

This is a pre-copy-editing, author-produced PDF of an article accepted for publication in Journal of Plankton Research following peer review. The definitive publisher-authenticated version is available online at: <http://plankt.oxfordjournals.org/cgi/content/abstract/31/4/399>

The significance of organic nutrients in the nutrition of *Pseudo-nitzschia delicatissima* (Bacillariophyceae)

Sofia Loureiro¹, Cécile Jauzein², Esther Garcés^{1,*}, Yves Collos², Jordi Camp¹ and Dolors Vaqué¹

¹ Departament de Biologia Marina i Oceanografia, Institut de Ciències del Mar, ICM-CSIC. Pg. Marítim de la Barceloneta 37-49, 08003 Barcelona, Catalonia, Spain

² Université Montpellier II, CNRS, Ifremer, Laboratoire Ecosystèmes Lagunaires (UMR 5119). CC093, 34095 Montpellier Cedex 5, France

*: Corresponding author : Garcés E., email address : esther@icm.es

Abstract:

The influence of organic nutrients on the evolution of *Pseudo-nitzschia delicatissima* cultures was investigated in an enrichment experiment with high-molecular-weight dissolved organic matter (HMWDOM) and in an uptake assay with ¹⁵N-ammonium and ¹⁵N-urea. HMWDOM was extracted from seawater collected at a nearby shore station during the decline of a diatom bloom. Four incubations were prepared: L1/5+DOM (*P. delicatissima* grown in L1 growth medium with 1/5 of the nitrate concentration of standard L1), (L1-N)+DOM (L1 without nitrate, i.e. nitrogen-deficient treatment), L1-DOM (control culture without added DOM) and BV+DOM (bacterial and viral control, free of microalgae). Incubations were carried out for 10 days. Chlorophyll *a* concentrations differed after day 4 and reached higher levels in the L1-DOM incubation by the end of the experiment; however, similar growth rates were observed in all incubations (1.64 ± 0.05 divisions day⁻¹). The persistently lower cellular chlorophyll content in (L1-N)+DOM during the experiment was consistent with N limitation conditions. The data suggested that the nitrogen needed for the growth of (L1-N)+DOM cells might have originated from the DOM. Based on the results of ¹⁵N uptake assays, it was concluded that *P. delicatissima* more readily acquires ammonium than urea. Nevertheless, under low N conditions, *P. delicatissima* may use urea as an alternative N source, and comparable photosynthetic rates are attained on either substrate. Taken together, our results suggest a positive effect of organic nutrients on the growth of *P. delicatissima*.

Keywords: Dissolved Organic Matter (DOM) · Urea · Nitrogen Uptake · Harmful Algal Blooms (HABs)

INTRODUCTION

The role of dissolved organic matter (DOM) in aquatic food webs has been widely explored in recent years, in contrast to earlier investigations where methodological limitations restricted chemical analyses. Nevertheless, comparisons of those earlier methods with recent analytical results have shown that the former were free of serious errors (Williams, 1995). While the importance of this active organic pool remains the focus of current investigations, much remains to be understood regarding the complex mechanisms and dynamics underlying the production and decay processes inherent in multitrophic levels (Hansell and Carlson, 2002). Interactions with the microbial loop involving processes of excretion, lysis, remineralisation and predation are but a few examples (Sondergaard *et al.*, 2004). Within DOM, the dissolved organic nitrogen (DON) is an important component of the total nitrogen present in marine systems (Berman and Bronk, 2003). The increasing amounts of urea being deposited in coastal waters account for the fact that it is a significant component of the total DON pool, motivating therefore the existent research of the dynamics of this organic nitrogen resource in order to better understand its effects on the structure of the food web (Glibert *et al.*, 2006).

Several algal species forming harmful algal blooms (HABs) are able to use mixotrophic mechanisms, rather than strict autotrophic ones, to dominate phytoplankton blooms (Glibert and Legrand, 2006). Knowledge of the nutritional preferences and acquisition mechanisms employed by potentially harmful species is therefore critical to better comprehend the development and maintenance of HABs. The ability of flagellate life forms to overcome low nutrient affinity by complementing their diet of inorganic nutrients with organic ones is a well-known survival strategy of this microalgae group (Collos *et al.*, 2007; Smayda, 1997). Although inorganic nutrients are regarded as the main source of diatom sustenance, past (Lewin and Hellebust, 1976) and present (Berman and Bronk, 2003; Seitzinger and Sanders, 1999) research has focused on the availability and uptake of organic substrates by diatoms as a means of diversifying from conventional trophic pathways.

Toxic blooms of *Pseudo-nitzschia* spp. are associated with the production of domoic acid, a toxin that may induce a form of neurological damage known as amnesic shellfish poisoning (ASP) in consumers of contaminated vectors such as shellfish and sardines (Bates *et al.*, 1998; Fryxell *et al.*, 1997). The danger posed by ASP has economic consequences as well arising from the precautionary closure of shellfish harvesting when high levels of toxin and/or toxin-producing diatom species are detected (Addison and Stewart, 1989). The 11 potentially toxic species that comprise *Pseudo-nitzschia* spp. include some reported as poisonous in some areas but harmless in others. For example, *Pseudo-nitzschia delicatissima* (Cleve) Heiden is a cryptic species complex

68 containing both toxic strains (Fryxell *et al.*, 1997) as well as non-toxic ones (Fehling *et al.*, 2005).
69 Genotype and environmental conditions, including nutrient status, have been suggested as triggers
70 of the variability in toxin production (GEOHAB, 2005). *P. delicatissima* is a ubiquitous and thus
71 physiologically versatile microalgae that forms blooms in diverse environments, from upwelling
72 regions (Kudela *et al.*, 2005) to areas subject to anthropogenic influences (Quijano-Scheggia *et al.*,
73 2008a). It is considered a r-selected diatom, owing to its fast growth rate and cell yield (Quijano-
74 Scheggia *et al.*, 2008b). Research on the *Pseudo-nitzschia* spp. trophic preferences has evidenced
75 the ability of this genus to use organic elements as a complementary source of nutrition (Hillebrand
76 and Sommer, 1996; Howard *et al.*, 2007) and as a dark survival strategy (Mengelt and Prézelin,
77 2002).

78 This investigation was composed of two experimental set-ups aimed at evaluating the
79 influence of organic elements on the growth of *Pseudo-nitzschia delicatissima*. The first involved
80 the addition of concentrated high-molecular-weight dissolved organic matter (HMWDOM) to a
81 culture of *P. delicatissima*, while in the second ¹⁵N-labeling was used to determine the *P.*
82 *delicatissima* uptake kinetics of ammonium and urea.

83

84

MATERIAL AND METHODS

85

86 **Experiment I. *Pseudo-nitzschia delicatissima* and DOM**

87 ***Pseudo-nitzschia delicatissima* culture.** A strain of *P. delicatissima* (ICMB-F2B2), isolated from
88 Arenys de Mar (Catalonia Coast, Spain), was grown in L1 medium (Guillard and Hargraves, 1993)
89 as a unialgal culture. A second culture of this strain was acclimatized to a low-nitrogen growth
90 medium, L1/5 (1/5 of the original L1 concentration of nitrate), three weeks before the experiment
91 by a total of three transfers to the new medium. Both cultures were grown at a salinity of 30 and at a
92 temperature of 20°C, with a 12:12 light:dark cycle. Illumination was provided by fluorescence tubes
93 (Gyrolux, Sylvania, Germany) with a photon irradiance of 100 $\mu\text{mol photons m}^{-2} \text{s}^{-1}$.

94

95 ***DOM extraction and concentration.*** Seawater was collected from the oligotrophic Blanes coastal
96 station (41°40'19"N, 2°47'11"E) in 28 May (2007), 60 km north of Barcelona, Spain, during the
97 decline of a diatom bloom (phytoplankton base data of the Microbial Observatory of Blanes Bay,
98 MOBB). Sampling during a period of low tourism and dry spring conditions (monthly mean rain <
99 1 mm; meteorological data from the area, Meteocat) additionally minimized allochthonous DOM
100 sources from inland flow, thus favoring the presence of autochthonous DOM of autotrophic origin.
101 HMWDOM (>1000 Da) was obtained by filtration of the seawater through a 0.2- μM cartridge (Pall

102 Corporation) and further concentration (50×) by tangential-flow filtration (Prep/Scale-TFF
103 cartridge, Millipore).

104

105 **Experimental design.** Exponentially growing L1 and L1/5 cultures were diluted to $\approx 1 \times 10^3$ cells
106 L^{-1} and distributed into 2.0 L polycarbonate containers (Nalgene) to a total volume of 1.8 L. Two
107 treatments and two controls were prepared in duplicate: L1/5+DOM, (L1-N)+DOM (N-deficient
108 with DOM addition), L1-DOM (control without added DOM), and BV+DOM (bacterial+viral
109 control, obtained by removing *Pseudo-nitzschia delicatissima* from L1/5 medium by filtration
110 through a 5- μ m pore size filter). The starting concentrations (prior to filtration) in *P. delicatissima*
111 culture were 8×10^3 bacteria ml^{-1} and 2×10^6 viruses ml^{-1} . Incubations were enriched with DOM in
112 the order of $129 \pm 50 \mu M$. The experiment was carried out in semi-continuous mode by replacing
113 13% of the volume every two days (dilution rate = $0.07 d^{-1}$). The outflow was used to estimate
114 chemical and biological parameters. Incubation containers were mounted on a plankton wheel and
115 rotated throughout the experiment (0.2 rpm) to promote homogeneous conditions and active growth.

116

117 **Experiment II. *Pseudo-nitzschia delicatissima*: ammonium and urea uptake**

118 **Experimental design.** *Pseudo-nitzschia delicatissima* cells growing exponentially in L1 were
119 collected by 5- μ m gravity filtration, which allowed the simultaneous exclusion of most bacteria
120 from the original culture (Doucette and Powell, 1998). Cells were further resuspended in N-
121 depleted growth medium for 6 h to ensure minimum ammonium concentrations at the beginnings of
122 the experiments. Two series of 1-h incubations were prepared by adding ^{13}C -labeled bicarbonate to 50
123 ml of N-starved cultures at a fixed concentration, followed by the addition of either ^{15}N -ammonium
124 or ^{15}N -urea in a range of nitrogen concentrations (0.1, 0.2, 0.5, 1, 2, 3, 5, and 10 μM -N), with a total
125 of eight experimental sets per series. Concurrently, separate, short-time, 50-ml incubations with a
126 fixed N source (^{15}N -ammonium or ^{15}N -urea) were established every 15 min to follow the short-time
127 evolution of N-uptake. After incubation, all cultures were filtered through pre-combusted (4 h at
128 $450^\circ C$) Gelman A/E filters (1- μ m equivalent pore size). The filters were subsequently dried ($60^\circ C$
129 for 24 h) and then stored at room temperature until analysis.

130

131 **N-uptake calculations.** ^{15}N -enrichments were converted to N-specific uptake rates of N-ammonium
132 ($V_{N\text{-ammonium}}$) and N-urea ($V_{N\text{-urea}}$), as described in Collos et al. (2005). Since N-specific uptake rates
133 vs. substrate concentrations followed saturation kinetics, the Michaelis-Menten nonlinear regression
134 was applied to estimate the half-saturation constant (K_s , μM -N), the maximum uptake rate (V_{max} , h^{-1}),
135 1), and the initial slope (α , $L h^{-1} \mu g at N^{-1}$). The latter was obtained from the uptake rate at 0.5 μM -N

136 concentration (Hurd and Dring, 1990). The results reflect the competitive ability of cells at low
137 substrate concentrations.

138

139 **Chemical and biological analysis.** Inorganic nutrients (ammonium, nitrite, nitrate, phosphate, and
140 silicate) samples were stored at -20°C and subsequently analyzed by means of an auto-analyzer
141 (Alliance Evolution II) using standard colorimetric techniques (Grasshoff *et al.*, 1983). Urea
142 concentrations were determined according to the method of Goeyens *et al.* (1998). Samples for the
143 analysis of organic elements were filtered through 0.7 µm filters (pre-combusted at 450°C for 6 h
144 and washed with Milli-Q water before sample collection). DOM was estimated by the distribution
145 of dissolved organic carbon, dissolved organic nitrogen, and dissolved organic phosphorus (DOC,
146 DON, and DOP, respectively). DOC was determined by high-temperature catalytic oxidation using
147 a Shimadzu TOC-V (Álvarez-Salgado and Miller, 1998) from samples fixed with H₃PO₄ and stored
148 at 4°C in 10 ml flame-sealed glass ampoules (pre-combustion: 450°C, 24h). Total dissolved nitrogen
149 (TDN) and total dissolved phosphorus (TDP) were determined using a Bran+Luebbe AA3-auto-
150 analyzer, after persulfate oxidation (Grasshoff *et al.*, 1999). DON was derived by subtracting
151 dissolved inorganic nitrogen (DIN= nitrate + nitrite + ammonium) from TDN, and DOP by
152 subtracting dissolved inorganic phosphate from TDP. Chlorophyll *a* (chl *a*) samples were filter-
153 extracted in acetone for 48 h and then analyzed in a Turner 10 AU fluorometer according to the
154 method of Yentsh & Menzel (1963). Particulate nitrogen (PN) and particulate carbon (PC) as well
155 as ¹³C/¹²C and ¹⁵N/¹⁴N isotopic ratios were estimated by means of an Integra CN elemental
156 analysis–mass spectrometry system (PDZ Europa, UK). *Pseudo-nitzschia delicatissima* cells were
157 maintained with Lugol-iodine solution, allowed to sediment (24 h), and quantified in a
158 representative area with a Leica-Leitz DM-IL inverted microscope (Andersen and Throndsen,
159 2003). Bacterial cells were fixed with 1% paraformaldehyde, frozen in liquid N, stored at -80°C and
160 quantified in a Becton-Dickinson FACScalibur flow cytometer emitting at 488 nm (Gasol and
161 delGiorgio, 2000). Further unfrozen samples were stained with 2.5 mM Syto 13 (Molecular
162 Probes), mixed with yellow-green latex beads (Polyscience) as an internal standard, and run at low
163 speed until 10,000 events had been registered. Bacteria were identified by their signature in a plot of
164 side scatter (SSC) versus green fluorescence (FL1). Virus-like particles were glutaraldehyde-fixed
165 (0.5% final concentration) and stored as described for the bacterial samples. Unfrozen samples were
166 further diluted in TE buffer, stained with SYBR Green I for 10 min at 80°C in the dark., cooled, and
167 estimated by cytometry (Brussaard, 2004).

168

169

RESULTS

170

Experiment I: *Pseudo-nitzschia delicatissima* and DOM

171
172
173
174
175
176
177
178
179
180
181
182
183
184
185
186
187
188
189
190
191
192
193
194
195
196
197
198
199
200
201
202
203
204
205

A general decreasing trend in ammonium concentration was observed in all incubations (Fig. 1a), reaching a minimum value in (L1-N)+DOM ($1.05 \pm 0.65 \mu\text{M}$). Nitrate consumption was accelerated at the end of the experiment in L1/5+DOM, in the DOM control (L1-DOM), and in the N-deficient incubation [(L1-N)+DOM], whereas in the BV+DOM control nitrate concentrations oscillated, with no clearly defined tendency (Fig. 1b, c). Phosphate slowly accumulated in (L1-N)+DOM (Fig. 1d), whereas the opposite pattern was observed in L1/5+DOM; in L1-DOM, phosphate peaked at day 4 and decreased thereafter. In the BV+DOM incubation, phosphate decreased at the beginning and end of the experiment, while in mid-experiment (days 4 – 8) it accumulated. In most incubations, there was an initial and final decrease in silicate concentrations (Fig. 1e) (days 0 – 2, and 8 – 10), with an increase during the middle stage of the experiment (days 2 – 8). The only exception was BV+DOM, in which no specific trend was observed. The extracted HMWDOM used for enrichment had an organic molar C:N ratio of 19. DOC concentrations were mostly stable throughout the course of the experiment (Fig. 2a). DON accumulated by the end of the experiment in L1/5+DOM and L1-DOM, with a marked increment by day 8 in the latter (Fig. 2b). In (L1-N)+DOM, DON remained constant throughout the incubation while in BV+DOM a peak was registered by day 2. DOP also peaked at day 4 in BV+DOM (Fig. 2c). In L1/5+DOM, DOP gradually increased after day 4, while in L1-DOM and (L1-N)+DOM an initial uptake period occurred, lasting until day 4 after which the DOP concentration did not vary greatly.

The growth of *Pseudo-nitzschia delicatissima* cells was similar in all experimental sets (1.64 ± 0.05 division day^{-1}) (Fig. 3). However, chl *a* concentrations markedly differed after day 4, reaching higher levels for L1-DOM and lower levels for (L1-N)+DOM by day 10 (32.6 ± 0.7 and $7.1 \pm 0.1 \mu\text{g L}^{-1}$, respectively), thus accounting for a difference in chl *a* content (320 and 207 fg chl *a* per cell on day 10, respectively). This 1.5-fold difference was already present at the beginning of the incubation period. Total bacteria (Fig. 4) rose steadily until day 6 (or day 8 in L1/5+DOM treatment), after which the number of cells declined. An exception was noted in the L1-DOM control, in which the initial increment until day 4 was followed by a second increase (days 6 – 10). Bacterial growth remained low during the experiment. Viral abundance (Fig. 5) had small variations in L1/5+DOM and (L1-N)+DOM, although in BV+DOM a peak was observed by day 4. In L1-DOM, low levels of virus were maintained until day 10, when a small increase was noted. Although the seawater sample used to extract HMWDOM was pre-filtered through a 0.2- μm filter and the concentrated HMWDOM immediately used for the enrichment experiment, the numbers of bacteria and virus in the DOM control (L1-DOM) vs. the remaining incubations at the beginning of the experiment differed, testifying to the presence of bacteria and viruses in the DOM extract.

206

207

Experiment II: *Pseudo-nitzschia delicatissima*: ammonium and urea uptake

208

209

210

211

212

213

214

215

216

217

218

219

220

221

222

223

224

225

226

227

228

229

230

231

232

233

234

235

DISCUSSION

236

237

Methodological aspects

238

239

240

C:N particulate molar ratios from *Pseudo-nitzschia delicatissima* cells were mainly stable during this experiment (19.4 ± 4.8 and 22.3 ± 5.0 for incubations enriched with ^{15}N -ammonium and ^{15}N -urea, respectively) due to the short incubation time (1 h). Resuspension of the cells for 6 h resulted in the development of a growth medium with very low ammonium levels ($< 0.1 \mu\text{M}$), as expected. Nitrate concentrations were low and varied only slightly ($3.8 \pm 0.3 \mu\text{M}$) in all of the experimental sets, indicating low uptake of this N source and either non-interference or a constant rate of interference of nitrate to either ammonium or urea uptake. The detected nitrate levels also excluded possible modifications to N metabolism by N-limiting conditions.

The linear increase in cellular ^{15}N -ammonium and ^{15}N -urea incorporation along a temporal axis (Fig. 6) reflected the constant uptake rate of both N sources during the time elapsed, with higher uptake values for ^{15}N -ammonium. The increase in cell ^{15}N for the ammonium experiment presents a slightly concave shape, with a pronounced decrease in uptake after 70 minutes. However, those trends should not lead to significant deformation of the kinetics curve shown in Fig. 7a, for which data were obtained with an incubation time of 1 h.

The $V_{\text{N-ammonium}}$ and $V_{\text{N-urea}}$ along the concentration gradient were fitted to the Michaelis-Menten model (Fig. 7a, b). There are some indications of multiphasic uptake with an initial phase transition between 3 and 5 $\mu\text{M-N}$ for both N sources. In particular, for ammonium, a single Michaelis-Menten model clearly tended to overestimate uptake between 1 and 5 $\mu\text{M-N}$ and to underestimate it at higher substrate concentrations. The existence of multiphasic uptake systems was confirmed by the discrepancy noted in the kinetic parameters values obtained from the two substrate concentration ranges, 0.1 – 3 and 0.1 – 10 $\mu\text{M-N}$ (Table I). For both V_{max} and K_s , values for ammonium were higher than those for urea. Still, similar inorganic carbon uptake rates were attained for cells grown in either ^{15}N source, implying comparable photosynthetic rates (Fig. 8). The C:N uptake ratios were nearly 8 for the ammonium series, indicating balanced growth, and nearly 100 for the urea series, clearly revealing a deficit in nitrogen supply in those samples.

Bioassays consist of manipulated systems in which a limited number of variables are controlled in order to clarify specific aspects of an entire ecosystem that are otherwise difficult to

241 address. It should be noted, however, that this approach is not without disadvantages. The
242 confinement of microorganisms and thus their disruption from natural physical (e.g., mixing,
243 advection, sedimentation), chemical (e.g., nutrient sources and feedback mechanisms), and
244 biological (e.g., grazing, competition, allelopathy) processes limits extrapolation of the outcome to
245 nature (Howarth, 1988). However, organic nutrient dynamics are too complex to study in
246 multispecific natural assemblages, where interactions between organisms and environmental
247 parameters constantly modify the availability and nature of organic matter (Hansell and Carlson,
248 2002). In this investigation, the use of unialgal cultures allowed us to follow the evolution of
249 *Pseudo-nitzschia delicatissima* cells in the presence of DOM. This approach also proved useful to
250 explore the organic nutritional traits of several microalgae species (Doblin *et al.*, 1999; Stolte *et al.*,
251 2006).

252 Bacteria and viruses are recognized as important elements of the microbial food web,
253 influencing nutrient biogeochemical cycles and availability to higher trophic levels (Azam *et al.*,
254 1983; Fuhrman, 1999). The fact that 0.2 μm filtration and DOM extraction techniques failed to
255 efficiently remove 100% of the bacterial and viral assemblages from concentrated seawater
256 highlights the need to follow up these variables in similar experimental sets in order to understand
257 both holistic population interactions and systemic evolution.

258

259

DOM enrichment: experiment I

260

261 Although the pulsed nitrate was readily consumed, resulting in a generally stable
262 concentration of this nutrient, an increase in uptake was apparent by the end of the experiment,
263 coincident with a decrease in the ammonium concentration to below 1.4 – 2.5 μM . This result can
264 be explained by ammonium-metabolite inactivation of nitrate uptake (Flynn, 1991) at ammonium
265 levels higher than this threshold value, as has been previously observed in several phytoplankton
266 species (Dortch, 1990). Inorganic N-sources may also have been used by the bacterial assemblage,
267 leading to variations in the N concentrations measured in the BV+DOM control, in which no
268 microalgae cells were present. Indeed, in agreement with C:N stoichiometric balance, organic
269 substrates with high C:N ratios indicate that bacteria are forced to use inorganic nutrients as a
270 nitrogen source (Goldman and Dennett, 2000). The uptake of dissolved inorganic nutrients (DIN) to
271 promote bacterial growth and the consequent competition between phytoplankton and bacteria for
272 DIN were demonstrated in previous experiments and often lead to low bacterial growth (Davidson
273 *et al.*, 2007; Joint *et al.*, 2002). In bacteria, nitrate is considered to serve only as a secondary N
274 source because most species lack nitrate reductase, implying a high energetic cost (extra organic
275 material) to assimilate this inorganic nutrient (Joint *et al.*, 2002). Phosphate and silicate were in

276 excess throughout the experiment, reasonably precluding nutrient-limiting conditions for these
277 nutrients.

278 The C:N molar ratio of the extracted HMWDOM (C:N = 19) was within the range of
279 autotrophic DOM (produced by phytoplankton) (Biddanda and Benner, 1997) and reflected the
280 production of carbon-rich organic matter associated with actively growing diatoms or diatoms at the
281 decline of a bloom (Nagata, 2000; Wetz and Wheeler, 2007). The ratio was also very similar to the
282 mean DOC/DON (17) measured just outside of Thau lagoon (Souchu *et al.*, 1997), where *Pseudo-*
283 *nitzschia* species also grow to bloom proportions. The constant level of organic resources observed
284 for most incubations provides evidence for a rapid coupling between the production and uptake of
285 labile (turnover times of days) organic matter. Nonetheless, there were also episodes in which
286 organic compounds accumulated. DOM accumulation generally occurs when production and
287 consumption mechanisms are decoupled (Biddanda & Benner 1997). Various conditions may arise,
288 simultaneously or exclusively, that lead to decoupling: (1) impaired increase of organic matter
289 release and microbial uptake, (2) increase in the presence of semi-labile (turnover times of months)
290 or refractory (turnover times of years) DOM, (3) inhibition of biological decomposition processes
291 by nutrient limitation, (4) bacterial biomass control through grazing or viral lysis (Williams 1995;
292 S ndergaard *et al.* 2004). In the present investigation, organic matter accumulated in distinct
293 contexts, - e.g., DON and DOP peaks in BV+DOM were concurrent with an increase in the
294 concentration of virus; the DON increase in L1/5+DOM on day 10 was coincident with a decrease
295 in bacterial abundance - thereby implying the occurrence of conditions (1), (2), (3), and (4) on
296 different occasions throughout the course of the experiment. However, the estimation of DON by
297 subtracting DIN from TDN produces the accumulation of analytical errors (with respect to TDN,
298 ammonium, nitrite and nitrate) that may lead to scattering of the data (Berman & Bronk 2003).
299 Indeed, the DON accumulation detected in the L1-DOM incubation from days 6 to 8 (Fig. 2) gave
300 rise to a low DOC:DON, which, according to published literature, is not feasible (Bronk, 2002). The
301 fact that, in L1-DOM, nitrate uptake per unit chl *a* during days 6 – 8 was above the maximum
302 possible (Collos *et al.*, 2005) points to an incorrect estimation of this parameter on this occasion.

303 Although the chl *a* concentrations in (L1-N)+DOM were lower than in the other treatments,
304 *Pseudo-nitzschia delicatissima* abundance was similar in every incubation. Throughout the
305 experiment, the chl *a* cell content in (L1-N)+DOM was lower than in L1-DOM, which is consistent
306 with DIN limitation (Caperon and Meyer, 1972; Thomas and Dodson, 1972). This may also indicate
307 mixotrophic growth, whereby some or most of the nitrogen (and carbon) would have come from
308 DOM, thus reducing the need for photosynthetic pigments while maintaining comparable cell
309 concentrations (Fig. 3). Similar trends were observed in Chrysophyceae (Lewitus and Caron, 1991),
310 in which glycerol addition led to a two-fold reduction in cellular chl *a* content while the growth rate

311 increased. Moreover, the fact that DON from (L1-N)+DOM was constant during the entire
312 experiment implies a fast uptake of organic nitrogen inputs. Overall, although unialgal cultures do
313 not allow the inference of straight-forward utilization of organic nutrient sources by microalgae, one
314 is tempted to suggest that organic nitrogen provides, directly or indirectly, a nutrient source to
315 maintain *P. delicatissima* at high concentrations.

316 In the DOM control (L1-DOM), virus-like particles remained at a minimum until bacteria
317 reached their maximum abundance, by the end of the experiment, implying that viruses need a
318 minimum host level to trigger an infection cycle (Weinbauer, 2004). The decrease in bacterial
319 numbers in DOM-enriched treatments over the last days (days 6 – 10) of this study likely reflected
320 the interplay between viral lysis and (as already discussed) nutrient limitation (Fuhrman, 1999;
321 Williams, 1995).

322

323 **Ammonium and urea uptake: experiment II**

324

325 Although cells were subject to a 6-h starvation period, no pre-conditioning (Dortch *et al.*,
326 1991) or surge uptake (Conway *et al.*, 1976) effects were visible in temporal incubations of ¹⁵N-
327 ammonium or ¹⁵N-urea (Fig. 6). The V_{\max} for ammonium (about 1.4 d⁻¹) was similar to the growth
328 rate (about 1.2 d⁻¹), confirming the absence of a surge uptake at the time scale of measurement used
329 here. A C:N uptake ratio near the Redfield ratio provided further evidence of balanced growth.
330 Indeed, the trend in *P. delicatissima* may be towards uptake rather than storage, which, in turn,
331 allows a rapid assimilation of nutrients into the build-up of biomass (Collos, 1986; Tilman *et al.*,
332 1982). These data concur with the classification of *Pseudo-nitzschia delicatissima* as an r-strategist
333 (Quijano-Scheggia *et al.*, 2008b).

334 The release of ammonium during urea assimilation was observed previously in microalgae
335 and often implies the presence of interacting mechanisms to regulate the uptake of these N
336 substrates (Jauzein *et al.*, 2008). Nevertheless, in this experiment, ammonium excretion was not
337 apparent during the assimilation of urea, likely excluding interactions between ammonium and urea
338 upon their uptake. The kinetic parameters were within the range determined for coastal
339 phytoplankton assemblages and cultured neritic diatoms (Eppley *et al.*, 1969; Kudela and Cochlan,
340 2000 ref. therein). More specifically, for ammonium, the K_s value was similar to that of *P. australis*

341 (Cochlan *et al.*, in press) when the entire concentration range was taken into account. There is also
342 some evidence for multiphasic uptake in data from Cochlan *et al.* (in press) describing ammonium
343 uptake by *Pseudo-nitzschia australis*, with phase transitions between 2 and 5 μM . Those values are
344 similar to the ones reported here, but the authors did not attempt to estimate K_s at the lower
345 concentration range. Concerning urea, the saturation kinetics observed here for *P. delicatissima*
346 were very different from the linear response reported by Cochlan *et al.* (in press) for *P. australis*.
347 The latter is more similar to that seen in some strains of *Alexandrium catenella* (Jauzein *et al.*
348 2008). At 10 $\mu\text{M-N}$, there was a 30-fold difference between ammonium and urea uptake in our
349 study vs. the five-fold difference reported by Cochlan *et al.* In both species of *Pseudo-nitzschia*,
350 ammonium is clearly preferred to urea.

351 The preference for ammonium is often related to transport mechanisms and the biological
352 energetic costs involved in the synthesis of urease, needed for the assimilation of urea (Berman and
353 Bronk, 2003). However, the K_s values suggested that during periods of low nitrogen concentration
354 *P. delicatissima* may turn to urea as an alternative N source, as already observed for other diatoms
355 (Rees and Syrett, 1979). In our case (Fig. 1a), the rather high K_s for ammonium indicated that
356 *Pseudo-nitzschia* spp. could not compete with bacteria for this N source (Fuhrman *et al.*, 1988) and
357 was thus probably forced to use organic N compounds such as urea for growth. Similar carbon-
358 uptake levels point to comparable photosynthetic rates in *P. delicatissima* growing on either
359 substrate, possibly due to the fact that urea provides both nitrogen and carbon to the cells (Berg *et al.*
360 *et al.*, 1997). Previous work evidenced the ability of *Pseudo-nitzschia australis* to grow equally well
361 on urea as on other inorganic N sources (Howard *et al.*, 2007).

362 363 **Summary and possible implications for the natural environment** 364

365 The results of experiments I and II suggest a potential use of organic sources for the growth
366 of *Pseudo-nitzschia delicatissima*. The presence of autotrophic DOM likely favors the development
367 of *Pseudo-nitzschia delicatissima* during shortage of mineral nutrients. In natural environments,
368 such conditions may be met after blooming species have depleted inorganic resources and a high
369 content of exudated DOM is available (Sondergaard *et al.*, 2004). Indeed, in situ proliferations of *P.*
370 *delicatissima* are often observed in the sequence of blooms of other diatoms, such as *Chaetoceros*
371 spp. and *Leptocylindrus* spp. (Casas *et al.*, 1999; Quijano-Scheggia *et al.*, 2008a), or they may be
372 monospecific, with blooms of *P. delicatissima* following those of *P. calliantha*, such as occurs in

373 Mediterranean areas (Caroppo *et al.*, 2005; Quijano-Scheggia *et al.*, 2008a), whereas in Korean
374 waters *P. delicatissima* blooms occur after *P. pungens* events (Cho *et al.*, 2002).

375 It has been reported that natural assemblages of *Pseudo-nitzschia* spp. growing on
376 predominantly urea-based regimens produce more domoic acid than when growing on inorganic
377 nitrogen compounds (Howard *et al.*, 2007). The increasing loads of urea deposited into coastal
378 waters (Glibert *et al.*, 2006) worldwide could therefore imply an increase in the toxicity of *Pseudo-*
379 *nitzschia delicatissima* blooms. Conclusive evidence is nevertheless only possible by direct tests on
380 toxic blooming strains, due to the diverse nutritional mechanisms of *P. delicatissima* cryptic
381 complexes.

382 The wide temporal blooming season of *Pseudo-nitzschia delicatissima* has been explained
383 by the succession of multiple genotypes of this species (Orsini *et al.*, 2004). It can therefore be
384 hypothesized that strains able to profit from the presence of organic substrates thrive during periods
385 of low inorganic nutrient concentrations, thus sustaining *P. delicatissima* blooms over prolonged
386 periods. Furthermore, the overwhelming presence of *P. delicatissima* in mucilaginous
387 macroaggregates (Totti *et al.*, 2005) together with the colonization by this diatom of large
388 *Phaeocystis* spp. colonies (Sazhin *et al.*, 2007) also suggests an opportunistic ability of *P.*
389 *delicatissima* to explore environments with a high organic nutrient content.

390 Overall, present and past research points to the ability of certain *Pseudo-nitzschia*
391 *delicatissima* strains to benefit from the presence of organic substrates. Organic sources could as
392 such represent the “missing link” to explain the growth of *P. delicatissima* when other factors fail to
393 do so (Fehling *et al.*, 2005; Quijano-Scheggia *et al.*, 2008b). In a broader context, increasing
394 evidence of the influence of organic substrates on the growth of several microalgae, including
395 potentially harmful species, calls for this parameter to be included in monitoring programs and
396 taken into account in the establishment of management and policy initiatives.

397

398 **Acknowledgements.** We are grateful to R. Ventosa and M. I. Abad for the nutrient analyses as well
399 as to Dr. E. Berdalet and V. Pérez for analyses of TOC. We acknowledge Dr. Sonia Quijano for
400 help during the ammonium and urea uptake experiment. We are also grateful to two anonymous
401 referees for critical review and useful comments. This study was partially funded by the EC-funded
402 Research Project SEED (Life cycle transformations among HAB species, and the environmental
403 and physiological factors that regulate them; GOCE-CT-2005-003875), the contract between ACA
404 and CSIC, and by PROCAVIR (CTM2004-04404-CO2-01) and MICROVIS (CTM2007-62140)
405 projects. S. Loureiro was supported by a FCT (Fundação para a Ciência e para a Tecnologia,
406 Portugal) grant within the III Quadro Comunitário de Apoio by the FSE; E. Garcés work was
407 supported by the Ramon y Cajal contract of the Spanish Ministry of Education and Science.

408 Meteorological data were kindly provided by the regional meteorological service METEOCAT.
409 Phytoplankton data for Blanes Bay were kindly provided by the Microbial Observatory of Blanes
410 Bay net (MOBB). C. Jauzein was supported by an Ifremer-Région Languedoc-Roussillon
411 scholarship. Y. Collos was supported by CNRS.
412

- 414 Addison, R. F. and Stewart, J. E. (1989) Domoic acid and the eastern Canadian molluscan shellfish
415 industry *Aquaculture*, **77**, 263-269.
- 416 Álvarez-Salgado, X. A. and Miller, A. E. J. (1998) Simultaneous determination of dissolved organic
417 carbon and total dissolved nitrogen in seawater by high temperature catalytic oxidation:
418 conditions for accurate shipboard measurements. *Mar. Chem.*, **62**, 325-333.
- 419 Andersen, O. and Throndsen, J. (2003) Estimating cell numbers. In Hallegraeff, G. M. and
420 Anderson, D. M. (eds), *Manual on harmful marine microalgae*. Unesco Publishing, Paris,
421 pp. 99-129.
- 422 Azam, F., Fenchel, T., Field, J. G., Gray, J. S., Meyer-Reil, L. A. and Thingstad, F. (1983) The
423 Ecological Role of Water-Column Microbes in the Sea. *Mar. Ecol. Prog. Ser.*, **10**, 257-263.
- 424 Bates, S. S., Garrison, D. L. and Horner, R. A. (1998) Bloom dynamics and physiology of domoic-
425 acid producing *Pseudo-nitzschia* species. In Anderson, D. M., Cembella, A. D. and
426 Hallegraeff, G. M. (eds), *Physiological ecology of harmful algal blooms*. Springer-Verlag,
427 Heidelberg, pp. 267-292.
- 428 Berg, G. M., Glibert, P. M., Lomas, M. W. and Burford, M. A. (1997) Organic nitrogen uptake and
429 growth by the chrysophyte *Aureococcus anophagefferens* during a brown tide. *Mar. Biol.*,
430 **129**, 377-387.
- 431 Berman, T. and Bronk, D. A. (2003) Dissolved organic nitrogen: a dynamic participant in aquatic
432 ecosystems. *Aquat. Microb. Ecol.*, **31**, 279-305.
- 433 Biddanda, B. and Benner, R. (1997) Carbon, nitrogen and carbohydrate fluxes during the
434 production of particulate and dissolved organic matter by marine phytoplankton. *Limnol.*
435 *Oceanogr.*, **42**, 506-518.
- 436 Bronk, D. A. (2002) Dynamics of DON. In Hansell, D. A. and Carlson, C. A. (eds),
437 *Biogeochemistry of marine dissolved organic matter*. Academic Press, Amsterdam, pp. 153-
438 247.
- 439 Brussaard, C. P. D. (2004) Optimization of Procedures for Counting Viruses by Flow Cytometry.
440 *Appl. Environ. Microbiol.*, **70**, 1506-1513.
- 441 Caperon, J. and Meyer, J. (1972) Nitrogen-limited growth of marine phytoplankton. 1. Changes in
442 population characteristics with steady-state growth rate. *Deep Sea Res.*, **19**, 601-618.
- 443 Caroppo, C., Congesti, R., Bracchini, L. and Albertano, P. (2005) On the presence of *Pseudo-*
444 *nitzschia calliantha* Lundholm, Moestrup et Hasle and *Pseudo-nitzschia delicatissima*
445 (Cleve) Heiden in the Southern Adriatic Sea (Mediterranean Sea, Italy). *J. Plankton Res.*,
446 **27**, 763-774.
- 447 Casas, B., Varela, M. and Bode, A. (1999) Seasonal succession of phytoplankton species on the
448 coast of A Coruña (Galicia, northwest Spain). *Bol. Inst. Esp. Oceanogr.*, **15**, 413-429.
- 449 Cho, E. S., Hur, H. J., Byun, H. S., Lee, S. G., Rhodes, L. L., Jeong, C. S. and Park, J. G. (2002)
450 Monthly monitoring of domoic acid producer *Pseudo-nitzschia multiseries* (Hasle) Hasle
451 using species-specific DNA probes and WGA lectins and abundance of *Pseudo-nitzschia*
452 species (Bacillariophyceae) from Chinhae Bay, Korea. *Bot. Mar.*, **45**, 364-372.
- 453 Cochlan, W. P., Herndon, J. and Kudela, R. M. (in press) Inorganic and Organic Nitrogen Uptake
454 by the Toxigenic Diatom *Pseudo-nitzschia australis* (Bacillariophyceae). *Harmful Algae*
- 455 Collos, Y. (1986) Time-lag algal growth dynamics: biological constraints on primary productivity in
456 aquatic environments. *Mar. Ecol. Prog. Ser.*, **33**, 193-206.
- 457 Collos, Y., Vaquer, A. and Souchu, P. (2005) Acclimation of nitrate uptake by phytoplankton to
458 high substrate levels. *J. Phycol.*, **41**, 466-478.
- 459 Collos, Y., Vaquer, A., Laabir, M., Abadie, E., Laugier, T., Pastoureaud, A. and Souchu, P. (2007)
460 Contribution of several nitrogen sources to growth of *Alexandrium catenella* during blooms
461 in Thau Lagoon, southern France. *Harmful Algae*, **6**, 781-789.

- 462 Conway, H. L., Harrison, P. J. and Davis, C. O. (1976) Marine Diatoms Grown in Chemostats
463 under Silicate or Ammonium Limitation. II. Transient Response of *Skeletonema costatum* to
464 a Single Addition of the Limiting Nutrient. *Mar. Biol.*, **35**, 187-199.
- 465 Davidson, K., Gilpin, L. C., Hart, M. C., Fouilland, E., Mitchell, E., Calleja, I. A., Laurent, C.,
466 Miller, A. E. J. and Leakey, R. J. G. (2007) The influence of the balance of inorganic and
467 organic nitrogen on the trophic dynamics of microbial food webs. *Limnol. Oceanogr.*, **52**,
468 2147-2163.
- 469 Doblin, M. A., Blackburn, S. I. and Hallegraeff, G. M. (1999) Growth and biomass stimulation of
470 the toxic dinoflagellate *Gymnodinium catenatum* (Graham) by dissolved organic substances.
471 *J. Exp. Mar. Biol. Ecol.*, **236**, 33-47.
- 472 Dortch, Q. (1990) The interaction between ammonium and nitrate uptake in phytoplankton. *Mar.*
473 *Ecol. Prog. Ser.*, **61**, 183-201.
- 474 Dortch, Q., Thompson, P. A. and Harrison, P. J. (1991) Short-term interaction between nitrate and
475 ammonium uptake in *Thalassiosira pseudonana*: effect of preconditioning nitrogen source
476 and growth rate. *Mar. Biol.*, **110**, 183-193.
- 477 Doucette, G. J. and Powell, C. L. (1998) Algal-Bacterial interactions: can they determine the PSP-
478 related toxicity of dinoflagellates? In Reguera, B., Blanco, J., Fernandez, M. L. and Wyatt,
479 T. (eds), *Harmful algae*. Xunta de Galicia & Intergov. Oceanog. Comm. UNESCO, pp. 406-
480 409.
- 481 Eppley, R. W., Rogers, J. N. and McCarthy, J. J. (1969) Half-saturation constants for uptake of
482 nitrate and ammonium by marine phytoplankton. *Limnol. Oceanogr.*, **14**, 912-920.
- 483 Fehling, J., Davidson, K. and Bates, S. S. (2005) Growth dynamics of non-toxic *Pseudo-nitzschia*
484 *delicatissima* and toxic *P. seriata* (Bacillariophyceae) under simulated spring and summer
485 photoperiods. *Harmful Algae*, **4**, 763-769.
- 486 Flynn, K. J. (1991) Algal carbon-nitrogen metabolism: a biochemical basis for modelling the
487 interactions between nitrate and ammonium uptake. *J. Plankton Res.*, **13**, 373-387.
- 488 Fryxell, G. A., Villac, M. C. and Shapiro, L. P. (1997) The occurrence of the toxic diatom genus
489 *Pseudo-nitzschia* (Bacillariophyceae) on the West Coast of the USA, 1920-1996: a review.
490 *Phycologia*, **36**, 419-437.
- 491 Fuhrman, J. A., Horrigan, S. G. and Capone, D. G. (1988) Use of ¹³N as tracer for bacterial and
492 algal uptake of ammonium from seawater. *Mar. Ecol. Prog. Ser.*, **45**, 271-278.
- 493 Fuhrman, J. A. (1999) Marine viruses and their biogeochemical and ecological effects. *Nature.*, **399**,
494 541-548.
- 495 Gasol, J. M. and delGiorgio, P. A. (2000) Using flow cytometry for counting natural planktonic
496 bacteria and understanding the structure of planktonic bacterial communities. *Sci. Mar.*, **64**,
497 197-224.
- 498 GEOHAB (eds) (2005) *Global Ecology and Oceanography of Harmful Algal Blooms, GEOHAB*
499 *Core Research Project: HABs in Upwelling Systems*. IOC and SCOR, Paris and Baltimore
- 500 Glibert, P. M., Harrison, J., Heil, C. and Seitzinger, S. (2006) Escalating worldwide use of urea - a
501 global change contributing to coastal eutrophication. *Biogeochemistry*, **77**, 441-463.
- 502 Glibert, P. M. and Legrand, C. (2006) The diverse nutrient strategies of Harmful Algae: Focus on
503 osmotrophy. In Granéli, E. and Turner, J. T. (eds), *Ecology of Harmful Algae*. . 189.
504 Springer pp. 163-175.
- 505 Goeyens, L., Kindermans, N., Abu-Yusuf, M. and Elskens, M. (1998) A room temperature
506 procedure for the manual determination of urea in seawater. *Est. Coast. Shelf Sci.*, **47**, 415-
507 418.
- 508 Goldman, J. C. and Dennett, M. R. (2000) Growth of marine bacteria in batch and continuous
509 culture under carbon and nitrogen limitation. *Limnol. Oceanogr.*, **45**, 789-800.
- 510 Grasshoff, K., Ehrhardt, M. and Kremling, K. (eds) (1983) *Methods of Seawater Analysis*. Verlag-
511 Chemie, Weinheim, Germany
- 512 Grasshoff, K., Kremling, K. and Ehrhardt, M. (eds) (1999) *Methods of Seawater Analysis* Wiley-
513 VCH, Weinheim

- 514 Guillard, R. R. L. and Hargraves, P. E. (1993) *Stichochrysis immobilis* is a diatom, not a
515 chrysophyte. *Phycologia*, **32**, 234-236.
- 516 Hansell, D. A. and Carlson, C. A. (eds) (2002) *Biogeochemistry of Marine Dissolved Organic*
517 *Matter*.
- 518 Hillebrand, H. and Sommer, U. (1996) Nitrogenous nutrition of the potentially toxic diatom
519 *Pseudo-nitzschia pungens* f. *multiseriis* Hasle. *J. Plankton Res.*, **18**, 295-301.
- 520 Howard, M. D. A., Cochlan, W. P., Ladizinsky, N. and Kudela, R. M. (2007) Nitrogenous
521 preference of toxigenic *Pseudo-nitzschia australis* (Bacillariophyceae) from field and
522 laboratory experiments. *Harmful Algae*, 206-217.
- 523 Howarth, R. W. (1988) Nutrient limitation of net primary production in marine ecosystems. *Ann.*
524 *Rev. Ecol.*, **19**, 89-110.
- 525 Hurd, C. L. and Dring, M. J. (1990) Phosphate uptake by intertidal algae in relation to zonation and
526 season. *Mar. Biol.*, **107**, 281-289.
- 527 Jauzein, C., Collos, Y., Garcés, E., Vila, M. and Masó, M. (2008) Short-term temporal variability of
528 ammonium and urea uptake by *Alexandrium catenella* (Dinophyta) in cultures. *J. Phycol.*,
529 **44**, 1136-1145.
- 530 Joint, I. P., Hendriksen, P., Fonnes, G. A., Bourne, D., Thingstad, T. F. and Riemann, B. (2002)
531 Competition for inorganic nutrients between phytoplankton and bacterioplankton in nutrient
532 manipulated mesocosms. *Aquat. Microb. Ecol.*, **29**, 145-159.
- 533 Kudela, R., Pitcher, G., Probyn, T., Figueiras, F., Moita, T. and Trainer, V. (2005) Harmful Algal
534 Blooms in Coastal Upwelling Systems. *Oceanography*, **18**, 185-197.
- 535 Kudela, R. M. and Cochlan, W. P. (2000) Nitrogen and carbon uptake kinetics and the influence of
536 irradiance for a red tide bloom off southern California. *Aquat. Microb. Ecol.*, **21**, 31-47.
- 537 Lewin, J. and Hellebust, J. A. (1976) Heterotrophic Nutrition of the Marine Pennate Diatom
538 *Nitzschia angularis* var. *affinis*. *Mar. Biol.*, **36**, 313-320.
- 539 Lewitus, A. J. and Caron, D. A. (1991) Physiological responses of phytoflagellates to dissolved
540 organic substrate additions. 1. Dominant role of heterotrophic nutrition in
541 *Poterioochromonas malhamensis* (Chrysophyceae). *Plant Cell Physiol.*, **32**, 671-680.
- 542 Mengelt, C. and Prézelin, B. B. (2002) A Potential Novel Link between Organic Nitrogen Loading
543 and *Pseudo-Nitzschia* spp. Blooms, Ed, Vol 20. Proc. California World Ocean Conf., Sta
544 Barbara California
- 545 Nagata, T. (2000) Production mechanisms of dissolved organic matter. In Kirchman, D. L. (ed),
546 *Microbial Ecology of the Oceans*. Wiley-Liss, New York, pp. 121-152.
- 547 Orsini, L., Procaccini, G., Sarno, D. and Montresor, M. (2004) Multiple rDNA ITS-types within
548 diatom *Pseudo-nitzschia delicatissima* (Bacillariophyceae) and their relative abundances
549 across a spring bloom in the Gulf of Naples. *Mar. Ecol. Prog. Ser.*, **271**, 87-98.
- 550 Quijano-Scheggia, S., Garcés, E., Flo, E., Fernández-Tejedor, M., Diogène, J. and Camp, J. (2008a)
551 Bloom dynamics of the genus *Pseudo-nitzschia* (Bacillariophyceae) in two coastal bays, NE
552 Spain (Mediterranean Sea). *Sci. Mar.*, **72**, 577-590.
- 553 Quijano-Scheggia, S., Garcés, E., Sampedro, N., van Lenning, K., Flo, E., Andree, K., Fortuño, J.
554 M. and Camp, J. (2008b) Identification and characterisation of the dominant *Pseudo-*
555 *nitzschia* species (Bacillariophyceae) along the NE Spanish coast (Catalonia, NW
556 Mediterranean). *Sci. Mar.*, **72** 343-359.
- 557 Rees, T. A. V. and Syrett, P. J. (1979) The uptake of urea by the diatom, *Phaeodactylum* *New*
558 *Phytol.*, **82**, 169-178.
- 559 Sazhin, A. F., Artigas, L. F., Nejstgaard, J. C. and Frischer, M. E. (2007) The colonization of two
560 *Phaeocystis* species (*Prymnesiophyceae*) by pennate diatoms and other protists: a significant
561 contribution to colony biomass. *Biogeochemistry*, **83**, 137-145.
- 562 Seitzinger, S. P. and Sanders, R. W. (1999) Atmospheric inputs of dissolved organic nitrogen
563 stimulate estuarine bacteria and phytoplankton. *Limnol. Oceanogr.*, **44**, 721-730.
- 564 Smayda, T. J. (1997) Harmful algal blooms: their ecophysiology and general relevance to
565 phytoplankton blooms in the sea. *Limnol. Oceanogr.*, **42**, 1137-1153.

- 566 Sondergaard, M., Thingstad, F., Stedmon, C., Kragh, T. and Cauwet, G. (2004) DOM sources and
567 microbes in lakes and coastal waters. In Sondergaard, M. and Thomas, D. N. (eds),
568 *Dissolved Organic Matter (DOM) in Aquatic Ecosystems. A Study of European Catchments*
569 *and Coastal Waters*. The Domaine project, pp. 23-36.
- 570 Souchu, P., Gasc, A., Cahet, G., Vaquer, A., Collos, Y. and Deslous-Paoli, J. M. (1997)
571 Biogeochemical composition of mediterranean waters outside the Thau lagoon. *Est. Coast.*
572 *Shelf Sci.*, **44**, 275-284.
- 573 Stolte, W., Balode, M., Carlsson, P., Grzebyk, D., Janson, S., Lips, I., Panosso, R., Ward, C. J. and
574 Granéli, E. (2006) Stimulation of nitrogen-fixing cyanobacteria in a Baltic Sea plankton
575 community by land-derived organic matter or iron addition. *Mar. Ecol. Prog. Ser.*, **327**, 71-
576 82.
- 577 Thomas, W. H. and Dodson, A. N. (1972) On nitrogen deficiency in tropical oceanic
578 phytoplankton. II. Photosynthetic and cellular characteristics of a chemostat-grown diatom. .
579 *Limnol. Oceanogr.*, **17**, 515-523.
- 580 Tilman, D., Kilham, S. S. and Kilham, P. (1982) Phytoplankton community ecology: the role of
581 limiting resources. *Annu. Rev. Ecol. Syst.*, **13**, 349-372.
- 582 Totti, C., Cangini, M., Ferrari, C., Kraus, R., Pompei, M., Puggnetti, A., Romagnoli, T., Vanucci, S.
583 and Socal, G. (2005) Phytoplankton size-distribution and community structure in relation to
584 mucilage occurrence in the northern Adriatic Sea. *Sci. Total Environ.*, **353**, 204-217.
- 585 Weinbauer, M. G. (2004) Ecology of prokaryotic viruses. *FEMS Microbiol. Rev.*, **28**, 127-181.
- 586 Wetz, M. S. and Wheeler, P. A. (2007) Release of dissolved organic matter by coastal diatoms.
587 *Limnol. Oceanogr.*, **52**, 798-807.
- 588 Williams, P. J. I. e. B. (1995) Evidence for the seasonal accumulation of carbon-rich dissolved
589 organic matter, its scale in comparison with changes in particulate material and
590 consequential effect on net C/N assimilation ratios. *Mar. Chem.*, **51**, 17-29.
- 591 Yentsch, C. S. and Menzel, D. W. (1963) A method for the determination of phytoplankton
592 chlorophyll and phaeophytin by fluorescence. *Deep Sea Res.*, **10**, 221-231.
- 593
594

594

595

596 Table I. Kinetic parameters of ammonium and urea uptake by *Pseudo-nitzschia delicatissima*
 597 cultures. Range: substrate concentrations, $\mu\text{M-N}$; K_s : half-saturation constant, $\mu\text{M-N}$; V_{max} :
 598 maximum uptake rate, h^{-1} ; r^2 : coefficient of determination; p: probability value; α initial slope, L h^{-1}
 599 $\mu\text{gat N}^{-1}$.

| | Range | K_s | V_{max} | r^2 | p | α |
|----------|----------|-------|------------------|-------|-------|----------|
| Ammonium | 0.1 - 3 | 0.38 | 0.030 | 0.978 | <0.01 | 0.034 |
| | 0.1 - 10 | 2.2 | 0.058 | 0.915 | <0.01 | 0.021 |
| Urea | 0.1 - 3 | 0.28 | 0.0017 | 0.978 | <0.01 | 0.002 |
| | 0.1 - 10 | 0.54 | 0.0021 | 0.922 | <0.01 | 0.002 |

600

601

602

602 Fig. 1. Dissolved ammonium (a), nitrate (b, c), phosphate (d), and silicate (e) concentrations during
603 the experiment. Due to the different scales of nitrate concentration, this nutrient is displayed in two
604 distinct plots (b and c). Error bars correspond to standard deviations. L1 and L1/5 are growth media;
605 BV+DOM is the bacteria+virus control.

606
607 Fig. 2. DOC (a), DON (b), and DOP (c) concentrations. (b) Left y-axis represents DOM-enriched
608 incubations (L1/5+DOM; [(L1-N)+DOM]; BV+DOM); right y-axis represents DOM control
609 incubation (L1-DOM). Error bars correspond to standard deviations; DOC = dissolved organic
610 carbon; DON = dissolved organic nitrogen; DOP = dissolved organic phosphorus. BV+DOM: DON
611 concentration at day 10 was not available.

612
613 Fig. 3. *Pseudo-nitzschia delicatissima* abundance (a) and chlorophyll *a* (b) concentrations. Error
614 bars correspond to standard deviations.

615
616 Fig. 4. Bacteria concentrations. Error bars correspond to standard deviations.

617
618 Fig. 5. Virus concentrations. Error bars correspond to standard deviations.

619
620 Fig. 6. ¹⁵N-incorporation by *Pseudo-nitzschia delicatissima* cells during short-term incubations over
621 a temporal axis (p<0.001).

622
623 Fig. 7. Michaelis-Menten kinetic curves of V_{N-ammonium} (a) and V_{N-urea} (b) in *Pseudo-nitzschia*
624 *delicatissima* cultures. Note the difference of scales between plots.

625
626 Fig. 8. ¹³C assimilation in cells as a function of elapsed time.

627

628















