at places where acoustic surveys had revealed the presence of a colony. We chose vantage points with a large view of the cliffs where calls were previously recorded.

For each bird detected, observers noted the species, flight direction, and behaviors (commuting, prospecting, or landing). Birds were observed until they left the vision field or landed. When a bird landed, suggesting the presence of a burrow, we located the landing place precisely on a picture of the site using visual landmarks (tree, rocks, etc.). Identification of the species was made on the basis of flight pattern and vocalizations. The only other nocturnal seabird known to be present at the places where we prospected is the Tropical Shearwater (*Puffinus bailloni*). Compared to the Mascarene Petrel, the Tropical Shearwater is smaller and has a very rapid flight with almost continuous wing flapping and short gliding. The two other nocturnal seabirds (Barau’s Petrel [*Pterodroma barau*] and Wedge-tailed Shearwater [*Ardenna pacifica*]) have never been recorded on the sites that we prospected. The Barau’s Petrel breeds higher in the mountains (Probst et al. 2000) and the Wedge-tailed Shearwater breeds mostly on coastal cliffs (Faulquier et al. 2017).

Field expeditions and description of the colonies

On the basis of the combined results of acoustic surveys and nocturnal observations, we selected sites with high acoustic activity and observations of landing birds. All of these sites were on vertical cliffs; thus, accessing the spots where landing birds were observed required abseiling. On one of these sites field workers were lowered by helicopter.

Once discovered, the colonies were carefully prospected and all nests were labeled, geo-localized, checked, and measured (height, width of the entrance, and minimum depth to the nearest 5 mm), and their contents noted. In order to start a long-term population monitoring, all accessible birds were banded, weighed (to the nearest 5 g), and measured (wing chord to the nearest mm; culmen

length, bill depth, crochet length, and tarsus length to the nearest 0.1 mm). We also described elevation, slope, vegetation type, vegetation structure, and orientation of the cliff of each colony.

Rapid assessment of the threats and conservation actions implemented at breeding colonies

In order to detect invasive predators (cats, rats) or burrow competitors such as the tenrec (*Tenrec ecaudatus*), we deployed five camera traps (Bushnell Trophy Cam Wireless HD 119598) at the first colony discovered. The tenrec is a hedgehog-like mammal endemic to Madagascar and introduced to Réunion Island. It frequently uses natural cavities and burrows of other species, including seabirds (M. Le Corre and J. Dubos, *personal observations*). Camera traps were configured with a night-only, motion-trigger sensor recording mode, with 30 seconds of video and sound recording for each triggering.

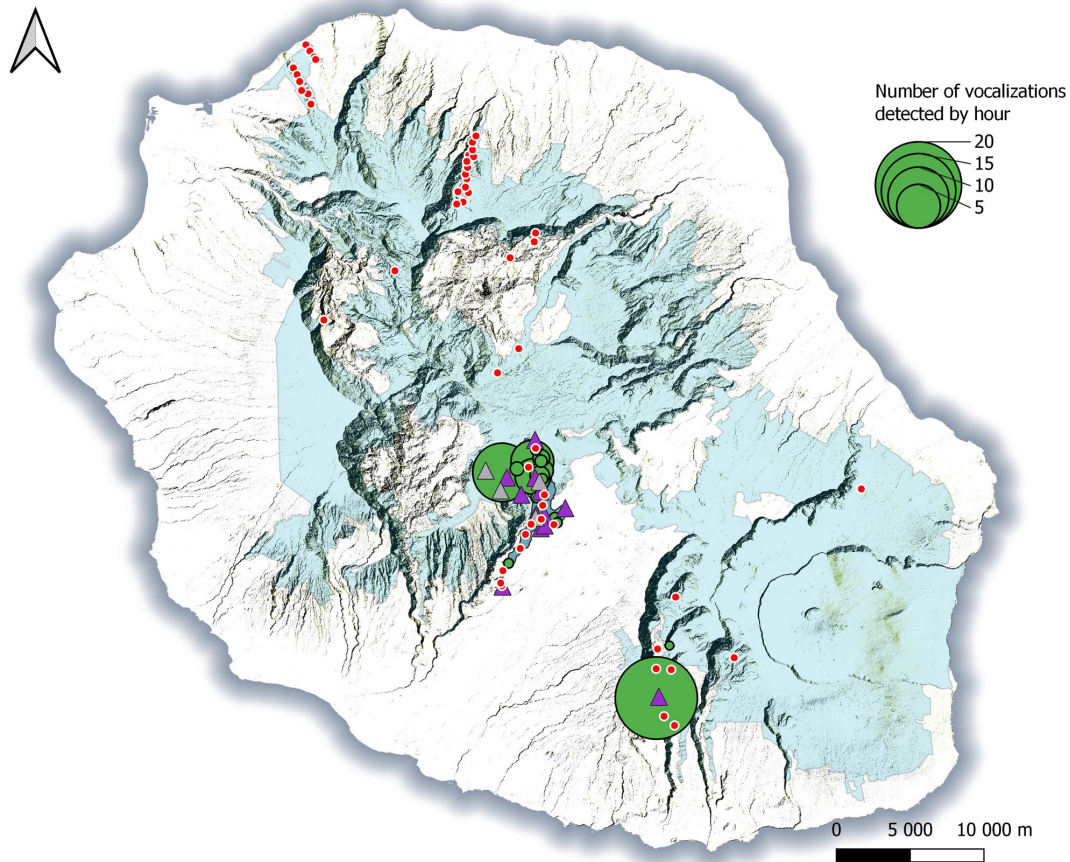
Because rats and cats were suspected to be present at the colony or in its immediate vicinity, we implemented active rat and cat control as soon as the colony was discovered. At the first colony, rat control was carried out by dispersing blocks of poison containing bromadiolone 0.005% (n°CAS 28772-56-7). To do so, 25 bait stations were set with two blocks of poison (30 g each) per station, 20 of which were along the path to access the colony (one bait station every 30 m) and five were at the colony. Bait consumption at each station was monitored at each visit to the colony and a consumption rate was calculated by dividing the number of consumed blocks by the total number of spread blocks. Consumed blocks were replaced on a regular basis. In total, 2.2 kg of bait were spread at the bait stations. In addition, 17 kg of rat poison was manually spread in the most inaccessible parts of the cliff (12 kg) and at the colony (5 kg). Two automated, rechargeable percussion traps (A24, Goodnature) equipped with a percussion counter were set at the colony and baited with peanut butter and chocolate. These two traps were deployed at places unreachable by a petrel. The number of percussions was checked during the last visit at the colony.

Indices of the presence of cats (e.g., scats) were observed near the first colony discovered. All scats were removed and cat control was carried out with two cages (Havahart x-large 1-door trap), baited with oiled sardines and deployed on access tracks. Cages were automatically monitored with wireless camera traps (Bushnell Trophy Cam Wireless HD 119598) configured with a motion-trigger sensor. A group of three pictures was recorded for each triggering. Traps were checked remotely and automatically every morning through a picture sent to cell phones. Because of the late discovery of the second breeding colony (see Results) and the presence of rats being confirmed with infrared thermal binoculars, we manually dispersed 10 kg of rat poison at this colony and in its vicinity.

Other methods that were rapidly abandoned

We deployed VHF beeper tags (Biotrack, 3.5 g) on adults stranded in towns as a consequence of light pollution. The tags were attached to three tail feathers with nylon thread and were set with a beep every two seconds. Four telemetry receivers (Lotek SRX800) were installed in Grand Bassin in order to detect the presence of the equipped birds. Because no birds were detected and other methods resulted in the finding of two colonies, we rapidly abandoned this method.

Fig. 1. Locations of autonomous recording units (ARUs; circles) used for acoustic prospection and sites of nocturnal visual inspections with infrared thermal binoculars (triangles), in 2015–2017, Réunion Island. Green and red circles represent sites with and without Mascarene Petrel (*Pseudobulweria aterrima*) calls recorded, respectively. The size of the green circles is proportionate to the vocalization rate during a month of records. Purple and gray triangles show sites with and without Mascarene Petrel observations, respectively. The blue area indicates the National Park of Réunion Island.



We also used two trained small-sized sniffer dogs (a dachshund and a mongrel dog, both about 10 kg and under three years old) to look for burrows at targeted sites. The dogs were trained with scents of Barau's and Mascarene petrels (linen impregnated with the smell of stranded birds). The dachshund was particularly lazy and inefficient, and the mongrel, though very active and determined, was totally inefficient in the vertical cliffs where Mascarene petrels were suspected to be. For these reasons, and because we found the colonies using the other methods, we rapidly abandoned this method.

RESULTS

Automated acoustic survey

From July 2015 to April 2017, ARUs were deployed at 68 sites (18 in 2015–2016 and 50 in 2016–2017), representing 14,903.5 h of recordings, of which 75.2% were analyzed (see details in Table 1 and Appendices 1–2). Mascarene Petrels were detected at 17 of these sites (25%; Fig. 1). Most sites with vocal activities ($n = 15$) were located at Grand Bassin, a remote and deep canyon with

almost no human settlement. Two other sites with vocal activities were located at Rivière des Remparts, another remote area with huge vertical cliffs and almost no habitation. On average, the vocal activity was low (mean \pm sd = 0.66 ± 2.03 voc/h; range [min; max] = 0.003; 8.410 voc/h) suggesting that most of these sites had small numbers of birds (Fig. 1). However, the vocal activity was greater at two sites (> 1 voc/h): Rein de Dimitile (RD) and Le Trou (see Appendix 3).

Temporal changes of vocal activity at Malabar site

We recorded 1290.75 h at the MB site (966.75 h in 2015–2016 and 324.00 h in 2016–2017; see Appendix 1B). In November and December 2015, the time of the night with the highest vocal activity was between 2 AM and 3 AM ($26.5 \pm 1.5\%$ of vocalizations; Fig. 2). At the scale of a year, the maximum vocal activity was observed in December and January (Fig. 3).

Infrared thermal binoculars survey

We used infrared thermal binoculars at 16 sites where Mascarene Petrels were detected during the acoustic surveys. This represented

Table 1. Characteristics of the automated acoustic survey conducted at Réunion Island for the detection of Mascarene Petrel (*Pseudobulweria aterrima*) breeding sites.

Breeding season	2015–2016	2016–2017	Total
Sampling period	1 Jul 2015–30 Apr 2016	1 Jul 2016–28 Apr 2017	-
Total recording duration (h)	9213.5	5690.0	14,903.5
Analyzed recordings (h)	5516.5 (59.87%)	5690.0 (100%)	11,206.5
Number of surveyed sites	18	50 (+ 3 same as 2015–2016)	68
Number of sites where Mascarene Petrels were detected	7	10 (+ 3 same as 2015–2016)	17

Fig. 2. Temporal change in vocal activity of Mascarene Petrel (*Pseudobulweria aterrima*) at Malabar, in November and December 2015, Réunion Island. Vocal activity is the proportion of calls recorded per hour after sunset (out of the total number of calls recorded during the period).

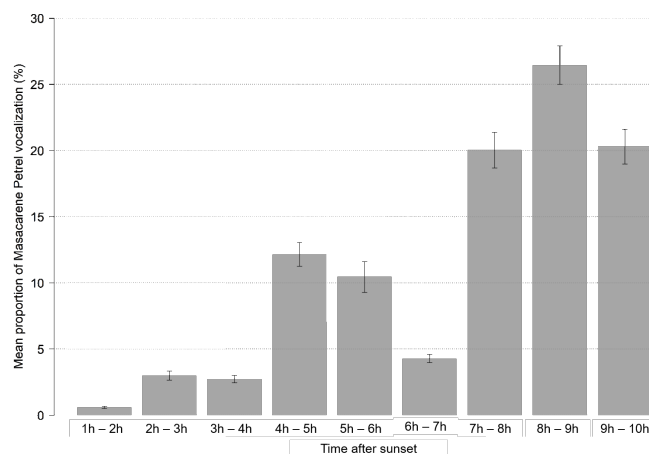
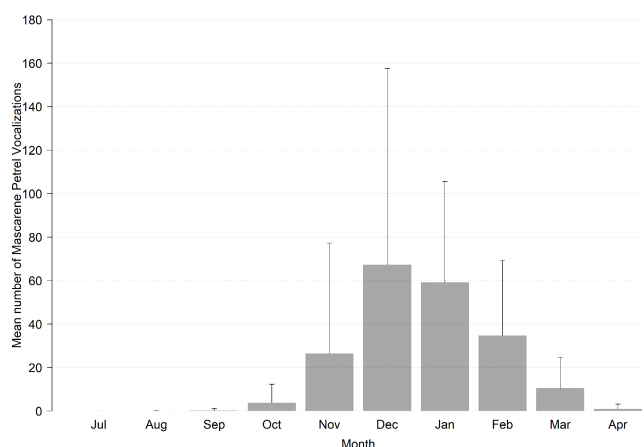


Fig. 3. Monthly variations in the mean number of calls per hour of Mascarene Petrel (*Pseudobulweria aterrima*) at Malabar (manual analysis) during the peak of daily vocal activity (eight to nine hours after sunset), from July 2015 to April 2017, Réunion Island.



50 sessions totaling 205 h of observations. All sessions took place during the breeding season (see details in Table 2; Fig. 1). At five sites no Mascarene Petrels were observed. In the remaining 11 sites, we made 94 observations of Mascarene Petrels, all of them in flight (Appendix 4). On the basis of these observations, we characterized three types of flight: (1) commuting flight, with the bird flapping almost continuously and flying in a straight line (Appendix 4A); (2) prospecting flight, with the bird flapping, gliding, and circling close to cliffs (Appendix 4B and C); and (3) landing flight, with the bird circling above an area until it landed (Appendix 4D).

Table 2. Characteristics of the infrared thermal binocular survey conducted at Réunion Island for the detection of Mascarene Petrel (*Pseudobulweria aterrima*) breeding sites.

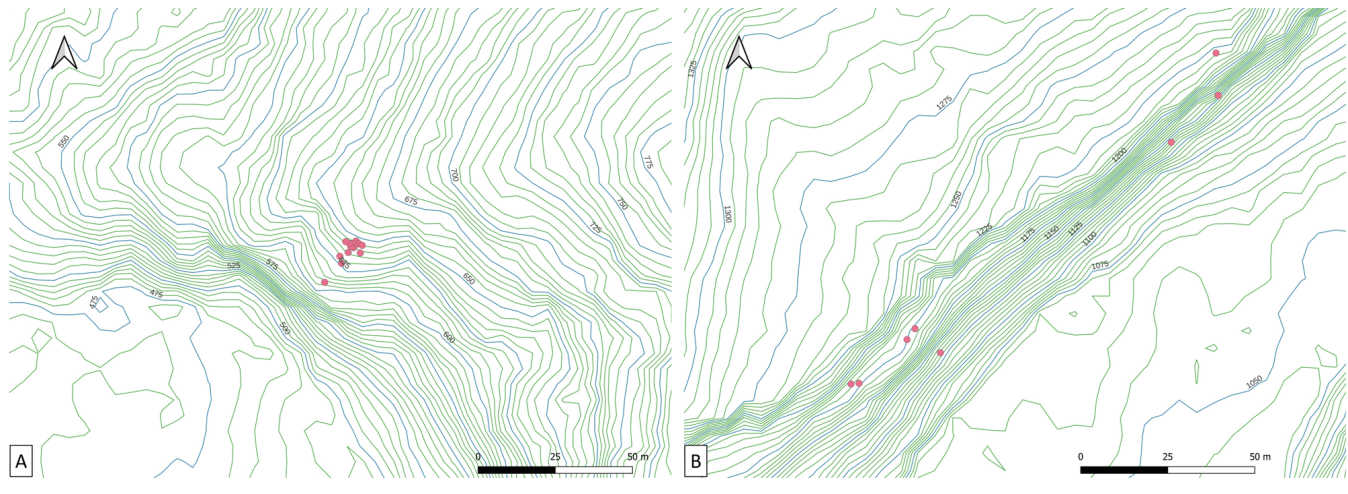
Breeding season	2015–2016	2016–2017	Total
Number of infrared thermal binocular sessions	18	32	50
Observations effort (h)	102	103	205
Number of surveyed sites	6	14	16
			(4 sites were surveyed during both seasons)
Sites with Mascarene Petrels	4	9	11
			(2 sites were surveyed during both seasons)
Sites with landing Mascarene Petrels	0	5	5

We observed landing petrels at five sites. On three of these sites, the number of observations was extremely low despite considerable observation effort, suggesting that very few, if any, birds bred there (Appendix 5). On the two other sites, 12 and 17 birds, respectively, were observed landing, suggesting a more active colony (Appendix 5). These two sites were considered accessible and were selected for cliff surveying by abseiling.

Discovery of the breeding colonies

An initial climbing expedition of three people was conducted on 15 November 2016 at RD to reach the landing area. A team placed at the base of the cliff guided the climbers with walkie-talkies and binoculars to help them to approach the landing area. The colony was found after three hours of searching and abseiling. After this discovery, two expeditions were conducted in December 2016 and February 2017 to complete the census and to monitor the colony. Fourteen burrows were found and 13 birds were banded and measured (Appendix 6). Two burrows contained an egg and one

Fig. 4. Topographical details of Mascarene Petrel (*Pseudobulweria aterrima*) breeding colonies located at Rein de Dimitile (A) and Rond des Chevrans (B). The red dots correspond to the location of the burrows.



contained a downy chick (wing chord = 31.3 mm; weight = 275 g) during the first visit in November. These three burrows were found empty in December, suggesting a breeding failure, presumably because of rat predation.

The second colony was discovered at Rond des Chevrans (RDC) on 28 February 2017. Four people were first lowered to the cliff by helicopter. They reached the landing area by abseiling 40 m down the cliff. Eight burrows were found containing nine adult birds. Seven of them were caught, banded, and measured (Appendix 6), although two were inaccessible. No chicks or eggs were found during this expedition. Because this discovery occurred toward the end of the breeding season, no other expedition was organized during that season.

Characteristics of the colonies and burrows

The breeding colony of RD is located on the lower part of a tall cliff, 690 m high, facing west. The colony is at an elevation of about 640 m a. s. l. and 100–150 m above a riverbed (Fig. 4A). The colony contains 14 burrows in an area of 800 m² (burrow density: 0.0175 burrow/m²; Table 3). The minimum and maximum distance between burrows is 0.65 m and 40 m, respectively.

The breeding colony at RDC is located on the lower part of a tall cliff, 1050 m high, facing southeast. The colony is at an elevation of 1200 m and 100–150 m above a riverbed (Fig. 4B). Eight burrows are spread into four patches of one to four burrows each, with a total surface of 10,000 m² (burrow density: 0.0008 burrow/m²; Table 3). The minimum and maximum distance between burrows is 0.5 m and 140 m, respectively.

For the two colonies, the habitat is characterized by young indigenous sub-humid forest of the megatherm hygrophilous stage (Cadet 1977), dominated by *Olea lancea* and the endemic *Monimia rotundifolia* (Huré 2019). Burrows are dug in humus, under tree roots or rocks. They typically have one to three entries and a single incubation chamber. The mean dimensions of the burrows are presented in Table 3.

Table 3. Characteristics of the two Mascarene Petrel (*Pseudobulweria aterrima*) colonies discovered in 2016–2017, Réunion Island.

Colony	Rein de Dimitile (RD)	Rond des Chevrans (RDC)
Elevation (m)	640	1200
Elevation from the riverbed (m)	100–150	100–150
Slope of the cliff	50° to 90°	50° to 90°
Vegetation height	Maximum 7 m	Maximum 7 m
Vegetation type	Hygrophile forest species of medium and high altitude, dominated by <i>Olea lancea</i> and <i>Monimia rotundifolia</i>	
Habitat structure	Continuous forest ridge	Patches of vegetation separated by rocky surfaces
Colony orientation	West	Southeast
Numbers of burrows found at the time of discovery	14	8
Mean width of burrow entrance (cm ± sd)	17.7 ± 4.5 (n = 12)	15.5 ± 1.9 (n = 4)
Mean height of burrow entrance (cm ± sd)	12.8 ± 3.4 (n = 12)	14.3 ± 4.2 (n = 4)
Mean depth of burrows (cm ± sd)	100.0 ± 38.1 (n = 11)	—
Distribution of the burrows	One single patch	4 small patches containing 1 to 4 burrows each
Substrate where burrows are dug	Humus under rocks or roots	Humus under rocks or roots

Rapid assessment of the threats

The two breeding colonies presented a low level of invasion by exotic plants. However, 15 exotic species were identified at the colonies or in their immediate vicinity, of which five are potentially invasive and worrying: the west Indian lantana (*Lantana camara*), the Brazilian pepper tree (*Schinus terebinthifolia*), the Cattley guava (*Psidium cattleianum*), the

coastal she-oak (*Casuarina equisetifolia*), and the giant cabuya (*Furcraea foetida*).

Camera traps at the RD colony recorded 616 videos between November 2016 and January 2017. Mascarene Petrels were the main species observed on videos (95.3%, $n = 587$), followed by rats (3.2%, $n = 20$; Appendix 7A and 8), tenrecs (1.3%, $n = 8$; Appendix 7B), and shrews (*Suncus murinus*; 0.2%, $n = 1$). The rats were observed foraging with no sign of interference with petrels. Tenrecs were observed entering active burrows occupied by breeding Mascarene Petrels on three occasions, suggesting possible interference and competition for burrows. Furthermore, one tenrec was observed attacking a Mascarene Petrel before entering a burrow containing a chick. No cats were detected at the breeding colony.

The consumption rate of poison at the bait stations between November 2016 and February 2017 was 83.11% at RD. One percussion from a A24 was recorded in February 2017. No cats were captured and no new feces were found. Cage-traps were retrieved during the last expedition at the RD colony.

VHF tracking of stranded birds and use of sniffer dogs at potential breeding sites

We found two stranded adults that were fit enough for being equipped with VHF transmitters (on 13 and 16 November 2015, at Le Tampon and Sainte-Marie, respectively). Although four receivers were simultaneously deployed at a suspected breeding site (Grand Bassin; Fig. 1), no detection was made. We had no success with the sniffer dog method because all suspected breeding sites were on vertical cliffs that dogs could not access.

DISCUSSION

Advantages of new technologies for seabird research and conservation

Our study shows that the combination of new technologies such as ARUs and infrared thermal binoculars is of particular interest by facilitating the discovery of remote breeding colonies of secretive nocturnal seabirds. It allowed us to locate colonies very efficiently with limited physical effort, searching time, financial costs, and human risks. Infrared thermal binoculars observations and camera traps at colonies were essential for the initial investigations of behaviors of birds and their predators, as well as for providing first assessments of the threats.

ARUs are used in avian research for species richness, abundance estimates (Shonfield and Bayne 2017, Pérez-Granados and Traba 2021), phenology (Blumstein et al. 2011), and behavioral ecology (Blumstein et al. 2011). This technology is also lauded as a useful tool for monitoring rare and elusive species (Blumstein et al. 2011, Holmes et al. 2015) and is recommended when conducting large-scale monitoring of birds in remote locations (see Hill et al. 2006, Venier et al. 2012). This non-invasive and easy-to-use tool reduces disturbance caused by human presence and increases the spatial and temporal scale of studies with minimum effort and time invested in the field. In addition, with current technological advances, ARUs are becoming more affordable with greater portability and longevity in the field (see Hill et al. 2006). However, several limitations have to be considered when using ARUs. First, for bird species that mainly vocalize from the ground (as Mascarene Petrels do; J. Dubos, *unpublished observations*), the distance and the quality of call detections may be reduced, leading

to possible false negatives or impeding the ability to locate the origin of a sound. Second, the very important quantity of data produced leads to painstaking and time-consuming data management and processing (Edney and Wood 2021). Although the use of automated recognition can shorten the process (Knight et al. 2017), the low number of good quality recordings may limit the efficiency of automated detection.

Infrared thermal binoculars are increasingly used in seabird ecology (e.g., Syposz et al. 2021). In comparison to classic nocturnal detection tools such as night vision goggles or radar (Swift and Burt-Toland 2009, Galase 2019), infrared thermal binoculars present strong advantages. For instance, they do not require minimal ambient light, as is the case with night vision goggles (Galase 2019), because detection is based on body temperature only. In contrast to radars, infrared thermal binoculars are a low-cost technology, easy to use and to carry in the field with low maintenance (Orben et al. 2019). Furthermore, at close range (< 500 m) most detected birds can be identified to the species level. Infrared thermal binoculars can be combined with an image recorder so that recorded videos can be reviewed for further analysis. However, detection range of infrared thermal binoculars can be reduced in poor environmental conditions such as foggy or rainy weather, rough topography, or dense vegetation cover. The accuracy and the portability of this technology continues to improve, making it increasingly suitable for rough fieldwork (Gade and Moeslund 2014). This tool was a key element in the success of the present study and we strongly recommend it.

Several other methods are commonly used when looking for secretive seabird species (Gummer et al. 2015, Rayner et al. 2015, 2019, Galase 2019). For instance, tracking birds can help to find a breeding site. Rayner et al. (2015) discovered the first breeding colony of the New Zealand Storm-Petrel (*Fregetta maoriana*) by radio-tagging 24 birds captured at sea, of which 11 were detected from land by using remote base stations and hand-held telemetry receivers. Two of these equipped birds were found in burrows (Rayner et al. 2015). We tentatively used this method during our project but no equipped birds were detected. This lack of success is probably because of the fact that too few birds were equipped. Because the other methods resulted in the finding of colonies, we rapidly abandoned this method.

Satellite telemetry (Argos transmitters or GPS) has been used more or less successfully to locate breeding colonies of elusive and rare seabirds, with tags deployed on birds captured at sea after having been attracted with chumming (see, e.g., Rayner et al. 2019). However, in our context, we quickly discarded these options because of the enormous difficulties in attracting birds close enough to the boat to be catchable (M. Le Corre, P. Pinet, and J. Dubos, *unpublished data*).

Conservation scent dogs (so-called “sniffer dogs”) are a powerful olfactory tool for the detection of rare species (Bennett et al. 2020), including cavity-nesting and nocturnal seabirds (reviewed in Bolton et al. 2021). In 2018, Galase identified the first colony of Band-rumped Storm-Petrel (*Oceanodroma castro*) on Hawaii Island after a conservation scent dog detected 18 potential burrows over an area of 3.4 km² (Galase 2019). In our context, it became apparent that sniffer dogs were unusable because of the topography of the breeding sites.

Implications for conservation

Before the discovery of the two breeding colonies, conservation actions targeting the Mascarene Petrel were limited to rescue campaigns to reduce light-induced mortality (Le Corre et al. 2003, Chevillon et al. 2022) and to small-scale cat controls at suspected breeding sites (Riethmuller et al. 2012; authors, *unpublished data*). The National Park of Réunion Island was established in 2007 (Fig. 1). Twelve of the 18 potential breeding sites (including the two discovered colonies) are located within this park.

Now that we have discovered two breeding colonies and know how to access them on a regular basis, we are able to implement specific conservation actions that more efficiently target threats at these colonies. Four species of introduced mammals have been detected at the colonies or in their immediate vicinity (rats, cats, tenrecs, and shrews). At least two of them are known to prey upon petrel chicks or adults (rats and cats), and the tenrec is suspected of competing with petrels for burrows. We implemented permanent cat and rat control with an early detection system, using a network of camera traps, to prevent any reinvasion. This strategy was very efficient and resulted in a rapid increase of the breeding success at both colonies. Rat control is now extended to other unmonitored and inaccessible breeding sites by using unmanned aircraft systems (Réunion National Park, *personal communication*).

Many studies have demonstrated the efficiency of artificial breeding colonies as an active conservation measure, especially for petrels (e.g., *Pterodroma cahow*, Wingate 1977; *Ardenna pacifica*, Byrd et al. 1983; *Pterodroma leucoptera*, Priddel et al. 2006; *Pterodroma axillaris*, Gummer et al. 2015). Artificial nest boxes limit intra- and inter-specific competition for nests (see Gummer et al. 2015) and reduce predation risk. In 2018, two artificial breeding colonies, consisting of 19 and 20 underground nesting boxes, respectively, were settled near each breeding site (Pinet 2020). The dimensions of these boxes were comparable to those of the natural burrows (see Table 3). These two artificial colonies were equipped with solar-powered acoustic social attraction systems and were permanently freed of rats and cats. The first Mascarene Petrels were observed prospecting at one of these sites within one year (Pinet 2020). We are now monitoring the two sites in order to detect the first breeding attempt.

In the future, the use of predator-proof fences could be explored to prevent colonies from being re-invaded by exotic mammals, but we expect to quickly encounter limitations because of (1) the characteristics of the breeding sites (i.e., high elevation, deep slope, and rough field), and (2) the cost of installing and maintaining the fences. In the longer term, we could translocate chicks from natural colonies impacted by introduced predators to artificial sites where predators have been permanently removed. However, to date this has been impossible because the natural colonies that would be accessible for translocation have produced too few chicks. Removing these chicks may have limited positive demographic impact while significantly jeopardizing existing colonies (Fischer et al. 2022). In addition, there are plans to establish other artificial colonies at places farther from light-polluted areas (Virion et al. 2020).

We found 18 other sites where birds were vocally active. Each of these potential breeding sites probably gathers a small number of pairs. All of these sites are on vertical cliffs, inaccessible to humans

but not to rats and cats. Furthermore, these sites are extremely unstable and subject to permanent erosion and collapse. This situation leads to a conservation paradox for this critically endangered species. On the one hand, protecting 18 sites simultaneously against introduced predators is extremely challenging. On the other hand, the fact that the species breeds on 18 discreet sites protects it from environmental stochasticity (cliff collapses, fire, cyclonic events). We intend to implement long-term acoustic monitoring to detect any drop of vocal activity, which may indicate a decline of a given colony. This may allow us to implement targeted conservation actions at these colonies.

Implications for applied research

Applied research is an essential step for a precise assessment of the threats and viability of a population. Since the discovery of the colonies, the phenology, biology, and population dynamics of the species are now studied by using acoustic and mark-recapture tools at the two breeding colonies. First estimations of demographic parameters, such as breeding success and adult survival, are now used to estimate population viability and to measure impacts of the initial conservation actions at colonies. We also described the at-sea distribution (Saunier 2019) and marine habitat selection (Fernandez 2021) of Mascarene Petrels during the non-breeding period using geolocators. Genetic analyses have allowed us to describe the genetic diversity, estimate contemporary effective population size, and search for evidence of population bottleneck (Lopez et al. 2021). Further genetic studies will look for potential genetic structures between colonies that will be used to identify adapted conservation units (Danckwerts et al. 2021).

CONCLUSION

To our knowledge, Mascarene Petrel is now the only critically endangered species of the genus *Pseudobulweria* whose burrows have been discovered. This discovery represents the cornerstone on which knowledge of the ecology of the Mascarene Petrel is now tremendously increasing, allowing the implementation of effective conservation actions, crucial for the protection of the species. In addition, this study demonstrates the efficiency and the complementarity of non-invasive methods for the discovery of one of the most elusive, rare, and endangered seabird species. Although searching for breeding sites of elusive nocturnal seabirds may present different local challenges, depending on each situation, we are convinced that our research strategy is reproducible and adaptable to other situations and would be of great interest for the conservation of other species facing the same gap of knowledge, particularly endemic nocturnal petrels and shearwaters.

Author Contributions:

PP & MLC conceived and designed the study. JD, YSM, MR, PS, CC, FJ, MLC, & PP collected data and performed the analyses. CCJ, JD, & MLC participated in the writing. All authors reviewed and contributed to the final version of the manuscript.

Acknowledgments:

This study was funded by the SMAC program, Seabird Multidisciplinary Applied Research for Conservation, co-funded by the European Union (ERDF) and the Région Réunion. This study is a production of the European project LIFE + Petrels (grant number: LIFE13 BIO/FR/000075) co-led by the National Park of Réunion Island, the Réunion University, the Société d'Etudes Ornithologiques de La Réunion (SEOR), and the Office National de la Chasse et de la Faune Sauvage, with financial support from the European Union, the Direction de l'Environnement l'Aménagement et du Logement (DEAL), and the Conseil Départemental of Réunion Island. We thank Gérard Millischer of National Park of Mercantour for lending us the infrared thermal binoculars and for his help in the field.

Data Availability:

Data/code sharing not applicable.

LITERATURE CITED

- Attie, C., J.-C. Stahl, and V. Bretagnolle. 1997. New data on the endangered Mascarene Petrel *Pseudobulweria aterrima*: a third twentieth century specimen and distribution. *Colonial Waterbirds* 20(3):406-412. <https://doi.org/10.2307/1521590>
- Bennett, E. M., C. E. Hauser, and J. L. Moore. 2020. Evaluating conservation dogs in the search for rare species. *Conservation Biology* 34(2):314-325. <https://doi.org/10.1111/cobi.13431>
- Blumstein, D. T., D. J. Mennill, P. Clemins, L. Girod, K. Yao, G. Patricelli, J. L. Deppe, A. H. Krakauer, C. Clark, K. A. Cortopassi, et al. 2011. Acoustic monitoring in terrestrial environments using microphone arrays: applications, technological considerations and prospectus. *Journal of Applied Ecology* 48(3):758-767. <https://doi.org/10.1111/j.1365-2664.2011.01993.x>
- Bolton, M., G. Morgan, S. E. Bolton, J. R. F. Bolton, S. Parmor, and L. Bambini. 2021. Teaching old dogs and young dogs new tricks: canine scent detection for seabird monitoring. *Seabird* 33:35-52. <http://www.seabirdgroup.org.uk/seabird-33-35>
- Byrd, G. V., D. I. Moriarty, and B. G. Brady. 1983. Breeding biology of Wedge-tailed Shearwaters at Kilauea Point, Hawaii. *Condor* 85(3):292-296. <https://doi.org/10.2307/1367063>
- Cadet, L. J. T. 1977. La végétation de l'île de La Réunion: Etude phytoécologique et phytosociologique. Thesis. Université Aix-Marseille III, Marseille, France.
- Chevillon, L., J. Tourmetz, J. Dubos, Y. Soulaïmana-Mattoir, C. Hollinger, P. Pinet, F.-X. Couzi, M. Riethmuller, and M. Le Corre. 2022. 25 years of light-induced petrel groundings in Reunion Island: retrospective analysis and predicted trends. *Global Ecology and Conservation* 38:e02232. <https://doi.org/10.1016/j.gecco.2022.e02232>
- Croxall, J. P., S. H. M. Butchart, B. Lascelles, A. J. Stattersfield, B. Sullivan, A. Symes, and P. Taylor. 2012. Seabird conservation status, threats and priority actions: a global assessment. *Bird Conservation International* 22(1):1-34. <https://doi.org/10.1017/S0959270912000020>
- Danckwerts, D. K., L. Humeau, P. Pinet, C. D. McQuaid, and M. Le Corre. 2021. Extreme philopatry and genetic diversification at unprecedented scales in a seabird. *Scientific Reports* 11:6834. <https://doi.org/10.1038/s41598-021-86406-9>
- Dias, M. P., R. Martin, E. J. Pearmain, I. J. Burfield, C. Small, R. A. Phillips, O. Yates, B. Lascelles, P. G. Borboroglu, and J. P. Croxall. 2019. Threats to seabirds: a global assessment. *Biological Conservation* 237:525-537. <https://doi.org/10.1016/j.biocon.2019.06.033>
- Edney, A. J., and M. J. Wood. 2021. Applications of digital imaging and analysis in seabird monitoring and research. *Ibis* 163(2):317-337. <https://doi.org/10.1111/ibi.12871>
- Faulquier, L., L. Solier, F. X. Couzi, J. Tourmetz, M. Saliman, and M. Le Corre. 2017. Evaluation du Plan De Conservation en faveur du Pétrel de Barau *Pterodroma barau* 2008-2016. UMR ENTROPIE, Université de La Réunion/Société d'Etudes Ornithologiques de La Réunion/Parc National de La Réunion, La Réunion, France.
- Fernandez, R. 2021. Modélisation de la distribution océanique du pétrel noir de Bourbon (*Pseudobulweria aterrima*) dans l'océan Indien pendant sa période internuptiale. Université de La Réunion, Saint-Denis, La Réunion, France.
- Fischer, J. H., H. U. Wittmer, C. F. Kenup, K. A. Parker, R. Cole, I. Debski, G. A. Taylor, J. G. Ewen, and D. P. Armstrong. 2022. Predicting harvest impact and establishment success when translocating highly mobile and endangered species. *Journal of Applied Ecology* 59(8):2071-2083. <https://doi.org/10.1111/1365-2664.14219>
- Gade, R., and T. B. Moeslund. 2014. Thermal cameras and applications: a survey. *Machine Vision and Applications* 25:245-262. <https://doi.org/10.1007/s00138-013-0570-5>
- Galase, N. K. 2019. First confirmed Band-rumped Storm Petrel *Oceanodroma castro* colony in the Hawaiian Islands. *Marine Ornithology* 47:25-28.
- Gangloff, B., H. Shirihai, D. Watling, C. Cruaud, A. Couloux, A. Tillier, E. Pasquet, and V. Bretagnolle. 2012. The complete phylogeny of *Pseudobulweria*, the most endangered seabird genus: systematics, species status and conservation implications. *Conservation Genetics* 13:39-52. <https://doi.org/10.1007/s10592-011-0261-6>
- Gineste, B., M. Souquet, F.-X. Couzi, Y. Giloux, J.-S. Philippe, C. Hoarau, J. Tourmetz, G. Potin, and M. Le Corre. 2017. Tropical Shearwater population stability at Reunion Island, despite light pollution. *Journal of Ornithology* 158:385-394. <https://doi.org/10.1007/s10336-016-1396-5>
- Gummer, H., G. Taylor, K.-J. Wilson, and M. J. Rayner. 2015. Recovery of the endangered Chatham Petrel (*Pterodroma axillaris*): a review of conservation management techniques from 1990 to 2010. *Global Ecology and Conservation* 3:310-323. <https://doi.org/10.1016/j.gecco.2014.12.006>
- Hill, G. E., D. J. Mennill, B. W. Rolek, T. L. Hicks, and K. A. Swiston. 2006. Evidence suggesting that Ivory-billed Woodpeckers (*Campephilus principalis*) exist in Florida. *Avian Conservation and Ecology* 1(3):2. <https://doi.org/10.5751/ACE-00078-010302>

- Holmes, S. B., K. Tuininga, K. A. McIlwrick, M. Carruthers, and E. Cobb. 2015. Using an integrated recording and sound analysis system to search for Kirtland's Warbler (*Setophaga kirtlandii*) in Ontario. *Canadian Field-Naturalist* 129(2):115-120. <https://doi.org/10.22621/cfn.v129i2.1688>
- Huré, M. 2019. Caractérisation et modélisation spatiale des habitats de reproduction des pétrels endémiques menacés de La Réunion, le pétrel de Barau (*Pterodroma barau*) et le pétrel noir de Bourbon (*Pseudobulweria aterrima*): implication pour la conservat. Université de La Réunion, Saint-Denis, La Réunion, France.
- International Union for Conservation of Nature (IUCN). 2021. The IUCN red list of threatened species. Version 2021-3. IUCN, Cambridge, UK. <https://www.iucnredlist.org>.
- Jouanin, C. 1970. Note taxonomique sur les Petits Puffins, *Puffinus lherminieri*, de l'Océan Indien occidental. *Oiseau et RFO* 40:303-306.
- Jouanin, C. 1987. Notes on the nesting of Procellariiformes in Réunion. Pages 359-363 in A. W. Diamond, editor. *Studies of Mascarene Island birds*. Cambridge University Press, Cambridge, UK.
- Knight, E. C., K. C. Hannah, G. J. Foley, C. D. Scott, R. M. Brigham, and E. Bayne. 2017. Recommendations for acoustic recognizer performance assessment with application to five common automated signal recognition programs. *Avian Conservation and Ecology* 12(2):14. <https://doi.org/10.5751/ACE-01114-120214>
- Le Corre, M., T. Ghestemme, M. Salamolard, and F.-X. Couzi. 2003. Rescue of the Mascarene Petrel, a critically endangered seabird of Réunion island, Indian Ocean. *Condor* 105(2):387-391. <https://doi.org/10.1093/condor/105.2.387>
- Le Corre, M., A. Ollivier, S. Ribes, and P. Jouventin. 2002. Light-induced mortality of petrels: a 4-year study from Réunion Island (Indian Ocean). *Biological Conservation* 105(1):93-102. [https://doi.org/10.1016/S0006-3207\(01\)00207-5](https://doi.org/10.1016/S0006-3207(01)00207-5)
- Lopez, J., N. Nikolic, M. Riethmuller, J. Dubos, P. Pinet, P. Souharce, F.-X. Couzi, M. Le Corre, A. Jaeger, and L. Humeau. 2021. High genetic diversity despite drastic bottleneck in a critically endangered, long-lived seabird, the Mascarene Petrel *Pseudobulweria aterrima*. *Ibis* 163(1):268-273. <https://doi.org/10.1111/ibi.12864>
- Orben, R. A., A. B. Fleishman, A. L. Borker, W. Bridgeland, A. J. Gladics, J. Porquez, P. Sanzenbacher, S. W. Stephensen, R. Swift, M. W. McKown, and R. M. Suryan. 2019. Comparing imaging, acoustics, and radar to monitor Leach's Storm-Petrel colonies. *PeerJ* 7:e6721. <https://doi.org/10.7717/peerj.6721>
- Paleczny, M., E. Hammill, V. Karpouzi, and D. Pauly. 2015. Population trend of the world's monitored seabirds, 1950-2010. *PLoS ONE* 10(6):e0129342. <https://doi.org/10.1371/journal.pone.0129342>
- Pérez-Granados, C., and J. Traba. 2021. Estimating bird density using passive acoustic monitoring: a review of methods and suggestions for further research. *Ibis* 163(3):765-783. <https://doi.org/10.1111/ibi.12944>
- Pinet, P. 2020. A new success story for Mascarene Petrel conservation, Réunion Island. *Indian Ocean Seabird Group Newsletter* 7:5-6.
- Priddel, D., N. Carlile, and R. Wheeler. 2006. Establishment of a new breeding colony of Gould's Petrel (*Pterodroma leucoptera leucoptera*) through the creation of artificial nesting habitat and the translocation of nestlings. *Biological Conservation* 128(4):553-563. <https://doi.org/10.1016/j.biocon.2005.10.023>
- Probst, J.-M., M. Le Corre, and C. Thébaud. 2000. Breeding habitat and conservation priorities in *Pterodroma barau*, an endangered gadfly petrel of the Mascarene archipelago. *Biological Conservation* 93(1):135-138. [https://doi.org/10.1016/S0006-3207\(99\)00114-7](https://doi.org/10.1016/S0006-3207(99)00114-7)
- Rayner, M. J., K. A. Baird, J. Bird, S. Cranwell, A. F. Raine, B. Maul, J. Kuri, J. Zhang, and C. P. Gaskin. 2019. Land and sea-based observations and first satellite tracking results support a New Ireland breeding site for the Critically Endangered Beck's Petrel *Pseudobulweria beckii*. *Bird Conservation International* 30(1):58-74. <https://doi.org/10.1017/S0959270919000145>
- Rayner, M. J., C. P. Gaskin, N. B. Fitzgerald, K. A. Baird, M. M. Berg, D. Boyle, L. Joyce, T. J. Landers, G. G. Loh, S. Maturin, et al. 2015. Using miniaturized radiotelemetry to discover the breeding grounds of the endangered New Zealand Storm Petrel *Fregetta maoriana*. *Ibis* 157(4):754-766. <https://doi.org/10.1111/ibi.12287>
- Riethmuller, M., F. Jan, Y. Giloux, and M. Saliman. 2012. Plan national d'actions en faveur du Pétrel noir de Bourbon *Pseudobulweria aterrima* (2012-2016). Ministère de l'Ecologie, du Développement durable et de l'Energie/Direction de l'Environnement de l'Aménagement et du Logement de La Réunion, La Réunion, France.
- Rodríguez, A., J. M. Arcos, V. Bretagnolle, M. P. Dias, N. D. Holmes, M. Louzao, J. Provencher, A. F. Raine, F. Ramírez, B. Rodríguez, et al. 2019. Future directions in conservation research on petrels and shearwaters. *Frontiers in Marine Science* 6:1-27. <https://doi.org/10.3389/fmars.2019.00094>
- Rodríguez, A., N. D. Holmes, P. G. Ryan, K.-J. Wilson, L. Faulquier, Y. Murillo, A. F. Raine, J. F. Penniman, V. Neves, B. Rodríguez, et al. 2017. Seabird mortality induced by land-based artificial lights. *Conservation Biology* 31(5):986-1001. <https://doi.org/10.1111/cobi.12900>
- Saunier, M. 2019. Etude de la phénologie, de la distribution et de l'activité en mer des pétrels noirs de Bourbon (*Pseudobulweria aterrima*). Thesis, Université de La Réunion, Saint-Denis, La Réunion, France.
- Shonfield, J., and E. M. Bayne. 2017. Autonomous recording units in avian ecological research: current use and future applications. *Avian Conservation and Ecology* 12(1):14. <https://doi.org/10.5751/ACE-00974-120114>
- Swift, R., and E. Burt-Toland. 2009. Surveys of procellariiform seabirds at Hawai'i Volcanoes National Park, 2001-2005. Pacific Cooperative Studies Unit, University of Hawai'i at Mānoa, Honolulu, Hawai'i, USA.

Syposz, M., O. Padget, J. Willis, B. M. Van Doren, N. Gillies, A. L. Fayet, M. J. Wood, A. Alejo, and T. Guilford. 2021. Avoidance of different durations, colours and intensities of artificial light by adult seabirds. *Scientific Reports* 11:18941. <https://doi.org/10.1038/s41598-021-97986-x>

Troy, J. R., N. D. Holmes, J. A. Veech, and M. C. Green. 2013. Using observed seabird fallout records to infer patterns of attraction to artificial light. *Endangered Species Research* 22:225-234. <https://doi.org/10.3354/esr00547>

Venier, L. A., S. B. Holmes, G. W. Holborn, K. A. McIlwrick, and G. Brown. 2012. Evaluation of an automated recording device for monitoring forest birds. *Wildlife Society Bulletin* 36(1):30-39. <https://doi.org/10.1002/wsb.88>

Virion, M.-C., L. Faulquier, M. Le Corre, F.-X. Couzi, M. Salamolard, B. Lequette, P. Pinet, J. Dubos, M. Riethmuller, Y. Soulainana Mattoir, et al. 2021. Plan national d'action en faveur des pétrels endémiques de La Réunion 2021-2030. UMR ENTROPIE, Université de La Réunion, Société d'Études Ornithologiques de La Réunion, Parc national de La Réunion, La Réunion, France.

Warham, J. 1990. *The petrels: their ecology and breeding systems*. Academic, London, UK.

Warham, J. 1996. *The behaviour, population biology and physiology of the petrels*. Academic, London, UK.

Wingate, D. B. 1977. Excluding competitors from Bermuda petrel nesting burrows. Pages 93-102 in S. A. Temple, editor. *Proceedings of symposium on management techniques for preserving endangered birds*. University of Wisconsin-Madison, Madison, Wisconsin, USA.

Appendix 1. Details about ARU deployment during the two seasons of prospection, 2015-2016 & 2016-2017 (A) without Malabar and (B) about Malabar.

(A)

Season	Site	Initial date of records	Final date of records	Number of days	Number of recorded hours	Number of analyzed hours	Type of analyse	Contact
2015-2016	Augustave	20/10/2015	18/11/2015	29	262	65 (24.80 %)	manual	no
		15/12/2015	11/02/2016	58				
		20/02/2016	29/02/2016	9				
2016-2017	Barrage BDP	08/10/2016	09/11/2016	32	66	66 (100 %)	software	no
2015-2016	Bras Caron	14/07/2015	31/07/2015	17	940	940 (100 %)	Manual / software	no
		07/08/2015	26/08/2015	19				
		05/09/2015	09/01/2016	126				
		04/02/2016	27/02/2016	23				
2016-2017	Bras Dimitile	10/09/2016	11/10/2016	31	64	64 (100 %)	software	no
2016-2017	Bras Sec 1 haut	07/07/2016	27/07/2016	20	190	190 (100 %)	software	no
		28/09/2016	02/10/2016	4				
2015-2016	Bras Sec 2 haut	28/04/2016	30/04/2016	3	20	20 (100%)	manual	no
2016-2017	Bras Sec 2 haut	28/09/2016	25/10/2016	27	271	271 (100 %)	software	yes
		26/10/2016	09/11/2016	14				
		10/11/2016	09/02/2017	91				
2015-2016	Bras Sec bas	27/01/2016	17/02/2016	21	185	185 (100 %)	manual	yes
2016-2017	Canalisation Pont d'Yves 2	17/03/2017	06/04/2017	20	42	42 (100 %)	software	no
2016-2017	Caverne Mussard	10/09/2016	03/10/2016	23	48	48 (100 %)	software	no
2016-2017	Coteau Sipec	12/10/2016	15/11/2016	34	70	70 (100 %)	software	no
2016-2017	Crête A coteau Maigre 1	30/08/2016	30/09/2016	31	64	64 (100 %)	software	no
2016-2017	Crête A coteau Maigre 2	01/10/2016	05/01/2017	96	194	184 (94.80 %)	software	yes
2016-2017	Crête B coteau Maigre 1	08/09/2016	30/11/2016	83	168	168 (100 %)	software	yes
2016-2017	Crête B coteau Maigre 2	01/12/2016	31/01/2017	61	278	278 (100 %)	software	yes
		10/02/2017	18/04/2017	67				
2016-2017	Crête Bras Sainte Suzanne	24/08/2016	17/02/2017	177	356	356 (100 %)	software	no
2016-2017	Crête C coteau Maigre	07/09/2016	03/01/2017	118	238	238 (100 %)	software	yes
2016-2017	Crête Chevron	25/08/2016	28/02/2017	187	376	376 (100 %)	software	yes
2015-2016	Dugain	06/10/2015	22/10/2015	16	786,5	534,5 (67.99 %)	manual	yes
		04/11/2015	15/01/2016	72				
		26/01/2016	15/02/2016	20				
2016-2017	Dugain	12/07/2016	03/08/2016	22	578	578 (100%)	software	yes
		18/09/2016	28/03/2017	191				
2016-2017	Echelle BDP	02/12/2016	31/12/2016	29	62	62 (100 %)	software	no
2016-2017	En face d'ilet Citron	17/01/2017	28/04/2017	101	192	192 (100 %)	software	no

		10/10/2015	04/11/2015	25				
		07/11/2015	27/11/2015	20				
2015-2016	Fleur jaune	04/12/2015	29/12/2015	25	363.5	126 (34.7 %)	manual	no
		07/01/2016	30/01/2016	23				
		04/02/2016	26/02/2016	22				
		05/03/2016	21/03/2016	16				
2016-2017	Fleur jaune les hauts	18/02/2017	22/03/2017	32	56	56 (100 %)	software	no
2016-2017	Grand Ilet	12/10/2016	05/11/2016	24	50	50 (100 %)	software	no
2016-2017	Grande Chaloupe 1	18/01/2017	24/01/2017	6	27	27 (100 %)	software	no
2016-2017	Grande Chaloupe 2	18/01/2017	24/01/2017	6	27	27 (100 %)	software	no
2016-2017	Grande Chaloupe 3	18/01/2017	24/01/2017	6	27	27 (100 %)	software	no
2016-2017	Grande Chaloupe 4	18/01/2017	24/01/2017	6	27	27 (100 %)	software	no
2016-2017	Grande Chaloupe 5	18/01/2017	24/01/2017	6	27	27 (100 %)	software	no
2016-2017	Grande Chaloupe 6	18/01/2017	24/01/2017	6	36	36 (100 %)	software	no
2016-2017	Ilet Aurelien dijoux	10/11/2016	01/12/2016	21	44	44 (100 %)	software	yes
2015-2016	Ilet Camille	04/11/2015	18/11/2015	14	467.5	205,5 (44.0 %)	manual	yes
		03/12/2015	19/01/2016	47				
		08/10/2015	02/11/2015	25				
		10/11/2015	24/11/2015	14				
		25/11/2015	03/12/2015	8				
2015-2016	Langevin Grand Pays	08/12/2015	02/01/2016	25	441,5	167 (37.8 %)	manual	no
		07/01/2016	30/01/2016	23				
		06/02/2016	28/02/2016	22				
		10/03/2016	18/03/2016	8				
		02/10/2015	08/10/2015	6				
2015-2016	Le Bloc	09/10/2015	03/11/2015	25	476,5	112,5 (23.6 %)	manual	no
		07/11/2015	05/02/2016	90				
		05/03/2016	28/03/2016	23				
2016-2017	Le Trou	17/09/2016	02/04/2017	197	367	367 (100 %)	software	yes
		06/10/2015	24/10/2015	18				
2015-2016	Le Trou	04/11/2015	08/01/2016	65	836	836 (100 %)	Manual/software	yes
		27/01/2016	11/02/2016	15				
		04/03/2016	19/03/2016	15				
2015-2016	Mahavel	09/10/2015	03/11/2015	25	503	253.5 (50.4 %)	manual	yes
		07/11/2015	01/03/2016	115				
		11/03/2016	01/04/2016	21				

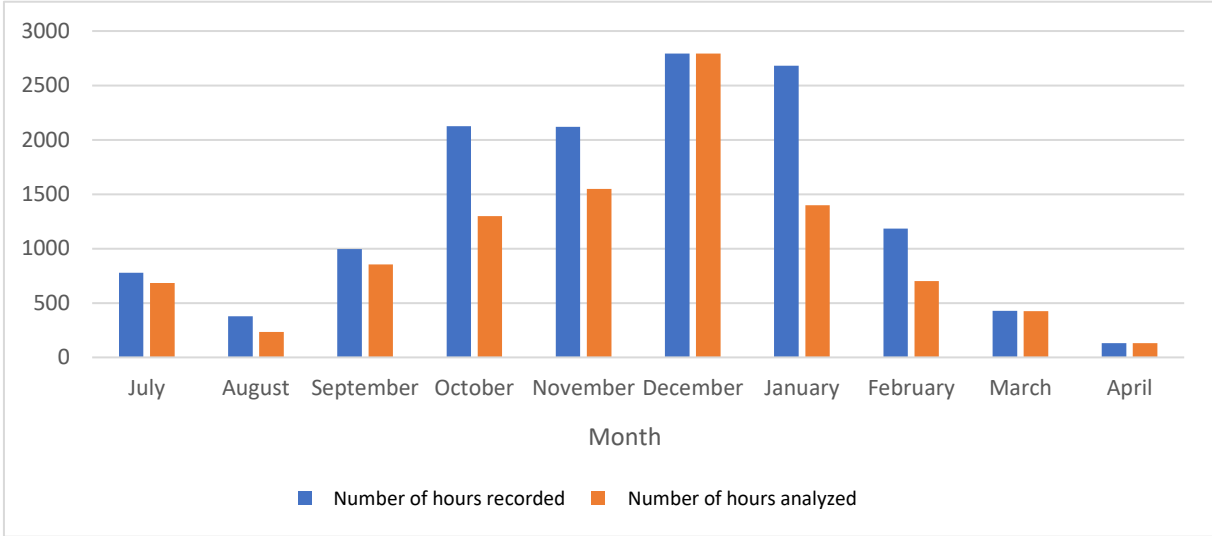
		14/07/2015	27/07/2015	13				
		06/08/2015	22/08/2015	16				
		05/09/2015	10/09/2015	5				
2015-2016	Mare à Vieille Place	17/09/2015	24/09/2015	7	815	255 (31.3 %)	manual	no
		10/10/2015	29/10/2015	19				
		07/11/2015	29/12/2015	52				
		01/01/2016	06/01/2016	5				
		20/01/2016	31/01/2016	11				
		23/07/2015	29/07/2015	6				
2015-2016	Mollaret	04/08/2015	16/09/2015	43	429.75	231.75 (53.9 %)	manual	no
		16/09/2015	04/10/2015	18				
2016-2017	Pic des Chèvres	04/10/2016	12/10/2016	8	16	16 (100 %)	software	no
2016-2017	Piton Dédé	21/02/2017	20/03/2017	27	56	56 (100 %)	software	no
2016-2017	Piton Petit Louis	03/02/2017	09/03/2017	34	70	70 (100 %)	software	no
		05/10/2015	25/10/2015	20				
		04/11/2015	11/11/2015	7				
2015-2016	Piton Rouge_bas	12/11/2015	18/11/2015	6	678	233.5 (34.4 %)	manual	yes
		03/12/2015	23/12/2015	20				
		29/12/2015	14/01/2016	16				
		26/01/2016	16/02/2016	21				
		11/11/2015	23/11/2015	12				
2015-2016	Piton Rouge_haut	02/12/2015	11/12/2015	9	286	286 (100 %)	manual	yes
		11/12/2015	17/12/2015	6				
		18/12/2015	08/01/2016	21				
2015-2016	Piton Villecourt	05/10/2015	22/10/2015	17	143	60 (42.0 %)	manual	no
2016-2017	Ravineà Jacques 1	24/01/2017	01/02/2017	8	36	36 (100 %)	software	no
2016-2017	Ravineà Jacques 2	24/01/2017	01/02/2017	8	36	36 (100 %)	software	no
2016-2017	Ravineà Jacques 3	24/01/2017	01/02/2017	8	36	36 (100 %)	software	no
2016-2017	Ravineà Jacques 4	24/01/2017	01/02/2017	8	36	36 (100 %)	software	no
2016-2017	Rein Dimitile	10/09/2016	11/10/2016	31	64	64 (100 %)	software	yes
2015-2016	Rivière de l'Est	13/11/2015	24/01/2016	72	327	129 (39.4 %)	manual	no
		10/02/2016	21/02/2016	11				
2015-2016	Roche Plate	18/10/2015	15/12/2015	58	286,5	64 (22.3 %)	manual	no
		14/01/2016	13/03/2016	59				

2016-2017	SM1_rivière des pluies	21/12/2016	28/12/2016	7	31.5	31.5 (100 %)	software	no
2016-2017	SM10_rivière des pluies	04/01/2017	11/01/2017	7	31.5	31.5 (100 %)	software	no
2016-2017	SM11_rivière des pluies	04/01/2017	11/01/2017	7	31.5	31.5 (100 %)	software	no
2016-2017	SM12_rivière des pluies	04/01/2017	11/01/2017	7	31.5	31.5 (100 %)	software	no
2016-2017	SM13_rivière des pluies	04/01/2017	11/01/2017	7	31.5	31.5 (100 %)	software	no
2016-2017	SM14_rivière des pluies	04/01/2017	11/01/2017	7	31.5	31.5 (100 %)	software	no
2016-2017	SM2_rivière des pluies	21/12/2016	28/12/2016	7	31.5	31.5 (100 %)	software	no
2016-2017	SM3_rivière des pluies	21/12/2016	28/12/2016	7	31.5	31.5 (100 %)	software	no
2016-2017	SM4_rivière des pluies	21/12/2016	28/12/2016	7	31.5	31.5 (100 %)	software	no
2016-2017	SM5_rivière des pluies	28/12/2016	04/01/2017	7	31.5	31.5 (100 %)	software	no
2016-2017	SM6_rivière des pluies	28/12/2016	04/01/2017	7	31.5	31.5 (100 %)	software	no
2016-2017	SM7_rivière des pluies	28/12/2016	04/01/2017	7	31.5	31.5 (100 %)	software	no
2016-2017	SM8_rivière des pluies	28/12/2016	04/01/2017	7	31.5	31.5 (100 %)	software	no
2016-2017	SM9_rivière des pluies	28/12/2016	04/01/2017	7	31.5	31.5 (100 %)	software	no
2016-2017	Source aux hirondelles 1	09/11/2016	21/12/2016	42	126	126 (100 %)	software	yes
		16/09/2016	15/10/2016	29				
2016-2017	Source aux hirondelles 2	12/07/2016	30/07/2016	18	384	384 (100 %)	software	yes
		22/12/2016	03/04/2017	102				
2016-2017	telepherique BDP	06/09/2016	07/10/2016	31	64	64 (100 %)	software	no
2016-2017	Vaquerois	06/01/2017	02/02/2017	27	56	56 (100 %)	software	no

(B)

Season	Site	Initial date of records	Final date of records	Number of days	Number of recorded hours	Number of analyzed hours	Type of analyse	Contact
2015-2016	Malabar	22/07/2015	07/09/2015	47	966.75	812.00 (84.0 %)	manual	yes
		07/09/2015	11/10/2015	34				
		26/10/2015	12/12/2015	47				
		31/12/2015	18/01/2016	18				

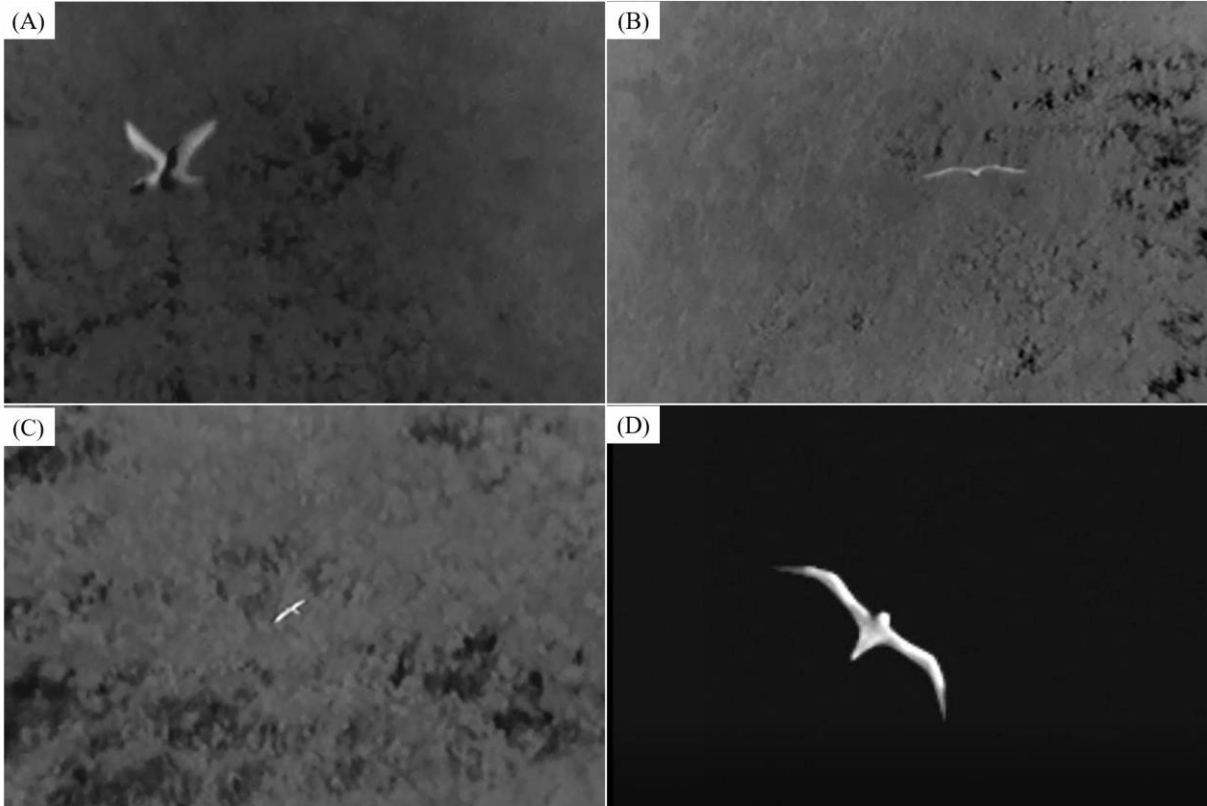
Appendix 2. Acoustic recording and analysis effort by month between July 2015 and April 2017 (without Malabar data).



Appendix 3. Sites with Mascarene petrel vocalizations detected during the automated bioacoustics survey, 2015-2016 & 2016-2017, Reunion island.

Site	Number of hours of records	Number of vocalizations	Number of hours of records analysed	Vocalization rate (voc/h)
Malabar (crête)		6801	1136.25	5.985
Le Trou		1769	1203.00	1.470
Rein Dimitile		538	64.00	8.406
Dugain		362	1364.50	0.265
Crête C coteau Maigre		44	238.00	0.185
Crête B coteau Maigre 1		42	168.00	0.250
Crête Chevron		29	376.00	0.077
Source aux hirondelles 1		28	96.00	0.292
Ilet Camille		13	205.50	0.063
Piton Rouge_haut		11	286.00	0.038
Bras sec bas		7	185.00	0.038
Crête B coteau Maigre 2		7	278.00	0.025
Crête A coteau Maigre 2		3	194.00	0.015
Piton Rouge_bas		3	233.50	0.013
Source aux hirondelles 2		3	414.00	0.007
Bras sec 2 haut		1	291.00	0.003
Ilet aurelien dijoux		1	44.00	0.023
Mahavel		1	253.50	0.004

Appendix 4. Examples of Mascarene Petrel observations with IR thermal camera during nocturnal observation sessions, in 2016-2017, Reunion Island. White and black colors are for cold and warm temperatures, respectively. Pictures display (A) a commuting flight, (B) a prospecting flight from the back, (C) prospecting flight from above and (D) an individual preparing its landing (view from under).



Appendix 5. Details about landing birds during nocturnal
prospections.

Site	Number of landing birds	Number of observation hours	Rate of landing birds
Bras Sec	1	22.0	0.05
DZ DropZone	1	3.5	0.29
Le Trou	2	24.25	0.08
RD	12	26.25	0.46
RDC	17	11.25	1.51

Appendix 6. Measurements of the ringed adults captured on the newly discovered breeding colonies of Mascarene Petrel in 2016-2017, Reunion Island.

Colony	Date	Nest id	Ring number	WC (cm)	CL (cm)	BD (cm)	CR (cm)	TL (cm)	W (g)
RD	15/11/2016	C101	GE62054	266.0	29.5	13.1	17.9	39.8	310.0
	15/11/2016	C103	GE62055	254.0	27.7	11.3	16.6	38.4	270.0
	15/11/2016	C105	GE54900	260.0	32.4	13.3	16.7	43.8	330.0
	15/11/2016	C105	GE62053	266.0	28.7	12.3	16.8	39.4	260.0
	15/11/2016	C107	GE54899	259.0	30.9	12.6	17.5	43.3	310.0
	16/11/2016	C102	GE62065	268.0	31.9	12.2	16.9	44.5	285.0
	16/11/2016	C113	GE62066	260.0	31.0	11.9	19.8	40.9	255.0
	06/12/2016	C110	GE62067	262.0	28.4	12.1	18.4	40.4	280.0
	06/12/2016	C110	GE62068	256.0	29.2	13.5	20.0	42.6	340.0
	17/02/2017	C108	FX26514	260.0	31.4	13.3	20.2	43.3	280.0
	17/02/2017	C109	FX26513	254.0	30.7	13.5	19.1	41.4	300.0
	17/02/2017	C113	FX26515	261.0	30.0	12.9	19.7	41.5	270.0
	17/02/2017	C114	FX26516	253.0	29.3	12.3	19.0	39.8	235.0
				<i>Mean (sd)</i>	259.9 (4.8)	30.1 (1.4)	12.6 (0.7)	18.4 (1.4)	41.5 (1.9)
RDC	28/02/2017	C201	FX20997	252.0	29.8	11.7	18.1	40.8	250.0
	28/02/2017	C201	FX26517	253.0	29.3	11.3	16.9	39.6	260.0
	28/02/2017	C204	GE62056	253.0	30.4	12.4	19.3	39.2	280.0
	28/02/2017	C205	FX26518	248.0	29.6	13.1	19.5	37.6	-
	01/03/2017	C206	GE62057	252.0	30.9	12.9	19.9	40.5	280.0
	01/03/2017	C207	GE62069	256.0	30.2	12.5	19.6	42.6	250.0
	01/03/2017	C208	FX20998	252.0	28.7	12.1	16.0	40.1	255.0
				<i>Mean (sd)</i>	252.3 (2.4)	29.8 (0.7)	12.3 (0.6)	18.5 (1.5)	40.1 (1.5)
			Total mean (sd)	257.3 (5.5)	30.0 (1.2)	12.5 (0.7)	18.4 (1.4)	41.0 (1.9)	278.9 (28.3)

Subscripts: RD, Rein de Dimitile; RDC, Rond des Chevrons; WC, wing chord; CL, culmen length; BD, bill depth; CR, crochet length; TL, tarsus length; W, weight

Appendix 7. Illustration of the presence of rats (*Rattus rattus*, A) and Tenrecs (*Tenrec ecaudatus*, B) inside the breeding colonies of Mascarene petrels, Reunion Island.

(A)



08-12-2016 00:51:13

(B)



29-11-2016 21:56:36

Appendix 8. Details of rat observations at the first colony discovered, from 16 November 2016 to 25 January 2017, Reunion Island.

Time range	Number of days with cameras on	Number of rat pictures	Rate of rat observation
16 Nov – 30 Nov	52	3	0.06
01 Dec – 31 Dec	135	17	0.13
01 Jan – 25 Jan	30	0	0.00